

Article

Bat Species Diversity and Abundance of Trophic Guilds after a Major Hurricane along an Anthropogenic Disturbance Gradient

Luz María Sil-Berra ¹, Cornelio Sánchez-Hernández ^{2,*}, María de Lourdes Romero-Almaraz ³
and Víctor Hugo Reynoso ²

¹ Posgrado en Ciencias Biológicas, Instituto de Biología, Universidad Nacional Autónoma de México, Ciudad de México C.P. 04510, Mexico

² Departamento de Zoología, Instituto de Biología, Universidad Nacional Autónoma de México, Ciudad de México C.P. 04510, Mexico

³ Escuinapa No. 92 bis. Col. Pedregal de Santo Domingo, Ciudad de México C.P. 04369, Mexico

* Correspondence: cornelio@unam.mx

Abstract: The frequency and intensity of hurricanes have increased with climate change, and their effects on most taxa are not known. We analyzed a species diversity of bats in three locations with different regimes of anthropic disturbance. We assessed the effect of the season and post-hurricane time on the abundance of trophic guilds in coastal Jalisco, México, during the two years following Hurricane Patricia (category 4). During a sampling effort of 15,629.76 m² of netting, we captured 790 bats of 21 species. The species diversity was higher in the site with the highest proportion of primary tropical deciduous forest and was higher in 2016 than in 2017; the species composition did not differ greatly between the two years. The abundance of bats in various trophic guilds varied relative to the four climatic seasons. The general abundance of bats, frugivores-omnivores, and insectivores showed a significant increasing trend over time after the hurricane, which may indicate a recovery of the ecosystem or an abundance of early-successional fruiting plants. The results also confirm that species diversity recovers faster in a conserved forest. Thus, it is important to conserve natural areas to mitigate the effects of major disturbances.

Keywords: Chiroptera; climate change; disturbances; diversity; trophic guild; tropical deciduous forest



Citation: Sil-Berra, L.M.; Sánchez-Hernández, C.; Romero-Almaraz, M.d.L.; Reynoso, V.H. Bat Species Diversity and Abundance of Trophic Guilds after a Major Hurricane along an Anthropogenic Disturbance Gradient. *Diversity* **2022**, *14*, 818. <https://doi.org/10.3390/d14100818>

Academic Editors: Luis Daniel Ávila Cabadilla and Mariana Yólotl Álvarez Añorve

Received: 1 September 2022

Accepted: 24 September 2022

Published: 29 September 2022

Publisher's Note: MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Copyright: © 2022 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (<https://creativecommons.org/licenses/by/4.0/>).

1. Introduction

Hurricanes are tropical cyclones of great dimensions (300–340 km in diameter and with a 119 km/h minimum velocity) that occur in the North Atlantic Ocean, the Caribbean, and the Northeast Pacific Ocean. They contribute to the transport of the warm air mass of the tropics from low to high latitudes and change the “nutrients transport” [1]. Based on their intensity, hurricanes can influence in diverse ways the biodiversity and functioning of ecosystems [2–4]. In addition, the frequency and intensity of major hurricanes has increased in the last decades due to global climate change [5–7].

Several studies have focused on evaluating the direct and indirect effects of hurricanes on ecological and biological traits [8]. The direct effects are mostly due to the direct exposure to winds and rain, and can result in mortality or displacement of individuals and propagules [1]. The indirect effects are induced by changes in the physical environment, ecosystem productivity, and availability of resources, which can result in effects such as increased vulnerability due to the scarcity of resources or a high predation risk. These effects can occur during or immediately after the hurricane; for example, the change in the availability of resources can force changes in the diet, foraging, day-roosting habitat, or reproductive patterns. Meanwhile, the long-term effects can last from few years to centuries [8,9] and influence the recovery and ecological succession of ecosystems which include changes in the species composition, species diversity, and population sizes [10]. The level of the effect and recovery time vary in accordance with intensity, dimension,

and time when hurricanes occur; the traits of the landscape; the structure of vegetation; topography; land use and management; and the natural susceptibility of species of plants and animals present to these events [4,9,11].

For terrestrial mammals, the exposure to extreme climate events, such as hurricanes and fires, has received some attention [12], as has their susceptibility as a result of their sensibility (vagility and territoriality) and their adaptive capacity (diet and habitat specialization) [13]. Previous studies indicate that Primates, Rodentia, and Chiroptera have the highest proportion of species affected by cyclones and fires worldwide [12].

Bats are an important group in ecosystems due to their wide ecological diversity and ecosystem services [14,15]. Their dispersion capacity provides a degree of resistance and resilience to disturbances in comparison with other terrestrial mammals [16,17]; however, some bat species have life-history traits or behaviors that make them vulnerable [18,19]. The effects of hurricanes on bats are multi-faceted and depend on habitat traits and their foraging, feeding, and roosting habits [20].

In October 2015, Hurricane Patricia (category 4, Saffir–Simpson scale) made landfall on the coast of Jalisco, Mexico, and affected the structure and phenology of vegetation [21–24], as well as the local distribution patterns, species diversity, and richness of various taxa (e.g., rodents and birds) [25,26]. In bats, an immediate effect was the decrease in the diversity of species and abundance of some functional groups, such as nectar-feeding bats. The effects were different among locations with distinct anthropogenic disturbance and hurricane impact regimes [20]. Understanding these assemblages is necessary to estimate the consequences of major natural events on the biodiversity and their interaction with anthropogenic disturbance in the longer term.

Land use can influence the effects, responses, and rate of recovery of ecosystems after hurricanes [8]. We evaluated the hurricane effects on bat assemblages in three locations with distinct vegetation traits, anthropogenic disturbance, and land use. Our purposes were: (1) to evaluate species diversity among locations and between two post-hurricane years; (2) to identify differences in seasonal abundance patterns among trophic guilds; and (3) to evaluate trends in abundance of each trophic guild during two years after Hurricane Patricia.

We expected to find a higher diversity in the second year after the hurricane in the most conserved areas and in locations less impacted by the hurricane. We also expected variation in bat abundance across seasons and an increasing trend of abundance of bats in general and in each trophic guild during the period after the hurricane. These changes would reflect the recovery in flowering and fructification phenology and in ecosystem functioning in general.

2. Materials and Methods

2.1. Study Sites

Our study was conducted in three locations in the coast of Jalisco, Mexico, in a heterogeneous landscape conformed mainly by primary and secondary tropical deciduous forest, primary sub-deciduous forest, farming and grazing areas, and underdeveloped human settlements [27]. The region includes a Natural Protected Area, the Chamela-Cuixmala Biosphere Reserve, that subsumes 131,142 ha, mainly of tropical deciduous forest [28]. A pristine study location (Chamela) was found within this area. A moderately disturbed location (La Fortuna), 15 km N of Chamela, contained primary and secondary tropical deciduous forest in addition to farming and grazing areas. The most anthropically disturbed location (Zapata) is 15 km S of Chamela and included secondary tropical deciduous forest along with farming and grazing areas. More information about the study areas can be found in Bullock [20]. The eye of Hurricane Patricia passed through Chamela and Zapata, but not through La Fortuna.

Traditionally, two climatic seasons are considered for this region: a rainy (July–October) and a dry season (November–June) [29]. However, based on the availability of bat resources and precipitation level, Stoner [30] proposed four climate seasons: (1) an early dry season

from January to March, with some precipitation events [31] and when chiropterophilic flowering and canopy fruiting season is at their peak; (2) a late dry season from April to June, with few chiropterophilic flowering and fruits, but by the end of this season the first rains of the year start; (3) an early rainy season from July to September, when the peak of rainfall and chiropterophilic flowering occurs, but during which in general there are few fruits; (4) a late rainy season from October to December, with the last rains of the year, with some chiropterophilic flowering, and with a greater number of trees having fruits.

2.2. Bat Sampling

We sampled bats with mist-nets (Avinet, Inc., Portland, ME, USA; 75 denier, 2 ply, 38-mm mesh) set in three sampling stations in each location. Sampling occurred every three months during 2016 and 2017, covering the four seasons of the Chamela region [30]. In each sampling period, we placed eight mist-nets, five at ground level (three of 9 [height] \times 2.68 m [length] and two of 12×2.68 m) and three at 3.0 – 5.68 m height (two of 9×2.68 m and one of 12×2.68 m). The distance between the adjacent nets varied from 20 to 50 m, and the nets remained open during five hours after sunset. The sampling effort of each sampling period was 217.08 m² net-nights (i.e., number of square meters of net set on a given night) with a total of $15,629.76$ m² for the study.

We determined the identity of the bats using identification keys [32,33], following the classification proposed by Pavan and Marroig [34,35] for *Pteronotus*, by Baird et al. [36] for Lasiurini, and by Calahorra-Oliart et al. [37] for *Glossophaga*. We recorded biological and reproductive data, including forearm length (Truper[®] \pm 0.05 mm) and mass (Pesola[®] \pm 0.05 g). To avoid measuring the same individual twice and in case we had across-year recaptures, we marked small bats with numbered metal rings (Alloy Split Rings 2.9 mm, Porzana Ltd., East Sussex, UK) and the largest ones with plastic collars with colored beads. The bats were released at their capture sites.

2.3. Data Analysis

To compare the species diversity between years and locations, we built individual-based and sampled-coverage rarefaction curves in iNext [38]. The species diversity (order $q = 0, 1,$ and 2) among locations was compared with the same number of individuals ($n = 182$) and the same coverage (0.98) throughout 1000 Bootstrap iterations and 95% CI.

To compare species composition between years and locations, we calculated similarity indices in SpadeR [39] with 1000 iterations in all cases. For comparison among locations in two years, we computed Bray–Curtis indices. This index is reliable when absolute abundances are compared under standardized sampling efforts across all of the communities. Additionally, we performed a paired comparison with Sørensen and Jaccard indices. For the comparison between years by location, we used the Chao–Jaccard and Chao–Sørensen indices. These indices consider the relative abundance of the species and are recommended when under-sampling is suspected and when it is likely that samples have many rare species [40], as occurs with bat assemblages.

To identify the drivers of differences in species diversity, we built rank-abundance curves for each location. We performed a Wilcoxon matched-pairs signed rank test on the number of individuals captured by species and by location, to identify the species with more changes in their abundance between years.

Bat species were classified into five trophic guilds: frugivore-omnivore, frugivore, nectarivore, insectivore, and hematophagous, according to relevant literature [33]. To evaluate the effect of the season and post-hurricane time on the abundance of bats by trophic guild, we used generalized linear models (GLM) with Poisson distribution. We performed two tests in each case: one with two classical seasons, dry and rainy (season1), and the other with the four seasons as described by Stoner [30] (season2). We evaluated four models that were compared with a null model with the anova function in R [41]:

Model 1: captures~season1 + post-hurricane time

Model 2: captures~season1

Model 3: captures~season2 + post-hurricane time

Model 4: captures~season2

Null model: captures~1

We selected the model with the lowest Akaike Information Criterion (AIC) and highest explained deviance that was significantly distinct from the null model.

Finally, to identify the trend direction in the abundance of bats, we transformed the data with $\sqrt{(x + 1)}$ to diminish the influence of extreme data due to seasonality. With these data, we performed Mann–Kendall trend tests [42–44] with all the data and by trophic guild in XLSTAT [45], and with the temporal series softened with the moving average model. The Mann–Kendall trend test or Kendall's τ (tau) is a non-parametric test that has been used to identify trends considering randomness and seasonality, mainly in climatological and hydrological analyzes (e.g., [46]).

3. Results

3.1. Species Diversity and Composition

We captured 790 bats from 21 species, 15 genera, and 6 families. The most frequently encountered species were the Jamaican fruit-eating bat, *Artibeus jamaicensis* ($n = 190$), common vampire bat, *Desmodus rotundus* ($n = 167$), Mexican mustached bat, *Pteronotus mexicanus* ($n = 131$), and great fruit-eating bat, *A. lituratus* ($n = 122$). The location with the most captures and the greatest number of species was Chamela ($n = 316$, 18 species), followed by La Fortuna ($n = 292$, 15 species) and finally Zapata ($n = 182$, 12 species; Table 1).

Table 1. Number of individuals captured at each study location.

Family/Species	Trophic Guild ^a	Chamela	La Fortuna	Zapata	Total
Mormoopidae					
<i>Pteronotus mexicanus</i>	IN	40	76	15	131
<i>Pteronotus psilotis</i>	IN	1	3		4
<i>Mormoops megalophylla</i>	IN	1			1
<i>Pteronotus davyi</i>	IN	1			1
Natalidae					
<i>Natalus mexicanus</i>	IN	1			1
Noctilionidae					
<i>Noctilio leporinus</i>	PI		1		1
Phyllostomidae					
<i>Artibeus jamaicensis</i>	FO	111	52	27	190
<i>Desmodus rotundus</i>	HE	89	50	28	167
<i>Artibeus lituratus</i>	FO	25	52	45	122
<i>Sturnira parvidens</i>	FO	1	10	41	52
<i>Glossophaga soricina</i>	NE	14	10	15	39
<i>Dermanura phaeotis</i>	FO	19	8	1	28
<i>Dermanura tolteca</i>	FR	2	13	4	19
<i>Leptoncteris yerbabuena</i>	NE	4	12		16
<i>Centurio senex</i>	FR	1	2	3	6
<i>Glossophaga commissarisi</i>	NE	2	1		3
<i>Glossophaga morenoi</i>	NE	1	1	1	3
Vespertilionidae					
<i>Rhogeessa parvula</i>	IN	2		1	3
<i>Lasiurus frantzii</i>	IN	1			1
<i>Dasypterus intermedius</i>	IN		1		1
Molossidae					
<i>Nyctinomops aurispinosus</i>	IN			1	1
Total species		18	15	12	21
Total individuals		316	292	182	790

^a FO, frugivore-omnivore; FR, frugivore; NE, nectarivore; IN, insectivore; HE, hematophagous; PI, piscivore.

Individual-based rarefaction analyses did not show significant differences in species richness ($q = 0$) among locations (Figure 1a). However, coverage-based rarefaction (Figure 1d)

indicated a higher richness in Chamela than in other locations. In contrast, order 2 diversity showed that the number of effective species was lower in Chamela (Figure 1c,f); the abundance among the most frequently encountered species was more uneven in Chamela. In the comparison between years, no significant difference was demonstrated in overall species diversity. By location, diversity was higher in 2016 than 2017 only in Chamela for order 0 and 1 species diversity (Figure 2).

The similarity of species composition among locations varied from 0.60 to 0.68 (Bray–Curtis index). In paired comparisons, the similarity was highest between Chamela and Zapata and lowest between La Fortuna and Zapata. However, the differences were not significant (95% CI; Table 2), and high values in the indices reflect the few changes in species compositions between years for the three locations (Table 3).

The rank–abundance curve for Chamela showed a higher richness in 2016 than 2017, with the abundance distribution for 2016 being more homogeneous among species and there being a greater number of rare species. The most frequently netted species in 2016 were *A. jamaicensis*, *D. rotundus*, and *A. lituratus*, while in 2017 they were *A. jamaicensis*, *D. rotundus*, and *P. mexicanus* (Figure 3). The Wilcoxon pair tests were not significant when considering the number of individuals ($p = 0.88$) or relative abundance ($p = 0.14$), which indicated that there were differences between the two years in the abundance rank of the species, showing an increase in 2017 of *D. rotundus* and *P. mexicanus*, and a decrease of *A. lituratus* (Figure 4).

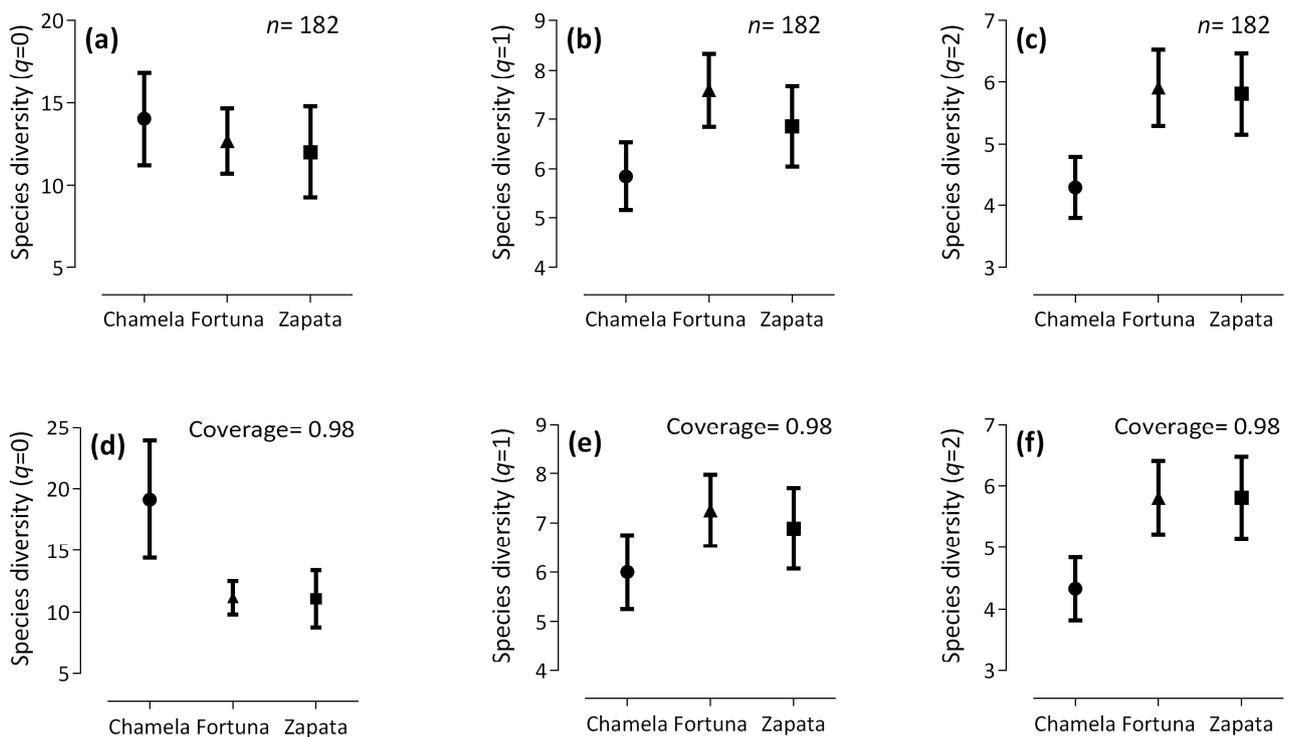


Figure 1. Diversity orders $q = 0, 1,$ and 2 based in number of individuals (a–c) and sampling coverage (d–f) for three locations during two years of fieldwork.

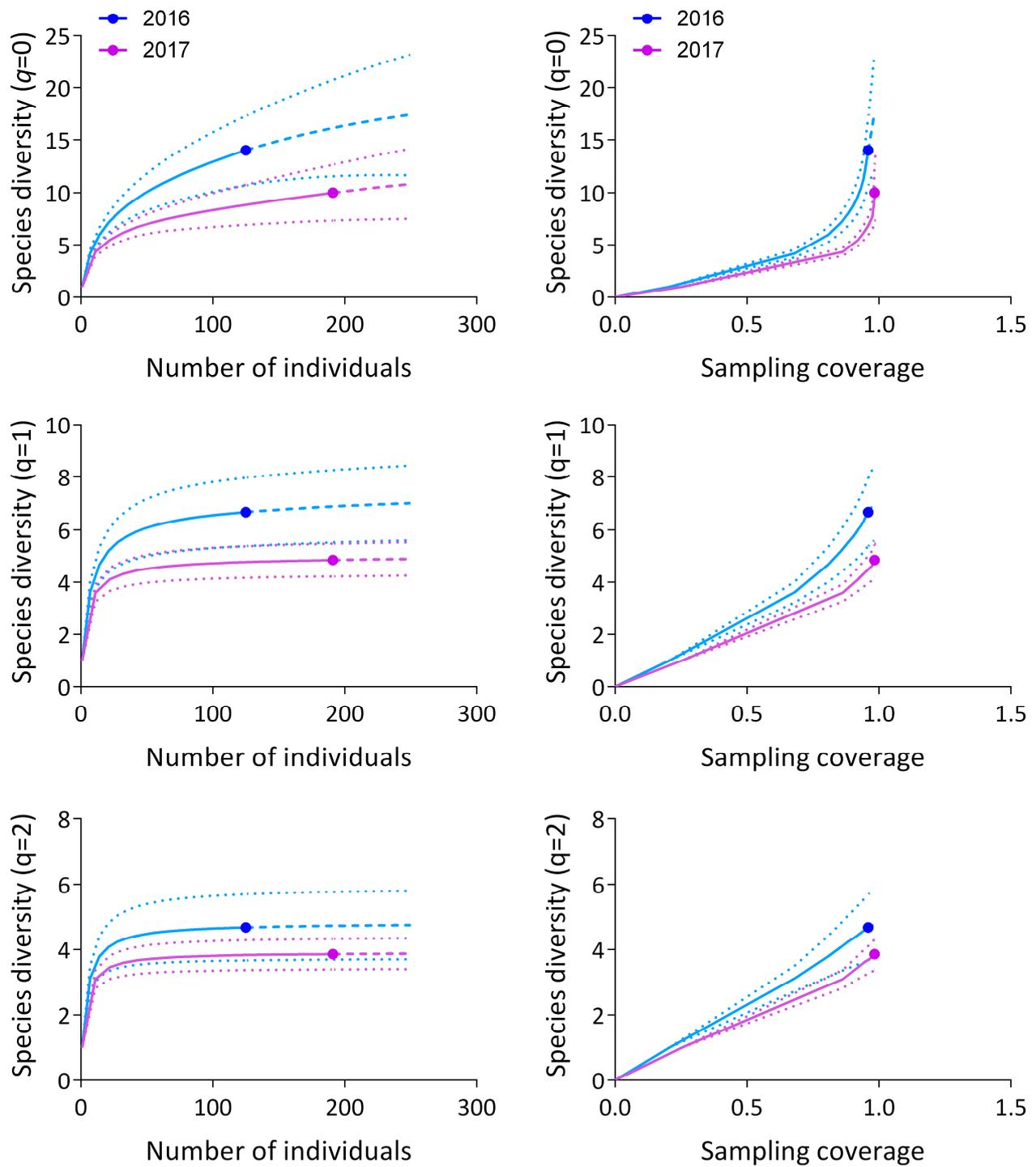


Figure 2. Interpolation (continuous lines) and extrapolation curves (discontinuous lines) of species diversity orders $q = 0, 1$ and 2 , based in number of individuals (**left panel**) and sampling coverage (**right panel**) for Chamela in 2016 (blue line) and 2017 (purple line).

Table 2. Bat species composition similarity among locations after two years of Hurricane Patricia and CI (95%).

Locations	Jaccard's Index	Sørensen's Index
Fortuna-Chamela	0.429 (0.214–0.644)	0.600 (0.361–0.840)
Fortuna-Zapata	0.367 (0.171–0.564)	0.537 (0.323–0.751)
Chamela-Zapata	0.474 (0.222–0.725)	0.643 (0.376–0.910)

Table 3. Bat species composition similarity by location between two post-hurricane years (2016 vs. 2017) and CI (95%).

Location	Chao-Jaccard's Index	Chao-Sørensen's Index
Fortuna	0.968 (0.899–1.040)	0.983 (0.957–1.010)
Chamela	0.875 (0.869–0.881)	0.933 (0.930–0.936)
Zapata	0.957 (0.765–1.150)	0.978 (0.847–1.110)

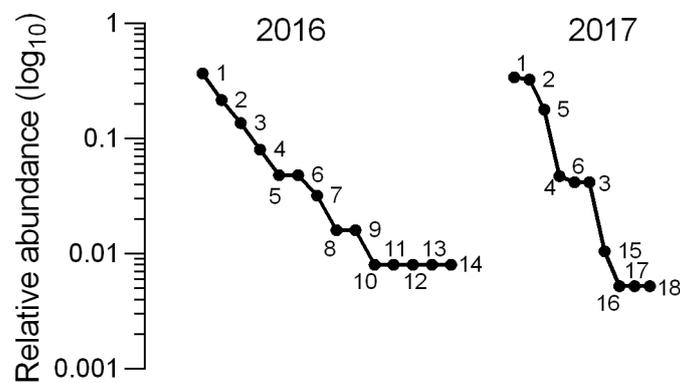


Figure 3. Rank-abundance curves of bat species captured in 2016 and 2017 in Chamela, Jalisco. 1: *Artibeus jamaicensis*, 2: *Desmodus rotundus*, 3: *A. lituratus*, 4: *Dermanura phaeotis*, 5: *Pteronotus mexicanus*, 6: *Glossophaga mutica*, 7: *Leptonycteris yerbabuena*, 8: *G. commissarisi*, 9: *D. tolteca*, 10: *Mormoops megalophylla*, 11: *P. psilotis*, 12: *P. fulvus*, 13: *G. morenoi*, 14: *Sturnira parvidens*, 15: *Rhogeessa parvula*, 16: *Natalus mexicanus*, 17: *Centurio senex*, 18: *Lasiurus frantzii*.

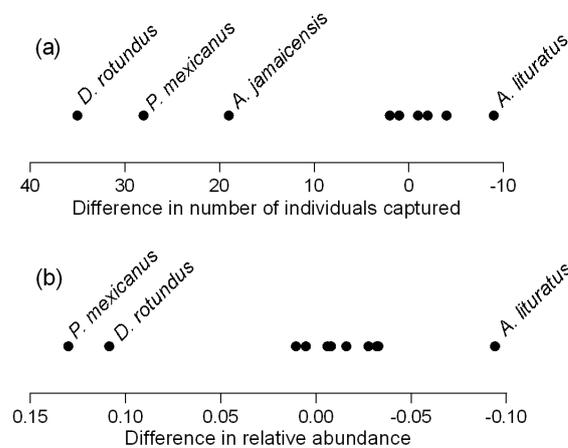


Figure 4. Differences in number of captures and relative abundance between 2016 and 2017 of captured species in Chamela. Only names of species with the greatest changes are included. Some of the dots in the cluster of points around zero represent multiple species.

3.2. Seasonal and Post-Hurricane Abundance Trends

GLM analyses considering the three locations showed that the number of bats captured was influenced by the season and the post-hurricane time. All evaluated models showed significant differences among the seasons. The best model was model 3, which included the four seasons and the post-hurricane time. This indicated that in the early rainy season more bats were captured, and in the early dry and late rainy seasons fewer bats were caught (Figure 5a).

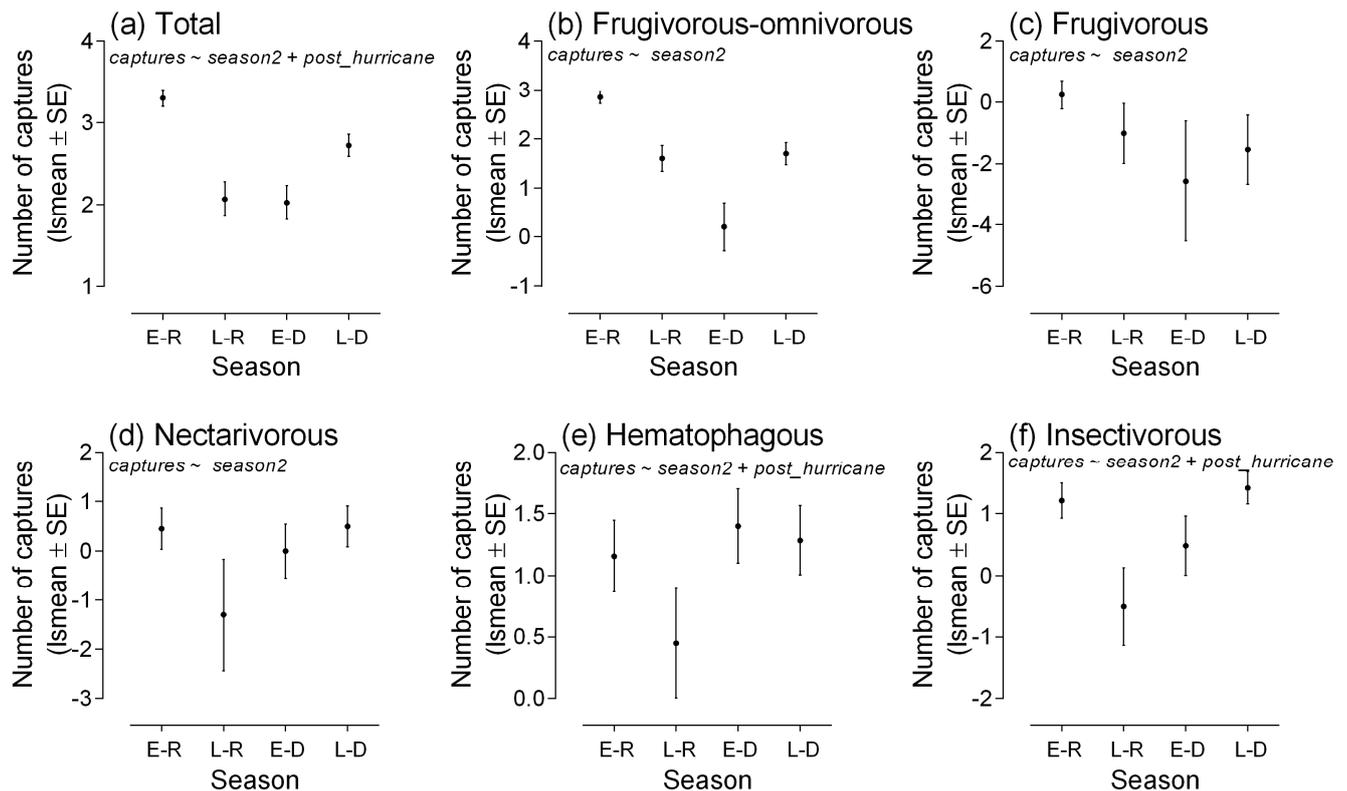


Figure 5. Seasonal variation in abundance of bats (number of captures), total (a) and by trophic guilds (b–f), in accordance with selected models throughout GLMs. E-R: Early rainy, L-R: Late rainy, E-D: Early dry, L-D: Late dry. Details are shown in Table S1 of Supplementary Materials.

For the frugivore-omnivore and frugivore bats, models 3 and 4 were the best. According to these models, more frugivore-omnivore and frugivore bats were captured in the early rainy season, fewer frugivore-omnivore bats in the late rainy and late dry seasons, and fewer frugivore bats in the early and late dry seasons (Figure 5b,c).

For the nectarivore guild, the differences among models were minimal and there was no post-hurricane effect. According to model 4, which is the simplest with a higher explained deviance and a lower AIC, more nectarivore bats were captured in the early rainy and late dry seasons and fewer in the late rainy season (Figure 5d). For hematophage bats, the explained deviance in all models was low. Model 3 showed an effect of the season, with fewer captures in the late rainy season and the post-hurricane time (Figure 5e).

For the insectivore guild, the explained deviance did not exceed 31% in all the models. Model 3 showed the highest deviance (31%) and the lowest AIC (274.2). This model showed that more insectivore bats were captured in the late dry and early rainy seasons ($p = 0.71$), and there was a post-hurricane effect (Figure 5f).

The Mann–Kendall trend test showed that the number of captures of all bats increased after the hurricane (Figure 6). By guilds, the abundance of frugivore-omnivore and insecti-

vore guilds increased ($p < 0.01$). Nectivore bats showed a decrease, but it was not significant ($p = 0.07$). Frugivore and hematophage bats did not show any trend (Figure 6).

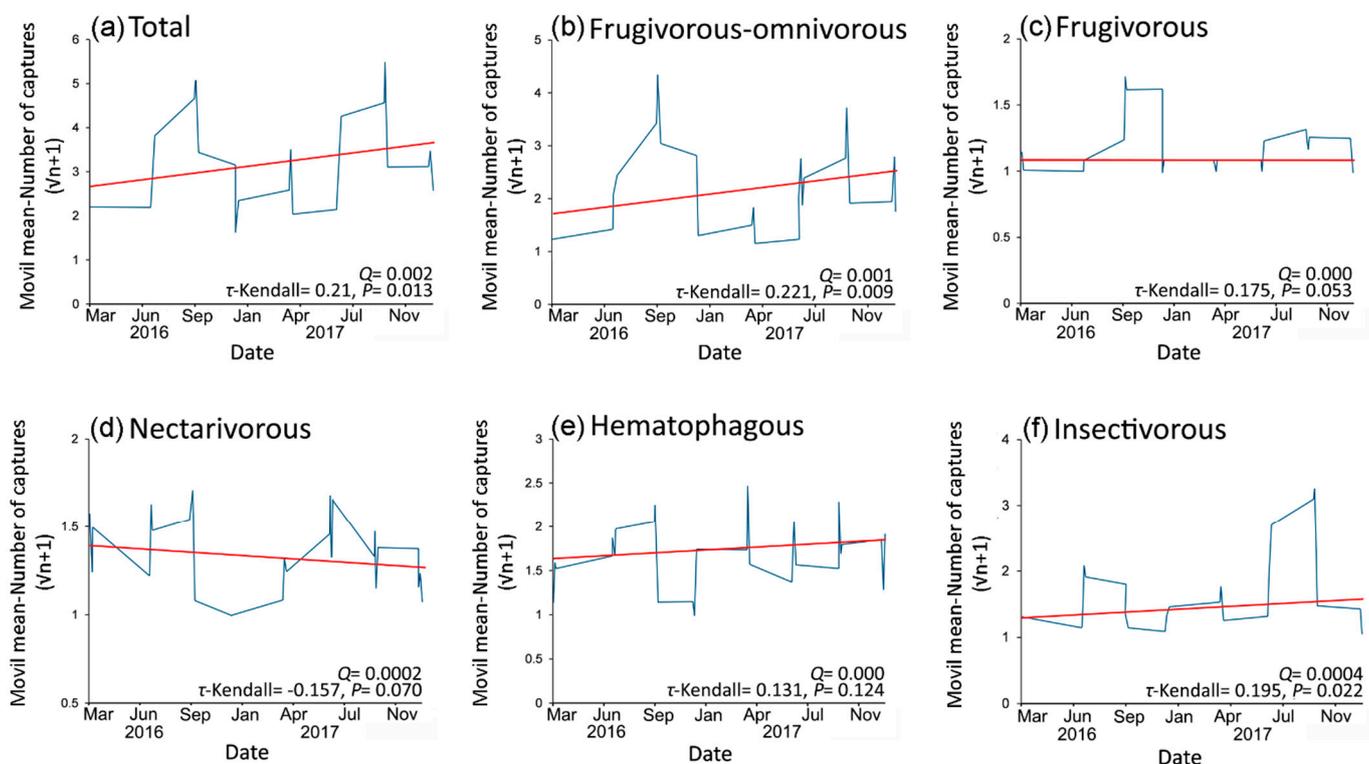


Figure 6. Post-hurricane trends in the abundance of bats, total (a) and by trophic guilds (b–f), from March 2016 to December 2017. Sen's slope (red line) adjusted to smoothed curve (blue line).

4. Discussion

As expected, two years after Hurricane Patricia, bat species richness was higher in Chamela than in the other locations, but it had a lower evenness among the most common species (lower diversity order 2). Previous reports indicated that eight months after Hurricane Patricia, the dominance and evenness were lower in Chamela and Zapata compared with pre-hurricane samplings, and the species richness was not different among locations [20]. These results suggest that areas with a high quantity and diversity of native vegetation such as Chamela, which contains primary deciduous and semideciduous forest, are more resilient when facing natural disturbances. This may indicate that there was a greater availability of resources, despite the disturbance, in a shorter amount of time with the end result being a higher bat species richness.

Previous research has indicated that the dry forest in Chamela has low resistance to hurricanes, but is highly resilient due to the quantity of moisture left. The recovery rate depends on wind magnitude, post-disturbance climate variations, and vegetation traits [22,47]. These factors, in addition to landscape traits and anthropic degradation regimen, may influence the recovery of populations and bat assemblages after hurricanes [48]. Hurricane Patricia caused more damage to tree structure in primary vegetation than in secondary vegetation, but primary vegetation was more resilient in terms of lower mortality and higher density of leaf outbreaks [22]. This may have been the result because primary vegetation has a greater height, diameter, and age [22], which in turn provide a greater ability for vegetation to obtain resources, and leads to a faster recovery of the bat assemblages in the most conserved locations such as Chamela. Previous reports have documented changes in the roosting habits after hurricanes, such as there being more roosts inside reserves than outside, and displacement to more permanent roosts such as caves [48]. Though tree damage can create new roosts in broken tops and cavities, usually

bats are more often found in older trees where they find an optimal microclimate [49]. Also, changes in the diet of frugivorous species occurred in that they consumed more leaves than fruits after hurricanes [48]. Leaf outbreaks and roosts in primary vegetation after the hurricane could have provided resources exploited by frugivorous bats in Chamela.

Lower order-2 diversity in Chamela suggests a low evenness in the abundance of the most frequently encountered species, *Artibeus jamaicensis* and *D. rotundus*. The recovery of resources may have been exploited mainly by frugivorous-omnivorous species, or by those that do not rely heavily on vegetation. Similarly, in avian assemblages, an increase in the capture rates of omnivores and open-area foragers has been documented after hurricanes [50].

Chamela also had the greatest differences in diversity and species composition between years. Contrary to expectation, diversity was higher in 2016 than in 2017 ($n = 14$ vs. 10 in species richness and $n = 5$ vs. 3 in rare species) in Chamela. Additionally, there was a greater evenness in the abundance among species in 2016, and the species-composition similarity decreased to a greater extent between years in comparison with the other locations. These changes could have been the result of several factors. First, the higher vegetation cover in Chamela compared with other locations may have functioned as a shelter area for generalists (e.g., *A. lituratus* and the pygmy fruit-eating bat, *Dermanura phaeotis*), specialists (e.g., the Commissaris's long-tongued bat, *Glossophaga commissarisi*, and the Toltec fruit-eating bat, *Dermanura tolteca*), common species (e.g., *A. jamaicensis* and *D. rotundus*), and rare species (e.g., the Thomas's naked-backed bat, *Peronotus fulvous*, and the western long-tongued bat, *Glossophaga morenoi*) in 2016 immediately after the hurricane. According to the intermediate-disturbance hypothesis and the recovery process, the most resistant species and those not so resistant that colonize new niches after disturbances were able to share the same area, thus increasing diversity.

Second, climate variation between years can influence diversity. The Chamela Weather Station, UNAM, indicated that 2016 was drier than 2017, which would possibly result in lower diversity in 2016. However, humidity favors the vegetation of the entire region and due to a more humid condition in 2017, some species of bats may have dispersed to other areas searching for food or habitat (e.g., *A. lituratus*), thus reducing the local diversity in this year. The greater difference in species composition between years was the result of an increase in 2017 in the abundance of *A. jamaicensis* and *D. rotundus*, and a decrease of *A. lituratus*; the number of the rare species also was lower in 2017. Although *A. jamaicensis* and *A. lituratus* are related species and have similar diets [51], *A. lituratus* feeds on a higher quantity of pioneer plants [52,53] (which should be more abundant after a storm), it has a larger body size and a larger home range, and forages in canopy and understory, as well as different successional stages [54]. Consequently, in 2017, *A. lituratus* could have extended its home range or moved to other locations with earlier successional stages.

The fact that the species composition between La Fortuna and Zapata was less similar than between Chamela and Zapata may be an effect of the decline in similarity by distance [55]. La Fortuna and Zapata are the furthest apart, and greater differences in composition could be expected with respect to the impact of Hurricane Patricia.

Despite the negative effects of the hurricane on the abundance of some bat functional groups [20], there was a seasonal variation in the abundance of trophic guilds, as suggested by Stoner [30]. This abundance was higher during the early rainy season (July to September) and lower in the late rainy season (October to December) and early dry season (January to March). During the early rainy season, there is a higher precipitation, greater abundance of chiropterophilic flowers (e.g., the orange flame vine, *Combretum fruticosum*, the morrito, *Crescentia alata*, the pochote, *Ceiba aesculifolia*, columnar cactus, *Stenocereus chrysocarpus*), and some fruits are more readily available, which benefits all trophic guilds. The decrease in abundance of bats in the late rainy season may result from a decrease in the rains and in chiropterophilic flowering (although many trees would still have fruits), which would explain the smaller decrease in the abundance of frugivore guilds. In the early dry season, the second peak of chiropterophilic flowering (e.g., *C. fruticosum*, *Ceiba aesculifolia*, the

morning glories, *Ipomoea ampullacea*, the orchid tree, *Bauhinia unguolata*, the agave, *Agave ortgiesiana*) and canopy fruiting occur [30,56]. The increase in the abundance of members of the nectarivore guild coincides with that, but not with the decrease in the abundance of species in the frugivore-omnivore guild. This could be the result of the impact of Hurricane Patricia on the phenology of plants [57], but more information is needed.

The representative species of the frugivore-omnivore guild were *A. jamaicensis*, *A. lituratus*, and *D. phaeotis*, so it would be important to research the response to disturbances of some plants that are an important part of their diet, such as figs (*Ficus* spp.), mombins (*Spondias* spp.), columnar cactus (*Stenocereus* spp.), guava (*Psidium* spp.), cecropias (*Cecropia* spp.), and nightshade (*Solanum* spp.) [33,51,58–61]. For example, it has been reported that some *Ficus* species are medium-low resistant to winds by hurricanes [62] which in Chamela is one of the genus with more resprouting capacity [22]; in contrast, columnar cactus are more vulnerable to these winds for their trunk shape and structure [63].

Another factor influencing variation in abundance of bats is reproduction patterns. The greater abundance of frugivore bats in the early wet season may be associated with our high capture rates of juvenile individuals ($n = 67$ for *A. jamaicensis* and $n = 26$ for *A. lituratus*) compared to other seasons all together ($n = 5$ for *A. jamaicensis* and $n = 0$ for *A. lituratus*). This may indicate that, although these bats have a continuous polyestrous reproductive pattern, they have a birth peak that coincides with the beginning of the fruiting peak. This may reflect a strategy to ensure the survival success of future adults, in which the fruiting peak (October to December) concurs with the time when the young reach postnatal development. Additionally, foraging time or distance also may change in accordance with the availability of resources or reproduction conditions [64–66], modifying the capture rate in mist-nets.

Nectarivorous and insectivorous bats have two peaks of abundance during the early rainy and late dry seasons. In contrast, Stoner [30] reported a higher abundance of nectarivorous bats in the early rainy and early dry seasons at Chamela. It could be that the peak abundance of chiropterophilic flowers during the early dry season (January to March) in 2016 was affected by Hurricane Patricia, which would have resulted in a delay in the peak abundance of nectarivorous bats. Other probable causes may be related with interannual climatic variation, for example winter rains influenced by El Niño Southern Oscillation (ENSO) [67], that could promote changes in the abundance pattern of bats.

The beginning of chiropterophilic flowering can occur during the late dry season, continuing until the early rainy season [30]. Important plants for the most frequently encountered nectarivore species (Merriam's long-tongued bat, *Glossophaga mutica*, and lesser long-nosed bat, *Leptonycteris yerbabuena*) are manjack (*Cordia alliodora*), white silk-cotton tree (*Ceiba pentandra*), *C. alata*, shavingbrush tree (*Pseudobombax ellipticum*), morning glory (*Ipomoea* sp.), and columnar cactus (*Stenocereus* sp.) [33,56,68]. For the genus *Cordia*, the main damages resulting from Hurricane Patricia were broken and inclined trunks. For *Crescentia*, it was secondary broken branches. Damages for *Ipomoea wolcottiana* were the most severe, with uprooted trees [22]. As mentioned earlier, columnar cacti were severely damaged [63].

The greater abundance of insectivorous bats during the late dry and early rainy seasons can be explained by the timing of the first rains, which increases the abundance and diversity of insects. The most abundant insectivore was *P. mexicanus*, a generalist species in terms of diet and habitat. The variation in its abundance paralleled insect availability [69,70]. In Chamela, *P. mexicanus* feeds mainly on Lepidoptera and Diptera. It does not show changes in dietary breadth with the season, and their diet overlaps with that of Dobson's lesser mustached bat, *P. psilotis* [71], although this latter species is scarce in the region.

Vampire bats were less abundant in the late rainy season. Temporal variations in *Desmodus rotundus* are mainly related to livestock management practices [72] and, to a lesser extent, to variation in the abundance of other prey [73] and to temperature variation due to seasonality [74,75]. The management of cattle by villagers varied seasonally

and with respect to the availability of resources. When resources are limited, villagers confine cattle or let them graze in smaller areas, which can influence the capture rate of hematophagous bats. In Chamela, the nights are colder from December to February than during the rest of the year [76], which may influence the activity of vampire bats. Variation in abundance of vampire bats is also related to reproduction. Although they exhibit a continuous asynchronous polyestrous reproduction pattern, births have been reported to peak in the rainy season [77–79]. More detailed studies could provide additional information on local variation in activity, local movements, foraging patterns, and population dynamics, as well as the association of these with climatic seasonality.

The increasing trend in abundance of bats may indicate recovery of the forest and ecosystem in general two years after the hurricane. However, by trophic guilds, this trend was found only in the frugivore-omnivore and insectivore bats. In addition, an analysis of both before and after the hurricane indicated that the abundance of nectarivore bats decreased after this event [20]. Altogether, results may indicate that this guild had not recovered by the end of our sampling in December 2017. This agrees with findings by Renton et al. [57], who reported that phenological patterns of flowering and fructification in general had not recovered from the impact of Hurricane Patricia a year later, but specific studies on chiropterophilic plants are needed. Similarly, in avian assemblages, frugivore and nectarivore guilds were the most affected two years after Hurricane Iris (category 4) in Belize [50]. *Glossophaga mutica* appears to be one of the most affected bats in that region. This bat was consistently the most frequently captured species in samplings prior to Hurricanes Jova (category 2; October 2011) and Patricia [30,61,80–82]. However, it was not in the most recent studies [26] and not even in our study, in which it represented only 6% of the abundance of all phyllostomid bats.

Among the recommendations for future studies, we urge that long-term data be gathered so that it would be possible to analyze the relationship among interannual climate variations and the effects of hurricanes and bat diversity. This would also allow analysis of long-term data to predict population-size changes and to distinguish variation in populations, due to the effects of hurricanes from natural patterns as a result of stochastic events or natural population dynamics. These studies are particularly needed for vulnerable species such as nectarivores, in order to decipher their ecological interactions and discover what factors are affecting the populations of these species.

5. Conclusions

Based on our findings, we conclude that: (1) the species diversity of bats recovered faster in areas with more primary forest; (2) despite the impact of the hurricane on forest phenology, the abundance of trophic guilds maintained seasonality, mainly for frugivore and nectarivore guilds; (3) an increasing trend in the abundance of frugivore-omnivore and insectivore bats may indicate a recovery of their resources; and (4) nectarivore bats were the most severely impacted by Hurricane Patricia, and it is likely that their resources had not yet recovered two years after the impact. The resilience of tropical deciduous and semideciduous forests facing extreme climate events such as hurricanes, promotes recovery of biodiversity at a higher rate. The high vagility of bats compared to other taxa may confer advantages after certain disturbances, allowing faster recovery of assemblages; however, basic resources required by bats must be available. Hence, it is important to conserve natural protected areas for their function as shelter areas and to increase the complexity in agricultural or abandoned areas, in order to mitigate the effects of extreme events such as hurricanes. Understanding the effects of seasons and storms on local assemblages may serve as a place-holder for making inferences about climate change as we accumulate more years and more data.

Supplementary Materials: The following supporting information can be downloaded at: <https://www.mdpi.com/article/10.3390/d14100818/s1>. Table S1: Models (GLM's with Poisson distribution) to assess the effect of seasonality and post-hurricane time on the number of bats captured from March 2016 to December 2017.

Author Contributions: Conceptualization, L.M.S.-B., C.S.-H. and V.H.R.; Methodology, L.M.S.-B. and M.d.L.R.-A.; Software, L.M.S.-B.; Validation, V.H.R.; Formal Analysis, L.M.S.-B.; Investigation, L.M.S.-B. and M.d.L.R.-A.; Resources, C.S.-H.; Data Curation, L.M.S.-B.; Writing—Original Draft Preparation, L.M.S.-B.; Writing—Review & Editing, L.M.S.-B., M.d.L.R.-A. and V.H.R.; Visualization, L.M.S.-B.; Supervision, C.S.-H. and V.H.R.; Project Administration, C.S.-H.; Funding Acquisition, L.M.S.-B. and C.S.-H. All authors have read and agreed to the published version of the manuscript.

Funding: Field research was supported by the Instituto de Biología, UNAM. LMSB thanks to the Consejo Nacional de Ciencia y Tecnología (CONACyT) for PhD scholarship (grant 400449) and the Southwestern Association of Naturalists (SWAN) for the Howard McCarley Award in 2016. Idea Wild Association donated mist-nets and other supplies for field research (ID: SILMEXI0315).

Institutional Review Board Statement: Sampling and marking procedures followed the guidelines of the American Society of Mammalogists [83]. Permits for conducting this research were provided to C. Sánchez-Hernández by the Instituto Nacional de Ecología, Dirección General de Vida Silvestre FAUT.0103.

Data Availability Statement: Not applicable.

Acknowledgments: Fieldwork was largely supported by F. Novoa, S. Ibarra, E. Morales, J.R. Carlin, A. Cruz, J.G. Rocha, S. Cano, A. Gallegos, and A. Lira. We especially thank Eva Robles J. and her family for their attention in Francisco Villa during fieldwork. We thank G. D. Schnell and V. Vratny for reviewing the manuscript in English, the personnel of the Estación de Biología Chamela of the Universidad Nacional Autónoma de México (UNAM) for the logistical support during fieldwork. Helpful comments were received from two anonymous reviewers.

Conflicts of Interest: The authors declare no conflict of interest. The funders had no role in the design of the study; in the collection, analyses, or interpretation of data; in the writing of the manuscript, or in the decision to publish the results.

References

1. Michener, W.K.; Blood, E.R.; Bildstein, K.L.; Brinson, M.M.; Gardner, L.R. Climate change, hurricanes and tropical storms, and rising sea level in coastal wetlands. *Ecol. Appl.* **1997**, *7*, 770–801. [[CrossRef](#)]
2. Brokaw, N.V.L.; Walker, L.R. Summary of the effects of Caribbean hurricanes on vegetation. *Biotropica* **1991**, *23*, 442–447. [[CrossRef](#)]
3. Lloyd, J.D.; Rimmer, C.C.; Salguero-Farías, J.A. Short-term effects of Hurricanes Maria and Irma on forest birds of Puerto Rico. *PLoS ONE* **2019**, *14*, e0214432. [[CrossRef](#)] [[PubMed](#)]
4. Waide, R.B. Summary of the response of animal populations to hurricanes in the Caribbean. *Biotropica* **1991**, *23*, 508–512. [[CrossRef](#)]
5. Emanuel, K. Increasing destructiveness of tropical cyclones over the past 30 years. *Nature* **2005**, *436*, 686–688. [[CrossRef](#)]
6. Elsner, J.B.; Kossin, J.P.; Jagger, T.H. The increasing intensity of the strongest tropical cyclones. *Nature* **2008**, *455*, 92–95. [[CrossRef](#)]
7. Knutson, T.R.; McBride, J.L.; Chan, J.; Emanuel, K.; Holland, G.; Landsea, C.; Held, I.; Kossin, J.P.; Srivastava, A.K.; Sugi, M. Tropical cyclones and climate change. *Nat. Geosci.* **2010**, *3*, 157–163. [[CrossRef](#)]
8. Lugo, A.E. Visible and invisible effects of hurricanes on forest ecosystems: An international review. *Aust. Ecol.* **2008**, *33*, 368–398. [[CrossRef](#)]
9. Ackerman, J.D.; Walker, L.R.; Scatena, F.N.; Wunderle, J. Ecological effects of hurricanes. *Bull. Ecol. Soc. Am.* **1991**, *72*, 178–180.
10. Wiley, J.W.; Wunderle, J.M. The effects of hurricanes on birds, with special reference to Caribbean Islands. *Bird Conserv. Int.* **1993**, *3*, 319–349. [[CrossRef](#)]
11. Schoener, T.W.; Spiller, D.A.; Losos, J.B. Variable ecological effects of hurricanes: The importance of seasonal timing for survival of lizards on Bahamian Islands. *Proc. Natl. Acad. Sci. USA* **2004**, *101*, 177–181. [[CrossRef](#)]
12. Ameica y Juárez, E.I.; Mace, G.M.; Cowlshaw, G.; Cornforth, W.A.; Pettorelli, N. Assessing exposure to extreme climatic events for terrestrial mammals. *Conserv. Lett.* **2013**, *6*, 145–153. [[CrossRef](#)]
13. Ameica, E.I.; Mace, G.M.; Cowlshaw, G.; Pettorelli, N. Relative vulnerability to hurricane disturbance for endangered mammals in Mexico: A call for adaptation strategies under uncertainty. *Anim. Conserv.* **2019**, *22*, 262–273. [[CrossRef](#)]
14. Fenton, M.B.; Acharya, L.; Audet, D.; Hickey, M.B.C.; Merriman, C.; Obrist, M.K.; Syme, D.M. Phyllostomid bats (Chiroptera: Phyllostomidae) as indicators of habitat disruption in the Neotropics. *Biotropica* **1992**, *24*, 440–446. [[CrossRef](#)]
15. Jones, G.; Jacobs, D.S.; Kunz, T.H.; Willig, M.R.; Racey, P.A. Carpe noctem: The importance of bats as bioindicators. *Endanger. Spec. Res.* **2009**, *8*, 93–115. [[CrossRef](#)]
16. Buchalski, M.R.; Fontaine, J.B.; Heady, P.A.; Hayes, J.P.; Frick, W.F. Bat response to differing fire severity in mixed-conifer forest California, USA. *PLoS ONE* **2013**, *8*, e57884. [[CrossRef](#)] [[PubMed](#)]

17. Holbech, L.H. Differential responses of bats and non-volant small mammals to habitat disturbances in two tropical forest types of southwest Ghana. In *Animal Diversity, Natural History and Conservation*; Gupta, V.K., Verma, A.K., Eds.; Daya Publishing House: New Delhi, India, 2013; Volume 2, pp. 273–297. [[CrossRef](#)]
18. Jones, K.; Barlow, K.; Jennings, N.; Rodríguez-Durán, A.; Gannon, M.R. Short-term impacts of extreme environmental disturbance on the bats of Puerto Rico. *Anim. Conserv.* **2001**, *4*, 59–66. [[CrossRef](#)]
19. Meyer, C.F.J.; Fründ, J.; Pineda, L.W.; Kalko, E.K.V. Ecological correlates of vulnerability to fragmentation in Neotropical bats. *J. Appl. Ecol.* **2008**, *45*, 381–391. [[CrossRef](#)]
20. Sil-Berra, L.M.; Sánchez-Hernández, C.; Romero-Almaraz, M.L.; Reynoso, V.H. Vulnerability to natural disturbance in communities of Neotropical bats: Short-term impact of Hurricane Patricia on the Mexican Pacific Coast. *For. Ecol. Manag.* **2021**, *479*, 118596. [[CrossRef](#)]
21. Gavito, M.E.; Sandoval-Pérez, A.L.; del Castillo, K.; Cohen-Salgado, D.; Colarte-Avilés, M.E.; Mora, F.; Santibáñez-Rentería, A.; Siddique, I.; Urquijo-Ramos, C. Resilience of soil nutrient availability and organic matter decomposition to hurricane impact in a tropical dry forest ecosystem. *For. Ecol. Manag.* **2018**, *426*, 81–90. [[CrossRef](#)]
22. Jiménez-Rodríguez, D.L.; Álvarez-Añorve, M.Y.; Pineda-Cortés, M.; Flores-Puerto, J.I.; Benítez-Malvido, J.; Oyama, K.; Ávila-Cabadilla, L.D. Structural and functional traits predict short term response of tropical dry forests to a high intensity hurricane. *For. Ecol. Manag.* **2018**, *426*, 101–114. [[CrossRef](#)]
23. Martínez-Yrizar, A.; Jaramillo, V.J.; Maass, M.; Búrquez, A.; Parker, G.; Alvarez-Yépiz, J.C.; Araiza, S.; Verduzco, A.; Sarukhán, J. Resilience of tropical dry forest productivity to two hurricanes of different intensity in western Mexico. *For. Ecol. Manag.* **2018**, *426*, 53–60. [[CrossRef](#)]
24. Parker, G.; Martínez-Yrizar, A.; Álvarez-Yépiz, J.C.; Maass, M.; Araizad, S. Effects of hurricane disturbance on a tropical dry forest canopy in western Mexico. *For. Ecol. Manag.* **2018**, *426*, 39–52. [[CrossRef](#)]
25. Martínez-Ruiz, M.; Renton, K. Habitat heterogeneity facilitates resilience of diurnal raptor communities to hurricane disturbance. *For. Ecol. Manag.* **2018**, *426*, 134–144. [[CrossRef](#)]
26. Tapia-Palacios, M.A.; García-Suárez, O.; Sotomayor-Bonilla, J.; Silva-Magaña, M.A.; Pérez-Ortíz, G.; Espinosa-García, A.C.; Ortega-Hurtado, M.A.; Díaz-Ávalos, C.; Suzán, G.; Mazari-Hiriart, M. Abiotic and biotic changes at the basin scale in a tropical dry forest landscape after Hurricanes Jova and Patricia in Jalisco, Mexico. *For. Ecol. Manag.* **2018**, *426*, 18–26. [[CrossRef](#)]
27. INEGI, Instituto Nacional de Estadística y Geografía. *Conjunto de Datos Vectoriales de la Carta de Uso de Suelo y Vegetación, Serie VI. Conjunto Nacional. Escala 1:250,000*; Instituto Nacional de Estadística y Geografía: Mexico City, Mexico, 2017.
28. Ceballos, G.; Szekely, A.; García, A.; Rodríguez, P.; Noguera, F. *Programa de Manejo de la Reserva de la Biosfera Chamela-Cuixmala*; Instituto Nacional de Ecología, SEMARNAP: Mexico City, Mexico, 1999; 141p.
29. Bullock, S.H. Climate of Chamela, Jalisco, and trends in the South Coastal Region of Mexico. *Arch. Meteorol. Geophys. Bioclimatol. Ser. B* **1986**, *36*, 297–316. [[CrossRef](#)]
30. Stoner, K.E. Phyllostomid bat community structure and abundance in two contrasting tropical dry forest. *Biotropica* **2005**, *37*, 591–599. [[CrossRef](#)]
31. Bullock, S.H.; Solís-Magallanes, J.A. Phenology of canopy trees of a tropical deciduous forest in Mexico. *Biotropica* **1990**, *22*, 22–35. [[CrossRef](#)]
32. Álvarez, T.; Álvarez-Castañeda, S.T.; López-Vidal, C.J. *Claves para Murciélagos Mexicanos*; Centro de Investigaciones Biológicas del Noroeste, S.C.: La Paz, Mexico; Escuela Nacional de Ciencias Biológicas, Instituto Politécnico Nacional: Mexico City, Mexico, 1995; 65p.
33. Sánchez-Hernández, C.; Romero-Almaraz, M.L.; Schnell, G.D.; Kenedy, M.L.; Best, T.L.; Owen, R.D.; González-Pérez, S.B. *Bats of Colima, Mexico*; University of Oklahoma Press: Norman, OK, USA, 2016; 321p.
34. Pavan, A.C.; Marroig, G. Integrating multiple evidences in taxonomy: Species diversity and phylogeny of mustached bats (Mormoopidae: *Pteronotus*). *Mol. Phyl. Evol.* **2016**, *103*, 184–198. [[CrossRef](#)]
35. Pavan, A.C.; Marroig, G. Timing and patterns of diversification in the Neotropical bat genus *Pteronotus* (Mormoopidae). *Mol. Phyl. Evol.* **2017**, *108*, 61–69. [[CrossRef](#)]
36. Baird, A.B.; Braun, J.K.; Mares, M.A.; Morales, J.C.; Patton, J.C.; Tran, C.Q.; Bickham, J.W. Molecular systematic revision of tree bats (Lasiurini): Doubling the native mammals of the Hawaiian Islands. *J. Mammal.* **2015**, *96*, 1255–1274. [[CrossRef](#)]
37. Calahorra-Oliart, A.; Ospina-Garcés, S.M.; León-Paniagua, L. Cryptic species in *Glossophaga soricina* (Chiroptera: Phyllostomidae): Do morphological data support molecular evidence? *J. Mammal.* **2021**, *102*, 54–68. [[CrossRef](#)]
38. Chao, A.; Ma, K.H.; Hsieh, T.C. iNEXT (iNterpolation and EXTrapolation). 2016. Available online: http://chao.stat.nthu.edu.tw/wordpress/software_download/ (accessed on 23 February 2021).
39. Chao, A.; Ma, K.H.; Hsieh, T.C.; Chiu, C.H. SpadeR (Species-Richness Prediction and Diversity Estimation in R). 2015. Available online: http://chao.stat.nthu.edu.tw/wordpress/software_download/ (accessed on 23 February 2021).
40. Chao, A.; Chazdon, R.L.; Colwell, R.K.; Shen, T.-J. A new statistical approach for assessing similarity of species composition with incidence and abundance data. *Ecol. Lett.* **2005**, *8*, 148–159. [[CrossRef](#)]
41. R Core Team. *R: A Language and Environment for Statistical Computing*; R Foundation for Statistical Computing: Vienna, Austria, 2016; Available online: <http://www.R-project.org> (accessed on 31 August 2022).
42. Mann, H.B. Non-parametric tests against trend. *Econometrica* **1945**, *13*, 163–171. [[CrossRef](#)]
43. Kendall, M.G. *Rank Correlation Methods*, 4th ed.; Charles Griffin: London, UK, 1975.

44. Gilbert, R.O. *Statistical Methods for Environmental Pollution Monitoring*; Wiley: New York, NY, USA, 1987.
45. Addinsoft. *XLSTAT Statistical and Data Analysis Solution*; Addinsoft: New York, NY, USA, 2021; Available online: <https://www.xlstat.com> (accessed on 22 March 2021).
46. Douglas, E.M.; Vogel, R.M.; Kroll, C.N. Trends in floods and low flows in the United States: Impact of spatial correlation. *J. Hydrol.* **2000**, *240*, 90–105. [[CrossRef](#)]
47. Álvarez-Yépiz, J.C.; Martínez-Yrizar, A.; Fredericksen, T.S. Special issue: Resilience of tropical dry forests to extreme disturbance events. *For. Ecol. Manag.* **2018**, *426*, 1–6. [[CrossRef](#)]
48. Pierson, E.; Elmqvist, T.; Rainey, W.; Cox, P. Effects of tropical cyclonic storms on flying fox populations on the South Pacific Islands of Samoa. *Conserv. Biol.* **1996**, *10*, 438–451. [[CrossRef](#)]
49. Kunz, T.H.; Lumsden, L.F. Ecology of cavity and foliage roosting bats. In *Bat Ecology*; Kunz, T.H., Fenton, M.B., Eds.; University of Chicago Press: Chicago, IL, USA, 2003; pp. 3–89. [[CrossRef](#)]
50. Johnson, A.B.; Winker, K. Short-term hurricane impacts on a neotropical community of marked birds and implications for early-stage community resilience. *PLoS ONE* **2010**, *5*, e15109. [[CrossRef](#)]
51. García-Estrada, C.; Damon, A.; Sánchez-Hernández, C.; Soto-Pinto, L.; Ibarra-Núñez, G. Diets of frugivorous bats in montane rain forest and coffee plantations in southeastern Chiapas, Mexico. *Biotropica* **2012**, *44*, 394–401. [[CrossRef](#)]
52. Lou, S.; Yurrita, C.L. Análisis de nicho alimentario en la comunidad de murciélagos frugívoros de Yaxhá, Petén, Guatemala. *Act. Zool. Mex.* **2005**, *21*, 83–94. [[CrossRef](#)]
53. Eusebio-Valdes, G. Análisis de la Hipótesis de Estratificación Vertical para la Coexistencia de Murciélagos Frugívoros del Bosque Tropical Seco. Master's Thesis, Universidad Michoacana de San Nicolás de Hidalgo, Morelia, Mexico, 2019.
54. García-García, J.L.; Santos-Moreno, A. Efectos de la estructura del paisaje y de la vegetación en la diversidad de murciélagos filostómidos (Chiroptera: Phyllostomidae) de Oaxaca, México. *Rev. Biol. Trop.* **2014**, *62*, 217–239. [[CrossRef](#)] [[PubMed](#)]
55. Sojininen, J.; McDonald, R.; Hillebrand, H. The distance decay of similarity in ecological communities. *Ecography* **2007**, *30*, 3–12. [[CrossRef](#)]
56. Stoner, K.E.; Salazar, K.A.O.; Fernández, R.C.R.; Quesada, M. Population dynamics, reproduction, and diet of the lesser long-nosed bat (*Leptonycteris curasoae*) in Jalisco, Mexico: Implications for conservation. *Biodivers. Conserv.* **2003**, *12*, 357–373. [[CrossRef](#)]
57. Renton, K.; Salinas-Melgoza, A.; Rueda-Hernández, R.; Vázquez-Reyes, L.D. Differential resilience to extreme climate events of tree phenology and cavity resources in tropical dry forest: Cascading effects on a threatened species. *For. Ecol. Manag.* **2018**, *426*, 164–175. [[CrossRef](#)]
58. Sánchez-Hernández, C. Los murciélagos de la Estación de Investigación, Experimentación y Difusión “Chamela”, Jalisco, México. In *Actas de la II Reunión Iberoamericana de Conservación y Zoología de Vertebrados*; Castroviejo, J., Ed.; Estación Biológica de Doñana: Cáceres, Spain, 1984; pp. 385–398.
59. Ramírez-Priego, N. Estudio de los Hábitos Alimentarios del Murciélago *Artibeus jamaicensis* Mediante la Determinación de Variaciones Estacionales en su Composición Isotópica de Carbono y Nitrógeno en la Bahía de Chamela, Jalisco. Bachelor's Thesis, Universidad Nacional Autónoma de México, Mexico City, Mexico, 2000. Available online: <https://repositorio.unam.mx/contenidos/337058> (accessed on 1 April 2022).
60. Díaz-Camacho, W.G. Identificación de los Hábitos Alimenticios del Murciélago *Dermanura phaeotis* (Chiroptera: Phyllostomidae) por medio de Isotopos Estables de Nitrógeno y Carbono en Los Tuxtlas, Veracruz. Bachelor's Thesis, Universidad Nacional Autónoma de México, Mexico City, Mexico, 2004. Available online: <https://repositorio.unam.mx/contenidos/291225> (accessed on 1 April 2022).
61. Zarazúa-Carbajal, M.; Ávila-Cabadilla, L.D.; Álvarez-Añorve, M.Y.; Benítez-Malvido, J.; Stoner, K.E. Importance of riparian habitat for frugivorous bats in a tropical dry forest in Western Mexico. *J. Trop. Ecol.* **2017**, *33*, 74–82. [[CrossRef](#)]
62. Duryea, M.; Kampf, E. *Wind and Trees: Lesson Learned from Hurricanes*; University of Florida: Gainesville, FL, USA, 2007.
63. Aranda-Pineda, J.A.; Fernández-Muñoz, T.; Marten-Rodríguez, S.; Quesada-Avendaño, M. Size matters? Study of the damage provoked by a hurricane into a population of an endemic columnar cactus of Mexico. *Cactáceas Suculentas Mex.* **2016**, *61*, 85–95.
64. Aldridge, H.D.J.N.; Brigham, R.M. Factors influencing foraging time in two aerial insectivores: The bird *Chordeiles minor* and the bat *Eptesicus fuscus*. *Can. J. Zool.* **1991**, *69*, 62–69. [[CrossRef](#)]
65. Rydell, J. Variation in foraging activity of an aerial insectivorous bat during reproduction. *J. Mammal.* **1993**, *74*, 503–509. [[CrossRef](#)]
66. Wilkinson, L.C.; Barclay, R.M.R. Differences in the foraging behaviour of male and female big brown bats (*Eptesicus fuscus*) during the reproductive period. *Écoscience* **1997**, *4*, 279–285. [[CrossRef](#)]
67. García-Oliva, F.; Camou, A.; Maass, J.M. El clima de la región central de la costa del Pacífico mexicano. In *Historia Natural de Chamela*; Noguera, F.A., Vega-Rivera, J.H., García-Aldrete, A.N., Quesada-Avendaño, M., Eds.; Instituto de Biología, Universidad Nacional Autónoma de México: Mexico City, Mexico, 2002; pp. 3–10.
68. Sánchez-Casas, N.; Álvarez, T. Palinofagia de los murciélagos del género *Glossophaga* (Mammalia: Chiroptera) en México. *Acta Zool. Mex.* **2000**, *81*, 23–62. [[CrossRef](#)]
69. Williams-Guillén, K.; Perfecto, I. Ensemble composition and activity levels of insectivorous bats in response to management intensification in coffee agroforestry systems. *PLoS ONE* **2011**, *6*, e16502. [[CrossRef](#)]
70. Oliveira, L.; Marciente, R.; Magnusson, W.; Bobrowiec, P.; Bobrowiec, D. Activity of the insectivorous bat *Pteronotus parnellii* relative to insect resources and vegetation structure. *J. Mammal.* **2015**, *96*, 1036–1044. [[CrossRef](#)]

71. Salinas-Ramos, V.B.; Herrera-Moltalvo, L.G.; León-Regagnon, V.; Arrizabalaga-Escudero, A.; Clare, E.L. Dietary overlap and seasonality in three species of mormoopid bats from a tropical dry forest. *Mol. Ecol.* **2015**, *24*, 5296–5307. [[CrossRef](#)] [[PubMed](#)]
72. Lanzagorta-Valencia, K.; Fernández-Méndez, J.I.; Medellín, R.A.; Rodas-Martínez, A.Z.; Ávila-Flores, R. Landscape and cattle management attributes associated with the incidence of *Desmodus rotundus* attacks on cattle. *Ecosistemas Recur. Agropecu.* **2019**, *7*, 1–10. [[CrossRef](#)]
73. Sánchez-Cordero, V.; Botello, F.; Magaña-Cota, G.; Iglesias, J. Vampire bats, *Desmodus rotundus*, feeding on white-tailed deer, *Odocoileus virginianus*. *Mammalia* **2011**, *75*, 91–92. [[CrossRef](#)]
74. Trajano, E. Movements of cave bats in southeastern Brazil, with emphasis on the population ecology of the common vampire bat, *Desmodus rotundus* (Chiroptera). *Biotropica* **1996**, *28*, 121–129. [[CrossRef](#)]
75. Zortéa, M.; Silva, D.A.; Calaça, A.M. Susceptibility of targets to the vampire bat *Desmodus rotundus* are proportional to their abundance in Atlantic Forest fragments? *Iheringia Série Zool.* **2018**, *108*, e2018037. [[CrossRef](#)]
76. Ayala-Berdon, J.; Schondube, J.E.; Stoner, K.E. Seasonal intake responses in the nectar-feeding bat *Glossophaga soricina*. *J. Comp. Physiol. B* **2009**, *179*, 553–562. [[CrossRef](#)]
77. Schmidt, C. Reproduction. In *Natural History of Vampire Bats*; Greenhall, A.M., Schmidt, U., Eds.; CRC Press, Inc.: Boca Raton, FL, USA, 1988; pp. 99–109.
78. Lord, R. Seasonal reproduction of vampire bats and its relation to seasonality of bovine rabies. *J. Wildl. Dis.* **1992**, *28*, 292–294. [[CrossRef](#)]
79. Núñez, H.A.; de Viana, M.L. Estacionalidad reproductiva en el vampiro común *Desmodus rotundus* (Chiroptera, Phyllostomidae) en el Valle de Lerma (Salta, Argentina). *Rev. Biol. Trop.* **1997**, *45*, 1231–1235.
80. Chávez, C.; Ceballos, G. Diversidad y abundancia de murciélagos en selvas secas de estacionalidad contrastante en el oeste de México. *Rev. Mex. Mastozool.* **2001**, *5*, 27–44. [[CrossRef](#)]
81. Stoner, K.E. Murciélagos nectarívoros y frugívoros del Bosque Tropical Caducifolio de la Reserva de la Biosfera Chamela-Cuixmala. In *Historia Natural de Chamela*; Noguera, F.A., Vega-Rivera, J.H., García-Aldrete, A.N., Quesada-Avedaño, M., Eds.; Instituto de Biología, Universidad Nacional Autónoma de México: Mexico City, Mexico, 2002; pp. 443–472.
82. Ávila-Cabadilla, L.D.; Stoner, K.E.; Henry, M.I.; Álvarez, M.Y. Composition, structure, and diversity of phyllostomid bat assemblages in different successional stages of a tropical dry forest. *For. Ecol. Manag.* **2009**, *258*, 986–996. [[CrossRef](#)]
83. Sikes, R.S.; The Animal Care and Use Committee of the American Society of Mammalogists. 2016 Guidelines of the American Society of Mammalogists for the use of wild mammals in research and education. *J. Mammal.* **2016**, *97*, 663–688. [[CrossRef](#)] [[PubMed](#)]