Utilization of Gaseous Carbon Dioxide and Industrial Ca-Rich Waste for Calcium Carbonate Precipitation: A Review

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Abstract: Technologies for the management of various types of waste and the production of useful products from them are currently widely studied. Both carbon dioxide and calcium-rich waste from various production processes are problematic wastes that can be used to produce calcium carbonate. Therefore, the purpose of this paper is to provide an overview about the state of the development of processes that use these two wastes to obtain a valuable $CaCO_3$ powder. The paper reviews the current research on the use of post-distillation liquid from the Solvay process, steelmaking slag, concrete, cement, and gypsum waste as well as some others industrial Ca-rich waste streams in the calcium carbonate precipitation process via carbonation route. This work is an attempt to collect the available information on the possibility of influencing the characteristics of the obtained calcium carbonate. It also indicates the possible limitations and implementation problems of the proposed technologies.

1. Introduction

Carbon dioxide is the main anthropogenic greenhouse gas [1]. The rapid economic growth resulted in an increase in energy demand and, consequently, an increase in the consumption of fuels, in particular fossil fuels such as oil, coal, and natural gas [2,3]. During their combustion, large amounts of CO_2 are emitted into the atmosphere, which has an adverse effect on the environment and is the cause of global warming. There is a strong pressure on the EU countries to make their economies more energy-efficient and zero-emission. However, despite this fact, the projected global coal production continues to increase by around 3%, and the global carbon-related CO_2 emissions from the energy sector will increase by 0.1% annually between 2015 and 2040 [4]. It is not surprising that numerous studies are focused on the management of waste gaseous carbon dioxide [5,6]. Carbon capture and storage (CCS) and carbon capture and utilization (CCU) technologies are the most popular processes to reduce CO_2 emissions [7–9] and play an important role in meeting the global targets specified at COP25 [10]. In the case of both technologies, three main CO_2 capture systems depending on the type of combustion process can be distinguished, post-combustion, pre-combustion, and oxyfuel combustion [6,8,9]. The choice of capture technology depends on the type of plant, i.e., the composition of the exhaust gas [11,12]. Post-combustion technology is the simplest to implement and is mainly based on chemical absorption [9,13,14]. Thus, this option is usually used as a modification to existing power plants. However, due to the low CO_2 content in flue gases (4% in the case of natural gas combustion, 7–14% in the case of coal combustion), it is relatively expensive to obtain a gas stream

with more than 95% CO_2 using this technology. This is due to the fact that low CO_2 concentration negatively affects the capture efficiency [15].

CCS technology is applicable to large CO_2 point emission sources and involves the capture of waste CO_2 from plants, including power plants or cement factories and then its transport to the storage site [11,16]. Most often, waste CO_2 is stored in geological formations. The advantage of CCS technology is its high CO_2 capture efficiency, which is usually above 80% [15]. However, large-scale deployment of CCS technology is not widespread so far due to both economic and technical barriers. First of all, the main obstacle is the unprofitability of such technology, as it requires a large financial investment [17]. Moreover, in some countries such as Norway, Great Britain, India, or Brazil, the geological possibilities for CO_2 storage are very limited, which increases the costs of transport and injection, making the CCS solution an unrealistic option [18].

In order for CO_2 capture to be a cost-effective process, either more efficient absorbents than commonly used should be applied, or alternative technological strategies should be implemented. In recent years, many alternatives to conventional CCS technologies, such as CCU, have been proposed [19–21]. Such technologies allow for reuse of captured CO_2, which may partially offset the total cost of carbon capture and storage [16]. CCU technology includes CO_2 absorption from anthropogenic emissions and its use as a substrate for the synthesis of valuable products such as concrete, plastics, fillers, or reagents for chemical synthesis [4,7,22–24]. One of the processes in which waste carbon dioxide from various sources can be applied as a substrate is carbon mineralization. An overview of mineral carbonation and its market potential has recently been presented [25] and an economic analysis of the various methods of CO_2 utilization has been prepared by Hepburn et al. [26]. They show that mineral carbonation is one of the methods that comes closest to large-scale commercialization due to thermodynamically favored and economic considerations. Furthermore, it is possible to use numerous industrial wastes as other substrates when the product of the mineral carbonation is calcium carbonate. Such solution ensures effective recycling of both gaseous CO_2 and liquid industrial waste, and in addition obtained final product can be applied in many industries [25].

This paper presents the current state of knowledge regarding the reuse of waste gaseous CO_2 and selected Ca-rich waste in the calcium carbonate precipitation process. Presented methods are mainly based on the applying of industrial residues for CO_2 capture by indirect methods and with the addition of additives. The use of post-distillation liquid from the Solvay process, steelmaking slag, concrete, cement, and gypsum wastes as well as some others industrial waste as calcium source has been discussed. The analysis of the conducted studies available in the literature focuses mainly on the characteristics of the obtained calcium carbonate particles and its purity. This is due to the fact that these factors are decisive when considering the commercial use of $CaCO_3$ and the purity of precipitated calcium carbonate (PCC) needs to be above 99.9% (wt.) if it is to be applied in any commercial application. The possible encountered implementation problems of the proposed technologies were also considered.

2. $CaCO_3$ Precipitation Methods

2.1. Carbonation Method

Carbonation is the main method used for the production of calcium carbonate on an industrial scale. This process involves the introduction of gaseous carbon dioxide into the reaction mixture being an aqueous solution containing calcium ions. Calcium hydroxide suspension (gas-suspension system) is used as the source of Ca^{2+} ions and in this case the alkaline environment causes CO_2 to be easily absorbed [27]. Equation (1) provides the overall reaction:

$$Ca(OH)_2 + CO_2 \rightarrow CaCO_3 \downarrow + H_2O, \quad \Delta H_{298}^0 = -108.40 \text{ kJ} \tag{1}$$

The alternative is a gas-liquid reaction with a Ca^{2+} ion source in the form of an aqueous calcium salt solution such as $CaCl_2$ and with the presence of absorption promoter such as amines or ammonia. In the presence of ammonia, the overall reaction presented in Equation (2) occurs in the solution:

$$CaCl_2 + 2NH_4OH + CO_2 \rightarrow CaCO_3 \downarrow + 2NH_4Cl + H_2O, \quad \Delta H_{298}^0 = -118.74 \text{ kJ} \qquad (2)$$

Generation of supersaturation in solution, and consequently induction of calcium carbonate crystals nucleation and growth is possible due to CO_2 absorption, which rate is the limiting factor. Ammonia hydrolysis results in the formation of hydroxyl ions. Bicarbonate ions are formed rapidly by the direct reaction between CO_2 and OH^- ions, after which they are transformed to carbonate ions. The last stage is the reaction between Ca^{2+} and $CO_3{}^{2-}$, whereby $CaCO_3$ is precipitated [28]. Changes in the composition of the aqueous phase cause a change in the solubility of carbon dioxide, which affects the driving force of absorption [29]. The rate of CO_2 absorption in aqueous solutions, in turn, has an impact on the rate of chemical reaction between carbonate and calcium ions, and depends on parameters such as mass transfer coefficient, equilibrium and actual CO_2 solubility, and contact liquid gas. When the absorption of carbon dioxide is accelerated, the rate of precipitation increases as well [29]. Temperature is an important parameter influencing the course of the calcium carbonate formation. The solubility of CO_2 in solution is described by Henry's law and the increase in temperature results in a decrease in CO_2 solubility, and thus in a decrease in the rate of CO_2 absorption [30]. On the other hand, the reactions of $CaCO_3$ formation with the use of gaseous CO_2 are exothermic, i.e., with the release of heat. Therefore, most of the research is conducted under moderate temperature conditions, which favors the CO_2 transfer from the gas phase to the liquid phase.

2.2. Liquid-Liquid Method

Another method of calcium carbonate production used in industry is the liquid-liquid process in which aqueous solutions containing calcium and carbonate ions are mixed [31]. It is often used in laboratory tests due to the ease of controlling process variables [32]. In the literature, a lot of studies can be found in which aqueous solutions of $CaCl_2$ and Na_2CO_3 [33], as well as $Ca(OH)_2$ and H_2CO_3 [34] are used. The source of carbonate ions can also be $(NH_4)_2CO_3$, while calcium acetate as the source of Ca^{2+} ions [35] may be applied. An example of a general reaction can be written as Equation (3):

$$CaCl_2 + Na_2CO_3 \rightarrow CaCO_3 \downarrow + 2NaCl, \quad \Delta H_{298}^0 = 389.75 \text{ kJ} \qquad (3)$$

3. Calcium-Rich Wastes

3.1. Post-Distillation Liquid from Solvay Process

The Solvay method used for the production of sodium carbonate (soda ash) is one of the processes in which waste is a serious environmental problem [36]. The main and most problematic waste generated during this process is post-distillation liquid. It contains mainly water (956 kg/m^3 of waste stream), calcium chloride (112 kg/m^3), and unconverted sodium chloride (56 kg/m^3), but also calcium carbonate (10 kg/m^3), calcium hydroxide (7 kg/m^3), silicon oxide (3 kg/m^3), calcium sulfate (1 kg/m^3) and ammonia (0.1–0.01 kg/m^3) [37,38]. The post-distillation liquid flows out the stripping columns and is directed to the ground settlements called "white seas" or directly to natural water reservoirs [37,39]. However, such activities contribute to the strong salinity of nearby groundwater with calcium chloride, as well as the pollution of natural water reservoirs and soil [39].

Due to the fact that Ca-rich waste from the Solvay process is liquid, it can be used directly as the reaction medium without carrying out a calcium ion extraction step. In this section, two possible paths of post-distillation liquid management for $CaCO_3$ precipitation are discussed: using (i) liquid-liquid method with aqueous Na_2CO_3 solution and (ii) carbonation, in which gaseous waste CO_2 can also be applied.

3.1.1. Precipitation Using Aqueous Carbonate Solution

In most studies, using a post-distillation liquid from the Solvay process containing calcium chloride, sodium carbonate is applied to precipitate $CaCO_3$ as a source of carbonate ions. In this case, the occurring reaction is described by Equation (3). The sodium chloride solution resulting from this reaction, after filtering the calcium carbonate particles, can be recycled and reused in the soda ash manufacturing process [40].

In the research conducted by Kasikowski et al. [37], precipitation of calcium carbonate using calcium ions present in the post-distillation liquid and carbonate ions derived from an aqueous soda ash solution has been proposed. Soda from the Solvay process that does not meet the standards for dense or light soda ash is a substrate for preparing an aqueous Na_2CO_3 solution. Calcium carbonate in calcite form (chalk) obtained by such precipitation is used in the pharmaceutical and cosmetics industry because it is characterized by very high quality, especially low content of ferrous or ferric ions and sodium chloride. In addition, by using well-chosen conditions of the precipitation process, it is possible to obtain pure semi-brine above the precipitated $CaCO_3$. After filtering off calcium carbonate, the semi-brine can be recycled and reused in the production of soda ash or evaporated salts after saturation. For example, about 4000 tons of calcium carbonate is produced annually in the Janikowo Soda Factory (Janikowo, Poland) using this method, and this amount depends on the demand. This technological solution results in a profit of $560,000 per year [37]. In another research, Białowicz et al. [41] used a post-distillation liquid from the Janikowo Soda Factory and sodium bicarbonate solution for the precipitation process of calcium carbonate in the presence of urea. $CaCO_3$ powder was obtained with very low packing density. Moreover, both the increase in reaction time and the increase in urea concentration resulted in an increase in bulk and packing densities [41].

Somani et al. [40] studied the effect of process temperature and sodium carbonate concentration on the crystallization of individual $CaCO_3$ polymorphs. A model post-distillation liquid with a composition of 1.25 M $CaCl_2$ and 0.85 M NaCl was used. Only calcite and aragonite are formed in such a system, but no crystallization of vaterite has been observed. In addition, the increase in sodium carbonate concentration resulted in an increase in the calcite content of the precipitated calcium carbonate. However, the use of diluted Na_2CO_3 solution caused the appearance of aragonite. The formed $CaCO_3$ particles had a rhombohedral and needle-like shape [40]. However, this work did not contain information on the efficiency of the precipitation reaction, as it is not known how many calcium ions present in the initial reaction mixture were consumed to precipitate $CaCO_3$. A diagram showing the calcium carbonate precipitation process by the liquid-liquid method using post-distillation liquid from the Solvay process and an aqueous solution of sodium carbonate is presented in Figure 1a.

Another method proposed by Trypuć et al. [42] is based on $CaCO_3$ precipitation using a post-distillation liquid containing calcium ions and post-filtration solution as a source of carbonate and ammonium ions. Figure 1b shows a diagram of the $CaCO_3$ precipitation process carried out by this method. Post-filtration solution, which is the waste generated after filtering the $NaHCO_3$ precipitate in the Solvay process, consists of CO_3^{2-} (0.69 mol/dm^3), NH_4^+ (4.962 mol/dm^3), Cl^- (4.389 mol/dm^3) and SO_4^{2-} (0.012 mol/dm^3). It was found that increasing the time of the post-filtration liquid dosage into post-distillation liquid causes the increase of both packing and bulk densities. Moreover, the formation of agglomerates and only a small amount of single calcium carbonate crystals was observed. In this method, the post $CaCO_3$ filtration liquid contains NH_4Cl, $(NH_4)_2CO_3$, and unreacted NaCl. NH_4Cl and NaCl can be recovered from this solution using evaporation and fractional crystallization processes. Valuable solid products are obtained in this way. For example, the recovered sodium chloride can be recycled to the Solvay process for brine preparation or used in other industries. Thanks to this it is possible to fully use of Cl^- ions, which in a conventional process are directed as waste to natural water reservoirs. In addition, the proposed precipitation of calcium carbonate at temperatures not exceeding 50 °C results in obtaining chalk with the desired physicochemical properties due to easy control of process conditions [42]. Gao et al. [43] also proposed a method for utilizing both post-distillation liquid and post-filtration solution to precipitate calcium carbonate in the presence of selected surfactants

(CTAB, SDBS and Brij 35). In this process, spherical ultrafine precipitated calcium carbonate was obtained. The results show that the presence of selected surfactants had a significant effect on the mean particle size. Without the addition of surfactants, the average particle size was about 5.5 μm, while the presence of these additives results in the formation of fine particles with a size of about 1.5 μm. Furthermore, the authors suggest that the use of such technology allows soda factories to obtain benefits such as the sale of $CaCO_3$, but also release from fees for environmental pollution [43].
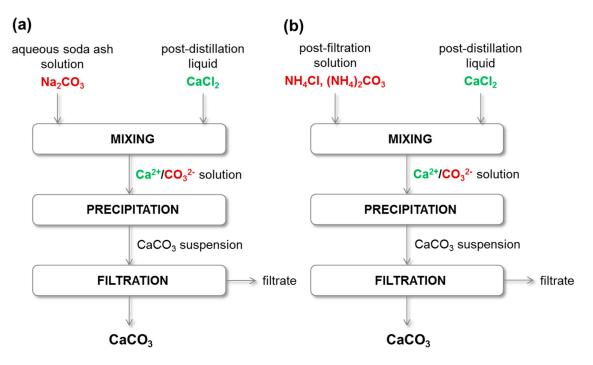

Figure 1. Diagram of the $CaCO_3$ precipitation process by the liquid-liquid method using post-distillation liquid and (**a**) an aqueous solution of Na_2CO_3 or (**b**) post-filtration solution from the Solvay process.

Mikhailova et al. [44] proposed the calcium carbonate precipitation method with the use of post-distillation liquid and excess of mother solution from Solvay process. The authors determined the optimal precipitation conditions leading to the obtaining of calcium carbonate with characteristics qualifying it for commercial applications. It has also been estimated that the application of the proposed process enables the production of 5000 tons of $CaCO_3$ per year, taking into account the production of liquid waste from the Solvay process in Ukraine [44].

Table 1 summarizes the process parameters of the selected calcium carbonate precipitation technologies using a post-distillation liquid and an aqueous sodium carbonate solution (liquid-liquid method). All presented methods were conducted under atmospheric pressure, while the temperature was in the range of 5 to 95 °C. The use of higher temperatures for liquid-liquid processes is due to the fact that the reaction (Equation (3)) is endothermic. In this case, heat is absorbed from the surroundings, so that elevated temperatures favor this reaction.

Table 1. Process parameters of CaCO$_3$ precipitation technologies for Solvay post-distillation liquid utilization using liquid-liquid method.

Ca^{2+} Source	CO$_3$$^{2-}$ Source	Reaction Conditions	Product Characteristic	Ref.
Post-distillation liquid (1.25 mol/dm^3 CaCl$_2$, 0.85 mol/dm^3 NaCl)	Aqueous Na$_2$CO$_3$ solution (0.4–3 mol/dm^3)	30–95 °C Atmospheric pressure	− Only calcite and aragonite are formed, − the ↑ in Na$_2$CO$_3$ concentration resulted in an ↓ in the calcite content of the CaCO$_3$, − the use of diluted Na$_2$CO$_3$ solutions caused the appearance of aragonite, − CaCO$_3$ particles with rhombohedral and needlelike shape.	[40]
Post-distillation liquid from Janikowo Soda Factory, Poland	Aqueous NaHCO$_3$ solution (1.158–1.226 mol/dm^3)	20, 70 °C Atmospheric pressure	− Calcite rhombohedral crystals, − Very low packing density < 1000 g/dm^3, − ↑ reaction time and ↑ urea concentration result in ↑ bulk and packing densities.	[41]
Post-distillation liquid (CaCl$_2$)	Post-filtration solution ((NH$_4$)$_2$CO$_3$)	30 °C Atmospheric pressure	− Precipitated CaCO$_3$ with 99% purity (Al$_2$O$_3$ 0.28–0.34%, SO$_3$ 0.39–0.54%, Fe$_2$O$_3$ 0.1–0.11%), − CaCO$_3$ in the form of calcite and vaterite, − formation of agglomerates.	[42]
Post-distillation liquid (15.28% CaCl$_2$, 0.11% NaCl)	Post-filtration solution (7.31% NaCl, 6.43% (NH$_4$)$_2$CO$_3$, 16.58% NH$_4$Cl)	5–40 °C Atmospheric pressure	− Spherical ultrafine precipitated calcium carbonate was formed, − the obtained products achieve the standards for filler of rubber, paint and pigment industries, − to obtain the spherical ultrafine CaCO$_3$, surfactant must be added to reaction system, − ↓ temperatures results in smaller particles, − ↑ reaction time results in larger particles, − using titrating mixing results in smaller particles compared to rapid dropping mixing.	[43]
Post-distillation liquid (1.35 mol/dm^3 CaCl$_2$, 1.19 mol/dm^3 NaCl)	Mother solution (1.73 mol/dm^3 NaHCO$_3$, 0.5 mol/dm^3 Na$_2$CO$_3$, 0.09 mol/dm^3 NaCl)	70–90 °C Atmospheric pressure	− Bulk weight 0.24 g/cm^3, − specific surface area > 23 m^2/g, − extent of sedimentation > 96.5%.	[44]

↑—increase, ↓—decrease.

3.1.2. Precipitation Using Flue Gas

Technology that allows the use of flue gases to precipitate calcium carbonate from post-distillation liquid has also been proposed [45]. It can be a variation of the wet-lime method, in which the absorption of combustion gases arising during combustion occurs in the overflow of distiller waste [37]. In the Solvay process, an excess of lime milk is added during the regeneration of ammonia from the filtration liquid. As a result, the distillation waste is strongly alkaline, which promotes the absorption of acid flue gas. However, it should be noted that several products are obtained, depending on the reacting gas component.

The absorption of carbon dioxide results in the formation of calcium carbonate, while the absorption of sulfur oxides causes the precipitation of calcium sulfate and sulfite. The reactions that occur during the absorption of the flue gas are presented below (Equations (4)–(9)):

$$Ca(OH)_2 + CO_2 \rightarrow CaCO_3 \downarrow + H_2O, \quad \Delta H_{298}^0 = -108.40 \text{ kJ} \tag{4}$$

$$Ca(OH)_2 + SO_2 \rightarrow CaSO_3 \downarrow + H_2O, \quad \Delta H_{298}^0 = -145.92 \text{ kJ} \tag{5}$$

$$CaCO_3 + SO_2 \rightarrow CaSO_3 \downarrow + CO_2, \quad \Delta H_{298}^0 = -49.52 \text{ kJ} \tag{6}$$

$$Ca(OH)_2 + SO_3 + H_2O \rightarrow CaSO_4 \cdot 2H_2O \downarrow, \quad \Delta H_{298}^0 = -336.73 \text{ kJ} \tag{7}$$

$$CaCO_3 + SO_3 + 2H_2O \rightarrow CaSO_4 \cdot 2H_2O \downarrow + CO_2, \quad \Delta H_{298}^0 = -240.33 \text{ kJ} \tag{8}$$

$$2CaSO_3 + O_2 + 4H_2O \rightarrow 2CaSO_4 \cdot 2H_2O \downarrow, \quad \Delta H_{298}^0 = -579.40 \text{ kJ} \tag{9}$$

It is also possible to simultaneously use waste gases containing carbon dioxide and post-distillation liquid from the Solvay process. However, of the individual components of distillation waste, only $Ca(OH)_2$ can react with CO_2 according to Equation (4). Calcium chloride, although its concentration in the liquid is still high, does not react with carbon dioxide. This is because $CaCl_2$ is the salt of a strong acid while carbonic acid is a weak acid. Therefore, to create conditions conducive to CO_2 absorption and the formation of carbonate ions in the solution, an absorption promoter, such as amines or ammonia, should be added to facilitate the CO_2 transfer from the gas phase into the liquid phase [37,46].

Barzagli et al. [22] proposed a semi-continuous technique for capturing carbon dioxide from a gas mixture applying a dilute aqueous Na_2CO_3 solution and $CaCl_2$ aqueous solution obtained as a by-product in the Solvay process. Firstly, sodium carbonate solution is saturated with CO_2 and then such mixture can react with calcium chloride solution. Using this method, 80% absorption efficiency is obtained and high-quality calcium carbonate is precipitated. This technique may allow to minimize the energy demand due to the fact that the process is carried out at room conditions. By using this method, the main disadvantages of processes using aqueous solutions of alkanolamines, ammonia, or alkali metal carbonates are avoided. There is no energy-consuming regeneration of the absorbent, or costs associated with the evaporative loss of ammonia and oxidative and thermal degradation of amines [22]. The scheme of this method is presented in Figure 2.

In another study, a method for CO_2 sequestration from flue gases using a model post-distillation liquid from Solvay process in the presence of ammonia has been proposed [47]. The scheme of this method is shown in Figure 3. As a result, calcium carbonate precipitates mainly in the form of vaterite and is a valuable product that can be used, among others, for paper production or as an adsorbent for heavy metal ions removal [48]. An increase in ammonia concentration favored CO_2 and Ca^{2+} consumption. About 80% of calcium ions from the post-distillation liquid and about 83% of carbon dioxide from a mixed air-CO_2 gas stream were consumed under the most favorable conditions of this one-stage precipitation process [47]. After filtering off the $CaCO_3$ particles, a waste solution containing ammonium chloride is formed. However, it can be used to recover metal ions or as a leaching agent [49,50]. The NH_4Cl solution can also be applied in the carbon capture and storage process [51].
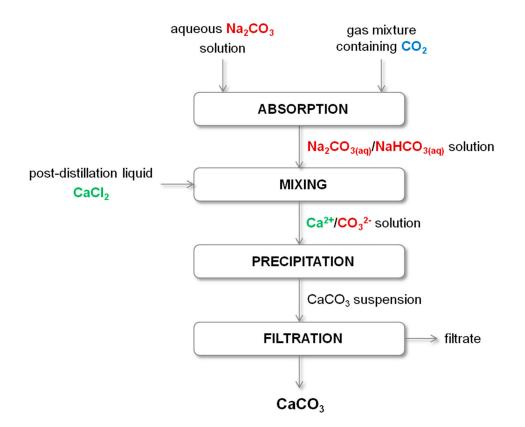
Figure 2. Diagram of the $CaCO_3$ precipitation process using post-distillation liquid from the Solvay process and aqueous Na_2CO_3 solution saturated with CO_2.
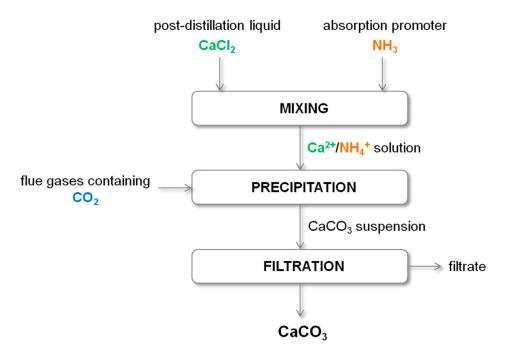

Figure 3. Diagram of the $CaCO_3$ precipitation process by the gas-liquid method using post-distillation liquid from the Solvay process and CO_2 from flue gases.

Li et al. [52] proposed a new technology for processing post-distillation liquid and waste gaseous carbon dioxide. Calcium chloride is converted into calcium carbonate and hydrogen chloride based on the reactive extraction and crystallization process. Isoamyl alcohol and commercial tertiary amine

N235 were used as the extraction system. The conversion rate of calcium chloride under optimal process conditions is 75%.

During this method, the overall reaction described by Equation (10) occurs. The addition of a tertiary amine results in a constant shift of equilibrium in Equation (10) in the right direction:

$$CaCl_2 + CO_2 + 2R_3N + H_2O \rightarrow CaCO_3 \downarrow + 2HCl + R_3N \cdot HCl \qquad (10)$$

An important issue to consider in processes using liquid waste are the additional compounds and ions present in them. In the case of post-distillation liquid from the Solvay process, they are, among others, Na^+, K^+, Mg^{2+}, Al^{3+} cations, and SO_4^{2-}, NO_3^- anions. The influence of these ions on $CaCO_3$ precipitated by carbonation was studied by Dong et al. [53]. It turns out that Na^+, K^+, and NO_3^- have a very weak effect on calcium carbonate, and therefore small amounts of these ions can remain in solution without consequences and no removal is necessary. However, Mg^{2+} and SO_4^{2-} can adsorb on the crystal surface at the early crystallization stage and inhibit the transformation of vaterite into calcite [54]. Magnesium ions also promote the growth of aragonite and inhibit calcite precipitation [55]. Therefore, the content of these ions should be reduced to 1% before using the liquid as a precipitation medium. In addition, a significant amount of very small particles with a diameter less than 1 μm are due to the presence of SO_4^{2-}. Accordingly, these anions must be removed before industrial production [53].

Table 2 summarizes the process parameters of the selected calcium carbonate precipitation technologies using a post-distillation liquid and gaseous carbon dioxide (carbonation process). When analyzing the process parameters of all presented methods, it is worth noting that in each case the process pressure is atmospheric, while in most cases the temperature is moderate, which corresponds to the conditions most favorable to exothermic CO_2 absorption.

Table 2. Process parameters of $CaCO_3$ precipitation technologies for Solvay post-distillation liquid utilization using carbonation method.

Ca^{2+} Source	Gas Mixture Composition	Temperature	Pressure		Product Characteristic	Ref.
Aqueous $CaCl_2$ solution	15% vol. CO_2 in air	20 °C	Atmospheric	–	High-quality calcium carbonate.	[22]
Post-distillation liquid	Fumes created during the combustion of 3 g coal samples	20, 40, 60 °C	Atmospheric	–	Calcite as main form of $CaCO_3$.	[37]
Post-distillation liquid (112 g/dm^3 $CaCl_2$, 56 g/dm^3 NaCl)	15 and 100% vol. CO_2 in air	Room	Atmospheric	– –	Fine $CaCO_3$ particles in vaterite form, the precipitated particles are characterized by a size below 5 μm.	[47]
Aqueous $CaCl_2$ solution (1 mol/dm^3)	100% vol. CO_2	20 °C	Atmospheric	– – – –	Very weak Na^+, K^+, NO_3^- effect on precipitated product, Mg^{2+} can adsorb on the surface of the $CaCO_3$ crystal grain or embedded in the lattice to form Mg-calcite, Mg^{2+} and SO_4^{2-} can adsorb on the crystal surface on the early crystallization stage and inhibit transformation of vaterite into calcite, average particle diameter of 22.354 μm.	[53]
Aqueous $CaCl_2$ solution (0.025–3 mol/dm^3)	100% vol. CO_2	5–55 °C	Atmospheric	– – –	Average particle diameter of 40 μm, rhombohedral calcite, high purity $CaCO_3$.	[52]

3.2. Steelmaking Slag

Steel production is one of the largest industries in the world where wastes are a serious problem [56,57]. Most of the steel is produced by an integrated technological path, in which limestone or burnt lime is applied to remove impurities from molten iron and crude steel. As a result of this treatment, steelmaking slags, which mainly consist of calcium silicates and calcium oxide,

are formed. Their specific composition depends on the production process. Steelmaking slags are used, among others, as lime in agriculture, in road construction as a substitute for gravel [58], or in cement and concretes as an additional cementing material [59]. However, it is not always possible to use them economically due to stringent environmental regulations. In such situation, they are problematic postproduction waste. Therefore, alternative methods of utilizing steelmaking slag are sought [60]. One of the proposed approaches is the use of this waste for the production of calcium carbonate by carbonation route (according to Equations (11) and (12)) due to the high content of calcium oxide. The resulting $CaCO_3$ could replace the limestone used for purification in the steel industry. However, if the manufactured product were of sufficiently high quality, then it could be sold and used in other industries [61]:

$$CaO + H_2O \rightarrow Ca(OH)_2, \quad \Delta H^0_{298} = -65.30 \text{ kJ} \tag{11}$$

$$Ca(OH)_2 + CO_2 \rightarrow CaCO_3 \downarrow + H_2O, \quad \Delta H^0_{298} = -108.40 \text{ kJ} \tag{12}$$

Potentially, calcium silicate (wollastonite) is the largest source of calcium ions in steelmaking slag. According to Kojima et al. [62], the rate of direct CO_2 reaction with $CaSiO_3$ is too slow to be practically used. Thus, to use calcium from $CaSiO_3$, as well as to prevent unwanted pollution, the separation of calcium from the slag must first be carried out. It results in an accelerated reaction rate between Ca^{2+} and gaseous CO_2. One of the methods reported in the literature is the dissolution of calcium silicates with acetic acid according to Equation (13). A solution of calcium acetate and solid silicon oxide are formed. Gaseous CO_2 is then introduced into the solution to precipitate calcium carbonate by carbonation according to Equation (14). The acetic acid added in the first stage is regenerated and can be reused [63]. Furthermore, flue gases from concentrated emission sources can be a source of gaseous CO_2, which allows for its sequestration and removal:

$$CaSiO_3 + 2CH_3COOH \rightarrow Ca^{2+} + 2CH_3COO^- + H_2O + SiO_2, \quad \Delta H^0_{298} = -74.52 \text{ kJ} \tag{13}$$

$$Ca^{2+} + 2CH_3COO^- + H_2O + CO_2 \rightarrow CaCO_3 \downarrow + 2CH_3COOH, \quad \Delta H^0_{298} = -19.58 \text{ kJ} \tag{14}$$

Eloneva et al. [61] investigated $CaCO_3$ precipitation process using gaseous CO_2 and waste steel converter slag, the diagram of which is presented in Figure 4. In this research, a modification of the method proposed by Kakizawa et al. [63] was used because acetic acid produced simultaneously with calcium carbonate prevents the precipitation of most calcium from the solution in the case of low acetic acid concentrations (30% or less). Therefore, the addition of sodium hydroxide was used to prevent the formation of acetic acid and favors the formation of $CaCO_3$. This method is described in Equations (15)–(17). When a higher concentration of acetic acid is used, a greater efficiency in leaching of calcium ions is obtained, but at the same time the dissolution of other undesirable impurities (like silicon and iron) is increased. 86% of Ca^{2+} ions present in the solution were consumed to precipitate calcium carbonate, while the obtained product had a purity of 99.5–99.8%. In addition, CO_2 concentration has no effect on either the quality or the amount of $CaCO_3$ precipitated. This is very beneficial because flue gases can be used directly in this process without prior CO_2 separation and purification. The lowest CO_2 concentration (10% vol.) resulted in longer precipitation time, but this effect can be eliminated by raising the temperature or increasing the flow rate. This is the solution to the problem of low concentration of carbon dioxide in the exhaust gas. It has been estimated that using the proposed method, a maximum of 42,000 t per year of pure $CaCO_3$ can be obtained from the annual production of waste steel converter slag from the Raahe Works factory (Raahe, Finland). Although this process is also suitable for other steel plants, the necessary consumption of NaOH and CH_3COOH results in too high process costs, so its implementation cannot be economical. This problem can be

solved by replacing acetic acid with a salt solution and sodium hydroxide with a certain alkaline liquid waste stream, which would reduce the cost of the process [61]:

$$CaSiO_3 + 2CH_3COOH \rightarrow Ca^{2+} + 2CH_3COO^- + H_2O + SiO_2, \quad \Delta H^0_{298} = -74.52 \text{ kJ} \qquad (15)$$

$$Ca(CH_3COO)_2 + 2NaOH \rightarrow 2CH_3COONa + Ca(OH)_2, \quad \Delta H^0_{298} = -3.90 \text{ kJ} \qquad (16)$$

$$Ca(OH)_2 + CO_2 \rightarrow CaCO_3 \downarrow + H_2O, \quad \Delta H^0_{298} = -108.40 \text{ kJ} \qquad (17)$$
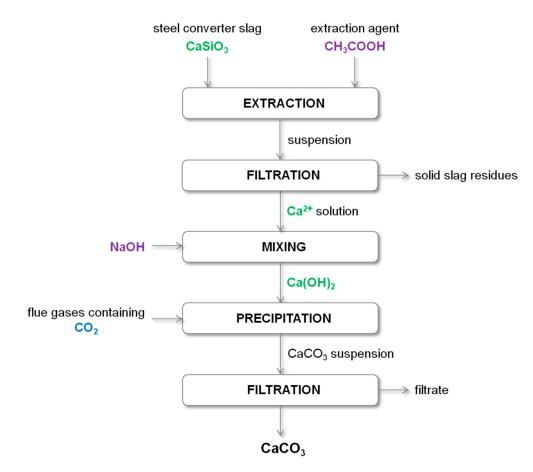

Figure 4. Diagram of the $CaCO_3$ precipitation process using steel converter slag and CO_2 from flue gases.

In another method of using steel converter slag for the production of $CaCO_3$ by carbonization, calcium ions are extracted from this waste with an aqueous solution of ammonium chloride, ammonium nitrate and ammonium acetate [58,64,65] using different slag to liquid ratios. To prevent the escape of ammonia, Ca^{2+} extraction is conducted in a closed reactor. The carbonization process is carried out in an open reactor at room temperature and atmospheric pressure, which reduces energy demand and costs [64]. The resulting calcium carbonate has high quality. The main problem of the proposed method is the formation of waste residual slag. Finding a method for recycling this waste can generate additional process costs [66]. It has been estimated that from 4.7 ton steel converter slag it is possible to produce 2.3 tons of $CaCO_3$ with simultaneous production of 3.4 tons of residual slag. The analysis was made on the example of Raahe Works, which produces about 200 kt steel slag per year. Assuming the consumption of all waste slag, 42,000 tons of CO_2 can be bound in the form of about 96,000 tons of calcium carbonate per year, which will generate an income of around 10 million euro per year [64]. However, it should be noted here that these estimates do not take into account other costs of running the process. Notwithstanding, this solution has great economic potential. Studies using an

aqueous solution of ammonium chloride as the extraction agent for Ca^{2+} ions have also been carried out [51,67,68]. Currently, it is the most commonly used method, because it turned out to give the highest Ca^{2+} selectivity [69].

Table 3 summarizes the reaction conditions for calcium carbonate precipitation depending on the type of slag and the extraction agent and presents the characteristics of the obtained product. Atmospheric pressure was used in most of the research, while the tested temperature range was 20 to 80 °C. Due to the fact that CO_2 absorption is an exothermic process, the increased temperature does not have a positive effect on it, although it has an influence on the polymorphism of the obtained $CaCO_3$. Moreover, the use of elevated temperature generates additional costs of the process. As the literature shows, carbonation processes using steelmaking slag as a source of calcium ions and gas streams containing CO_2 allow to obtain highly pure calcium carbonate, which can be used as a substrate in other industries. However, there are many economic issues to consider as to whether the proposed methods are worth implementing.

Table 3. Process parameters of $CaCO_3$ precipitation technologies for steelmaking slag utilization using carbonation method.

Type of Slag	Extraction Agent	Gas Mixture Composition	Temperature	Pressure	Product Characteristic	Ref.
Oxygen furnace steelmaking slag	NH_4Cl aqueous solution	n.d.	40, 60, 80 °C	10, 20, 30, 40 bar	− 95% $CaCO_3$, − impurities: crystallite $MgCO_3$, Al_2O_3, Fe_2O_3, SiO_3, − formation of agglomerates, − ↑ of the carbonation temperature results in changes from scalenohedral to rhombohedral morphology.	[51]
Steel converter slag	Aqueous acetic acid solution (1 M)	10, 25, 50, 100% vol. CO_2 in N_2	30, 70 °C	Atmospheric	− $CaCO_3$ in calcite form, − very high ISO-brightness (98.7%), − mean particle size of 0.6 μm, − rhombohedral shape.	[61]
Steel converter slag	Aqueous ammonium salt solution (NH_4Cl, NH_4NO_3, CH_3COONH_4)	10, 25, 50, 100% vol. CO_2 in N_2	20–70 °C	Atmospheric	− Formed product is a 98.7% pure $CaCO_3$, − vaterite form.	[65]
Steel converter slag	1 M NH_4Cl aqueous solution	n.d.	20, 45 °C	Atmospheric	− $CaCO_3$ with 99.53% purity (impurities: Si, V, Mg, Cl), − rhombohedral calcite (20 °C), needle-like aragonite (45 °C).	[67]
Steelmaking slag	1 M NH_4Cl aqueous solution	99.999% vol. CO_2	25–80 °C	Atmospheric	− Round vaterite-phase particles (25 °C), − ↑ of temperature results in obtaining needle-like aragonite.	[68]
Argon oxygen decarburisation slag	2 M aqueous ammonium salt solution (NH_4Cl, NH_4NO_3, CH_3COONH_4)	22% vol. CO_2	60 °C	Atmospheric	− $CaCO_3$ in aragonite form, − 60% particles with size < 1 μm, − precipitated $CaCO_3$ moisture of 30% wt.	[70]

n.d.—no data, ↑—increase.

3.3. Concrete Wastes

Concrete as the main building material in the world, is manufactured in an amount of 4000 Mt per year (2017) and its market increased by 2.5% annually, while producing approximately 7% of global carbon dioxide emissions [71,72]. For example, the production of 1 kg of Portland cement generates more than 0.8 kg of CO_2. Waste concrete from demolished buildings and other constructions is a

serious problem, because their quantity is constantly increasing. Therefore, possible technologies for its utilization or reuse are sought [73]. There is a lot of research regarding the use of waste concrete for the production of PCC [74]. The potential amount of calcium that can be recovered from waste concrete depends on the content of this element in cement. For example, Portland cement contains 46.53% wt. of calcium [75], whereas concrete consists of 37% wt. coarse aggregate, 33% sand, 19% cement, and 11% water. Therefore, 88 kg of calcium can be recovered from a ton of Portland cement-based concrete [76]. Commonly used processes for the chemical recovery of calcium from waste concrete include three energy-intensive stages: crushing, milling, and leaching [74]. For the production of PCC from waste concrete, numerous studies have been conducted on various leaching agents. Water [77–79], acetic acid [80], ammonium salts [69,81], and nitric acid [82] have already been used. These solvents showed a relatively low calcium leaching efficiency of less than 60%, although higher pressure (about 10 bar) was used to improve the water leach. In addition, when producing $CaCO_3$, it is recommended to use an even lower yield (around 30%) to avoid contamination of the final product. An alternative solution is to use a stronger acid, e.g., hydrochloric acid [83,84], which maximizes the leaching efficiency. However, carbonization does not occur in a solution with a low pH, so when using an acid as an extraction agent, a base, such as ammonia, must be added to the resulting Ca-rich solution [85,86].

Jo et al. [87] examined the impact of the type of extraction agent on the waste concrete carbonation process. In this research, aqueous solutions of ammonium chloride, acetic acid, hydrochloric acid, and pure deionized water were used, and the processes were carried out at room temperature and atmospheric pressure. The results indicate that the type of extraction agent has no effect on calcium carbonate precipitation, because in each case the resulting precipitate consisted mainly of $CaCO_3$ in the form of a mixture of rhombohedral calcite and spherical vaterite. However, the weight fractions of individual polymorphs were not determined. However, when NH_4Cl solution was used, the highest extraction and carbonization efficiency were obtained.

Van der Zee and Zeman [74] proposed a method of producing precipitated $CaCO_3$ from waste concrete, where HCl was used as the leaching agent. Precipitation of calcium carbonate was carried out in a separate reactor by mixing the calcium-rich solution obtained by leaching and aqueous sodium carbonate solution. Na_2CO_3 solution can be produced by introducing any gas stream containing carbon dioxide into an aqueous NaOH solution. The overall reaction is shown in Equation (18) and the diagram of this process is shown in Figure 5. The HCl and NaOH used in this process were regenerated by membrane bipolar electrodialysis, which is an additional step requiring energy input. The authors suggest that reducing the costs of this stage can be achieved by using HCl concentration lower than 0.5 M:

$$(CaO)SiO_2 + 2HCl + 2NaOH + CO_2 \rightarrow CaCO_3 \downarrow + 2NaCl + SiO_2 + 2H_2O,$$
$$\Delta H_{298}^0 = -213.52 \text{ kJ} \tag{18}$$

Another important problem in the concrete industry is the formation of concrete sludge (waste cement) when fresh concrete is used to manufacture construction or concrete elements [88]. Concrete sludge is a fine fraction recovered during the recycling of waste concrete as a coarse aggregate [76] and has a composition very similar to fresh concrete diluted with water, because it is calcium-rich and strongly alkaline. An example composition of concrete sludge is water (77.1% wt.), CaO (17.2% wt.), SiO_2 (3.2% wt.), Fe_2O_3 (1.9% wt.) and Al_2O_3 (0.6% wt.) [89]. Usually, the utilization of this waste consists in its reuse as an additive to road materials or its storage in landfills [89]. However, the first solution is an expensive process because it requires the separation of solids from liquids, resulting in a strongly alkaline liquid that must be neutralized with acid. On the other hand, the storage of concrete sludge in landfills is environmentally harmful. Therefore, alternative methods are sought to dispose of problematic cement waste [89].
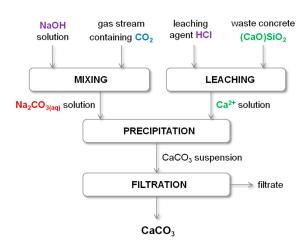
Figure 5. Diagram of the $CaCO_3$ precipitation process using waste concrete and CO_2 from flue gases.

One of the solutions is the use of concrete sludge as a raw material in the carbonation process, which results in the precipitation of calcium carbonate. Cement hydration products, such as calcium hydroxide, calcium silicate hydrate, and calcium aluminate hydrate, react with the gaseous carbon dioxide introduced and dissolved in the aqueous solution [90,91]. Occurring reactions are described in Equations (19)–(21):

$$Ca(OH)_2 + CO_2 \rightarrow CaCO_3 \downarrow + H_2O, \quad \Delta H^0_{298} = -108.40 \text{ kJ} \tag{19}$$

$$3CaO \cdot 2SiO_2 \cdot 3H_2O + 3CO_2 \rightarrow 3CaCO_3 \downarrow + 2SiO_2 + 3H_2O, \quad \Delta H^0_{298} = -265.56 \text{ kJ} \tag{20}$$

$$4CaO \cdot Al_2O_3 \cdot 13H_2O + 4CO_2 \rightarrow 4CaCO_3 \downarrow + 2Al(OH)_3 + 10H_2O,$$
$$\Delta H^0_{298} = -345.86 \text{ kJ} \tag{21}$$

Iizuka et al. [92] conducted research on bench-scale installation for the concrete sludge recycling process using gaseous CO_2. Industrial concrete sludge and flue gas from oil combustion boilers were applied. Authors proposed a two-stage recycling method for concrete sludge, the diagram of which is presented in Figure 6. The first step is the extraction of calcium ions, which involves diluting the concrete sludge with water (hydration) and extracting Ca^{2+} ions into the water phase assisted by mixing (Equation (22)). In another work, Ca^{2+} extraction from waste concrete was carried out by carbonic acid solution prepared using pressurized carbon dioxide [93,94]. The remaining precipitate is then separated from the extract. Gaseous carbon dioxide is introduced into the resulting calcium-rich solution, thereby precipitating calcium carbonate. Importantly, the proposed method can directly use fumes from the combustion of fossil fuels without purifying or compressing them. The solution formed after filtering the $CaCO_3$ particles is neutralized with carbonic acid. In addition, the resulting calcium carbonate can be used in many industries. The feasibility of the proposed technology was also assessed. However, commercialization of this process has not been shown to be economically viable [92]:

$$2(3CaO \cdot SiO_2) + 7H_2O \rightarrow 3CaO \cdot 2SiO_2 \cdot 4H_2O + 3Ca(OH)_2, \quad \Delta H^0_{298} = -157.82 \text{ kJ} \tag{22}$$
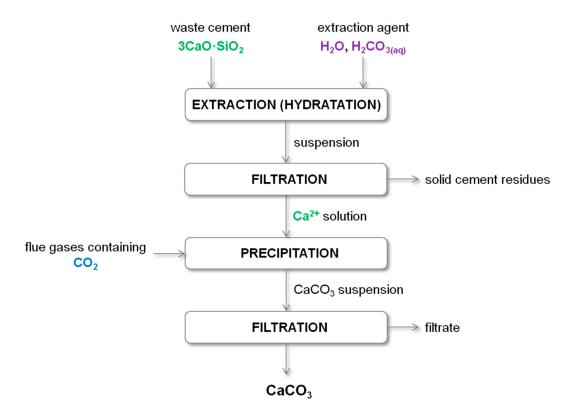
Figure 6. Diagram of the $CaCO_3$ precipitation process using waste cement and CO_2 from flue gases.

The same research team designed and built a pilot installation for recycling concrete sludge and CO_2 waste to precipitate calcium carbonate [89]. Concrete sludge from the production of piles and poles, boiler gas, and groundwater were used. The process had the same two stages as for bench scale [92]. Using this technology, it is possible to obtain high-quality $CaCO_3$ with a purity above 97%, as well as to reduce CO_2 emissions from waste gases. Importantly, this process is energetically beneficial because it is carried out at standard temperatures and atmospheric pressure, and moreover there is no need to purify the exhaust gas or increase its pressure.

Table 4 summarizes the process parameters of the selected calcium carbonate precipitation technologies using concrete waste or concrete sludge and gaseous carbon dioxide (carbonation process). All listed methods were carried out under atmospheric pressure. In the case of temperature, most of the processes were carried out at room temperature, while the influence of elevated temperature (30, 50, 70 °C) on the precipitation process was also investigated. When CO_2 absorption is accompanied by an exothermic reaction (Equation (22)), the lower process temperature favors such a transformation.

Table 4. Process parameters of $CaCO_3$ precipitation technologies for concrete and cement waste utilization using carbonation method.

Raw Material	Extraction Agent	Gas Mixture Composition	Temperature	Pressure	Product Characteristic	Ref.
Waste concrete	NH_4Cl, HCl, CH_3COOH, water	99.9% vol. of CO_2	Room	Atmospheric	– Mixture of rhombohedral calcite and spherical vaterite, – agent extraction type does not affect $CaCO_3$ precipitation.	[87]
Concrete sludge (waste cement)	Water (hydration)	6–13% vol. of CO_2	Room	Atmospheric	– High-purity $CaCO_3$ (>97%), – contaminants: SO_3 (1.2% wt.), Al_2O_3 (0.09% wt.), SiO_2 (0.08% wt.), Fe_2O_3 (0.04% wt.), MgO (<0.01% wt.).	[89]

Raw Material	Extraction Agent	Gas Mixture Composition	Temperature	Pressure	Product Characteristic	Ref.
Concrete sludge (waste cement)	Water (hydration)	8 and 13% vol. of CO_2	Room	Atmospheric	– Calcite form, – high-purity $CaCO_3$ (>99%), – contaminants: silicon, magnesium, – particles size in the range of 3–30 μm, – average particle size of 10 μm.	[92]
Concrete sludge (waste cement)	Carbonic acid solution (pressurized CO_2)	n.d.	30, 50, 70 °C	Atmospheric	– With addition of seed crystals: high-purity $CaCO_3$ (>98%), – without addition of seed crystals: $CaCO_3$ with 80% purity.	[94]

n.d.—no data.

3.4. Gypsum Wastes

There are three types of gypsum waste, phosphogypsum (PG), desulfurization gypsum (DG), and red gypsum (RG). PG is produced during phosphoric acid production, DG is generated in power plants through a reaction between sulfur gas and lime or limestone sorbents in the desulphurization unit, while RG is the waste from processing ilmenite ore to obtain TiO_2 particles. The waste PG and DG consist mainly of CaO, SO_3, and water, as well as minor impurities such as Fe_2O_3, SiO_2, and Al_2O_3 [95]. Due to the high calcium content, gypsum wastes have great potential to be used for CO_2 sequestration and calcium carbonate production. The mineral carbonation of gypsum obtained during flue gas desulfurization (DG) proceeds according to the reaction in Equation (23). According to the reaction stoichiometry, in such a process for every 1 ton of produced ammonium sulphate, 0.9 tons of $CaCO_3$ is generated and 1.6 tons of gypsum and 0.4 tons of CO_2 are consumed [96]:

$$CaSO_4 \cdot 2H_2O + CO_2 + 2NH_4OH \rightarrow CaCO_3 \downarrow + (NH_4)_2SO_4 + 3H_2O,$$
$$\Delta H_{298}^0 = -106.43 \text{ kJ} \tag{23}$$

Song et al. [97] proposed the direct carbonation of DG process, which was carried out in ambient conditions using an aqueous ammonia solution. A diagram of this process is shown in Figure 7. DG is characterized by high carbonation reactivity at room temperature and atmospheric pressure [97]. The polymorphic form of $CaCO_3$ obtained during direct and indirect carbonation was also compared. In the case of the direct process, the formation of a mixture of rhombohedral calcite and spherical vaterite was observed, with the vaterite content being dependent on the carbonation time. The longer the time, the greater the amount of vaterite in the resulting product. However, during the two-stage (indirect) process, the precipitated calcium carbonate was mainly in the form of calcite. Moreover, what is important, in the proposed direct carbonation process, obtained calcium carbonate was practically free of impurities, which allows for further use of such a product [97]. In another work, Song et al. [98] focused on the possibility of controlling the polymorphic form and morphology of calcium carbonate obtained during direct carbonation of DG using an aqueous ammonia solution. The controlling factor was the addition of ethanol to the reaction mixture under stoichiometric conditions and with an excess of ammonia. During both processes investigated by Song et al. [97,98], a pure and valuable product from industrial DG is produced. However, the problem is to obtain a small amount of $CaCO_3$ crystals. Additionally, these research are focused mainly on the polymorphic form and morphology of calcium carbonate and do not contain any discussion about the efficiency of CO_2 sequestration.
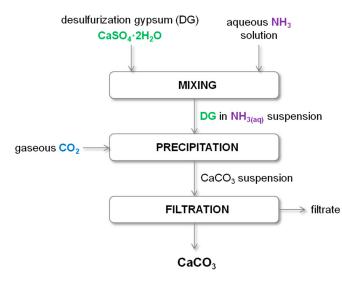
Figure 7. Diagram of the $CaCO_3$ precipitation process by direct carbonation of desulfurization gypsum.

Table 5 summarizes the process parameters of the selected calcium carbonate precipitation technologies using waste gypsum and gaseous carbon dioxide (carbonation process). All presented processes were carried out under atmospheric pressure and at moderate temperatures (20–40 °C). Such conditions favor both CO_2 absorption and the exothermic reaction (Equation (23)).

Table 5. Process parameters of $CaCO_3$ precipitation technologies for waste gypsum utilization using carbonation method.

Raw Material	Extraction Agent	Gas Mixture Composition	Temperature	Pressure	Product Characteristic	Ref.
PG	Aqueous NH_3	Pure CO_2	Room	Atmospheric	− Scalenohedral calcite, − mean size of 3.4 μm in diameter, − 95% gypsum conversion efficiency, − PCC impurities: Cd, Hg, Ni (<1 mg/kg), Pb (56 mg/kg), As (21 mg/kg).	[96]
DG	Aqueous NH_3	99.99% vol. of CO_2	Room	Atmospheric	− Pure $CaCO_3$, − carbonation efficiency of 5.3%, − mixture of rhombohedral calcite and spherical vaterite.	[97]
DG	Aqueous NH_3	99.99% vol. of CO_2	Room	Atmospheric	**Regardless of the amount of NH_3 in the solution:** − product consisting of $(NH_4)_2SO_4$ and $(NH_4)_2Ca(SO_4)_2·H_2O$. **Stoichiometric amount of NH_3:** − only rhombohedral calcite, − 30 and 50% vol. ethanol causes an increase in vaterite content. **Excess of NH_3:** − mainly spherical vaterite, − 30 and 50% vol. ethanol results in the formation of aragonite.	[98]
DG	Aqueous NH_3	Pure CO_2	20, 30, 40 °C	Atmospheric	− Amorphous nanosized particles with an aggregate structure, − average particle diameter of 86−104 nm at 30−40 °C.	[99]

3.5. Other Ca-Rich Waste

There are several other examples of alkaline liquid waste in the literature that can be used as Ca-rich raw material for the production of $CaCO_3$ by the carbonization method. Paper sludge ash (PSA) from paper manufacturing has a very small particle size, high pH (>12) [100] as well as high calcium content of about 22% wt. [101,102]. Kim et al. [102] conducted research on CO_2 storage by indirect carbonation with PSA at room temperature and atmospheric pressure. In this work, various types of

solvents for the extraction of calcium ions from PSA were investigated: acetic acid, hydrochloric acid, ammonium acetate, ammonium chloride, sodium citrate and water. Gaseous CO_2 was introduced into the obtained extract to precipitate calcium carbonate. To increase the CO_2 storage efficiency, additional steps such as pH swing or changing the amount of supplied CO_2 were used. The results indicate that the highest CO_2 capture efficiencies were obtained using acetic acid (324 kgCO_2/ton PSA) and hydrochloric acid (297 kgCO_2/ton PSA) by treating the extract with sodium hydroxide. In the case of the ammonium salts, efficiency in the range of 168–175 kgCO_2/ton PSA was obtained. The lowest efficiency was recorded for the processes with the use of sodium citrate and water, 136 and 29.1 kgCO_2/ton PSA, respectively. In addition, sodium citrate and ammonium salts can be recovered and reused. This is an important factor that increases the economic feasibility of the proposed carbonation processes. According to Zevenhoven et al. [101], PSA is a very suitable material for carbonation as pure calcium carbonate (90–98.5% wt.) is obtained compared to steelmaking slag and wollastonite. However, the water consumption in the PSA process is relatively high, making wastewater treatment problematic.

Another raw material is oil shale ash (OAS) which, depending on combustion technology, contains 10–25% wt. of free lime [103]. Some of the shale oil ash is used in agriculture, road construction, and as raw material for the production of construction materials, while the majority of this waste is stored in piles located near the power plant [104]. Due to the high content of free lime and anhydrite, which form alkaline leachate with a pH of 12–13, the disposal of this type of waste is a large environmental problem [105]. Tamm et al. [106] investigated the feasibility of $CaCO_3$ precipitation from shale ash using process modeling. According to the obtained results, such technology can be used in the processes of CO_2 mineralization and waste stabilization and valorization. Uibu et al. [103] constructed and tested a continuously operating reactor for the precipitation of $CaCO_3$ from oil shale ash-water suspensions with model gases containing CO_2 on a laboratory scale. The obtained products contained 17–20% wt. bound carbon dioxide and 0.6–2% wt. unreacted lime. According to the authors, the use of this technology allows for the capture of about 1.2 million tons of CO_2 from flue gases annually, estimated for the SC Narva Power Plant in Estonia (2007). However, further optimization of the process conditions is required to enable the production of PCC with a fine and uniform particle size [103].

4. Recommendations for Future Research

The state of the art of the presented methods using various calcium-rich wastes to capture CO_2 indicates that additional research is needed to implement these technologies. Most of the presented studies lack clearly defined optimal process conditions. The use of concrete and cement waste is the most advanced. For these solid wastes, indirect two-stage $CaCO_3$ precipitation methods have been proposed and tested on the bench and pilot scale [89,92,107]. The production of calcium carbonate is carried out at atmospheric pressure and at ambient temperature and there is no need to purify the exhaust gas.

Another significant problem is the generation of secondary waste, therefore the direction of their management should also be indicated. Only some studies suggest solutions to this problem [37,42,47,101,102].

Maintaining efficiency and energy demand for processing are important issues for future commercialization [25]. A useful approach to assessing a technology under development in terms of its energy efficiency is an exergy analysis [108–110]. Such attempts have been made for some CO_2 absorption processes [111,112] and oxy-fuel technologies [113–116]. The exergy analysis is an effective tool to improve the efficiency of the installation, reduce waste heat, and indicate the possibility of heat integration. Therefore, it is reasonable to apply this method to validate and optimize design concepts prior to implementation in the next generation carbon capture technologies using Ca-rich waste.

A further limitation is the small amount of research into the scaling up the process. Pilot-scale studies for carbonization of waste concrete and cement have been published [89,107].

The consequence of the above problems is a limited number of techno-economic analysis (TEA) of the proposed methods. The available data show that the mineral carbonation CCU technology has

an economic advantage in energy penalty, power plant construction, and operating cost over other proposed CCS and CCU technologies [117]. Recommendations for a techno-economic analysis of CCS and CCU processes have recently been published [118], and the application of this TEA guideline may improve the comparability of TEA studies.

5. Conclusions

Nowadays, a lot of attention is paid to the issue of the increasing amount of produced industrial waste, which is a serious environmental hazard. Numerous attempts are made to develop and implement technologies aimed at both disposal and reuse of this waste. Carbon dioxide emitted during the combustion of fuels or released during the manufacture of many products is indicated as one of the main greenhouse gas that contributes to rapid global warming. Another group of nuisance pollutants is calcium-rich wastes, such as distillation liquid from the Solvay process, metallurgical slag from steel production, cement industry waste, gypsum waste, ash from paper sludge and oil shale ash. The alkalinity of individual liquid waste and the liquid after leaching of solid waste allows them to be used to capture carbon dioxide from flue gas. Therefore, an interesting idea for the management of CO_2 and Ca-rich waste is to develop a technology for producing calcium carbonate via carbonation route. Such technologies help reduce harmful emissions while producing a valuable product that can be sold and used in many industries, such as paper, rubber, pharmaceuticals, and many others. Therefore, an important criterion in the evaluation of these technologies is the possibility of producing $CaCO_3$ meeting numerous criteria, such as high purity, adequate humidity, particle size, and shape, or polymorphic form. In addition, for the technology of calcium carbonate precipitation from waste liquids and CO_2 to be successfully implemented in production, this process must be characterized by low electricity demand. In general, the technology must be economically viable, so that the benefit from the sale of precipitated $CaCO_3$ should be more profitable higher than the cost of its production. Therefore, most of the technologies proposed and described in this work are carried out at room temperature and atmospheric pressure, so as not to generate additional costs associated with heating or maintaining the desired pressure of the reaction system. However, there are many other obstacles, such as the necessary reagent regeneration processes, which are expensive, or the generation of other waste streams for which further utilization processes should be implemented. An important issue is also the fact that most potential Ca-rich waste is in solid form, which requires the extraction of calcium ions into the solution. Therefore, the efficiency of the Ca ion leaching process is another important factor influencing the profitability of the designed process. In addition, the implementation of a given waste management technology may also be determined by political factors, when legal regulations of a given country force entrepreneurs to reduce emissions and pollutants.

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