

Article

Surface Heterogeneity-Involved Estimation of Sample Size for Accuracy Assessment of Land Cover Product from Satellite Imagery

Huiqun Ren¹, Guoyin Cai^{1,2,*} and Mingyi Du^{1,2}

- ¹ School of Geomatics and Urban Spatial Informatics, Beijing University of Civil Engineering and Architecture, Beijing 100044, China; 2108521517043@stu.bucea.edu.cn (H.R.); dumingyi@bucea.edu.cn (M.D.)
- ² Beijing Advanced Innovation Center for Future Urban Design, Beijing University of Civil Engineering and Architecture, Beijing 100044, China
- * Correspondence: cgyin@bucea.edu.cn

Received: 3 September 2019; Accepted: 9 October 2019; Published: 12 October 2019



Abstract: Sample size estimation is a key issue for validating land cover products derived from satellite images. Based on the fact that present sample size estimation methods account for the characteristics of the Earth's subsurface, this study developed a model for estimating sample size by considering the scale effect and surface heterogeneity. First, we introduced a watershed with different areas to indicate the scale effect on the sample size. Then, by employing an all-subsets regression feature selection method, three landscape indicators describing the aggregation and diversity of the land cover patches were selected (from 14 indicators) as the main factors for indicating the surface heterogeneity. Finally, we developed a multi-level linear model for sample size estimation using explanatory variables, including the estimated sample size (n) calculated from the traditional statistical model, size of the test region, and three landscape indicators. As reference data for developing this model, we employed a case study in the Jiangxi Province using a 30 m spatial resolution global land cover product (Globeland30) from 2010 as a classified map, and national 30 m land use/cover change (LUCC) data from 2010 in China. The results showed that the adjusted square coefficient of R^2 is 0.79, indicating that the joint explanatory ability of all predictive variables in the model to the sample size is 79%. This means that the predictability of this model is at a good level. By comparing the sample size N_S obtained by the developed multi-level linear model and n as calculated from the statistics model, we find that N_S is much smaller than n, which mainly contributes to the concerns regarding surface heterogeneity in this study. The validity of the established model is tested and is proven as effective in the Anhui Province. This indicates that the estimated sample size from considering the scale effect and spatial heterogeneity in this study achieved the same accuracy as that calculated from a probability statistical model, while simultaneously saving more time, labour, and money in the accuracy assessment of a land cover dataset.

Keywords: sample size estimation; land cover products; satellite images; surface heterogeneity; Jiangxi Province

1. Introduction

Land cover provides basic geospatial information for applications in the fields of global environmental change, natural resources management, carbon and nitrogen cycle, and ecological monitoring [1–3]. Because of the continuous earth surface scanning and the correspondingly long-term data archives, satellite remote sensing is proven as an effective way in mapping global land cover and measuring land cover dynamic change [4]. Currently, national and international agencies have successfully created no less than ten global scale land cover datasets with spatial resolutions of 1 km,



500 m, 300 m, 30 m, and 12 m. These existing land cover datasets provide basic geographic information for studying climate, hydrology, environment, ecology, and urban regions [5–8]. Their accuracies are undoubtedly one of the most concerning issues for the potential users. Although data accuracy or uncertainty information is given at the same time as the land cover product releases [9,10], users in different application fields generally need to verify the accuracy before making a decision in using the land cover product [11,12].

Sampling inspection is a commonly used method for verifying the accuracy of land cover products. It provides reliable information on product quality and uncertainty [13,14]. Determining the sample size, as a key procedure in the sampling scheme, lays the foundation for the later stages of sample displacement and verification. A reasonably-estimated sample size is an effective way to avoid the phenomena of over-sampling or under-sampling [15,16]. In addition, the sample size also affects the number of investigators, and the cost and time of the field investigation. Therefore, the estimation of sample size is not only important theoretically, but also plays a guiding role in scientific research and field work investigation.

Currently, there are three major sample size estimation methods: empirical values, fixed-grid sampling, and calculation from a statistical model. The empirical values are the sample numbers required by researchers to test the accuracy of each category of classified satellite images. For example, Hay provided at least 50 empirical values for samples in each category initially [17], although the sample size could be enlarged with an increase in spatial regions and/or the amount of categories involved in the image classification. Congalton provided 75–100 sample data for each category of classified image, which are common empirical values for testing classified images [18]. Empirical sample sizes are simple and can achieve the goal of sample representation through spatial optimization in the process of sample placement. It is usually used to determine the sample size in remote sensing products, especially in an accuracy assessment of classified images. However, with the sprawl of spatial regions, the spatial heterogeneity can vary to a large extent around the entire region(s).

A fixed grid is another commonly used method for determining a sample size for performing an accuracy assessment. It divides the study area into regular grids with a certain size, for instance, 1 km or 10 km, and a sample from each grid is required [19,20]. Ridder used a 10 km \times 10 km grid and randomly selected 9000 samples to assess the accuracy of a global forest dataset [21]. Stehman designed a 5 km \times 5 km grid, in which 500 grids were randomly selected by synthesizing information on global climate zones and population density [22,23]. Through designing a sample encryption algorithm, a dataset for the verification of global land cover products was generated. This dataset can be used to validate other 1 km or 500 m resolution land cover datasets. The fixed grid sampling method is easy to implement, but the determination of grid specifications relies more on expert knowledge.

A statistical model is widely used to calculate a sample size. This method is based on theories of traditional probability and statistics [24,25]. Tong calculates sample sizes from two scales, one on the level of map divisions, and the other on the level of map elements [26]. By establishing probability distribution functions in each scale, the calculated sample size overcomes the problem of 'too strict in map divisions and too wide in map elements', which commonly exists in classical sampling schemes. Based on probability and statistics theory, Olofsson calculated a sample size for simple random sampling and stratified sampling methods respectively, using a statistical model [27]. As compared with the empirical or fixed-grid method, the statistical model is based on theories of statistics and probability. However, the sample size obtained by this method relies on the expected classification accuracy and sampling error information of the products. One of the main problems is that the same remote sensing product in different test sites with different areas will obtain the same numbers of samples, indicating that the surface spatial characteristics, such as patch numbers, aggregation, and diversity, from different study sites, are not involved in the calculation of sample size.

Until now, the determination of a sample size has primarily relied on expert knowledge or conditional assumptions. This often makes it difficult to ensure the rationality of the sample size.

Therefore, this study addresses how to determine a number of samples while considering surface spatial heterogeneity.

Statistical theory is the foundation for determining the sample size. This study derived a sample size estimation using a stratified sampling approach. Then, a multi-level linear sample size estimation model was developed by considering the scale effect and surface spatial heterogeneity, with emphasis on two aspects of these issues. First, a watershed unit with ecological and geographical significance was introduced in this study as the basic spatial unit for performing the accuracy assessment, avoiding the objectivity issues existing in current spatial units of pixels or polygons [28]. Second, landscape indicators were employed to describe the surface heterogeneity and complexity. As the characteristics of the spatial heterogeneity would inevitably affect the sample size used for validating the land cover dataset, this study computed several major landscape indicators and assessed their impacts on the surface heterogeneity in watershed units, thereby reaching the goal of this study (to develop a reasonable model to estimate the sample size).

The remainder of the paper is organised as follows. Section 2 introduces the study area, data sources, and data pre-processing methods. Section 3 describes the commonly used statistical model of sample size estimation, the selection of scale factor and landscape indicators, and the development of the multi-level linear regression model. Section 4 presents results and an analysis of the developed model, and Section 5 provides preliminary conclusions.

2. Study Area and Data

2.1. Study Area

The Jiangxi Province is located in south-eastern China, with a total area of 169,900 square kilometres. It belongs to a subtropical humid climate with abundant rainfall. The landform is dominated by mountains (36%) and hills (42%). The main land use types are forest lands and crop lands, with the forest coverage rate reaching 60%, ranking as first in China. The typical geomorphological and climatic features cause Jiangxi to be covered by various types of land covers. Therefore, we selected the Jiangxi Province as the study site to develop the sample size estimation model [29,30]. The Anhui province which is adjacent to the Jiangxi Province is selected to testify the developed model (Figure 1).



Figure 1. The study site (a) locations in China of (b) the Jiangxi Province and the Anhui Province.

2.2. Data Sources

In this study, the global land cover dataset 'Globeland30' from 2010 was selected as the test dataset (http://www.globeland30.org). Globeland30 is a high-resolution land cover mapping product developed by the National Geomatics Centre of China, with a spatial resolution of 30 m. It has attracted the attention of researchers at home and abroad. The Globeland30 dataset includes 10 first-level classes. An approach based on the integration of pixel- and object-based methods with knowledge (POK) was used to extract land categories, effectively improving the classification accuracy [31].

The reference data is the national land use/cover change (LUCC) dataset from 2010 with a scale of 1:100,000, which was provided by the Institute of Geographic Sciences and Natural Resources Research, Chinese Academy of Sciences (CAS). The data was produced mainly by visual interpretation from remote sensing experts in China and was updated every five years. The LUCC data has six first-level classes and 25 second-level classes. Field investigation has shown that the LUCC classification accuracy is very high, with more than 90% overall accuracy (OA) for the second-level class. Both of these two datasets use the same major data sources of Landsat imagery [32,33]. In addition, the digital elevation model (DEM) dataset of the 'Advanced Spaceborne Thermal Emission and Reflection Radiometer Global Digital Elevation Model' (ASTGTM) (http://datamirror.csdb.cn) in the study area was downloaded to perform the division of the watershed unit.

2.3. Data Processing

2.3.1. Classification System Transformation

The Globeland30 and LUCC datasets must be pre-processed before accuracy assessment, owing to their differences in projection and classification systems. For avoiding area deformation, the Globeland30 dataset was projected to the Albert equal area projection (European Petroleum Survey Group (EPSG) code: 3857), which is the projection system employed for the LUCC dataset. Then, these two products were unified in the classification system. Based on the LUCC definition for each product, the classification transformation is presented in Table 1 [34,35]. Finally, we obtained the test and reference data, as shown in Figure 2.

C	Globeland30	Land Use/C	Cover Change (LUCC)
Index	Class Name	Index	Class Name
10		11	Paddy Land
10	Cultivated Land	12	Dry Land
		21	Forest
20	Forest	23	Woods
		24	Others
		31	Dense Grass
30	Grassland	32	Moderate Grass
		33	Sparse Grass
40	Shrubland	22	Shrub
FO		46	Bottom Land
50	vvetland	64	Swampland
		41	Stream and Rivers
60	Water Bodies	42	Lakes
		43	Reservoir and Ponds
		51	Urban Built-up
80	Artificial Surfaces	52	Rural Settlements
		53	Others

Table 1. Land cover dataset reclassification and corresponding relation.

Globeland30 Land Use/Cover Cha Index Class Name Index C	inge (LUCC)
Index Class Name Index C	0
	ass Name
90 Bare Land 65	Bare soil

67

Others



Table 1. Cont.

Figure 2. Globeland30 dataset and the reclassified land cover from land use/cover change (LUCC).

2.3.2. Digital Elevation Model (DEM) Processing

The DEM was projected into the Albert equal area projection system. A hydrological analysis was implemented by using ArcGIS 10.2 to calculate the watershed units. The watershed unit number was decided by the threshold value of the catchment surface [36–38]. After repeated experiments and visualization on the division of the watershed unit, we used a threshold value of flow accumulation of 4,000,000 for the catchment surface to obtain 48 basin units. Some units with small areas emerged adjacent to the unit that has the largest area. Finally, we collected 30 basin units as shown in Figure 3b, to perform the following study.



Figure 3. (a) 30 m 'ASTER Global Digital Elevation Model' (ASTGTM) digital elevation model (DEM) data and (b) the watershed units (calculated from a).

3. Methodology

3.1. Sample Size Determination from Probability Statistical Model

The estimation of the sample size is dependent on the sampling design method. For a land cover dataset with multiple categories, a stratified random sampling design has often been recommended in related research [39–41]. Cochran provided a set of rigorous calculation equations to obtain a sample size from the perspective of statistical theory, which is the foundation of latter research on sampling techniques [27,42,43]. According to the good practices for estimating accuracy recommended by Olofsson et al. [27], the commonly used equation for calculating sample size is:

$$n = \frac{(\sum W_h S_h)^2}{V(\overline{y}_{st}) + (1/N) \sum W_h S_h^2}$$
(1)

where, $V(\overline{y}_{st})$ is the standard error of the estimated OA, and is generally designated as 0.01, W_h is the stratum weight (proportion of area of class i in the map), $S_h = \sqrt{p_i(1-p_i)}$ [43], S_h represents the level *i* standard deviation, and p_i is the user accuracy, which can be obtained through experiments [42,44].

3.2. Determination of Variables in a Multi-Level Linear Model

On the basis of the estimated sample size from the statistical model, this study attempted to develop a sample size estimation model by considering the scale effects of spatial units and the spatial heterogeneity reflected from the land cover product. As at least three factors, i.e., sample size from the statistical model, spatial effect, and heterogeneity characteristics, impacted the number of samples, this study employed a multi-level linear regression method to develop the sample size estimation, expressed as follows:

$$N_{S} = A_{0} + nA_{1} + CA_{2} + LA_{3} = A_{0} + \frac{(\sum W_{h}S_{h})^{2}}{V(\bar{y}_{st}) + (1/N)\sum W_{h}S_{h}^{2}}A_{1} + CA_{2} + LA_{3}$$
(2)

In the above, N_S is the estimated sample size, n is the sample size calculated by the probability sampling method (Equation (1)), and C represents the sampling constraint on each watershed unit. Theoretically, the larger the test region is, the more samples are needed, and vice versa. In that regard, the 'best' sample constraints for regions with different areas should be discussed in this study. L is a set of spatial heterogeneity factors, including a number of landscape indicators that can indicate information on the fragmentation, diversity, and stability of the watershed units. A_0 , A_1 , A_2 , and A_3 are regression coefficients, and will be discussed in Section 4.

With the help of watershed units, we obtained Globeland30 and LUCC data as test and reference data respectively, for each single watershed. This following part mainly describes the calculation and determination of n, N_S , C, L, and the other variables needed in Equation (2). The main procedures are shown in Figure 4.



Figure 4. Procedures of parameter determination used in a multi-level linear regression method.

3.2.1. Determining N_S , n, and C

Equation (1) in Section 3.1 is used to calculate the sample size n. We can see that all parameters in Equation (1) are available except for S_h . To ensure the optimization and specificity of the n value in each watershed unit, S_h is determined by stratified random sampling for the land cover products of Globeland30. The main steps are as follows (Figure 5): First, each basin unit in the study area is sampled by the stratified random sampling method. Sample N_S values are assigned by 50, 100, 150, ..., 1000, i.e., with an interval of 50. Second, a user's accuracy p_i from different samples N_S in each basin unit is calculated from the constructed error matrix, while using LUCC as a reference data. Third, S_h is obtained for each single basin unit according to the evaluated results.



Figure 5. Flowchart of determining *N*_S, *n*, and *C*.

As an example, Table 2 describes the S_h results for the Number 2 basin unit. The value of h ranges from 1 to 8, representing eight different strata or land types of Globeland30. W_h is the area weight information of each stratum in the basin unit. The S_h for each basin unit can be obtained by the above-mentioned stratified random sampling.

After obtaining the parameter S_h , the sample size n of each N_S can be calculated from Equation (1). As there are 10 different values of N_S in each basin unit, there are 10 different corresponding values of n. The determination of n used in the simulation of the multi-level linear sample size model for a single basin unit should be based on the determination of N_S .

Figure 6 shows broken-line maps between the N_S values obtained from the successive stratified random sampling of 30 basin units and the OA information. The blue line is the result from the stratified random sampling method, the green line represents the result of the accuracy assessed by the whole sample, the black dotted line gives the allowable precision range under the premise of an absolute error of 0.01, and the red circled part is the selected region where the OA varies with N_S in the acceptable precision range. The first N_S value in the circled region is regarded as the estimated sample size [45], aiming to obtain a reasonable result in accuracy assessment by using as few samples as possible. The sample size *n* corresponds to N_S . Once N_S is determined, *n* can also be obtained. Finally, we obtained 30 records of N_S and *n* from the 30 basin units.

Sample constraint *C* describes how large of a region can allocate a sample, which is a factor affecting the number of samples. In this study, the sample constraints of each basin unit were obtained by the following expression:

$$C = N_S / S \tag{3}$$

Here, *S* is the basin unit area (km^2).

	Wı										N	s									
	••••	50	100	150	200	250	300	350	400	450	500	550	600	650	700	750	800	850	900	950	1000
S_1	0.23	0.47	0.41	0.42	0.49	0.49	0.41	0.45	0.41	0.39	0.46	0.45	0.47	0.43	0.44	0.39	0.46	0.47	0.40	0.45	0.44
S_2	0.67	0.41	0.29	0.32	0.33	0.25	0.28	0.34	0.30	0.35	0.31	0.29	0.32	0.30	0.34	0.32	0.32	0.31	0.29	0.31	0.33
S_3	0.05	-	0.00	0.33	0.33	0.27	0.00	0.00	0.25	0.00	0.35	0.28	0.20	0.26	0.40	0.29	0.35	0.31	0.22	0.33	0.25
S_4	0.00	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-
S_5	0.02	0.47	0.50	0.47	0.49	0.00	0.35	0.50	0.33	0.33	0.40	0.49	0.50	0.48	0.43	0.49	0.48	0.46	0.48	0.49	0.43
S_6	0.02	-	0.00	0.47	0.50	0.35	0.50	0.50	0.45	0.47	0.46	0.50	0.49	0.48	0.50	0.39	0.50	0.45	0.50	0.50	0.47
S_7	0.02	0.00	0.00	0.49	0.47	0.50	0.49	0.50	0.47	0.48	0.49	0.49	0.47	0.49	0.48	0.50	0.36	0.39	0.45	0.46	0.43
S_8	0.00	-	-	-	-	-	-	-	-	-	0.00	-	0.00	-	0.00	-	0.00	-	-	0.00	-

 Table 2. Parameter values from watershed unit Number 2.

Note: S_1, \ldots, S_8 corresponds to 10, 20, 30, 40, 50, 60, 80 and 90 land cover categories in Table 1 respectively, - represents that the stratum is not sampled in Globeland30.



Figure 6. Determination of sample size Ns.

The landscape index, an index for quantitative analysis of landscape patterns, can measure the type, quantity, shape, spatial distribution, and complexity of the analysis units [46,47]. In recent years, an increasing number of studies have used the landscape index to describe spatial heterogeneity information, although their focus is not on the estimation of sample size, but on the layout of sample points [45,48] or land cover extraction [49,50]. According to the target of estimating the sample size of the surface coverage data, 14 landscape indicators were selected to describe the spatial heterogeneity information of the landscape levels in the watershed units from seven categories: area metrics, contrast metrics, edge metrics, shape metrics, proximity metrics, aggregation metrics, and diversity indexes. The ecological significance and descriptions of these indicators are presented in Table 3.

Class	Name	Unit	Range	
Area Metrics	Largest Patch Index (LPI)	%	(0,1)	
Contrast Metrics	Mean Patch Size (MPS)	hm ²	>0	
Edge Metrics	Edge Density (ED) Patch density (PD)	m/hm ² –	- >0	
	Landscape shape index (LSI)	-	≥1	
Shape Metrics	Area-weighted mean shape index (AWMSI)	-	(1,2)	
Ĩ	Area-weighted Mean patch fractal dimension (AWMPFD)	ed Mean patch fractal – [1, ion (AWMPFD) – [1,		
Proximity Metrics	Mean proximity index (MPI)	-	>0	
	Shannon's diversity index (SHDI)	-	≥0	
Diversity Metrics	Patch richness density (PRD)	-	>0	
	Shannon's evenness index (SHEI)	-	(0,1)	
Aggregation Metrics	Interspersion and Juxtaposition index (IJI)	%	(0,100)	
00 0	Contagion index (CONTAG)	%	(0,100)	

Table 3. Landscape indicators used in this study.

We need to identify the most representative indicators from the 14 landscape indexes in the 7 categories. All-subsets regression, a commonly used method for feature selection, was employed to select the indicators. By adjusting the values of R^2 , the 'best' model was selected to determine the variables of the fitting model. As shown in Figure 7, we found that the adjusted R^2 value of the 'best' model was 0.78, and the corresponding landscape indices were the landscape shape index (LSI), contagion index (CONTAG), Shannon's evenness index (SHEI), area-weighted mean shape index (AWMSI), area-weighted mean patch fractal dimension (AWMPFD), and patch richness density (PRD). In addition, the R^2 value of the sample size from the probability sampling theory and sample constraints in the fitting model is greater than 0.7, indicating that both of them are independent in the model for sample size estimation. Therefore, the 6 indicators of LSI, CONTAG, SHEI, AWMSI, AWMPFD, and PRD were selected for the following multi-level regression analysis.



Figure 7. Result of all-subsets regression method.

3.2.3. Regression Analysis

Equation (2) indicates that the sample size estimation is supposed to be a multi-level linear function of sample size from probability sampling theory, sample constraints, and spatial heterogeneity. Among them, coefficients A_0 , A_1 , A_2 , and A_3 are determined by regression analysis. In this study, an ordinary least squares (OLS) regression model was used to determine these regression coefficients [51,52]. OLS is one of the most commonly used core methods in multi-level linear regression models. Its form is as follows:

$$\hat{Y}_i = \hat{B}_0 + \hat{B}_1 X_{1i} + \dots + \hat{B}_k X_{ki}$$
(4)

where, i = 1, 2, ..., n, \hat{Y}_i is the predicted value of the dependent variable corresponding to the i^{th} observation, X_{ki} denotes the value of the j^{th} predictive variable corresponding to the i^{th} observation, and \hat{B}_0 denotes the intercept term, i.e., the predicted value of Y when all of the predicted variables are zero. \hat{B}_k is the regression coefficient of the predictive variable j, i.e., the change of Y caused by X_j changing a unit.

The OLS model obtains regression coefficients by reducing the difference between the real values of response variables and the predicted values, i.e., by minimizing the sum of squares of residual errors:

$$\sum_{i=1}^{n} (Y_i - \hat{Y}_i)^2 = \sum_{i=1}^{n} (Y_i - (\hat{B}_0 + \hat{B}_1 X_{1i} + \dots + \hat{B}_k X_{ki}))^2 = \sum_{i=1}^{n} \varepsilon_i^2$$
(5)

4. Result and Analysis

4.1. Multi-Level Regression

As shown in Figure 3b, there are 30 basin units in the study area. Therefore, we obtained a total of 30 records with optimised sample size N_S . To show the correlations between sample size N_S and the explanatory variables, a scatterplot matrix was obtained, and is presented in Figure 8. The diagonal area is the density map of the variables, whereas the blue line in the non-diagonal area represents the linear and smooth-fitting curves. From the matrix, every predictive variable has a tilt trend to some extent. The sample size N_S decreases with the increase of sample constraints (*C*), SHEI, and AWMPFD, whereas it increases with an increase of CONTAG and PRD. This means that the relationship between the sample size N_S and the independent variables is not a phenomenon of simple positive or negative correspondence.



Figure 8. Scatterplot matrix between sample size N_S and the predictive variables.

After analysing the scatterplot matrix, the multi-level linear regression analysis was implemented, and the coefficients are shown in Table 4.

	Estimate	Standard Error	t Value	Pr (> t)	Significance Codes			
(Intercept)	-17,880	5846	-3.058	5.97×10^{-3}	0.01			
n	-0.278	0.076	-3.658	1.47×10^{-3}	0.01			
С	-9.392	1.344	-6.987	6.72×10^{-7}	0.001			
LSI	5.479	1.026	5.34	2.70×10^{-5}	0.001			
CONTAG	109.900	23.380	4.699	1.22×10^{-4}	0.001			
SHEI	7163	1465	4.89	$7.78 imes 10^{-5}$	0.001			
AWMSI	-7.424	4.932	-1.505	1.47×10^{-1}	1			
AWMPFD	5803	3092	1.877	7.45×10^{-2}	0.1			
PRD	48,490	20,160	2.405	2.55×10^{-2}	0.05			
Residual standard error		95.28 0	on 21 degree	es of freedom				
Multiple R ² squared	0.8384							
Adjusted R ² squared		0.7768						
F-statistic		1						
<i>p</i> -value								

Table 4. Multi-level regression coefficients of sample size *N_S*.

The column for Pr(>t) shows the significance of the regression coefficients of the independent variables, and the column of 'Significance codes' represents the degree of significance. The regression coefficients of AWMSI, AWMPFD, and PRD are not significant enough to pass the *t* test, and the multiple R^2 of the model is different from the adjusted R^2 . This reflects a problem of instability. Therefore, the model needs to be improved. From additional experiments, we found that the regression coefficients of the independent variables are more significant when AWMSI, AWMPFD, and PRD are removed. Although the R^2 value decreases, the stability performance improves (Table 5). As a result, only 3 indicators were involved in the final estimation of sample size.

	Estimate	Standard Error	t Value	Pr(> t)	Significance Codes
(Intercept)	-7159	1632	-4.386	1.98×10^{-4}	0.001
n	-0.255	0.070	-3.627	$1.34 imes 10^{-3}$	0.01
С	-9.261	1.366	-6.781	5.16×10^{-7}	0.001
LSI	4.210	0.762	5.523	1.11×10^{-5}	0.001
CONTAG	77.900	16.960	4.592	1.17×10^{-4}	0.001
SHEI	5085	960.500	5.294	1.98×10^{-5}	0.001
Residual standard error		100.7 c	on 24 degree	es of freedom	
Multiple R ² squared			0.793	5	
Adjusted R ² squared			0.750	5	
F-statistic	1				
<i>p</i> -value			1.58×1	0^{-7}	

Table 5. Modified multi-level regression coefficient of sample size N_S .

Therefore, the multi-level linear regression equation can be expressed as follows:

$$N_{\rm S} = -7159 - 0.255n - 9.261C + 4.21LSI + 77.9CONTAG + 5085SHEI$$
(6)

By substituting Equation (1) into Equation (6), we finally obtain the estimation of the sample size by considering the scale effect and spatial heterogeneity characteristics, and it is expressed as follows:

$$N_{S} = -7159 - 0.255 \frac{(\sum W_{h}S_{h})^{2}}{V(\bar{y}_{st}) + (1/N)\sum W_{h}S_{h}^{2}} - 9.261C + 4.21LSI + 77.9CONTAG + 5085SHEI$$
(7)

This relationship shows that LSI, CONTAG, and SHEI contribute positively to the sample size N_S , whereas the sample size n of probability sampling theory and the sample constraint C contribute negatively to the sample size N_S . CONTAG and SHEI contribute the most and play a vital role in the change of sample size. All of the predictive variables explain the variance of 79% of the sample size N_S .

4.2. Model Verification

This study used a cross-validation method to test the generalization ability of the established OLS regression model. Cross-validation, as a commonly used model validation technology, has the advantage of high prediction accuracy [53]. Considering the small amount of data used to fit the model, three-fold cross-validation was used. Three-fold cross-validation was used to divide the original sample into three equal-sized sub-samples, one of which was retained as test data for model validation, while the other two sub-samples were used as training data. This process was repeated three times, and each of the three sub-samples was used as the validation data. The average R^2 of the three cross-validation results was taken as the final estimation solution, as shown in Table 6. The results show that there are some differences between the original R^2 and the three-fold cross-validation R^2 . The R^2 value changes greatly after cross-validation, indicating that the stability between the variables and the generalization ability of the models is less than satisfactory.

Table 6. Cross-validation output results.

Original R-square	0.7935414
Three-fold cross-validated R-square	0.634445
Change	0.1590964

In addition to the cross-validation, we applied the method of sample size estimation to the Anhui Province to test the developed model. Using the same method as mentioned above, we divided the Anhui province into 32 watersheds and selected 14 units to perform the model validation (Figure 9). The same datasets as those used in the Jiangxi Province were used to test this model. First, we

calculated the sample size using a statistical model, i.e., Equation (1), and the developed model in this study, i.e., Equation (7), respectively. Then, we assessed the accuracy of Globeland30 by using the two above-calculated sample sizes. Finally, we compared the accuracies and the results are presented in Table 7. We can see that the difference of OA obtained from two different approaches is very small in most watersheds, the maximum and minimum difference is 4% and 0.1% respectively, with an average of 1.21%. In contrast to the similar accuracy, the sample sizes calculated by the developed model are smaller than that from the statistical model. For the employed 14 watersheds in the Anhui Province, there are 24,425 sample points calculated by the statistical model, our developed model decreases the sample size by 49%. This indicates that a smaller sample size can achieve the same performance as the statistical model by considering the scale effect and surface heterogeneity.



Figure 9. The divided watershed units in the Anhui Province.

Table 7. Comparison of sample size and accuracy calculated from statistical model and this study (No. = Number).

Method	Multi-LEVEI	Model	Statistical N	Aodel	Absolute Value of	Sample Size
Units Code	Ns	OA (%)	n	OA (%)	OA Difference (%)	Difference (N _S -n)
No.1_Ah	475	79.62	1511	79.56	0.06	1035
No.2_Ah	675	79.56	1388	79.54	0.02	714
No.3_Ah	1288	56.06	1561	56.63	0.57	274
No.4_Ah	1110	68.38	1620	68.40	0.02	510
No.5_Ah	485	77.11	1759	77.37	0.26	1274
No.6_Ah	956	75.10	1556	76.56	1.45	601
No.7_Ah	922	73.64	1591	69.94	3.71	669
No.8_Ah	879	66.78	2045	67.63	0.85	1166
No.9_Ah	978	69.19	1979	67.66	1.53	1002
No.10_Ah	1106	59.19	1902	59.94	0.75	796
No.11_Ah	696	52.65	2186	48.86	3.80	1490
No.12_Ah	1128	64.69	1384	66.16	1.47	256
No.13_Ah	912	61.95	2006	63.24	1.29	1093
No.14_Ah	790	63.29	1935	64.50	1.20	1145
	Sum = 12,399		Sum= 24,425		Average $= 1.21$	Sum = 12,026

The OLS regression method is used to analyse the influence of the sample size n of the probability sampling theory, the sample constraints C, and spatial heterogeneity factors on the estimated sample size N_S , and the relationship between them is obtained in Equation (7). The standardised regression coefficient method and the relative weight method are employed to evaluate the relative importance of the predictor variables in the multi-level regression analysis. The results are shown in Table 8 and Figure 10.

 Table 8. Standardised regression coefficients result.

Independent Variable	n	С	LSI	CONTAG	SHEI
Standardised regression coefficients	-3.60×10^{-1}	-8.13×10^{-1}	9.90×10^{-1}	1.97	2.31
Relative importance level	5	4	3	2	1

The standardised regression coefficient is the simplest method for predicting the relative importance of variables. It represents the expected change of response variables caused by the change of one standard deviation of a predictor variable when other predictor variables remain unchanged. Table 8 shows that when the other variables remain unchanged, SHEI changes by a standard deviation, and that the sample size will increase by 2.13 standard deviations, i.e., the most important relative to the sample size N_S . In contrast, the sample size n of the probability sampling theory has the least relative importance to the sample size.

As compared with the standardized regression coefficient, the relative weight is a more promising method for predicting the relative importance of factors [54]. It ranks variables according to their contribution to R^2 . Figure 10 shows the relative importance of each factor. The results show that the sample constraint *C* accounts for 44% of R^2 , which is of the greatest relative importance to the sample size N_S . It shows that the regional scale cannot be neglected in the sample size determination. SHEI and CONTAG, as spatial heterogeneity factors describing the aggregation and diversity of watershed units, explain 25% and 15% of R^2 , respectively. The remaining factors are LSI and *n*. Therefore, in terms of relative importance, the regional scale effect and spatial heterogeneity have an influence on the determination of the sample size in accuracy assessment of remotely sensed land cover products.



Figure 10. Relative importance of the predictor variables used in this study.

4.4. Analysis

Figure 11 is a histogram of the sample size N_S obtained from the multi-level linear model and sample size n and based on probability and statistics theory. Figures 12 and 13 are polygonal maps of the landscape index and sample constraint factor C, respectively. LSI reflects the complexity of the overall landscape shape of the watershed units. The larger the value, the simpler the overall landscape shape. CONTAG describes the degree of agglomeration of patch types in the watershed units. The higher the value, the better the degree of agglomeration. SHEI describes the diversity of the patch distribution in the watershed units. The smaller the value, the simpler the patch type, and the smaller the diversity.



Figure 11. Histogram of sample sizes n and N_S .

 N_S and n are obtained under the premise that the OA standard deviation $V(\overline{y}_{st})$ equals 0.01. According to Figure 11, the value of n is often more than 1000, and most of them are distributed near 1750, whereas the values of N_S are much smaller. This shows that the sample size N_S , as obtained by the multi-level linear model considering spatial heterogeneity, can save more manpower and material resources in the accuracy assessment of a land cover dataset.

The green squares in Figure 11 indicate the relative size of the basin unit area, i.e., the scale difference. According to Figure 3b, there are evident scale differences among the basin units. In theory, the larger the scale of the watershed unit, the more samples are extracted, and the *n* values obtained based on the probability and statistics theory should show similar laws. However, the larger the scale of the unit, the smaller the N_S values needed, such as the basin unit number (No.) 11, No. 20, and No. 35, and the smaller the unit, the larger the $N_{\rm S}$ values, e.g., basin unit No. 14 and No. 17. According to Figures 12 and 13, we can explain why N_S changes with areas of the spatial unit. The values of the LSI, CONTAG, and SHEI in units No. 11, No. 20, and No. 35 are higher, indicating that these few units have a high patch aggregation, uniform patch distribution, and simple landscape shape, and thus weak spatial heterogeneity. Although the scale is large, the required sample size N_S is small. However, the values of LSI and CONTAG in units No. 14 and No. 17 are low, and the SHEI values are high. They need a large sample size of N_S . Therefore, by analysing the relationship between the sample size N_S of the watershed units No. 11, No. 14, No. 17, No. 20, No. 35 and the landscape index, it is further demonstrated that the values calculated from the multi-level sample size estimation model (considering the spatial heterogeneity of land cover) are more reasonable than those calculated from the probability statistical model.



Figure 12. Diagrams of landscape shape index (LSI), contagion index (CONTAG), and C.

As noted above, *C* indicates how large of a sample is taken. If the land cover in different areas is homogeneous and the patch types and distribution are the same, the *C* values are similar. However, in Figure 12, the trend of *C* variation is similar to that of LSI, because the spatial distribution of the patches in different basin units is different. In an area with simple spatial distribution and weak spatial heterogeneity, N_S is small, and the *C* value is high, i.e., *C* has a negative correlation to N_S , whereas when N_S is large and the *C* value is low, *C* has a positive correlation to N_S .

Where the *C* values are the same, e.g., in basin units No. 16 and No. 17, because the LSI value of N_O. 16 is high, the SHEI value is low, and the spatial heterogeneity of the land cover is weaker than that of No. 17. Because the two scales are similar, the sample size of No. 17 is larger. Although the C values in basin units No. 38 and No. 39 are the same, the LSI values of No. 38 are large, but the CONTAG and SHEI values are less than those of No. 39. Therefore, the difference in spatial heterogeneity between No. 38 and No. 39 is smaller than that between No. 16 and No. 17, resulting in a small sample size gap between No. 38 and No. 39. As a result, *C*, as a factor affecting the sample size of N_S , is also influenced by the surface spatial heterogeneity.



Figure 13. Diagram of Shannon's evenness index (SHEI).

5. Conclusions

The accuracy of a dataset is often the first problem to be considered in scientific research and field applications. Sample size calculation is the first step in performing an accuracy assessment of land cover products from satellite imagery. On the basis of a statistical model for the estimation of sample size, this study establishes a multi-level linear model for estimation of sample size by considering the scale effect and spatial heterogeneity. A watershed unit was introduced to obtain a spatial analysis unit, to avoid subjectivity in the selection of assessment units. Landscape indices were selected to indicate the spatial heterogeneity of the region.

The multi-level linear sample size estimation model shows that:

(1): All predictive variables can explain 79% of the variance of the sample size N_S . The coefficients of the predictive variables of the model are significant, indicating that there is a strong relationship between the sample size N_S and the independent variables. By comparing the sample size N_S from the multi-level linear model with a sample size n based on probability and statistics theory, we see that the sample size of N_S is much smaller than that of n. The smaller sample size can achieve the same performance as the statistical model and it contributes to the consideration of surface heterogeneity. The relative importance of the predicted variables in the model is calculated by using standardised regression coefficients and relative weights. The results show that the CONTAG and SHEI indicators (describing the diversity and dispersion of basin units, respectively) are relatively important, followed by LSI, sample constraint C, and sample size n, as calculated from probability sampling theory. According to the validation of the developed model, we can conclude that the smaller sample size from the developed estimation model can achieve the same performance as the statistical model while saving more time, cost, and energy in the accuracy assessment of land cover products.

(2): After performing three-fold cross-validation, the R^2 value changes from 0.79 to 0.63. This means that the generalization of the sample size estimation model is still a problem, although the test of the model in the Anhui Province proved the validity of this estimation of sample size. Therefore, we need more work on the improvement and testing of the developed model for sample size estimation.

For a specific work on accuracy assessment, although the model established in this study cannot be directly applied, it is expected to provide an approach for the determination of sample size, by considering the study areas and the characteristics of the surface heterogeneity. In the future, we need to improve the developed model by applying this surface heterogeneity-concerned sample size estimation model to other study sites, aiming to assess the accuracy of a land cover dataset with as few samples as possible.

Author Contributions: H.R., and G.C. designed the research method, processed the land cover products, performed the model development and wrote the manuscript, M.D. proposed the requirement and revised the manuscript.

Funding: This work is financially supported by National Key Research and Development Program of China (No. 2016YFB0501404) and the National Natural Science Foundation (NSFC) (Key Project # 41930650).

Conflicts of Interest: The authors declare no conflict of interest.

References

- 1. Maria, B.; Monia, M.; Eman, H.; Jun, C.; Ran, L. The First Comprehensive Accuracy Assessment of GlobeLand30 at a National Level: Methodology and Results. *Remote Sens.* **2015**, *7*, 4191–4212.
- 2. Gallego Pinilla, F.J. Comparing CORINE Land Cover with a more detailed database in Arezzo (Italy). *Towards Agric.-Environ. Indic.* **2001**, *1*, 118–125.
- 3. Ren, H.; Cai, G.; Zhao, G.; Li, Z. Accuracy assessment of the globeland30 dataset in jiangxi province. *ISPRS Int. Arch. Photogramm. Remote Sens. Spat. Inf. Sci.* **2018**, *3*, 1481–1487. [CrossRef]
- Chen, J.; Jin, C.; Liao, A.; Xin, C.; Chen, L.; Chen, X.; He, C.; Gang, H.; Shu, P.; Miao, L. Global land cover mapping at 30 m resolution: A POK-based operational approach. *ISPRS J. Photogramm. Remote Sens.* 2015, 103, 7–27. [CrossRef]
- Global Urban Footprint. Available online: http://www.dlr.de/eoc/en/desktopdefault.aspx/tabid-9628/16557_ read-40454/ (accessed on 5 August 2018).
- Esch, T.; Heldens, W.; Hirne, A.; Keil, M.; Marconcini, M.; Roth, A.; Zeidler, J.; Dech, S.; Strano, E. Breaking new ground in mapping human settlements from space—The Global Urban Footprint. *ISPRS J. Photogramm. Remote Sens.* 2017, 134, 30–42. [CrossRef]
- 7. Ban, Y.; Gong, P.; Giri, C. Global land cover mapping using Earth observation satellite data: Recent progresses and challenges. *ISPRS J. Photogramm. Remote Sens.* **2015**, *103*, 1–6. [CrossRef]
- Yang, Y.; Xiao, P.; Feng, X.; Li, H. Accuracy assessment of seven global land cover datasets over China. *ISPRS J. Photogramm. Remote Sens.* 2017, 125, 156–173. [CrossRef]
- Friedl, M.A.; Mciver, D.K.; Hodges, J.C.F.; Zhang, X.Y.; Muchoney, D.; Strahler, A.H.; Woodcock, C.E.; Gopal, S.; Schneider, A.; Cooper, A. Global land cover mapping from MODIS: Algorithms and early results. *Remote Sens. Environ.* 2002, *83*, 287–302. [CrossRef]
- Mayaux, P.; Eva, H.; Gallego, J.; Strahler, A.H.; Herold, M.; Agrawal, S.; Naumov, S.; De Miranda, E.E.; Di Bella, C.M.; Ordoyne, C. Validation of the global land cover 2000 map. *IEEE Trans. Geosci. Remote Sens.* 2006, 44, 1728–1739. [CrossRef]
- 11. Tsendbazar, N.E.; Bruin, S.D.; Herold, M. Assessing global land cover reference datasets for different user communities. *ISPRS J. Photogramm. Remote Sens.* **2014**, *103*, 93–114. [CrossRef]
- 12. Foody, G. Harshness in image classification accuracy assessment. *Int. J. Remote Sens.* **2008**, 29, 3137–3158. [CrossRef]
- Herold, M.; Mayaux, P.; Woodcock, C.E.; Baccini, A.; Schmullius, C. Some challenges in global land cover mapping: An assessment of agreement and accuracy in existing 1 km datasets. *Remote Sens. Environ.* 2008, 112, 2538–2556. [CrossRef]
- 14. Curran, P.J.; Williamson, H.D. Sample size for ground and remotely sensed data. *Remote Sens. Environ.* **1986**, 20, 31–41. [CrossRef]

- 15. Foody, G.M. Sample size determination for image classification accuracy assessment and comparison. *Int. J. Remote Sens.* **2009**, *30*, 5273–5291. [CrossRef]
- Martínez-Abraín, A. Are there any differences? A non-sensical question in ecology. *Acta Oecol.* 2007, 32, 203–206. [CrossRef]
- 17. Hay, A.M. Sampling Designs to Test Land-use Map Accuracy. *Photogramm. Eng. Remote Sens.* **1979**, 45, 529–533.
- 18. Congalton, R.G. A review of assessing the accuracy of classifications of remotely sensed data. *Remote Sens. Environ.* **1991**, *37*, 270–279. [CrossRef]
- 19. Liu, H. Sampling Method with Remote Sensing for Monitoring of Cultivated Land Changes on Large Scale. *Trans. Chin. Soc. Agric. Eng.* **2001**, *17*, 168–171.
- 20. Stehman, S.; Selkowitz, D. A spatially stratified, multi-stage cluster sampling design for assessing accuracy of the Alaska (USA) National Land Cover Database (NLCD). *Int. J. Remote Sens.* **2010**, *31*, 1877–1896. [CrossRef]
- 21. Ridder, R.M. *Options and Recommendations for a Global Remote Sensing Survey of Forests;* Forest Resources Assessment Programme Working Paper 141; FAO: Rome, Italy, 2007.
- 22. Stehman, S.V.; Olofsson, P.; Woodcock, C.E.; Herold, M.; Friedl, M.A. A global land-cover validation data set, II: Augmenting a stratified sampling design to estimate accuracy by region and land-cover class. *Int. J. Remote Sens.* **2012**, *33*, 6975–6993. [CrossRef]
- 23. Stehman, S.V. Impact of sample size allocation when using stratified random sampling to estimate accuracy and area of land-cover change. *Remote Sens. Lett.* **2012**, *3*, 111–120. [CrossRef]
- 24. Fitzpatrick-Lins, K. Comparison of sampling procedures and data analysis for land-use and land-cover map. *Photogramm. Eng. Remote Sens.* **1981**, 47, 343–351.
- 25. Congalton, R.G.; Green, K. Assessing the Accuracy of Remotely Sensed Data—Principles and Practice, 2nd ed.; Lewis Publishers: New York, NY, USA, 1999.
- 26. Tong, X.; Wang, Z.; Xie, H.; Dan, L.; Jiang, Z.; Li, J.; Li, J. Designing a two-rank acceptance sampling plan for quality inspection of geospatial data products. *Comput. Geosci.* **2011**, *37*, 1570–1583. [CrossRef]
- 27. Olofsson, P.; Foody, G.M.; Herold, M.; Stehman, S.V.; Woodcock, C.E.; Wulder, M.A. Good practices for estimating area and assessing accuracy of land change. *Remote Sens. Environ.* **2014**, *148*, 42–57. [CrossRef]
- 28. Stehman, S.V.; Wickham, J.D. Pixels, blocks of pixels, and polygons: Choosing a spatial unit for thematic accuracy assessment. *Remote Sens. Environ.* **2011**, *115*, 3044–3055. [CrossRef]
- 29. Yin, S.; Chen, X.; Chuanqing, W.U.; Yao, Y.; Wang, X. Spatial-temporal analysis on the variations of the vegetation in Jiangxi Province based on NDVI time series. *J. Huazhong Norm. Univ.* **2013**, 47, 129–135.
- 30. Zhao, G.; Cai, G.; Mingyi, D.U. Accuracy Assessment for Land Cover Remote Sensing Mapping Product Based on Landscape Shape Index. *Beijing Surv. Mapp.* **2017**, *1*, 271–277.
- 31. Chen, J.; Chen, J.; Gong, P.; Liao, A.P.; Chao-Ying, H.E. Higher Resolution Global Land Cover Mapping. *Geomat. World* **2011**, *2*, 12–14.
- 32. Wang, S.; Liu, J.; Zhang, Z.; Zhou, Q.; Wang, C. Spatial Pattern Change of Land Use in China in Recent 10 Years. *Acta Geogr. Sin.* **2002**, *57*, 523–530.
- Wang, S.; Liu, J.; Zhang, Z.; Zhou, Q.; Wang, C. Study on spatial pattern and change of land use in recent ten years, China. In Proceedings of the IEEE International Geoscience and Remote Sensing Symposium, Toronto, ON, Canada, 24–28 June 2002; pp. 2369–2371.
- 34. Ma, J.; Sun, Q.; Xiao, P.; Wen, B. Accuracy Assessment and Comparative Analysis of GlobeLand30 Dataset in Henan Province. *J. Geogr.-Inf. Sci.* 2016, *18*, 1563–1572.
- 35. Huang, Y.; Liao, S. Regional accuracy assessments of the first global land cover dataset at 30-meter resolution: A case study of Henan province. *Geogr. Res.* **2016**, *35*, 1433–1446.
- 36. Mao, N.; Liu, W.; Wang, H.; Dai, H. *Arcgis 10 Tutorial: From Beginner to Master*; Surveying and Mapping Publisher: Beijing, China, 2012.
- 37. Yi, W.; Yang, P. Determination of drainage area threshold for extraction of DEM-based digital drainage network. *Jiangxi Hydraul. Sci. Technol.* **2008**, *34*, 259–262.
- Zongmei, L.I.; Zhouqin, M.A.; Nie, Q.; Man, W.; Huang, Y. Method of Ecological Watershed Partitioning. J. China Hydrol. 2017, 37, 27–30.
- 39. Stehman, S.V.; Zhang, J.; Foody, G.M. Sampling designs for accuracy assessment of land cover. *Int. J. Remote Sens.* **2009**, *30*, 5243–5272. [CrossRef]

- 40. Cakir, H.I.; Khorram, S.; Nelson, S.A.C. Correspondence analysis for detecting land cover change. *Remote Sens. Environ.* **2006**, *102*, 306–317. [CrossRef]
- 41. Olofsson, P.; Kuemmerle, T.; Griffiths, P.; Knorn, J.; Baccini, A.; Gancz, V.; Blujdea, V.; Houghton, R.A.; Abrudan, I.V.; Woodcock, C.E. Carbon implications of forest restitution in post-socialist Romania. *Environ. Res. Lett.* **2011**, *6*, 45–202. [CrossRef]
- 42. Badjana, H.M.; Olofsson, P.; Woodcock, C.E.; Helmschrot, J.; Wala, K.; Akpagana, K. Mapping and estimating land change between 2001 and 2013 in a heterogeneous landscape in West Africa: Loss of forestlands and capacity building opportunities. *Int. J. Appl. Earth Obs. Geoinf.* **2017**, *63*, 15–23. [CrossRef]
- 43. Cochran, W.G. Sampling Techniques; John Wiley & Sons.: New York, NY, USA, 1977.
- 44. Olofsson, P.; Foody, G.M.; Stehman, S.V.; Woodcock, C.E. Making better use of accuracy data in land change studies: Estimating accuracy and area and quantifying uncertainty using stratified estimation. *Remote Sens. Environ.* **2013**, *129*, 122–131. [CrossRef]
- Chen, J.; Chen, J.; Hao, W.; Hou, D.Y.; Zhang, W.W.; Zhang, J.; Zhou, X.G.; Chen, L.J. A landscape shape index-based sampling approach for land cover accuracy assessment. *Sci. China Earth Sci.* 2016, 59, 1–12. [CrossRef]
- 46. Crews-Meyer, K.; Hudson, P.; Colditz, R.R. Landscape Complexity and Remote Classification in Eastern Coastal Mexico: Applications of Landsatâ-7 ETM⁺ Data. *Geocarto Int.* **2004**, *19*, 45–56. [CrossRef]
- 47. Neel, M.C.; Mcgarigal, K.; Cushman, S.A. Behavior of class-level landscape metrics across gradients of class aggregation and area. *Landsc. Ecol.* **2003**, *19*, 435–455. [CrossRef]
- Huang, D.; Ke, C.; Wang, Z.; Shuang, L.; Information, S.O.; University, S.O. Accuracy assessment method for remote sensing image classification results based on spatial sampling theory. *Comput. Appl. Softw.* 2016, 33, 190–194.
- 49. Sun, W.; Du, B.; Xiong, S. Quantifying Sub-Pixel Surface Water Coverage in Urban Environments Using Low-Albedo Fraction from Landsat Imagery. *Remote Sens.* **2017**, *9*, 428. [CrossRef]
- 50. Sun, W.; Halevy, A.; Benedetto, J.J.; Czaja, W.; Li, W.; Liu, C.; Shi, B.; Wang, R. Nonlinear Dimensionality Reduction via the ENH-LTSA Method for Hyperspectral Image Classification. *IEEE J. Sel. Top. Appl. Earth Obs. Remote Sens.* **2014**, *7*, 375–388. [CrossRef]
- 51. Cai, G.; Du, M.; Gao, Y. City block-based assessment of land cover components' impacts on the urban thermal environment. *Remote Sens. Appl. Soc. Environ.* **2019**, *13*, 85–96. [CrossRef]
- 52. Kabacoff, R. *R in Action: Data Analysis and Graphics with R*, 2nd ed.; Manning Publications: Shelter Island, NY, USA, 2011.
- 53. Kohavi, R. A Study of Cross-Validation and Bootstrap for Accuracy Estimation and Model Selection. *Int. Jt Conf. Artif. Intell.* **1995**, *2*, 1137–1143.
- 54. Johnson, J.W. Factors Affecting Relative Weights: The Influence of Sampling and Measurement Error. *Organ. Res. Methods* **2004**, *7*, 283–299. [CrossRef]



© 2019 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (http://creativecommons.org/licenses/by/4.0/).