

Article

Exploring the Genetic Diversity and Population Structure of *Daphnia cucullata* Sars, 1862 in Boreal Lakes (Latvian Lakeland) Based on Microsatellites

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Abstract: We have used *Daphnia cucullata* (Crustacea: Cladocera) as a model organism for the first time in the four deepest Latvian lakes from the Boreal biogeographical region in order to find the genetic diversity of these populations. During the research, we detected the most appropriate microsatellite markers for future genetic studies of *Daphnia cucullata* populations of lakes Svente, Riča, Dridzis and Gerāņimovas-Ilzas in the Boreal biogeographical region. Based on these microsatellite markers, we determined the genetic diversity of these populations. The loci Dgm105 and Dgm101 had the maximum number of alleles and the maximum number of private alleles. The specific locus Dgm105 had five private alleles (62% of all detected alleles), and locus Dgm101 had four private alleles (57% of all detected alleles) in these loci. We determined the observed heterozygosity (H_{obs}) and the expected heterozygosity (H_{exp}) level (via Hardy–Weinberg equilibrium), the number of polymorphic loci, the number of detected alleles in each analyzed microsatellite locus, the average number of alleles at the locus (N_a), the average effective number of alleles at the locus (N_e), the F_{ST} of the population's genetic differentiation, the genetic distance (D) (following Nei) and the significance (χ^2 -test) of differences between the levels of observed and expected heterozygosity. It was shown that *Daphnia cucullata* populations from lakes with a low number of zooplankton taxa (Riča and Gerāņimovas-Ilzas) have a higher genetic diversity compared to lakes with a high number of zooplankton taxa (Dridzis and Svente). It was found that *Daphnia cucullata* populations from lakes Dridzis and Svente have the least genetic distance, and these populations form a single genetic group, as confirmed via clustering.

Keywords: Cladocera; *Daphnia cucullata*; genetic diversity; microsatellite-PCR; Boreal biogeographical region lakes



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1. Introduction

Zooplankton (e.g., Cladocera) plays an essential role in the transformation of substances and energy in water bodies and is an important stage in the food chain. Zooplankton representatives regulate bacterial and detrital quantity, and they are an important component of the feed for juvenile fish, adult plankton-feeding fish and many other aquatic animals [1–4]. Some Cladocera species, such as *Daphnia magna*, *Daphnia pulex* and *Daphnia cucullata*, are often used as bioindicators of water pollution and a good model organism of freshwater ecology. Research from Poland suggests that the body size of *Daphnia cucullata* seems to be a useful indicator of the ecological status of lakes [5–11]. *Daphnia* has been considered to be a control organism in freshwater as a kind of convergence model with adaptive features in radically different habitats [12,13]. To understand the adaptive processes in *Daphnia* populations under changing environmental conditions, investigations of the genetic structure of the population are necessary [14–18]. Cladocera are dominant in

continental water bodies in all climatic zones, and also in lakes Svente, Riča, Dridzis and Geranišimovas-Ilzas of the Daugava River basin [19–24].

Microsatellite markers are useful for population genetics studies because they typically have a high allelic variation within and between populations, thus increasing the probability of distinguishing between populations and detecting changes over time [25]. Also, microsatellites are neutral markers (in the non-coding regions of the genome) and do not affect the fitness of the organism. This provides an opportunity to understand gene flow and to analyze the degree of connection among contemporary populations [26] or of populations through time from different waterbodies [27–30]. The microsatellite analysis method can be used to determine the genetic variability and genetic monitoring of populations of different Cladoceran species as well [14,27,31–34]. Microsatellite loci are frequently used in studies of the genetic structure of different species in the *Daphnia* genus (*Daphnia pulex*, *Daphnia magna*, *Daphnia longispina*, *Daphnia galeata*, *Daphnia hyalina*, *Daphnia rosea*, and *Daphnia curvirostris*) [14,15,27,30,31,35–38]. But there are relatively few studies investigating the genetic structure of *Daphnia cucullata* [14]. However, this study was in laboratory conditions.

Therefore, the aim of this study to describe for the first time the genetic structure of native *Daphnia cucullata* populations in the four deepest Latvian lakes of the Boreal zone using microsatellite markers.

2. Material and Methods

2.1. Sampling Sites and Material Collection of *Daphnia cucullata*

The material for our study was collected in four deep lakes, which belong to the deep, well-transparent mesotrophic and mesoeutrophic Latvian lakes of the Daugava River basin. Lakes Dridzis, Svente and Riča are mesotrophic, but Lake Geranišimovas-Ilzas is mesoeutrophic [39]. Samples were taken in the deepest zone in each lake.

These lakes are relatively similar in terms of their morphometric characteristics. Lake Dridzis is the deepest lake in Latvia and in Baltia [40]. Lake Geranišimovas-Ilzas is the fifth deepest, but Lake Svente is the tenth deepest lake in Latvia [40–42], and Lake Riča is the ninth deepest Lake in Latvia [43]. The investigated lake characteristics and locations are presented in Figure 1 and Table 1. Geographical data positions were obtained via echo sounders with GPS receiver LOWRANCE LMS-522C and TRIMBLE Juno SB.

Table 1. Geographical, hydrological and morphological parameters of investigated lakes.

Lakes	Coordinates X/Y	Elevation of Lakes above Sea Level, m	Surface Area with Island, km ²	Surface Area without Island, km ²	Max. Depth, m *	Mean Depth, m *	Catchment Basin, km ²	Shore Length, km
Dridzis	705,390.852/ 208,462.077	159.8	7.72	7.56	64	12.8	46	42
Riča	670,715.594/ 175,721.067	145.8	13.12	13.07	39	9.7	123 */130 **	34
Svente	647,412.511/ 192,388.091	136.9	7.06	7.03	38	7.8	20	26
Geranišimovas-Ilzas	696,251.015/ 228,167.042	150.7	3.17	3.17	46	9.8	66	24

* Catchment basin in the territory of Latvia. ** Catchment basin in the territory of Belarus.

Zooplankton samples were taken with a *Hydro-bios* Apstein-type plankton net with an opening–closing mechanism (mesh size: 64 µm) to filter the water column, which was taken from the deep water to the surface. The collection of the zooplankton samples was performed using the APHA standard methods procedure [44,45]. The collected samples were preserved immediately after collection through adding 98% ethanol to the water sample; hence, the final concentration in the sample is ±70%. Samples were stored at −20 °C. After splitting, the collected specimens of *D. cucullata* were stored in 96% ethanol.

The samples had to be preserved immediately after harvesting to prevent individuals from biochemical and molecular degradation [17].



Figure 1. Location of lakes Svente, Riča, Dridzis and Geranimovas-Ilzas (map author: E. Iliško; a basemap was used from Open Sources).

2.2. Genetic Analysis

2.2.1. DNA Extraction

Genomic DNA extraction from adult *Daphnia cucullata* individuals (20–30 specimens from each population) was performed using slightly modified ‘salting out’ extraction methodology earlier described by Fitzsimmons and Innes [46]. The method consists of the following steps: Zooplankton samples were transferred to 1.5 mL reaction tubes containing 100 μ L of buffer A (100 mM Tris–HCl (pH 7.5), 100 mM ethylenediaminetetraacetic acid (EDTA), 100 mM NaCl and 0.5% SDS). Tubes were incubated at 70 °C for 35 min. Two hundred microliters of LiCl–KAc solution (one part 5 M KAc by volume with 2.5 parts 6 M LiCl) was added before tubes were incubated on ice for 15–20 min. Samples were spun at 13,700 \times g for 15 min. Supernatant was transferred into new tubes. One hundred and sixty microliters of cold (–20 °C) isopropanol was added, and the sample was mixed and then spun for 15 min. We aspirated away the supernatant via vacuum, spun, and then aspirated the remaining liquid. Samples were washed twice with cold (4 °C) 70% ethanol, being spun for 2 min before the supernatant was aspirated away each time. DNA was resuspended in 35 μ L of double-distilled water and left at 4 °C overnight [46,47].

2.2.2. Determination of the Quantity and Quality of Isolated DNA

The concentration of DNA samples and quantity, quality and suitability for PCR were determined using the spectrophotometer BioSpec- Nano (Shimadzu, Japan). The dry DNA samples were dissolved in dd H₂O for quantifying DNA. The ratio of absorbance at 260 and 280 nm ($A_{260}/A_{280} > 1.8$) and A_{260}/A_{230} were used to assess the purity of nucleic acids. The quality and suitability of the isolated DNA samples for PCR were checked on 1.5% agarose gel [17,47] with ethidium bromide.

2.2.3. Microsatellites Analysis

We used nine primers for the microsatellite loci of nuclear DNA (DaB10/15; Dp512; Dp519; DaB17/16; DaB17/17; SwiD1; Dgm101; Dgm105; Dgm109) [14] for genetic research on the Latvian *Daphnia cucullata* population, but six microsatellite primers with good representativity (SwiD1; Dgm105; Dgm101; DaB17/17; Dgm109; Dp519) were selected for the analysis. Three of them were dinucleotide microsatellite primers (SwiD1; Dgm101; Dp519), and two were trinucleotide microsatellite primers (Dgm105; Dgm109).

Microsatellite amplification was performed using the Eppendorf Mastercycler[®] proS (Eppendorf, Hamburg, Germany) automated polymerase chain reaction (PCR) system (Version 2.1, Free Software Foundation). PCR was performed in 10 µL for 0.2 mL PCR tubes. The PCR mixture components were 2.3 µL of genomic DNA sample (20 ng) in Dilution buffer (Thermo Fisher Scientific Baltics, Vilnius, Lithuania), 5 µL 2 × Phire Animal Tissue PCR Buffer (with dNTPs and MgCl₂) (Thermo Fisher Scientific Baltics, Vilnius, Lithuania), 0.2 µL Phire Hot Start II DNA polymerase (Thermo Fisher Scientific Baltics, Vilnius, Lithuania), 1.25 µL–8 µmol/µL primers F, 1.25 µL–8 µmol/µL primers R. Primers were obtained with fluorescently-4-labeled TMR, HEX and FAM. PCR was performed using the thermal cycling program under the following amplification cycle: denaturation at 98 °C for 5 min and 40 cycles at 98 °C for 5 s (denaturation), X °C or 55 °C for 10 s (solicitation or primer annealing), 72 °C for 20 s (synthesis), 72 °C for 1 min and 4 °C (cooling). Amplification was repeated three times with each primer, including a positive and negative control. After the PCR amplification, the products were maintained at 4 °C until analytical separation using the automated sequencer GeneScan[®] Analysis ABI PRISM 3100 (Applied Biosystems, Foster City, CA, USA) as the international size standard.

Based on the obtained results, for future genetic diversity studies of *D. cucullata* populations, we selected only primers that provide high levels of good amplifications and informative DNA fragments.

2.2.4. Statistical Processing and Analysis of the Obtained Data

The obtained data were processed and analyzed using the computer software GeneAlex 6.41 [48] and POPGENE 1.32 [49]. The number of alleles per locus, the frequency, the private alleles in each population [50] Nei, 1987 and the observed (H_{obs}) and expected (H_{exp}) heterozygosity level in polymorphic loci (according to Hardy–Weinberg) [51] were measured, and their differences among *Daphnia cucullata* individuals from different sampling places were calculated, and their significance with χ^2 criteria was calculated using GeneAlex 6.41 and POPGENE 1.32. The genetic relatedness of *D. cucullata* populations was estimated with genetic distance (D) [52]. Genetic differentiation among populations was estimated via principal component analysis (PCA) and pairwise F_{ST} values [50]. To estimate and visualizes the genetic structure and differentiation of the studied *D. cucullata* populations, STRUCTURE 2.3.4 [53] and STRUCTURE HARVESTER [54] were used.

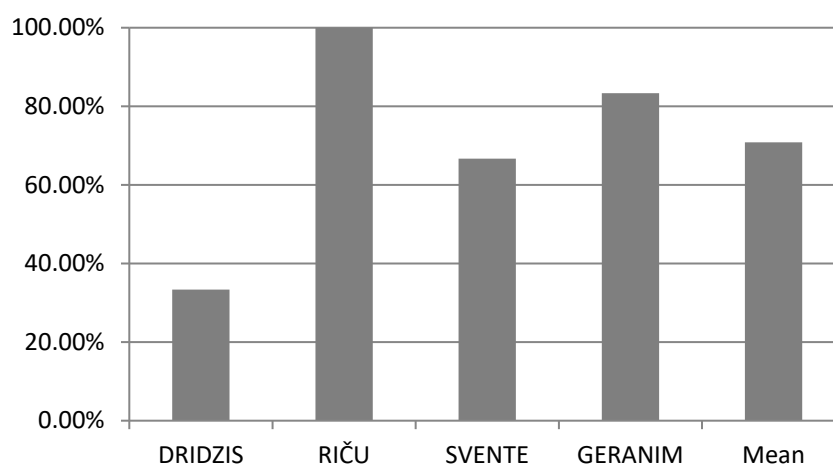
3. Results

The size of the scored polymorphic DNA fragments ranged from 122 bp to 303 bp (Table 2). The highest number of base pairs were found in loci Dgm109 (250–303 bp) and Dgm105 (165–240 bp), but the lowest were in loci DaB17/17 (100–106 bp) and SwiD1 (122–127 bp) (Table 2).

The average level of polymorphism in all studied *Daphnia cucullata* populations was the same and amounted to 100%, because all six analyzed microsatellite loci were polymorphic in all studied *D. cucullata* populations. The number of polymorphic loci of *D. cucullata* populations in the lakes Svente, Riča, Dridzis and Geraniimovas-Ilzas ranged from 33% to 100%. The lowest number of polymorphic microsatellite loci of the *D. cucullata* population was found in the Lake Dridzis (33%), while the highest number of polymorphic microsatellite loci was found in the lakes Riča (100%) and Geraniimovas-Ilzas (83%) (Figure 2).

Table 2. Characteristics of the 9 microsatellites: locus name, primer sequences, repeat motif, modification, fragment size range and annealing temperature (T_a).

Locus	Primer Sequences (5'–3')	Repeat Unit	Label Dye	Size Range (bp) (Our Data)	Size Range (bp) (Data after Brede et al.)	T (°C)
SwiD1	F:GCCGTGTTTCGAAAGCTAGTC R: AGCCGAACGAAAAACATGC	(TG) ₁₈	5'TAM	122–127	116–142	59.4
Dgm105	F:ATGTGAGCGCGGAGCATT R:GTCCAGCCGGCCATTTTCAGTT	(CAG) ₈ AG	5'FAM	165–240	172–197	59.4
Dgm101	F: TCTTGCTCGAATTCTCTCC R: CCTGTCTCACACGGAGC	(GA) ₁₀ AGA	5'HEX	165–180	162–177	54.5
DaB17/17	F:GAGAACCTTTTATCAGCTTCG R:ACTCATCTGGTGAGATGGATC	T ₉	5'TAM	100–106	100–109	55.9
Dgm109	F: CCAGCTGTTGACCACCTG R: TGCGCGAGGATTTCCAACAC	(ACC) ₇ AC	5'FAM	250–303	247–266	58.2
Dp519	F:AGTCGCGACGACATAAAGC R:GTGGTAGTTGTGGAATCCG	(TG) ₆ (GA) ₇	5'HEX	140–142	144–160	56.7
DaB10/15	F:AGAGAAGTGTGCGTTTC R:TGTTTCTATATCCCTCGG	TC ₆	5'TAM	No result	75–89	52.4
Dp512	F:TTTCGTTCTACCCAGGGAAG R:TTTGCTCGTCTGTGATAGGC	(TG) ₄ ...(GT) ₈	5'HEX	No result	125–141	57.3
DaB17/16	F: AGGGAACGAGCGGCGATAAG R:TCTTTGGCAGGCCACTGCCAAGG	GA ₁₀	5'FAM	No result	189–195	61.4

**Figure 2.** Portion of polymorphic loci of *Daphnia cucullata* populations in lakes Svete, Riča, Dridzis and Geranimovas-Ilzas. GERANIM—Lake Geranimovas-Ilzas.

Analyzing the general parameters of the abundance of the obtained alleles (Table 3), it can be seen that the detected number of alleles in the investigated locus differs in each population. Also, the number of detected alleles in each analyzed microsatellite locus was different. The maximum number of alleles was found in the loci Dgm105 (eight alleles) and Dgm101 (seven alleles); moreover, it should be noted that the maximum number of private alleles was also found in these loci, where the Dgm105 locus had five private alleles (62% of all detected alleles), while the Dgm101 locus had four private alleles (57% of all detected alleles) in these loci (Table 3). On the other hand, the lowest number of alleles was found in locus Dp519 (two alleles). It is also characteristic that private alleles were not detected in this locus at all (Table 3).

Table 3. Abundance of alleles in the studied microsatellite loci.

Locus	Total Number of Alleles in the Locus	Number of Private Alleles in the Locus	Proportion of Private Alleles (%)	Number of Populations in Which Private Alleles Have Been Detected
SwiD1	5	1	20	1
Dgm105	8	5	62	3
Dgm101	7	4	57	3
DaB17/17	4	2	50	2
Dgm109	5	3	60	2
Dp519	2	0	0	0

Alleles in the investigated *D. cucullata* populations (Figure 3) show that the alleles' abundance was different, but this difference is not statistically significant at ($p > 0.05$). The largest number of detected alleles per locus was for the *D. cucullata* population of Lake Geraņimovas-Ilzas (3.17), followed by the *D. cucullata* population of Lake Riča (3.00). The number of detected alleles per locus is relatively lower for the *D. cucullata* populations of lakes Dridzis (1.67) and Svente (1.83) (Figure 3).

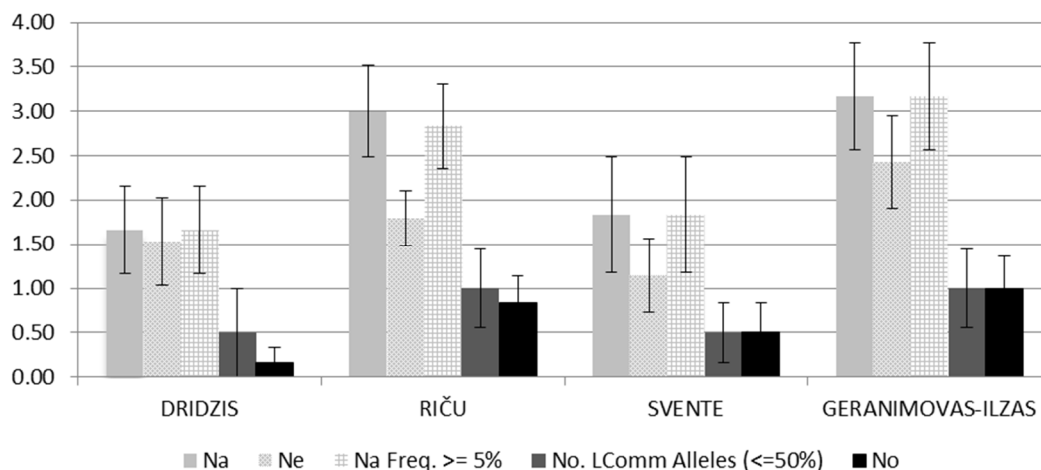


Figure 3. Allelic patterns across *Daphnia cucullata* populations between lakes Svente, Riča, Dridzis and Geraņimovas-Ilzas using microsatellites-PCR analysis (Na—the average number of alleles at the locus; Ne—the average effective number of alleles at the locus; $Na \geq 5\%$ —average number of alleles with an incidence of more than 5%; $No \leq 50\%$ —average number of alleles with an incidence of less than 50%; No—average number of private alleles; \pm standard deviation).

The average number of alleles per locus with a frequency of more than 5% is equal to the average number of alleles per locus in all studied *D. cucullata* populations (Figure 3). The average number of rare alleles per loci of the *D. cucullata* specimens under research, which was found to be less than 50%, is the same for the populations of lakes Geraņimovas-Ilzas and Riča, respectively—it is 1—while for the *D. cucullata* populations of lakes Dridzis and Svente, respectively, it is 0.5; but, overall, these differences are not significant ($p > 0.05$) (Figure 3).

The number of average effective alleles per locus differs significantly from the average observed number of alleles in *D. cucullata* populations of lakes Riča and Svente ($p < 0.05$), while these differences are not significant for the *D. cucullata* populations of lakes Dridzis and Geraņimovas-Ilzas ($p > 0.05$) (Figure 3).

The average level of the observed heterozygosity (H_o) was high in all studied *D. cucullata* populations, ranging from 1.67 to 3.17. The minimum value of H_o was 1.67 in Lake Dridzis, and the maximum was 3.17 in Lake Geraņimovas-Ilzas. Meanwhile, the average level of the expected heterozygosity (H_e) ranged from 1.15 to 2.43. The minimum

value of H_e was 1.15 in Lake Svente, and the maximum was 2.43 in Lake Geranımovas-Ilzas. In general, in all *D. cucullata* populations under research, the average observed level and the average expected level of heterozygosity (according to Hardy–Weinberg) was different, but these differences were insignificant ($p < 0.001$) (Figure 3).

The analyzed loci (SwiD1; Dgm105; Dgm101; DaB17/17; Dgm109; Dp519) were polymorphic in the investigated *D. cucullata* populations, and the level of polymorphism was very high. Genetic diversity across the studied *D. cucullata* samples found in each studied loci and each location are presented in Table 4. The greatest number of alleles (19) was found at loci DaB17/17 and Dp519 in the population of Lake Riča, and the minimum number of alleles (1) at locus Dgm101 was found in the population of Lake Dridzis. It should be noted that no alleles were detected at loci Dgm101 and Dgm109 in the population of Lake Svente. Private alleles were found at the loci SwiD1, Dgm105, Dgm101, DaB17/17 and Dgm109.

Table 4. Genetic diversity across studied *Daphnia cucullata* samples found in each studied locus and each studied lake.

Sample		SwiD1	Dgm105	Dgm101	DaB17/17	Dgm109	Dp519
Dridzis	N	4	4	1	14	4	14
	Na	4	1	1	2	1	1
	No	0	0	1	0	0	0
	Ho	0	0	0	0	0	0
	He	0.75	0	0	0.13	0	0
Riča	N	13	12	12	19	11	19
	Na	2	4	5	2	3	2
	No	1	1	2	0	1	0
	Ho	0	0.25	0	0	0	0
	He	0.14	0.51	0.68	0.46	0.31	0.1
Svente	N	4	8	0	15	0	16
	Na	3	4	0	2	0	2
	No	0	2	0	1	0	0
	Ho	0	0.25	0	0	0	0
	He	0.62	0.33	0	0.12	0	0.37
Geranimovas-Ilzas	N	7	6	7	14	6	8
	Na	2	5	4	3	4	1
	No	0	2	1	1	2	0
	Ho	0	0.17	0	0	0.17	0
	He	0.24	0.74	0.73	0.36	0.68	0

Na, the average number in a locus; Ne, the average effective number of alleles in a locus; No, the average number of private alleles; Ho, observed heterozygosity; He, expected heterozygosity.

A significant homozygote excess was observed in the *D. cucullata* population from Lake Dridzis at one locus, DaB17/17, $p < 0.001$; from Lake Riča at five loci (SwiD1, Dgm101, DaB17/17, Dgm109 and Dp519, $p < 0.001$); from Lake Svente at four loci (SwiD1 and Dgm105, $p < 0.05$; DaB17/17 and Dp519, $p < 0.001$) and from Lake Geranimovas-Ilzas at three loci (SwiD1 and Dgm101, $p < 0.01$; DaB17/17, $p < 0.001$) (Table 5). Microsatellite locus DaB17/17 has maximal differentiation ($p < 0.001$) between the level of observed and expected heterozygosity in all investigated lakes. In addition, microsatellite loci SwiD1, Dgm101, Dgm109, Dp519 and DaB17/17 have maximal differentiation ($p < 0.001$) in Lake Riča (Table 5). It should be noted that microsatellite loci Dgm105, Dgm101, Dgm109

and Dp519 were monomorphic in Lake Dridzis. Whereas microsatellite loci Dgm101 and Dgm109 were monomorphic in Lake Svente, microsatellite locus Dp519 also was monomorphic in Lake Geraņimovas-Ilzas (Table 5).

Table 5. Significance (χ^2 -test) of differences between the levels of observed and expected heterozygosity in studied *Daphnia cucullata* populations in lakes Svente, Riča, Dridzis and Geraņimovas-Ilzas.

Population/Microsatellite Loci	SwiD1	Dgm105	Dgm101	DaB17/17	Dgm109	Dp519
Dridzis	ns	M	M	***	M	M
Riča	***	ns	***	***	***	***
Svente	*	*	M	***	M	***
Geraņimovas-Ilzas	**	ns	**	***	ns	M

ns—not significant, M—monomorphic loci, * $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$.

The smallest genetic distance (D) [55] was observed between lakes Riča and Geraņimovas-Ilzas (0.16), while the greatest genetic distance was found between lakes Dridzis and Geraņimovas-Ilzas (0.70) and between lakes Geraņimovas-Ilzas and Svente (1.35) (Table 6).

Table 6. Genetic distance (D) [55] and genetic differentiation (after F_{ST} values) among *Daphnia cucullata* populations between lakes Svente, Riča, Dridzis and Geraņimovas-Ilzas.

Population	Dridzis	Riča	Svente	Geraņimovas-Ilzas
Dridzis		0.29	0.45	0.37
Riča	0.56		0.50	0.08
Svente	0.50	1.14		0.49
Geraņimovas-Ilzas	0.70	0.16	1.35	

genetic distance (D) values below diagonal; genetic differentiation (F_{ST} values) over diagonal.

By contrast, F_{ST} values for different *D. cucullata* populations under research ranged from 0.08 to 0.50. The highest values were between the *D. cucullata* populations of lakes Riča and Svente (0.50) and lakes Svente and Geraņimovas-Ilzas (0.49) (Table 6). The lowest F_{ST} values were between the *D. cucullata* populations of lakes Riča and Geraņimovas-Ilzas (0.08) (Table 6).

Principal component analysis (PCA), a graph of genetic structuring among four *D. cucullata* populations in lakes Svente, Riča, Dridzis and Geraņimovas-Ilzas, clearly showed the genetic structuring into different genetic groups (Figure 4). Stable groups of *D. cucullata* populations were formed between lakes Dridzis and Svente and between lakes Riča and Geraņimovas-Ilzas. In the principal component analysis plot, PC 1 and PC 2 explained 77.43% and 15.71% of the total genetic diversity. A similar result was obtained using Bayesian clustering analysis (STRUCTURE 2.3.4) [53] (Figure 5) and number of clusters of individuals using Evano et al.'s clustering approach (Figure 6).

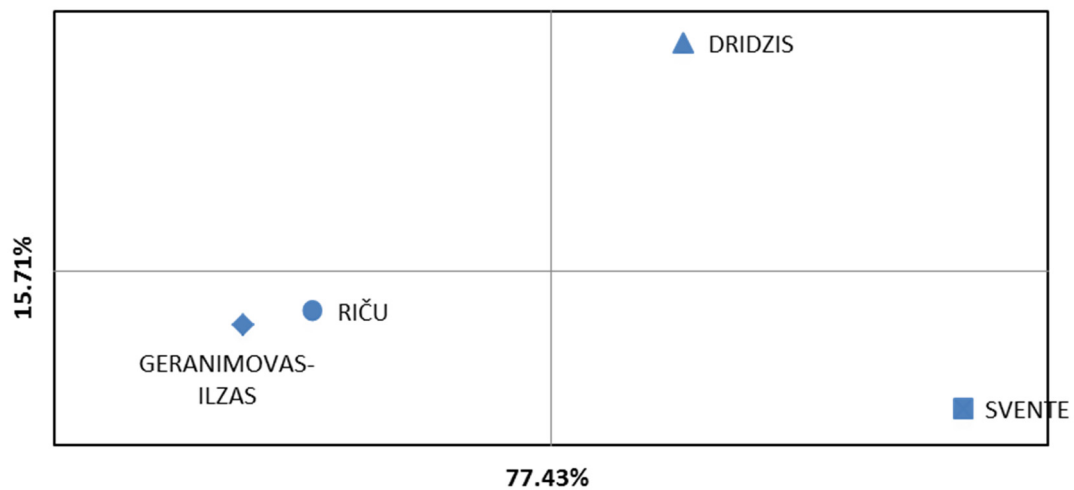


Figure 4. Principal component analysis (PCA). Plot of genetic structuring after Nei genetic distance data among *Daphnia cucullata* populations in lakes Svente, Riča, Dridzis and Geraņimovas-Ilzas.

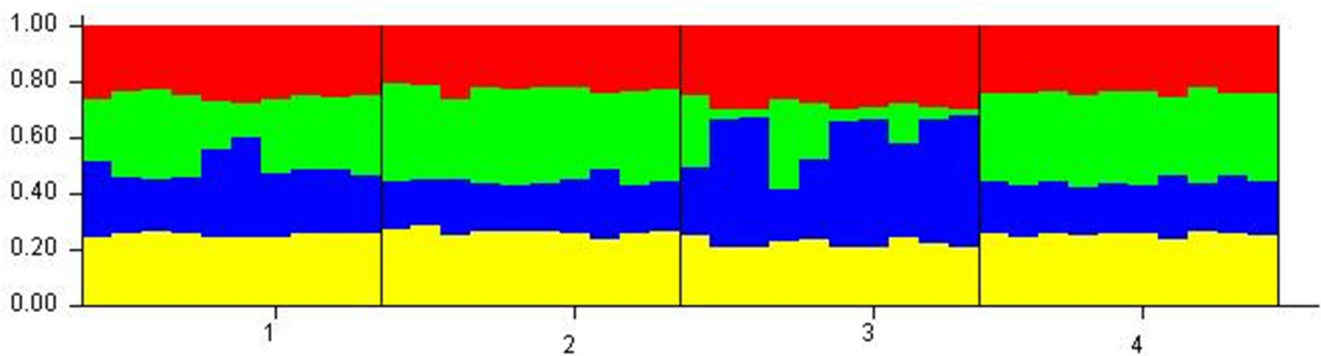


Figure 5. Bayesian clustering of individuals using STRUCTURE 2.3.4 [53] In the STRUCTURE analysis, color lines separate individuals from different sampling sites, and each individual is represented by a vertical line, which is partitioned into K-colored segments representing an individual’s estimated membership in K clusters (1—Dridzis, 2—Riča, 3—Svente, 4—Geraņimovas-Ilzas).

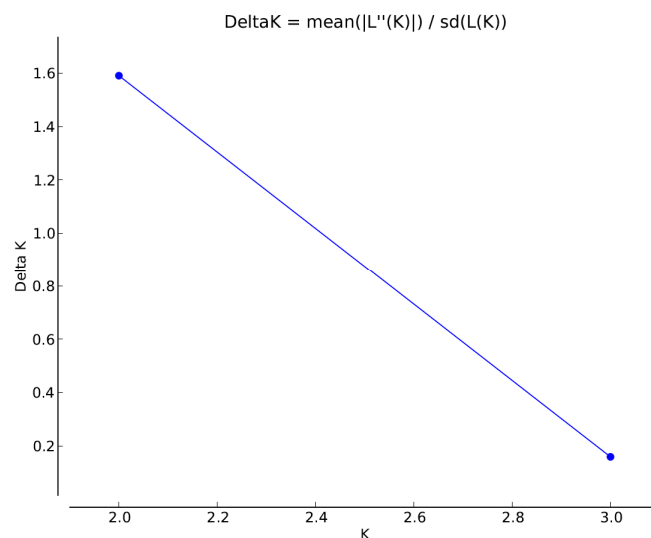


Figure 6. The number of clusters of individuals using Evano et al.’s clustering approach [56], assuming two genetic clusters ($K = 2$; $\Delta K = 1.59$; $\ln P(K) \pm SD = -298.76 \pm 78.73$).

4. Discussion

Daphnia cucullata is one of the most common specimens in the species composition of Lakeland (Latvia) in the Boreal biogeographical region of Europe [19,20,22,24,47,57]. The analysis of the genetic structure of a typical representative of the zooplankton, i.e., *Daphnia*, is important in terms of predicting the impact of anthropogenic and climatic factors both on zooplankton and aquatic ecosystems in general [14,27,31–34].

We used microsatellite loci, which are frequently used in studies of the genetic structure of other species in the *Daphnia* genus [14,27]. However, some primers that have been successfully used in the study of European populations have not been amplified in Latvian populations. For example, three of the microsatellite loci (DaB10/17; Dp512; DaB17/16), which were presented in *D. cucullata* populations from lakes in Switzerland and the Netherlands, did not appear in the *D. cucullata* population from these Latvian lakes. This may be indicated a significant difference between the genotypes of *D. cucullata* populations from Continental (Switzerland and Netherlands) and Boreal (Latvia) biogeographical regions in Europe.

The sizes of amplified loci in *D. cucullata* populations from Continental and Boreal (Latvia) regions were similar. A few microsatellite loci of *D. cucullata* populations in the Boreal region are generally slightly longer than those in the Continental region (for example, microsatellite loci Dgm105 (165–240 bp) and Dgm 109 (250–303 bp)). We assume that this can be explained by the fact that the samples of *D. cucullata* from the continental region were taken from cultured material in the laboratory, which was not exposed to the influence of various local and anthropogenic factors [14]. In our study, natural *Daphnia* populations were used directly from the lakes, and they had been regularly exposed to various anthropogenic factors, mainly to the impact of agriculture. Some relative differences between allele lengths among the *D. cucullata* populations in Boreal and Continental biogeographical regions in Europe most probably are the result of accidental genetic drift, but not of mutation [58].

The highest number of alleles was found in loci Dgm105 and Dgm101 (8 and 7); these loci also had the highest number of private alleles (62% and 57%) However, the number of alleles in loci Dgm105 and Dgm101 was much lower in *D. cucullata* populations in lakes in Switzerland and the Netherlands (2 and 3). These microsatellite loci can be considered the best to use in future studies of the genetic structure of *D. cucullata* populations in Latvia and maybe in Boreal biogeographical regions.

The number of alleles in loci DaB17/17 and Dgm109 was the same (4 and 5), but the number of alleles in locus Dp519 was lower compared to data from Switzerland and the Netherlands [14,27]. *D. cucullata* populations had the largest number of alleles per locus in lakes Geraniimovas-Ilzas and Riča (3.17 and 3.00, respectively), using microsatellites. The largest number of alleles was also in lakes Riča and Geraniimovas-Ilzas, and in Svete and Dridzis, using RAPD analysis [47]. The observed small differences between the allele lengths in the *D. cucullata* populations in the studied Latvian lakes and those found in Switzerland and the Netherlands are most likely the result of random genetic drift and not mutations [58]. It is possible that the increase in allelic diversity is influenced by various chemical compounds in the water. One of the main influencing factors may be different changes in temperature conditions in our studied lakes and lakes in Switzerland and the Netherlands [4,58–61]. The immigration of new individuals may introduce new genetic material and lead to the selection of closely related hybrids, which may lead to an increase in the frequency of “immigrant” alleles, thereby leading to an increase in genetic diversity in populations [47,62].

In our research, the highest level of genetic polymorphism in *D. cucullata* populations according to microsatellites in Lakeland (Latvia) of the Boreal region during the summer season was observed in lakes Riča (100%) and Geraniimovas-Ilzas (83%), and the lowest, in Lake Dridzis (33%). However, according to RAPD analysis, the highest level of genetic polymorphism was observed in lakes Dridzis (50%) and Geraniimovas-Ilzas (33%) [47]. The obtained results can probably be explained by the specificity of the selected nuclear DNA markers (RAPD and microsatellites).

Heterozygosity serves as an indicator of evolutionary potential and is important in determining population dynamics as well as population viability. A decrease in heterozygosity may lead to a decrease in adaptation in the population [6,34]. The average level of heterozygosity in the studied *D. cucullata* populations via microsatellites in lakes Svente, Riča, Dridzis and Geranišimovas-Ilzas was relatively high and ranges from 1.67 to 3.17. The average level of heterozygosity, based on RAPD analysis, in the studied *D. cucullata* populations in these lakes ranges from 0.18 to 0.20 [47]. However, a substantial difference between the observed and expected level of heterozygosity was found only in the locus SwiD1 in the *D. cucullata* population from Lake Geranišimovas-Ilzas ($p < 0.01$). In populations of *D. cucullata* that showed an extremely high level of heterozygosity, they were mostly hybrids, while morphological changes were not observed [6,63,64].

The data we obtained about the genetic distance and genetic differentiation of *D. cucullata* populations according to microsatellites in our studied lakes confirm our previous data about the genetic distance and genetic differentiation of *D. cucullata* populations via RAPD [47] and show that the studied populations of *D. cucullata* are different among themselves. In our study, the Bayesian and Evano clustering approaches showed that the populations of *D. cucullata* (Riča and Geranišimovas-Ilzas), which are relatively far from each other, form a separate genetic group. It is difficult to explain the fact that *Daphnia* populations in lakes that are distant from each other and are not connected to each other are the most similar. It has been assumed that migrating waterfowl can carry zooplankton ephippia up to 50 km per day when flying between feeding or roosting sites. The maximum distance that waterfowl can fly from one water body to another is 1500 km [65–67]. The distribution of *Daphnia* taxa coincides with the flight directions of migratory waterfowl [65]. A different spreading mechanism was shown in *Bythotrephes* (Cladocera), as their ephippia are less viable in the intestinal tract of birds than *Daphnia lumholtzi* [65]. Ephippia can also be carried by wind [68–71]. Ephippia can withstand harsh environmental conditions (freezing, desiccation), and in spring, under favorable conditions, young parthenogenetic females hatch from winter eggs [4,59–61,72–74].

The characteristics of *Daphnia* biology may influence the genetic diversity of populations. *D. cucullata* is a cyclic parthenogenetic organism. Parthenogenetic reproduction continues until adverse weather conditions occur, when some eggs develop into males and others develop into haploid eggs that require fertilization [4,58,74]. Males appear when there is a high population density or a rapid depletion of nutrients. In this case, diploids produce winter eggs, or ephippia [15,26]. Depending on the relative importance of recombination and parthenogenetic reproduction, populations of *D. cucullata* will have different local diversity and genetic population structures [58,73]. The length of the clonal phase and the frequency of sexual reproduction appear also to be related to the size and permanency of the habitat [4,58,60,61,71,75–80]. In cyclic parthenogenetic zooplankton, large habitats will have a larger stock of resting eggs than smaller habitats and, hence, a higher recruitment of sexual eggs from the resting egg bank at the beginning of the growing season. The recruitment of sexual eggs from the resting egg bank will increase genetic diversity and, thus, have a profound impact on the genetic structure of cyclical parthenogenetic *Daphnia* populations [71,77,80,81].

A very interesting fact in our previous study [24] was found: lakes Riča and Geranišimovas-Ilzas have a low number of zooplankton taxa (47 and 43, respectively), but the lakes Dridzis and Svente have a large number of zooplankton taxa (72 and 69, respectively) [24]. In the present study, we found that *D. cucullata* populations from lakes Riča and Geranišimovas-Ilzas have a higher genetic diversity compared to populations from lakes Dridzis and Svente, with a large number of zooplankton taxa. Thus, a higher genetic diversity in populations was observed in lakes which have a low number of zooplankton taxa. A small number of taxa possibly affects the ecological existence niche of *D. cucullata* taxa. Since larger biotopes often contain more ecological niches than smaller ones, this may favor the coexistence of ecologically distinct genotypes, because in this case, *D. cucullata* has less interaction and competition with other taxa, it adapts more to different conditions, it

interbreeds, etc. Therefore, it has more genetic diversity. It was shown that the number of sampled exemplars per population and the number of scored microsatellite DNA alleles are correlated with some of the population genetics parameters. The geographical distance between sampled populations may also have a strong positive effect on among-population diversity [76–78,80,82–84].

Thus, we used these microsatellite markers, especially Dgm105 and Dgm101, for the analysis of natural populations of *D. cucullata* in Latvian Lakeland in the Boreal region for the first time.

Our use of markers turned out to be very successful. Molecular genetic markers are widely used in population studies, and their utility depends on the degree of sequence variation. The nuclear and mitochondrial rRNA genes were the best possible genetic markers for molecular systematics, whereas the mitochondrial protein-coding and rRNA genes were suitable for molecular identification. Estimates derived from the dominantly inherited nuclear markers are very similar and may be directly comparable. Mitochondrial DNA (mtDNA) has been a marker for reconstructing the evolution patterns of populations, admixture, biogeography and speciation. Microsatellites have a higher mutation rate than others, leading to high genetic diversity because the microsatellites can be used to analyze rapid changes in the genetic structure of populations over a short period of time. We were able to detect subtle differences between populations that exist in hydrologically similar conditions of natural reservoirs with their help. Therefore, these microsatellites can be useful for larger-scale studies on the geographic ranges of *D. cucullata* in the Boreal region.

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References

1. Hebert, P.D.N. Competition in zooplankton communities. *Ann. Zool. Fenn.* **1982**, *19*, 349–356.
2. Malone, B.J.; McQueen, D.J. Horizontal patchiness in zooplankton populations in two Ontario kettle lakes. *Hydrobiologia* **1983**, *99*, 101–124. [[CrossRef](#)]
3. Pinel-Alloul, B. Spatial heterogeneity as a multiscale characteristic of zooplankton community. *Hydrobiologia* **1995**, *300*, 17–42. [[CrossRef](#)]
4. Wetzel, R.G. Limnology. In *Lake and River Ecosystems*, 3rd ed.; Academic Press: Cambridge, MA, USA, 2001; p. 1006.
5. Karpowicz, M.; Sługocka, Ł.; Kozłowska, J.; Ochocka, A.; López, C. Body size of *Daphnia cucullata* as an indicator of the ecological status of temperate lakes. *Ecol. Indic.* **2020**, *117*, 1–10. [[CrossRef](#)]
6. Adamowicz, S.J.; Hebert, P.D.N.; Marinone, M.C. Species diversity and endemism in the *Daphnia* of Argentina: A genetic investigation. *Zool. J. Linn. Soc. Lond.* **2004**, *140*, 171–205. [[CrossRef](#)]
7. Colbourne, J.K.; Hebert, P.D.N. The systematics of North American *Daphnia* (Crustacea: Anomopoda): A molecular phylogenetic approach. *Philos. Trans. R. Soc. Lond. Ser. B* **1996**, *351*, 349–360.
8. Colbourne, J.K.; Wilson, C.C.; Hebert, P.D.N. The systematics of Australian *Daphnia* and *Daphniopsis* (Crustacea: Cladocera): A shared phylogenetic history transformed by habitat-specific rates of evolution. *Biol. J. Linn. Soc.* **2006**, *89*, 469–488. [[CrossRef](#)]

9. Dussart, B.H.; Defaye, D. *Introduction to the Copepoda. Guides to the Identification of the Microinvertebrates of the Continental Waters of the World*, 2nd ed.; Backhuys Publishers: Leiden, The Netherlands, 2001; p. 344.
10. Harris, K.D.M.; Bartlett, N.J.; Lloyd, V.K. *Daphnia* as an emerging epigenetic model organism. *Genet. Res. Int.* **2012**, *2012*, 147892. [[CrossRef](#)]
11. Lampert, W. *Daphnia*: Model herbivore, predator and prey. *Pol. J. Ecol.* **2006**, *54*, 607–620.
12. Colbourne, J.K.; Hebert, P.D.N.; Taylor, D.J. Evolutionary origins of phenotypic diversity in *Daphnia*. In *Molecular Evolution and Adaptive Radiation*; Givnish, T.J., Sytsma, K.J., Eds.; Cambridge University Press: London, UK, 1997; pp. 163–188.
13. Dodson, S.I.; Hanazato, T. Commentary on effects of anthropogenic and natural organic-chemicals on development, swimming behavior and reproduction of *Daphnia*, a key member of aquatic ecosystems. *Environ. Health Perspect.* **1995**, *103*, 7–11.
14. Brede, N.; Thielsch, A.; Sandrock, C.; Spaak, P.; Keller, B.; Streit, B.; Schwenk, K. Microsatellite markers for European *Daphnia*. *Mol. Ecol. Notes* **2006**, *6*, 536–539. [[CrossRef](#)]
15. Cousyn, C.; De Meester, L.; Colbourne, J.K.; Brendonck, L.; Verschuren, D.; Volckaert, F. Rapid, local adaptation of zooplankton behavior to changes in predation pressure in the absence of neutral genetic changes. *Proc. Natl. Acad. Sci. USA* **2001**, *98*, 6256–6260. [[CrossRef](#)] [[PubMed](#)]
16. De Meester, L.; Weider, L.J.; Tollrian, R. Alternative antipredator defences and genetic polymorphism in a pelagic predator-prey system. *Nature* **1995**, *378*, 483–485. [[CrossRef](#)]
17. Harris, R.P.; Wiebe, P.H.; Lenz, J.; Skjoldal, H.R.; Huntley, M. *Zooplankton Methodology Manual*; Elsevier Academic Press: Cambridge, MA, USA, 2005; pp. 533–570.
18. Hellsten, M.E.; Sundberg, P. Genetic variation in two sympatric European populations of *Bosmina* spp. (Cladocera) tested with RAPD markers. *Hydrobiologia* **2000**, *421*, 157–164. [[CrossRef](#)]
19. Brakovska, A. *Daphnia cucullata* Sars, 1862 (CRUSTACEA: CLADOCERA) distribution and location in composition of zooplankton cenosis in Lake Dridzis. *Acta Biol. Univ. Daugavp.* **2014**, *14*, 1–19.
20. Brakovska, A.; Škute, R. Ecological characterization of zooplankton groups in the deepest lakes of East Latvia. *Acta Biol. Univ. Daugavp.* **2007**, *7*, 165–174.
21. Brakovska, A.; Škute, R. Ecological evaluation of zooplankton groups in Lake Geranimovas-Ilzas and Lake Garais. *Proc. 7th Int. Sci. Pract. Conf. Environment. Technol. Resour.* **2009**, *2*, 43–50. [[CrossRef](#)]
22. Brakovska, A.; Škute, R.; Škute, A. Heterogeneity of distribution and community composition of zooplankton in upper layers of Lake Svente. *Zool. Ecol.* **2012**, *22*, 172–180. [[CrossRef](#)]
23. Brakovska, A.; Škute, N. Optimisation of DNA extraction and RAPD-PCR amplification for population genetic analysis of *Daphnia cucullata* Sars, 1862 (Crustacea: Cladocera). *Acta Biol. Univ. Daugavp.* **2013**, *13*, 11–20.
24. Brakovska, A.; Paidere, J.; Škute, A. Dynamics and factors influencing zooplankton in the lakes Svente, Riča, Dridzis and Geranimovas-Ilzas (Eastern Latvia). *Acta Biol. Univ. Daugavp.* **2020**, *20*, 71–94.
25. Selkoe, K.A.; Toonen, R.J. Microsatellites for ecologists: A practical guide to using and evaluating microsatellite markers. *Ecol. Lett.* **2006**, *9*, 615–629. [[CrossRef](#)] [[PubMed](#)]
26. Haag, T.; Santos, A.S.; Sana, D.A.; Morato, R.G.; Cullen, L., Jr.; Crawshaw, P.G., Jr.; De Angelo, C.; Di Bitetti, M.S.; Salzano, F.M.; Eizirik, E. The effect of habitat fragmentation on the genetic structure of a top predator: Loss of diversity and high differentiation among remnant populations of Atlantic Forest jaguars (*Panthera onca*). *Mol. Ecol.* **2010**, *22*, 4906–4921. [[CrossRef](#)] [[PubMed](#)]
27. Colbourne, J.K.; Robison, B.; Bogart, K.; Lynch, M. Five hundred and twenty-eight microsatellite markers for ecological genomic investigations using *Daphnia*. *Mol. Ecol. Notes* **2004**, *4*, 485–490. [[CrossRef](#)]
28. Frisch, D.; Morton, P.K.; Roy Chowdhury, P.; Culver, B.W.; Colbourne, J.K.; Weider, L.J.; Jeyasingh, P.D. A millennial-scale chronicle of evolutionary responses to cultural eutrophication in *Daphnia*. *Ecol. Lett.* **2014**, *17*, 360–368. [[CrossRef](#)] [[PubMed](#)]
29. Forest, F.; Chase, M.W.; Persson, C.; Crane, P.R.; Hawkins, J.A. The role of biotic and abiotic factors in evolution of ant dispersal in the milkwort family (polygalaceae). *Evolution* **2007**, *61*, 1675–1694. [[CrossRef](#)] [[PubMed](#)]
30. Pálsson, S. Microsatellite variation in *Daphnia pulex* from both sides of the Baltic Sea. *Mol. Ecol.* **2000**, *9*, 1075–1088. [[CrossRef](#)]
31. Ender, A.; Schwenk, K.; Stadler, T.; Streit, B.; Schierwater, B. RAPD identification of microsatellites in *Daphnia*. *Mol. Ecol.* **1996**, *5*, 437–441. [[CrossRef](#)]
32. Fox, J.A. New microsatellite primers for *Daphnia galeata mendotae*. *Mol. Ecol. Notes* **2004**, *4*, 544–546. [[CrossRef](#)]
33. Lynch, M.; Spitze, K. Evolutionary genetics of *Daphnia*. In *Ecological Genetics*; Real, L., Ed.; Princeton University Press: Princeton, NJ, USA, 1994; pp. 109–128.
34. Schwenk, K.; Spaak, P. Evolutionary and ecological consequences of interspecific hybridization in cladocerans. *Experientia* **1995**, *51*, 465–481. [[CrossRef](#)]
35. Lynch, M.; Pfrender, M.; Spitze, K.; Lehman, N.; Hicks, J.; Allen, D.; Latta, L.; Ottene, M.; Bogue, F.; Colbourne, J. The quantitative and molecular genetic architecture of a subdivided species. *Evolution* **1999**, *53*, 100–110. [[CrossRef](#)]
36. Mergeay, J.; Verschuren, D.; De Meester, L. Cryptic invasion and dispersal of an American *Daphnia* in East Africa. *Limnol. Ocean.* **2005**, *50*, 1278–1283. [[CrossRef](#)]
37. Silva, R.M.; Pereira, F.; Carneiro, J.; Sobral, O.; Ribeiro, R.; Amorim, A.; Soares, A.M.V.M.; Lopes, I. Microevolution in a Natural Population of *Daphnia longispina* Exposed to Acid Mine Drainage. In *Interdisciplinary Studies on Environmental Chemistry—Biological Responses to Contaminants*; Hamamura, N., Suzuki, S., Mendo, S., Barroso, C.M., Iwata, H., Tanabe, S., Eds.; Terrapub: Pozzuoli, Italy, 2010; pp. 213–218.

38. Reid, V.A.; Carvalho, G.R.; George, D.G. Molecular genetic analysis of *Daphnia* in the English Lake District: Species identity, hybridisation and resting egg banks. *Freshw. Biol.* **2020**, *44*, 247–253. [CrossRef]
39. Urtāne, L. Cladocera as types and trophic status indicators of Latvian Lakes. *Thesis* **1998**, *168*. Not published (In Latvian)
40. Available online: <https://www.ezeri.lv/database> (accessed on 24 January 2023).
41. Eipurs, I. Lake Gerāņimovas-Ilzas. The Nature of Latvia. *Riga Encycl. Latv.* **1995**, *2*, 106. (In Latvian)
42. Tidriķis, A. Lake Svente. In *The Nature of Latvia*; Riga, Publishing House: Riga, Latvia, 1998; Volume 5, p. 180. (In Latvian)
43. Tidriķis, A. Lake Riču. In *The Nature of Latvia*; Riga, Publishing House: Riga, Latvia, 1997; Volume 4, pp. 243–244. (In Latvian)
44. APHA. *Standart Methods for the Examination of Waters and Wastewater*, 21st ed.; American Public Health Association: Washington, DC, USA, 2005.
45. Wetzel, R.G.; Likens, G.E. *Limnological Analyses*; Springer Science, Business Media: New York, NY, USA, 2000; p. 429.
46. Fitzsimmons, J.M.; Innes, D.M. No evidence of Wolbachia among Great Lakes area populations of *Daphnia pulex* (Crustacea: Cladocera). *J. Plankton Res.* **2005**, *27*, 121–124. [CrossRef]
47. Brakovska, A.; Paidere, J.; Škute, R.; Škute, N.; Škute, A. Occurrence of Cladocera and genetic diversity of *Daphnia cucullata* in pelagic zone of the Latvian salmonid lakes. *Est. J. Ecol.* **2013**, *62*, 244–264. [CrossRef]
48. Peakall, R.; Smouse, P.E. GENALEX 6: Genetic analysis in Excel. Population genetic software for teaching and research. *Mol. Ecol. Notes* **2006**, *6*, 288–295. [CrossRef]
49. Yeh, F.C.; Yang, R.C.; Boyle, T. *Population Genetics Software*, POPGENE 32-version 1.31: Microsoft Window-Based Freeware for Population Genetics Analysis; University of Alberta: Edmonton, AB, Canada, 1999.
50. Nei, M. *Molecular Evolutionary Genetics*; Columbia University Press: New York, NY, USA, 1987; p. 512.
51. Nei, M. Analysis of gene diversity in subdivided populations. *Proc. Natl. Acad. Sci. USA* **1973**, *70*, 3321–3323. [CrossRef]
52. Nei, M.; Tajima, F.; Tateno, Y. Accuracy of estimated phylogenetic trees from molecular data. 2 Genefrequency data. *J. Mol. Evol.* **1983**, *19*, 153–170. [CrossRef]
53. Hubisz, M.J.; Falush, D.; Stephens, M.; Pritchard, A.J. Interring weak population structure with the assistance of sample group information. *Mol. Ecol. Resour.* **2009**, *9*, 1322–1332. [CrossRef]
54. Earl, D.A.; vonHoldt, B.M. STRUCTURE HARVESTER: A website and program for visualizing STRUCTURE output and implementing the Evanno method. *Conserv. Genet. Resour.* **2012**, *4*, 359–361. [CrossRef]
55. Nei, M. Estimation of Average Heterozygosity and Genetic Distance from a small number of Individuals. *Genetics* **1978**, *89*, 583–590. [CrossRef] [PubMed]
56. Evanno, G.; Regnaut, S.; Goudet, J. Detecting the number of clusters of individuals using the software STRUCTURE: A simulation study. *Mol. Ecol.* **2005**, *14*, 2611–2620. [CrossRef] [PubMed]
57. Biogeographical Regions in Europe—European Environment Agency. Available online: europa.eu (accessed on 24 January 2023).
58. De Meester, L.; Vanoverbeke, J.; De Gelas, K.; Ortells, R.; Spaak, P. Genetic structure of cyclic parthenogenetic zooplankton populations—A conceptual framework. *Arch. Für Hydrobiol.* **2006**, *167*, 217–244. [CrossRef]
59. De Meester, L.; Gómez, A.; Simon, J.C. Evolutionary and ecological genetics of cyclical parthenogens. In *Evolution from Molecules to Ecosystems*; Moya, A., Font, E., Eds.; Oxford University Press: Oxford, UK, 2004; pp. 122–134.
60. Kalf, J. *Limnology: Inland Water Ecosystems*; Prentice-Hall: Upper Saddle River, NJ, USA, 2002; p. 592.
61. Lampert, W.; Sommer, U. *Limnoecology: The Ecology of Lakes and Streams*; Oxford University Press: New York, NY, USA; Oxford, UK, 1997; p. 382.
62. Haag, C.R.; Riek, M.; Hottinger, J.W.; Pajunen, V.I.; Ebert, D. Genetic diversity and genetic differentiation in *Daphnia* metapopulations with subpopulations of known age. *Genetics* **2005**, *170*, 1809–1820. [CrossRef] [PubMed]
63. Taylor, D.J.; Hebert, P.D.N. Habitat dependent hybrid parentage and differential introgression between neighboring sympatric *Daphnia* species. *Proc. Natl. Acad. Sci. USA* **1993**, *90*, 7079–7083. [CrossRef]
64. Taylor, D.J.; Hebert, P.D.N. A reappraisal of phenotypic variation in *Daphnia galeata mendotae*: The role of interspecific hybridization. *Can. J. Fish. Aquat. Sci.* **1993**, *50*, 2137–2146. [CrossRef]
65. Green, A.J.; Figuerola, J. Recent advances in the study of longdistance dispersal of aquatic invertebrates via birds. *Divers. Distrib.* **2005**, *11*, 149–156. [CrossRef]
66. Figuerola, J.; Green, A.J.; Santamaria, L. Passive internal transport of aquatic organisms by waterfowl in Doñana, south-west Spain. *Glob. Ecol. Biogeogr.* **2003**, *12*, 427–436. [CrossRef]
67. Figuerola, J.; Green, A.J.; Michot, T.C. Invertebrate eggs can fly: Evidence of waterfowl-mediated gene flow in aquatic invertebrates. *Am. Nat.* **2005**, *165*, 274–280. [CrossRef]
68. Hebert, P.D.N.; Moran, C. Enzyme variability in natural populations of *Daphnia carinata* King. *Heredity* **1980**, *45*, 313–321. [CrossRef] [PubMed]
69. Crease, T.J.; Lynch, M.; Spitze, K. Hierarchical analysis of population genetic variation in mitochondrial and nuclear genes of *Daphnia pulex*. *Mol. Biol. Evol.* **1990**, *7*, 444–458. [PubMed]
70. Gomez, A.; Carvalho, G.R. Sex, parthenogenesis and genetic structure of rotifers: Microsatellite analysis of contemporary and resting egg bank populations. *Mol. Ecol.* **2000**, *9*, 203–214. [CrossRef] [PubMed]
71. Vanoverbeke, J.; De Meester, L. Among-population genetic differentiation in the cyclical parthenogen *Daphnia magna* (Crustacea: Anomopoda) and its relation to geographic distance and clonal diversity. *Hydrobiologia* **1997**, *126*, 135–142. [CrossRef]

72. Decaestecker, E.; De Meester, L.; Mergeay, J. Cyclical Parthenogenesis in *Daphnia*: Sexual Versus Asexual Reproduction. *Lost Sex* **2009**, 295–316.
73. Hughes, R.N. *A Functional Biology of Clonal Animals*; Chapman & Hall: London, UK; New York, NY, USA, 1989.
74. Hobæk, A.; Larsson, P. Sex determination in *Daphnia magna*. *Ecology* **1990**, *71*, 2255–2268. [[CrossRef](#)]
75. Hebert, P.D.N. Genetics of *Daphnia*. In *Daphnia*; Peters, R.H., De Bernardi, R., Eds.; Mem Inst Ital Idrobiol: Verbania Pallanza, Italia, 1987; Volume 45, pp. 439–460.
76. De Meester, L. Local genetic differentiation and adaptation in freshwater zooplankton populations: Patterns and processes. *Ecoscience* **1996**, *3*, 385–399. [[CrossRef](#)]
77. Vanoverbeke, J.; De Gelas, K.; De Meester, L. Habitat size and the genetic structure of a cyclical parthenogen, *Daphnia magna*. *Heredity* **2007**, *98*, 419–426. [[CrossRef](#)]
78. Deng, H.W.; Lynch, M. Change of genetic architecture in response to sex. *Genetics* **1996**, *143*, 203–212. [[CrossRef](#)]
79. Hairston, N.G.; Holtmeier, C.L.; Lampert, W.; Weider, L.J.; Post, D.M.; Fischer, J.M.; Cáceres, C.E.; Fox, J.A.; Gaedke, U. Natural selection for grazer resistance to toxic cyanobacteria: Evolution of phenotypic plasticity? *Evolution* **2001**, *55*, 2203–2214. [[CrossRef](#)]
80. Lynch, M.; Deng, H.W. Genetic slippage in response to sex. *Am. Nat.* **1994**, *144*, 242–261. [[CrossRef](#)]
81. Michels, E.; Audenaert, E.; Ortells, R.; De Meester, L. Population genetic structure of three pond-inhabiting *Daphnia* species on a regional scale (Flanders, Belgium). *Freshw. Biol.* **2003**, *48*, 1825–1839. [[CrossRef](#)]
82. Spaak, P.; Denk, A.; Boersma, M.; Weider, L.J. Spatial and temporal patterns of sexual reproduction in a hybrid *Daphnia* species complex. *J. Plankton Res.* **2004**, *26*, 625–635. [[CrossRef](#)]
83. Wolf, H.G. Interspecific hybridization between *Daphnia hyalina*, *D. galeata* and *D. cucullata* and seasonal abundances of these species and their hybrids. *Hydrobiologia* **1987**, *145*, 213–217. [[CrossRef](#)]
84. Nybom, H. Comparison of different nuclear DNA markers for estimating intraspecific genetic diversity in plants. *Mol. Ecol.* **2004**, *13*, 1143–1155. [[CrossRef](#)]

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