

Article



# Insect Abundance and Richness Response to Ecological Reclamation on Well Pads 5–12 Years into Succession in a Semi-Arid Natural Gas Field

Michael F. Curran <sup>1,\*</sup>, Jasmine Allison <sup>2</sup>, Timothy J. Robinson <sup>3</sup>, Blair L. Robertson <sup>4</sup>, Alexander H. Knudson <sup>5</sup>, Bee M. M. Bott <sup>1,6</sup>, Steven Bower <sup>1</sup> and Bobby M. Saleh <sup>2</sup>

- <sup>1</sup> Abnova Ecological Solutions, Cheyenne, WY 82001, USA
- <sup>2</sup> PureWest Energy Partners, Denver, CO 80202, USA
- <sup>3</sup> Department of Statistics, University of Wyoming, Laramie, WY 82071, USA
- <sup>4</sup> Department of Mathematics and Statistics, University of Canterbury, Christchurch 8140, New Zealand
  - Plant Pest Diagnostics Laboratory, Utah State University, Logan, UT 84322, USA
- <sup>6</sup> Department of Soil and Crop Sciences, Colorado State University, Fort Collins, CO 80521, USA
- \* Correspondence: mike@abnovaecology.com

Abstract: Natural gas extraction is a critical driver of the economy in western North America. Ecological reclamation is important to ensure surface disturbance impacts associated with natural gas development are not permanent and to assist native biota. Previous studies in semi-arid natural gas fields within Sublette County, Wyoming, USA have shown insects respond favorably to 1–3-year-old well pads undergoing reclamation compared to older successional reference vegetation communities dominated by Wyoming big sagebrush (Artemisia tridentata spp. Wyomingensis). Here, we examined well pads which were initially seed 5, 8, 10, 11, and 12 years prior to our study. We used a free, image-based software called SamplePointv. 1.60 to quantify vegetation on these well pads and adjacent reference areas from cell phone camera photographs. Insects were collected with a sweep net and identified to the family and morphospecies level. Statistical analyses were conducted to compare both vegetation and insect communities between reclamation sites and their paired reference area. We found little statistical difference between vegetation communities across our study but found significantly more insect abundance on reclaimed well pads than reference areas in 3 of 5 years and significantly higher family and morphospecies richness on reclaimed well pads in 4 of 5 years. A total of 2036 individual insects representing 270 species from 71 families across 11 orders were identified across this study. A total of 1557 individuals (76.5%) were found on reclamation sites, whereas 479 (23.5%) were found in reference areas across the entire study. A total of 233 species (86.3% of total) were found on reclamation sites, whereas 121 species (44.8% of total) were found in reference areas across the entire study. A total of 67 families (94.4% of total) were found on reclamation sites, whereas 45 families (63.4% of total) were found in reference areas across the entire study. All 11 orders found in the study were found on reclamation sites, whereas 9 orders were found in reference areas across the entire study. Our results suggest reclamation of natural gas well pads within an old successional stand of sagebrush continues to support higher levels of insect biodiversity and abundance for at least 12 years. As insects are the most diverse group of animals on Earth and because they provide a wide array of ecosystem services, our findings suggest ecological reclamation plays an important role in returning biodiversity and ecosystem functionality to a semi-arid and old successional sagebrush-steppe ecosystem.

Keywords: biodiversity; ecosystem services; natural gas; sagebrush; SamplePoint

# 1. Introduction

Ecological reclamation is a necessary process to aid native biota after land development requiring surface disturbance has taken place [1]. Although drylands make up  $\sim$ 40% of



Citation: Curran, M.F.; Allison, J.; Robinson, T.J.; Robertson, B.L.; Knudson, A.H.; Bott, B.M.M.; Bower, S.; Saleh, B.M. Insect Abundance and Richness Response to Ecological Reclamation on Well Pads 5–12 Years into Succession in a Semi-Arid Natural Gas Field. *Diversity* 2024, *16*, 324. https://doi.org/10.3390/ d16060324

Academic Editors: Merav Seifan and Michael Wink

Received: 3 May 2024 Revised: 25 May 2024 Accepted: 27 May 2024 Published: 29 May 2024



**Copyright:** © 2024 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). the Earth's surface [2] and are home to ~39% of the human population [3], research from drylands only accounts for <6% of published terrestrial ecology studies [4] and <5% of published studies related to restoration ecology [5]. The expanse of drylands is predicted to increase with changing climate conditions and a growing human population will continue to rely on natural resources for the foreseeable future [6,7]. Wyoming, the ninth largest state in the United States by land mass, has climates ranging from arid to semi-arid across its vast expanse [8]. Although it is the least populated state, Wyoming is rich in natural resources and home to a wide array of biodiversity and wildlife [8]. Since biodiversity loss is known to have negative impacts on humanity and because natural resource extraction is often linked to land surface disturbance [8,9], understanding how biodiversity responds to land surface disturbance and reclamation activities associated with resource extraction is paramount.

Natural gas provides tangible benefits to humans and the production of the resource is a major driver of Wyoming's economy [8]. In order to extract natural gas, well pads are typically constructed by stripping and stockpiling soil and removing vegetation from an area to provide a safe working environment and to house equipment necessary to build a well [10]. After well construction is complete, the construction area is typically decompacted before soil from stockpiles is respread across the non-active area (~75%) of the well pad and then seeded with native vegetation to begin the process of interim reclamation [10,11]. In the Pinedale Anticline natural gas field (Sublette County, Wyoming, MI, USA), operators are required to comply with reclamation regulatory criteria, which vary between and among State and Federal Government agencies [12,13]. These criteria are focused on erosion control, minimizing invasive and noxious vegetation species, and establishing native vegetation communities which are similar to adjacent reference ecosystems which have not been directly impacted by natural gas activities [11–13]. In the past decade, efforts to exceed minimum regulatory criteria efforts towards restoring wildlife habitat are common practice in Wyoming [8].

Efforts to reclaim dryland ecosystems are challenged by a variety of environmental factors including invasive vegetation [14], low and unpredictable precipitation [8], a lack of available seed resources [15], poor soils [16], and extreme temperatures [17]. Situated in a sagebrush–steppe ecosystem with elevations above 2100 m in an area receiving 254–356 mm mean annual precipitation (MAP) and experiencing 42 frost-free days per year, many of these challenges are exacerbated in the Pinedale Anticline natural gas field. Aside from environmental and climatic challenges, operators in the area must deal with multiple land use goals (e.g., domestic livestock grazing, recreation, wildlife habitat) and abide by aforementioned regulatory criteria. Aside from erosion control and lack of noxious or invasive vegetation, operators in the Pinedale Anticline are required to achieve >50% of shrub density compared to an adjacent reference area. Typically, reference sites in the area are comprised old stands of Wyoming big sagebrush (Artemisia tridentata spp. Wyomingensis) which have been subject to domestic livestock grazing since the 1870s [18]. Research has shown sagebrush establishment in arid and semiarid environments may take decades [19]. As soil alterations associated with natural gas well pad development result in a pulse of plant-available nitrogen which sagebrush are not well adapted to [20,21], operators in the area have achieved better results combating annual invasive weed species and stabilizing locations using diverse seed mixes comprised native annual forb and perennial forb, grass and shrub species [22]. This can be looked at as a form of 'assisted succession' [23], setting a pathway for the eventual establishment of mature sagebrush. While careful reclamation planning, soil management, and seed mix application are required for initial vegetation establishment in arid environments [17,20], establishing plant-pollinator networks is critical to ensure these vegetation communities are self-sustaining over time [24–27].

While some plants in rangeland ecosystems rely mainly on wind pollination [28], over 75% of flowering plants rely on animal pollination to successfully reproduce and maintain genetic diversity [29]. Aside from their critical role as pollinators, insects are the most abundant and diverse group of animals on Earth and provide numerous important

ecosystem services [30–33]. In addition to pollination and biodiversity, other ecosystem services insects provide include nutrient cycling and decomposition [34,35], nutrition of higher food levels [36], and pest control [37]. Despite their importance, terrestrial insects have been declining globally [38] and are often overlooked in terrestrial wildlife studies [39]. Recent research suggests reductions in invertebrates are associated with declining ecosystem performance [40].

Giving cause for hope, a 2018 meta-analysis suggests habitat restoration efforts are beneficial to wild pollinators [41]. Measuring insect richness and diversity has shown to be a sound indicator of ecosystem functionality and restoration success since insects respond rapidly to environmental change, provide more ecosystem services than other animals, and because statistically valid samples can be obtained in a short time period [42,43]. Efforts to understand insect response to restoration efforts are heavily focused on pollinators in crop agriculture ecosystems [44] and very few studies have been conducted to determine how insects respond to reclamation associated with oil and natural gas development [45–47]. Previous studies examining how insects respond to reclamation practices on Sublette County natural gas well pads show well pads exhibiting early successional vegetation (1–3 years post reclamation initiation) contain higher insect abundance and diversity than adjacent reference ecosystems [46,47]. To date, no studies have examined how insect abundance and diversity fare on reclaimed well pads with vegetation communities in later successional stages.

The objective of this study is to examine differences between vegetation communities on natural gas well pads on which reclamation was initiated 5–12 years ago compared to adjacent reference communities as well as to examine differences in insect communities on those sites and their paired reference sites. A total of 15 natural gas well pads (3 each in which reclamation was initiated 5, 8, 10, 11, and 12 years ago) were examined along with a paired reference site. All reference areas are dominated by old stands of Wyoming big sagebrush with understories containing few grass and flowering species, whereas reclaimed well pad locations contained more diverse vegetation communities. Based on previous research examining vegetation and insects in this area [46,47], we hypothesized that vegetation communities on reclaimed well pads would be different than reference sites and that insect abundance and diversity would be higher on reclaimed well pads than reference sites. As information about insects in the area is limited, we also sought to determine if different insect families were distributed equally among reclaimed sites and reference areas.

# 2. Materials and Methods

#### 2.1. Study Area

The Pinedale Anticline natural gas field is located near the town of Pinedale in Sublette County, WY, USA (Figure 1). The entire gas field is located above 2100 m and is considered semi-arid, with a MAP of 254–365 mm. The area has an average of 42 frost-free days per year and experiences highly variable daily temperatures during the short growing season, with daily minimum temperatures ranging between -1 and 6 °C and daily maximum temperatures ranging between 19 and 26 °C. Despite the challenging environment, the Pinedale Anticline is home to wildlife including Greater sage-grouse (*Centrocercus urophasianus*), mule deer (*Odocoileus hemionus*), pronghorn antelope (*Antilocapra americana*), and a wide variety of songbirds and other sagebrush obligate species.





STEWART POINT

Figure 1. A map depicting the area of the Pinedale Anticline natural gas field with study locations color coded by seeding year.

Natural gas development began in the Pinedale Anticline ca. 2000, with predominantly 5 different upstream (i.e., exploration and production) operating companies managing different sections of the gas field. Reclamation regulatory criteria are set in place by the Wyoming Department of Environmental Quality (WDEQ) and the DOI BLM 2008 Record of Decision (ROD) for Supplemental Environmental Impact Statement (SEIS) Pinedale Anticline Project Area (PAPA). The BLM 2008 ROD SEIS PAPA created the Pinedale Anticline Project Office (PAPO) to oversee and manage mitigation and monitoring and the funding for mitigation and monitoring. The PAPO is made up of interagency governmental consortium consisting of the WDEQ, Wyoming Game and Fish Department, Wyoming Department of Agriculture, Sublette County Conservation District, and the Bureau of

Land Management. The PAPO oversees Pinedale Anticline Data Management System (PADMS) which is a reclamation data monitoring tool. The WDEQ Stormwater Pollution Prevention Plan requires reclaimed well pad sites to be stabilized as evidenced by 70% vegetation cover compared to an adjacent, undisturbed reference system as well as a lack of erosion features present in the reclamation area. The 2008 ROD SEIS PAPA reclamation regulatory criteria suggest reclaimed well pad locations must (1) be in stable condition with no erosion features, (2) have a minimum of 3 native perennial grass species of which at least 2 are bunchgrasses, (3) be free from all species listed as noxious in WY, (4) have a resilient vegetation community as evidenced by well-developed root systems, flowers and seed heads, (5) and meet several criteria based on comparison to an adjacent, undisturbed reference area. These criteria include that the reclamation site during interim reclamation which is the stage prior to final reclamation, must have the following in comparison to a reference site: (1) an average density or frequency of forbs which is >75% within 5 years, (2) an average density of frequency of shrubs which is >50% within 5 years, and (3) a percentage of bareground cover less than or equal. While all operators in the area are subject to the same criteria, initial reclamation plans and practices were unique to different operators. More recently, the natural gas well pad locations in the Pinedale Anticline have been acquired by one operating company who actively manages all locations.

Soils throughout the Pinedale Anticline are predominantly sandy loams. While reference sites are unique to each well pad location, the majority of reference sites throughout the field are dominated by old stands of Wyoming big sagebrush with a mixed grass understory which matches the United States Department of Agriculture—Natural Resources Conservation Service's ecological site description characterization of the historic plant climax community. According to definitions by the Society for Ecological Restoration, reference areas in the Pinedale Anticline natural gas field can be considered cultural reference ecosystems [48] since they have been managed and grazed by domestic livestock long before natural gas development began in the area [18].

#### 2.2. Site Selection

The Pinedale Anticline natural gas field was stratified by management area to assist with site selection (Figure 1). This gas field was developed by different operators and development was not consistent between 2000 and 2023 (e.g., some years saw more well pads constructed than others). Three sites were randomly selected from five management units after each management unit was grouped by year of initial reclamation and seed mix used in effort to obtain replicates based on similar vegetation communities at the time of sampling (Figure 1). Sites in the Boulder management area were seeded in 2011, sites in the Riverside management area were seeded in 2012, sites in the Mesa management area were seeded in 2013, sites in the Warbonnet management area were seeded in 2015, and sites in the Stewart Point management area were seeded in 2018. While seed mixes varied across management units, they were consistent within management units and all species within each seed mix are native to Sublette County, WY and approved by the BLM with goals of supporting multiple land uses for wildlife and domestic livestock (seed mixes unique to each area are available upon request). The locations selected within the Boulder management area ranged from 2125 to 2197 m elevation, whereas sites in the Riverside management area ranged from 2137 to 2150 m elevation, sites in the Mesa management area ranged from 2273 to 2277 m elevation, sites in the Warbonnet management area ranged from 2163 to 2231 m elevation, and sites in the Stewart Point management area ranged from 2204 to 2356 m elevation.

Vegetation and insect sampling was conducted concurrently during peak flowering between 14 and 16 August 2023. All sampling was conducted by author M.F.C. in sunshine conditions with wind speeds < 6 kph between 0900 and 1200. Temperatures during sampling events ranged between 13 and 21 °C on 14 August, between 13 and 22 °C on 15 August and between 14 and 23 °C on 16 August.

#### 2.3. Vegetation Sampling

A matched pair study design was used in this effort with each well pad and its paired reference site having two 40 m transects sampled. These transects were located 10 m and 20 m away from the well pad edge on both the reclamation area and the reference area. This design is similar to previous research [46,47] and is in accordance with the 2008 ROD SEIS PAPA. Along each transect 0.5 m<sup>2</sup> images were taken at 5 m intervals, resulting in 9 images per transect or 18 images per each well pad and paired reference location. All images were taken by hand with a 50 mega-pixel camera on a Samsung Galaxy A14 taking images perpendicular to the ground (freehand, nadir technique) [49]. The ground sample distance in each photo was ~0.2 mm, which is similar to previously published research using image analysis to identify vegetation on reclamation sites and in rangeland areas [13,50,51]. A  $5 \times 5$  grid was placed inside of each photo using a free image analysis software called 'SamplePoint' [52] to examine 25 points per photo as recommended by Ancin-Murguzur et al. [53], resulting in 225 data points per transect or 450 data points for each well pad and paired reference area. Each data point was classified as bareground (i.e., soil or rock), litter, or vegetation to the species-specific level. Vegetation at the species-specific level was classified as native or non-native and then grouped by life form (i.e., grass, forb, or shrub) prior to statistical analysis.

# 2.4. Insect Sampling

Insect sampling was conducted along the same transects during the same days and times as vegetation sampling. Forty sweeps were taken along each 40 m transect location using a canvas sweep net with a 38 cm diameter and 50 cm handle. These sweeps were conducted prior to vegetation sampling to minimize potential disruption to insects from observer being on site. After each transect sweep was completed, insects were placed in a Zip-Lock<sup>®</sup> bag and placed in a cooler containing dry ice which was a minimum of 200 m from the sampling locations. A minimum of 10 min was spent at this distance before walking back to the sampling area to complete another sweep net sample to minimize any 'flushing' effects on insects [54].

After insect sampling was complete, all insects were shipped in coolers to the Utah State University Plant Pest Diagnostics Laboratory in Logan, Utah, United States. Author A.H.K., who has a PhD in Entomology, identified insects to the family level and organized each family into morphospecies. All morphospecies correspond to individual species, but due to time and budgetary constraints it was not possible to key out each individual species to name across the entire study. The majority of insects were identified to family using Triplehorn and Johnson [55], with the Diptera identified to family using McAlpine et al. [56]. Diverse superfamilies (i.e., Chalcidoidea and Gellichoidea) were kept to superfamily save the nominal families. Insect species were then assigned individual morphospecies numbers and individuals of each morphospecies per site. Voucher specimens have been de-posited to the University of Wyoming Insect Museum.

#### 2.5. Statistical Analysis

As each management area varied in elevation and proximity to the river in the southern portion of the Pinedale Anticline natural gas field (Figure 1) and because insect sampling may be heavily influenced by a number of factors aside from vegetation (e.g., elevation, proximity to other resources, time of day, temperature, etc.) [57–59], we did not attempt to correlate insect abundance and diversity with vegetation cover. Instead, the vegetation and insect analyses were conducted independently among study areas described in Section 2.2.

Vegetation analysis was conducted using automatically generated .csv files from SamplePoint. Pairwise t-comparisons with Bonferroni adjustments were used to compare each cover class category across reclaimed and reference sites for each of the five locations. The Bonferroni adjusted *p*-values are provided in Table 1.

Cover Type	Boulder	Riverside	Mesa	Warbonnet	Stewart Point
Bareground	0.201	0.028	0.025	0.063	0.575
Forb	0.130	0.253	0.291	0.202	0.057
Grass	0.109	0.211	0.555	0.324	0.004
Litter	0.109	0.483	0.505	0.375	0.850
Non-Native	0.704	0.270	0.423	0.742	NA
Shrub	0.223	0.047	0.603	0.065	0.104

**Table 1.** Paired t- comparisons with Bonferroni adjustments of reclaimed vs. reference at each site (note that there was no recorded 'non-native' observations for Stewart Point).

For insect data, generalized linear model regressions were run to determine potential relationships between the response (insect abundance, species richness, and family richness, respectively) and site location, reclamation status and the interaction of site location and reclamation status. Poisson models were run for species richness and family richness and due to overdispersion, a negative binomial regression was utilized for modeling insect abundance. Parametric bootstrapping was then implemented to provide 95% confidence intervals on the predicted mean response for each combination of site type and reclamation status. Additionally, the top three insect families (in terms of abundance) at each study location/reclamation status combination were plotted to show differences in dominant families. In the case of the reference site at Mesa, only the top two ranked insect families were included because the overall number of insects in the third ranked category was less than 5. The negative binomial regressions and associated bootstrapped confidence intervals were run in Program R [60] using the MASS and ciTools libraries, respectively.

## 3. Results

#### 3.1. Vegetation Sampling

The percent cover for each cover class was not statistically different between reclaimed and reference for any class/site combination except for bareground (Riverside and Mesa locations), shrub (Riverside location), and grass (Stewart Point location), as demonstrated by the paired t- comparisons with Bonferroni adjustments in Table 1. Grouped bar plots showing the relative percent cover each of cover category/site combination for reference and reclaimed areas are provided in Figure 2.

#### 3.2. Insect Sampling

A total of 2036 individual insects representing 270 species from 71 families across 11 orders were identified across this study (see Supplementary Material File S1). A total of 1557 individuals (76.5%) were found on reclamation sites, whereas 479 (23.5%) were found in reference areas across the entire study. A total of 233 species (86.3% of total) were found on reclamation sites, whereas 121 species (44.8% of total) were found in reference areas across the entire study. A total of 7 families (94.4% of total) were found on reclamation sites, whereas 45 families (63.4% of total) were found in reference areas across the entire study. All 11 orders found in the study were found on reclamation sites, whereas 9 orders were found in reference areas across the entire study.

Insect abundance was greater on reclamation sites than reference sites in three of the five study areas (seeded in 2011, 2013, and 2018) as evidenced by the non-overlapping confidence intervals comparing reclaimed to reference in Figure 3, Panel a. While the reclaimed sites seeded in 2012 and 2015 had a higher total abundance than reference sites, their 95% CIs overlapped with those of their corresponding reference sites, suggesting no difference was plausible. Family and species richness were greater on reclamation sites than reference sites in four of five study areas as evidenced by the non-overlapping confidence intervals in Figure 3, Panels b,c. However, for sites seeded in 2012 family and



species richness, although greater on reclaimed sites vs. reference locations, their 95% CIs overlapped, thus suggested no difference was plausible.

**Figure 2.** Grouped bar plots showing the relative percent cover each of cover category/site combination for reference and reclaimed areas.





**Figure 3.** Panels ( $\mathbf{a}$ - $\mathbf{c}$ ) depict mean and 95% bootstrap confidence intervals on insect abundance, family richness, and species richness, respectively, for each study area. Estimates and intervals were obtained using the results from Poisson and Negative Binomial regressions and subsequent parametric bootstrapping.

Dominant insect families varied across management units (Figure 4). In the Boulder management unit (seeded 2011), the dominant families in both the reclamation area and reference sites were Chalcidoidea, Cicadellidae, and Miridae, with all three families found at higher levels in the reclamation area. In the Riverside management unit (seeded 2012), the dominant insect families on the reclamation sites were Chrysomelidae, Formicidae, and Miridae, whereas the dominant insect families in the reference sites were Miridae, Cicadellidae, and Tephritidae. The Mesa management unit (seeded 2013) saw Miridae, Tephritidae, and Rhyparochromidae as the dominant insect families on reclaimed sites with Cicadellidae and Miridae dominant on the reference areas. In the Warbonnet man-agement unit (seeded 2015), the dominant families in both the reclamation area and ref-erence sites were Chalcidoidea, Cicadellidae, and Miridae, with Cicadellidae being more abundant in

the reference areas and Chalcidoidea and Miridae more abundant on re-claimed sites. In the Stewart Point management unit (seeded 2018), Cicadellidae, Formi-cidae, and Miridae were the dominant insect families on the reclaimed sites while Cicadellidae, Miridae and Chalcidoidea were dominant in the reference areas—of the common families, more Cicadellidae and Miridae were found in the reclaimed areas than reference sites.



**Figure 4.** Panel (**a**) shows insect family abundance for the only 3 families which made up 62% of the total number of insects in the Boulder (2011 seeded) study area. Panel (**b**) shows insect family abundance for the only 5 families which made up >70% of the total number of insects in the Riverside (2012 seeded) study area. Panel (**c**) shows insect family abundance for the only 4 families which made up >66% of the total number of insects in the Mesa (2013 seeded) study area. Panel (**d**) shows insect family abundance for the only 3 families which made up >63% of the total number of insects in the Warbonnet (2015 seeded) study area. Panel (**e**) shows insect family abundance for the only 3 families which made up >65% of the total number of insects in the Stewart Point (2018 seeded) study area.

## 4. Discussion

While little statistically significant differences were found among vegetation cover class groups across this study, it may be important to note that grass, forb, and total vegetation cover had higher mean cover percentages in all study areas. As older stands of sagebrush found in the reference system of the Pinedale Anticline natural gas field have not been shown to host large amounts of insect diversity [46,47], it is perhaps not surprising that insect abundance was significantly greater on well pad groups compared to reference sites in three of the five areas with diversity at the morphospecies and family levels being significantly higher on reclaimed sites in four of the five areas studied.

This study was limited to one growing season and conducted only during peak flowering. However, its findings are important as it is the first to examine how reclaimed well pads which are greater than 3 years old compare to adjacent reference areas in a sagebrushsteppe ecosystem. Previous research in Sublette County natural gas fields clearly showed well pads undergoing reclamation which were 3 years old and under had more insect abundance and diversity than the old stands of reference area sagebrush [46,47], which is consistent with successional theory suggesting biodiversity peaks in early successional stages [61]. While successional processes develop over long periods of time in cold, arid environments, knowing that reclaimed well pads from 5 to 12 years old are able to support enhanced biodiversity during peak growing season is encouraging as understanding biodiversity loss and ecosystem restoration have been considered as two of the top four opportunities for future ecological successional studies [62]. The foci of the recent conservation of pattern and process paradigm in rangeland management is on fire and grazing as tools to create patchy mosaics of vegetation communities at the landscape scale, and suggests healthy rangeland landscapes are those which contain vegetation communities at different successional stages [63]. Our findings show promise that biodiversity at the insect level has the potential to remain significantly higher on the well pads in Sublette County natural gas fields undergoing reclamation for at least 12 years, which is valuable to the growing body of literature suggesting patchiness in vegetation communities across the landscape is important to rangeland management.

As Greater sage-grouse habitat management and conservation is of utmost importance across Wyoming and the western US [8], our findings are also promising. It is well-documented that insects and forbs play critical roles in the diet of sage-grouse during non-winter months [15,64,65]. In addition to insects being important to the sage-grouse diet, it is estimated that 96% of terrestrial birds rear their young solely or primarily on insect protein with declines in insect populations leading to declines in bird populations [66]. As the reclaimed well pads in our study predominantly comprised native vegetation, it should be noted that non-native vegetation typically has negative impacts on insects [67]. Therefore, those responsible for reclamation should strive to establish diverse stands of native vegetation.

In addition to our study being limited to only peak growing season during 1 year, our study was also limited in that we used sweep netting to capture insects. It is likely that we would have found different insect communities had we used pitfall traps or other sampling methods [68]. Despite these limitations, our results clearly suggest that natural gas well pads undergoing reclamation within the Pinedale Anticline natural gas field are capable of hosting significantly more insects than reference sites dominated by late successional stands, primarily comprising sagebrush, for at least 12 years. Future studies to capture ground-dwelling insects would be beneficial, as would longer-term studies and studies at different points in the growing season.

#### 5. Conclusions

As human populations continue to grow, our reliance on natural resources is likely to continue. Ecological reclamation will continue to play an increasingly important role in assisting the recovery of native biota. As insects provide a plethora of critical ecosystem services, studies to understand how they respond to ecological reclamation efforts will continue to be important. Our results corroborate findings from previous research suggesting that reclamation sites within old successional stands of sagebrush have the ability to support greater insect diversity and abundance, but add to the literature by demonstrating these impacts may last for at least 12 years post initial seeding efforts. Based on our findings, it is recommended that reclamation practitioners in arid and semi-arid sagebrush–steppe ecosystems of western North America continue to use diverse native seed mixes containing forbs and grasses as well as sagebrush. Additionally, these results may help inform habitat restoration within sagebrush–steppe ecosystems as there is a growing body of literature suggesting native vegetation species other than sagebrush are critical to overall ecosystem functionality [69].

**Supplementary Materials:** The following supporting information can be downloaded at: https://www.mdpi.com/article/10.3390/d16060324/s1, File S1: Raw insect data.

Author Contributions: Conceptualization, M.F.C.; methodology, M.F.C., J.A., B.L.R. and T.J.R.; reclamation oversight, J.A.; insect identification, A.H.K.; data curation, M.F.C., A.H.K. and S.B.;

formal analysis, M.F.C., B.L.R. and T.J.R.; resources, J.A.; figure creation, T.J.R. and B.M.S.; writing original draft preparation, M.F.C.; writing—review and editing, M.F.C., J.A., A.H.K., B.L.R., T.J.R., B.M.M.B., S.B. and B.M.S.; funding acquisition, M.F.C., B.L.R. and T.J.R. All authors have read and agreed to the published version of the manuscript.

Funding: This research was funded by PureWest Energy.

Institutional Review Board Statement: Not applicable.

**Data Availability Statement:** Insect row data can be found in File S1. Other requests should be forwarded directly to M.F.C.

Acknowledgments: We are grateful to PureWest Energy for providing funding, site access and records related to their reclamation practices. We thank Jesse Dillion (Cedar Creek Associates) for reviewing the proposal for funding and discussions about reclamation best practices. We are also grateful to Curt Yanish (Aster Canyon Consulting) and Mike Henn (Sublette County Conservation District) for providing information related to reclamation within the Pinedale Anticline natural gas field. We thank Emily Kress and Timothy Miller for processing, mounting, and labeling insect specimens.

**Conflicts of Interest:** Authors J.A. and B.M.S. are employees of PureWest Energy. Both assisted with manuscript editing and provided information about the Pinedale Anticline natural gas field and reclamation practices and BS provided assistance with map/figure creation. PureWest did not influence the decision to conduct this study, participate in data collection, analysis, interpretation of results, or decision to publish. Other authors declare no conflicts of interest.

# References

- 1. Gerwing, T.G.; Hawkes, V.C.; Gann, G.D.; Murphy, S.D. Restoration, reclamation, and rehabilitation: On the need for, and positing a definition of, ecological reclamation. *Restor. Ecol.* **2022**, *30*, e13461. [CrossRef]
- Reynolds, J.F.; Smith, D.M.S.; Lambin, E.F.; Turner, B.L.; Mortimore, M.; Batterbury, S.P.; Downing, T.E.; Dowlatabadi, H.; Fernández, R.J.; Herrick, J.E.; et al. Global desertification: Building a science for dryland development. *Science* 2007, *316*, 847–851. [CrossRef] [PubMed]
- 3. Plaza, C.; Zaccone, C.; Sawicka, K.; Méndez, A.M.; Tarquis, A.; Gascó, G.; Heuvelink, G.B.; Schuur, E.A.; Maestre, F.T. Soil resources and element stocks in drylands to face global issues. *Sci. Rep.* **2018**, *8*, 13788. [CrossRef] [PubMed]
- 4. Okin, G.S.; Heras, M.M.; Saco, D.L.; Throop, P.M.; Vivoni, H.L.; Parsons, E.R.; Wainwright, A.J.; Peters, D.P. Connectivity in dryland landscapes: Shifting concepts of spatial interactions. *Front. Ecol. Environ.* **2015**, *13*, 20–27. [CrossRef] [PubMed]
- Svejcar, L.N.; Kildisheva, O.A. The age of restoration: Challenges presented by dryland systems. *Plant Ecol.* 2017, 218, 1–6. [CrossRef]
- Yao, J.; Liu, H.; Huang, J.; Gao, Z.; Wang, G.; Li, D.; Yu, H.; Chen, X. Accelerated dryland expansion regulates future variability in dryland gross primary production. *Nat. Commun.* 2020, 11, 1665. [CrossRef] [PubMed]
- 7. Pîrlogea, C. The human development relies on energy. Panel data evidence. Procedia Econ. Financ. 2012, 3, 496–501. [CrossRef]
- Stahl, P.D.; Curran, M.F. Collaborative Efforts Towards Ecological Habitat Restoration of a Threatened Species, Greater Sage-Grouse in Wyoming, USA. In *Land Reclamation in Ecological Fragile Areas*; CRC Press: London, UK, 2017; pp. 251–254.
- 9. Cardinale, B.J.; Duffy, J.E.; Gonzalez, A.; Hooper, D.U.; Perrings, C.; Venail, P.; Narwani, A.; Mace, G.M.; Tilman, D.; Wardle, D.A.; et al. Biodiversity loss and its impact on humanity. *Nature* **2012**, *486*, 59–67. [CrossRef]
- 10. United States Department of the Interior; United States Department of Agriculture. *The Gold Book: Surface Operating Standards and Guidelines for Oil and Gas Exploration and Development*, 4th ed.; BLM/WO/ST-06/021+3071/REV 07; Bureau of Land Management: Denver, CO, USA, 2007.
- 11. Curran, M.F.; Stahl, P.D. Database management for large scale reclamation projects in Wyoming: Developing better data acquisition, monitoring, and models for applications to future projects. J. Environ. Solut. Oil Gas Min. 2015, 1, 31–43. [CrossRef]
- 12. Curran, M.F.; Wolff, B.J.; Stahl, P.D. Demonstration Study: Approaching oil and gas pad reclamation with data management: A framework for the future. *J. Am. Soc. Min. Reclam.* **2013**, *2*, 195–204.
- Curran, M.F.; Cox, S.E.; Robinson, T.J.; Robertson, B.L.; Rogers, K.A.; Sherman, Z.A.; Adams, T.A.; Strom, C.F.; Stahl, P.D. Spatially balanced sampling and ground-level imagery for revegetation monitoring on reclaimed well pads. *Restor. Ecol.* 2019, 27, 947–980. [CrossRef]
- 14. Davies, K.W.; Boyd, C.S.; Svejcar, L.N.; Clenet, D.R. Long-Term Effects of Revegetation Efforts in Annual Grass—Invaded Rangeland. *Rangel. Ecol. Mgmt.* **2024**, *92*, 59–67. [CrossRef]
- Curran, M.F.; Crow, T.M.; Hufford, K.M.; Stahl, P.D. Forbs and greater sage-grouse habitat restoration efforts: Suggestions for improving commercial seed availability and restoration practices. *Rangelands* 2015, 37, 211–216. [CrossRef]
- 16. Hoover, D.L.; Bestelmeyer, B.; Grimm, N.B.; Huxman, T.E.; Reed, S.C.; Sala, O.; Seastedt, T.R.; Wilmer, H.; Ferrenberg, S. Traversing the wasteland: A framework for assessing ecological threats to drylands. *BioScience* **2020**, *70*, 35–47. [CrossRef]

- Shackelford, N.; Paterno, G.B.; Winkler, D.E.; Erickson, T.E.; Leger, E.A.; Svejcar, L.N.; Breed, M.F.; Faist, A.M.; Harrison, P.A.; Curran, M.F.; et al. Drivers of seedling establishment success in dryland restoration efforts. *Nat. Ecol.* 2021, *5*, 1283–1290. [CrossRef] [PubMed]
- 18. Sommers, J. *Green River Drift: A History of the Upper Green River Cattle Association;* Falcon Press Publishing Co., Inc.: Helena, MT, USA, 1994.
- 19. Monroe, A.P.; Aldridge, C.L.; O'Donnell, M.S.; Manier, D.J.; Homer, C.G.; Anderson, P.J. Using remote sensing products to predict recovery of vegetation across space and time following energy development. *Ecol. Indic.* 2020, *110*, 105872. [CrossRef]
- Curran, M.F.; Sorenson, J.R.; Robinson, T.J. Soil stockpile age does not impact vegetation establishment in a cold, arid natural gas field. *Recl. Sci.* 2024, 1, 41–47. [CrossRef]
- Mason, A.; Driessen, C.; Norton, J.; Strom, C. First year soil impacts of well-pad development and reclamation on Wyoming's sagebrush steppe. *Nat. Resour. Environ. Issues* 2011, 17, 29–34.
- Curran, M.F.; Sorenson, J.; Stahl, P.D. Rocky mountain beeplant aids revegetation in an arid natural gas field. *Environ. Connect* 2019, 14, 27–29.
- Walker, L.R.R.; Walker, J.; del Moral, R. Forging a new alliance between succession and restoration. In *Linking Restoration and Ecological Succession*; Walker, L.R.R., Walker, J., Hobbs, R.H., Eds.; Springer: New York, NY, USA, 2007; pp. 1–18.
- 24. Archer, S.; Pyke, D.A. Plant-animal interactions affecting plant establishment and persistence on revegetated rangeland. *Rangel. Ecol. Manag.* **1991**, *44*, 558–565. [CrossRef]
- 25. Kearns, C.A.; Inouye, D.W. Pollinators, flowering plants, and conservation biology. Bioscience 1997, 47, 297–307. [CrossRef]
- 26. Dixon, K.W. Pollination and restoration. Science 2009, 325, 571–573. [CrossRef] [PubMed]
- 27. Cusser, S.; Goodell, K. Diversity and distribution of floral resources influence the restoration of plant-pollinator networks on a reclaimed strip mine. *Restor. Ecol.* **2013**, *26*, 713–721. [CrossRef]
- Harmon, J.P.; Ganguli, A.C.; Solga, M.A. An overview of pollination in rangelands: Who, why, and how. *Rangelands* 2011, 33, 4–9. [CrossRef]
- 29. National Research Council. Status of Pollinators in North America; National Academic Press: Washington, DC, USA, 2007; 322p.
- 30. Wilson, E.O. The little things that run the world (the importance and conservation of invertebrates). *Conserv. Biol.* **1987**, *1*, 344–346. [CrossRef]
- 31. Losey, J.E.; Vaughan, M. The economic value of ecological services provided by insects. *Bioscience* 2006, 56, 311–323. [CrossRef]
- 32. Noreiga, J.A.; Hortal, J.; Azcarate, F.M.; Berg, M.P.; Bonada, N.; Briones, M.J.; Del Toro, I.; Goulson, D.; Ibanez, S.; Landis, D.A.; et al. Research trends in ecosystem services provided by insects. *Basic Appl. Ecol.* **2018**, *26*, 8–23. [CrossRef]
- Dangles, O.; Casas, J. Ecosystem services provided by insects for achieving sustainable development goals. *Ecosyst. Serv.* 2019, 35, 109–115. [CrossRef]
- Seibold, S.; Rammer, W.; Hothorn, T.; Seidl, R.; Ulyshen, M.D.; Lorz, J.; Cadotte, M.W.; Lindenmayer, D.B.; Adhikari, Y.P.; Aragon, R. The contribution of insects to global forest deadwood decomposition. *Nature* 2021, 597, 77–81. [CrossRef]
- Belovsky, G.E.; Slade, J.B. Insect herbivory accelerates nutrient cycling and increases plant production. *Proc. Natl. Acad. Sci. USA* 2000, 97, 14412–14417. [CrossRef]
- 36. Tallamy, D. Bringing Nature Home: How Native Plants Sustain Wildlife in Our Gardens; Timber Press: Portland, OR, USA, 2009.
- Dainese, M.; Martin, E.A.; Aizen, M.A.; Albrecht, M.; Bartomeus, I.; Bommarco, R.; Carvalheiro, L.G.; Chaplin-Kramer, R.; Gagic, V.; Garibaldi, L.A. A global synthesis reveals biodiversity-mediated benefits for crop production. *Sci. Adv.* 2019, *5*, eaax0121. [CrossRef] [PubMed]
- 38. Van Klink, R.; Bowler, D.E.; Gongalsky, K.B.; Swengel, A.B.; Gentile, A.; Chase, J.M. Meta-analysis reveals declines in terrestrial but increases in freshwater insect abundances. *Science* **2020**, *368*, 417–420. [CrossRef] [PubMed]
- Grodsky, S.M.; Iglay, R.B.; Sorenson, C.E.; Moorman, C.E. Should invertebrates receive greater attention in wildlife research journals? J. Wildl. Manag. 2018, 79, 529–536. [CrossRef]
- 40. Eisenhauer, N.; Ochoa-Hueso, R.; Huang, Y.; Barry, K.E.; Gebler, A.; Guerra, C.A.; Hines, J.; Jochum, M.; Andraczek, K.; Bucher, S.F.; et al. Ecosystem consequences of invertebrate decline. *Curr. Biol.* **2023**, *33*, 4538–4547. [CrossRef] [PubMed]
- 41. Tonietto, R.K.; Larkin, D.J. Habitat restoration benefits wild bees: A meta-analysis. J. Appl. Ecol. 2018, 55, 582–590. [CrossRef]
- 42. Kimberling, D.N.; Karr, J.R.; Fore, L.S. Measuring human disturbance using terrestrial invertebrates in the shrub-steppe of eastern Washington (USA). *Ecol. Indic.* 2001, *1*, 63–81. [CrossRef]
- 43. Longcore, T. Terrestrial arthropods as indicators of ecological restoration success in coastal sage scrub (California, USA). *Restor. Ecol.* **2003**, *11*, 397–409. [CrossRef]
- 44. Menz, M.M.H.; Phillips, R.D.; Winfree, R.; Kremen, C.; Aizen, M.A.; Johnson, S.D.; Dixon, K.W. Reconnecting plants and pollinators: Challenges in the restoration of pollination mutualisms. *Trends Plant Sci.* **2011**, *16*, 4–12. [CrossRef]
- Sylvain, Z.A.; Espeland, E.K.; Rand, T.A.; West, N.M.; Branson, D.H. Oilfield reclamation recovers productivity but not composition of arthropod herbivores and predators. *Environ. Entomol.* 2019, 48, 299–308. [CrossRef]
- 46. Curran, M.F.; Robinson, T.J.; Guernsey, P.; Sorenson, J.; Crow, T.M.; Smith, D.I.; Stahl, P.D. Insect abundance and diversity respond favorably to vegetation communities on interim reclamation sites in a semi-arid natural gas field. *Land* **2022**, *11*, 527. [CrossRef]
- Curran, M.F.; Sorenson, J.R.; Craft, Z.A.; Crow, T.M.; Robinson, T.J.; Stahl, P.D. Ecological Restoration Practices within a Semi-arid Natural Gas Field Improve Insect Abundance and Diversity during Early and Late Growing Season. *Animals* 2022, *13*, 134. [CrossRef] [PubMed]

- 48. Society for Ecological International Science and Policy Working Group. *The SER International Primer on Restoration Ecology;* Society for Ecological Restoration International: Tucson, AZ, USA, 2004.
- 49. Cagney, J.; Cox, S.E.; Booth, D.T. Comparison of point intercept and image analysis for monitoring rangeland composition and trend. *Rangel. Ecol. Manag.* **2011**, *64*, 309–315. [CrossRef]
- Curran, M.F.; Cox, S.E.; Robinson, T.J.; Strom, C.F.; Stahl, P.D. Combining spatially balanced sampling, route optimization and remote sensing to assess biodiversity response to reclamation practices on semi-arid well pads. *Biodiversity* 2020, 21, 171–181. [CrossRef]
- 51. Curran, M.F.; Hodza, P.; Cox, S.E.; Lanning, S.G.; Robertson, B.L.; Robinson, T.J.; Stahl, P.D. Ground-level unmanned aerial system imagery coupled with spatially balanced sampling and route optimization to monitor rangeland vegetation. *J. Vis. Exp.* **2020**, *160*, e61052. [CrossRef]
- 52. Booth, D.T.; Cox, S.E.; Berryman, R.D. Sampling digital imagery with 'SamplePoint'. *Environ. Monit. Assess* **2006**, 123, 97–108. [CrossRef]
- Ancin-Murguzur, F.J.; Munoz, L.; Monz, C.; Fauchald, P.; Hausner, V. Efficient sampling for ecosystem service supply assessment at a landscape scale. *Ecosyst. People* 2019, 15, 33–41. [CrossRef]
- 54. Wenninger, E.J.; Inouye, R.S. Insect community response to plant diversity and productivity in a sagebrush-steppe ecosystem. *J. Arid Environ.* **2008**, *72*, 24–33. [CrossRef]
- 55. Triplehorn, C.A.; Johnson, N.F. *Borror and DeLong's Introduction to the Study of Insects*, 7th ed.; Thomson Brooks/Cole: Belmont, CA, USA, 2005.
- 56. McAlpine, J.F.; Peterson, B.V.; Shewell, G.E.; Teskey, H.J.; Vockeroth, J.R.; Wood, D.M. *Manual of Nearctic Diptera. Volume* 1; CABI Digital Library: Wallingford, UK, 1981; No. 27.
- 57. McCall, C.; Primack, R.B. Influence of flower characteristics, weather, time of day, and season on insect visitation rates in three plant communities. *Am. J. Bot.* **1992**, *79*, 434–442. [CrossRef]
- 58. Solga, M.J.; Harmon, J.P.; Ganguli, A.C. Timing is everything: An overview of phenological changes to plants and their pollinators. *Nat. Areas J.* **2014**, *34*, 227–235. [CrossRef]
- Curran, M.F.; Summerfield, K.; Alexander, E.-J.; Lanning, S.G.; Schwyter, A.R.; Torres, M.L.; Schell, S.; Vaughan, K.; Robinson, T.J.; Smith, D.I. Use of 3-dimensional videography as a non-lethal way to improve visual insect sampling. *Land* 2020, *9*, 340. [CrossRef]
- 60. R Core Team. *R: A Language and Environment for Statistical Computing;* R Foundation for Statistical Computing: Vienna, Austria, 2016. Available online: https://www.Rproject.org/ (accessed on 1 March 2024).
- 61. Walker, L.R.; del Moral, R. Primary Succession and Ecosystem Rehabilitation; Cambridge University Press: Cambridge, UK, 2003.
- 62. Prach, K.; Walker, L.R. Four opportunities for studies of ecological succession. Trends Ecol. Evol. 2011, 26, 119–123. [CrossRef]
- 63. Fuhlendorf, S.D.; Engle, D.M.; Elmore, R.D.; Limb, R.F.; Bidwell, T.G. Conservation of pattern and process: Developing an alternative paradigm of rangeland management. *Rangel. Ecol. Mgmt.* **2012**, *65*, 579–589. [CrossRef]
- 64. Klebenow, D.A.; Gray, G.M. Food habits of juvenile sage grouse. *Rangel. Ecol. Manag. J. Range Manag. Arch.* **1968**, 21, 80–83. [CrossRef]
- Drut, M.S.; Pyle, W.H.; Crawford, J.A. Diets and food selection of sage grouse chicks in Oregon. J. Range Mgmt. 1994, 47, 90–93.
  [CrossRef]
- 66. Tallamy, D.W.; Shriver, W.G. Are declines in insects and insectivorous birds related? Condor 2021, 123, duaa059. [CrossRef]
- 67. Tallamy, D.W. Do alien plants reduce insect biomass? Conserv. Biol. 2004, 18, 1689–1692. [CrossRef]
- 68. Spafford, R.D.; Lortie, C.J. Sweeping beauty: Is grassland arthropod community composition effectively estimated by sweep netting? *Ecol. Evol.* 2013, *3*, 3347–3358. [CrossRef]
- 69. Copeland, S.M.; Davies, K.W.; Boyd, C.S. Sagebrush Ecosystems are More Than Artemisia: The Complex Issue of Degraded Understories in the Great Basin. *Rangel. Ecol. Manag.* **2024**, *94*, 184–194. [CrossRef]

**Disclaimer/Publisher's Note:** The statements, opinions and data contained in all publications are solely those of the individual author(s) and contributor(s) and not of MDPI and/or the editor(s). MDPI and/or the editor(s) disclaim responsibility for any injury to people or property resulting from any ideas, methods, instructions or products referred to in the content.