



Article Synthesis of Fe-Modified g-C₃N₄ Nanorod Bunches for the Efficient Photocatalytic Degradation of Oxytetracycline

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Abstract: Antibiotic residues have been found to have potentially harmful effects on ecological and human health. Carbon nitride-based photocatalysts have widely focused on antibiotic photocatalytic degradation. Herein, we prepared Fe-modified g- C_3N_4 nanorod bunches (FCNBs) using chemical vapor co-deposition. Specifically, through the process of calcination, a blend of urea and chlorophyllin sodium iron salt underwent an intriguing transformation, resulting in the integration of Fe into the framework of the g- C_3N_4 nanorod cluster. The resulting photocatalyst exhibited remarkable stability and superior dispersibility. The prepared FCNBs had a unique structure, which was beneficial for increasing light absorption. Furthermore, the Fe species formed a chemical coordination with the g- C_3N_4 matrix, thereby altering the electronic structure of the matrix. This modification facilitated charge transfer, prolonged the carrier lifetime, and enhanced light absorption, all of which significantly increased the photocatalytic activity. The oxytetracycline degradation efficiency of FCNBs was 82.5%, and they demonstrated outstanding stability in cycle trials. This work introduces a promising photocatalyst for the degradation of antibiotics.

Keywords: g-C₃N₄; Fe; chemical vapor co-deposition; photocatalytic degradation; oxytetracycline

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Oxytetracycline (OTC), an antibiotic with strong antibacterial action, has been extensively utilized in aquaculture, animal husbandry, and human and animal medicine in recent years [1-4]. Excessive OTC is discharged into wastewater or soil, causing harm to ecosystems and human health [5]. The ubiquity of OTC in water severely pollutes the water supply and induces microbial resistance. More and more attention has been paid to adopting effective methods to eliminate OTC in water. Currently, a multitude of methods, including physical adsorption, chemical oxidation, biodegradation, and photocatalytic degradation, have been employed for the elimination of OTC [6–8]. Among them, the photocatalytic degradation of OTC has attracted wide attention because of its green and sustainable characteristics. In principle, the photocatalytic degradation of oxytetracycline involves the selection of appropriate semiconductors and solar energy as the input to promote a chemical reaction through four main steps, as follows: 1. The catalyst needs to absorb photons whose energy is greater than or equal to the semiconductor band gap to achieve the necessary conditions for photogenerated electron and hole separation. 2. Electrons are excited from the valence band (VB) of the semiconductor to the conduction band (CB). 3. Optically excited charge carriers migrate to the surface without futile recombination. 4. The accumulated photoexcited electrons and holes participate in the redox reaction on the surface to realize the degradation of oxytetracycline [2]. The key to photocatalytic degradation technology is the rational design of a photocatalyst to improve the photocatalytic degradation ability.

As an N-type semiconductor, graphitic carbon nitride $(g-C_3N_4)$ has led to great progress in boosting the photocatalytic degradation of oxytetracycline, estrone hormone, and dye and water splitting due to its visible-light response and excellent physicochemical properties [2,9–12]. However, there is still a need for improvement regarding its flaws, including the short visible-light absorption range, quick photogenerated carrier recombination, and low conductivity [10]. Doping, heterostructure construction, dye sensitization, and morphology control have been used to optimize $g-C_3N_4$ to enhance its photocatalytic performance [13–16]. Significantly, morphology regulation has been regarded as a simple and effective modification strategy. Jun et al. developed mesoporous $g-C_3N_4$, which enhances its ability to absorb light across a wider spectrum and extends the duration of the photoexcited charge carriers [17]. Similarly, Dou et al. prepared $g-C_3N_4$ nanosheets, significantly enhancing its photocatalytic degradation ability [18]. However, further alterations to its electronic structure are necessary in order to augment the optical, conductive, and other physical characteristics of $g-C_3N_4$.

The electronic structures of $g-C_3N_4$ can be successfully improved, and its photocatalytic degradation performance can be enhanced, by adding the highly conductive metal Fe [19]. Furthermore, lone pairs of electron, which can combine with Fe to generate Fe-N, are abundant in $g-C_3N_4$ [20]. Xu et al. utilized a one-step calcination technique for the production of Fe-doped $g-C_3N_4$ and this synthesis exhibited notable efficacy in the degradation of TC under visible-light-driven photodegradation reactions [21]. To enhance photocatalytic activity by encouraging the separation of electron–hole pairs, Sudrajat et al. incorporated Fe into $g-C_3N_4$ [22]. Van et al. used a simple and economical heat-stirring technique to create Fe-doped $g-C_3N_4$ high-performance photocatalysts [23]. However, these methods lead to the formation of Fe-particle-modified $g-C_3N_4$, which reduces the dispersion of Fe.

Herein, urea and a combination of chlorophyll sodium iron salt were utilized as precursors to create the uniformly distributed, Fe-modified, $g-C_3N_4$ nanorod bunches (FCNBs), which were then used to break down OTC under irradiation with visible light. The outcomes demonstrated the great stability and superior photocatalytic degradation performance of FCNBs, indicating their potential utility in the removal of organic pollutants such as OTC.

2. Experimental Section

2.1. Materials

Sinopharm Chemical Reagent Co., Ltd. (Shanghai, China) provided urea, ethanol, isopropanol (IPA), and ethylenediaminetetraacetic acid disodium (EDTA-2Na). Macklin Reagent Co., Ltd. (Shanghai, China) supplied chlorophyllin sodium iron salt, while benzoquinone (BQ) and OTC were procured from Shanghai Aladdin Biochemical Technology Co., Ltd. (Shanghai, China). We used all analytical-grade chemical reagents directly, without any purification.

2.2. Preparation of FCNBs and g- C_3N_4

Firstly, the filter paper made from cellulose was cut to a specific size and placed inside a nitrogen muffle furnace. Cellulose-based carbon paper was produced by gradually heating it at a rate of 3 °C/min in a nitrogen environment until it reached 700 °C, where it was held for 1 h. Afterwards, a mixture of 50 g urea and 0.5 g chlorophyll sodium iron salt was dissolved in a combination of 100 mL ethanol and deionized water at a temperature of 95 °C, resulting in the formation of a dark green solution. This solution was then heated under agitation for 2 h to obtain an evenly distributed solid mixture. The cellulose-based carbon paper was positioned on top of an alumina crucible, with its bottom covered by 10 g of the prepared mixture. Subsequently, the crucible underwent heating in a nitrogen atmosphere at a rate of 3 °C/min until reaching up to 550 °C for 2 h. As a result, the iron-modified g-C₃N₄ material adhered to the surface of the carbon paper, which is referred to as FCNBs. The FCNB on the carbon paper surface was collected into the sample tube by

brushing. $g-C_3N_4$ was obtained by heating 20 g urea up to 550 °C at an identical heating rate under nitrogen atmosphere for 2 h.

2.3. Characterizations

The samples underwent analyses via the utilization of a Hitachi S-4800 scanning electron microscope (SEM) and a Japan Electronics Co., Ltd. (Amagasaki, Japan) JEM-2100 transmission electron microscope (TEM) to characterize their surface morphology and internal structure. Furthermore, functional groups of the composites were assessed through testing, analyzed using a Perkin-Elmer Spectrum One Fourier transform infrared spectrometer (FT-IR) that was equipped to operate within a range of $500-4000 \text{ cm}^{-1}$. Measurements for X-ray diffraction (XRD) patterns were taken using a Bruker D8 Advance X-ray diffractometer, scanning within a 2θ range of 5– 80° at a rate of 5°/min. The chemical state of Fe in the FCNBs, as well as the surface element composition and corresponding content distribution, were determined by VG ESCALAB 250Xi (Thermo Fisher Scientific, Waltham, MA, USA) X-ray photoelectron spectroscopy (XPS). To measure the composite material's light absorption range, a Perkin-Elmer Lambda 950 UV-Vis spectrophotometer was utilized for analyses within a measuring range of 200-800 nm. Fluorescence intensity was measured using an FLS 920 fluorescence spectrophotometer (PL) to assess changes in the recombination efficiency of electron-hole pairs before and after modification, with an excitation wavelength set at 360 nm and measurement range from 300 to 700 nm. The Fe content in the samples was quantified using an Agilent 720 inductively coupled plasma optical emission spectrometer (ICP-OES).

2.4. Photoelectrochemical Measurements

Utilizing the CHI 760E electrochemical workstation, photoelectrochemical measurements were conducted. The working electrode was fabricated through the attachment of a sample to indium tin oxide glass. A platinum electrode served as the opposing pole, while the reference electrode, Ag/AgCl, was utilized. The utilized electrolyte was a 0.5 M Na₂SO₄ solution, with a 300 W xenon lamp serving as the illumination source. Analyses of the transient photocurrent response and electrochemical impedance spectroscopy were subsequently performed.

2.5. The Performance Measure of Catalysts

To assess how well the prepared samples degrade OTC under visible light, we dispersed 30 mg of the catalyst into 100 mL of a 10 mg/L OTC aqueous solution. To simulate visible light ($\lambda > 420$ nm), we used a 300 W xenon lamp (PLS-SXE300UV, Beijing Perfectlight Technology Co., Ltd., Beijing, China). Before irradiation, we stirred the mixture in darkness for 30 min to achieve an adsorption–desorption equilibrium. Throughout the irradiation process, we collected samples from the reaction solution at specific time intervals and then filtered them to remove the catalyst. We determined the OTC concentration by measuring its absorbance at 354 nm using a TU-1901 spectrophotometer. Additionally, we collected, washed, and dried samples for cycling experiments to evaluate the catalyst's stability.

3. Results and Discussion

3.1. Characterization of Catalysts

FCNBs were prepared through chemical vapor co-deposition (Scheme 1). SEM and TEM were used to examine the shape and structure of FCNBs. As depicted in Figure 1 and Figure S1, FCNBs exhibit a radial bundle structure, which can be further confirmed by TEM. Figure 2 illustrates the rod-like structural unit of FCNBs, which has a diameter of around 350 nm. Through the corresponding EDS (Figure 2c–e), it can be judged that its constituent elements contain C, N, and Fe. The g-C₃N₄ boasts a high concentration of N atoms, encompassing six lone pairs of electrons. This unique molecular structure proves to be highly conducive to the bonding of iron, thereby introducing Fe into g-C₃N₄. The iron



content of FCNBs was measured using ICP-OES and found to be 0.6 wt.% (Table S1), and it presented a uniform distribution across the entire surface of the $g-C_3N_4$.

Scheme 1. Flow chart for the synthesis of FCNBs.



Figure 1. SEM image (a) and the magnified SEM image (b) of FCNBs.

Using XRD and FTIR spectroscopy, the materials' structural and chemical characteristics were examined. Figure 3a displays the FTIR spectra of FCNBs and g-C₃N₄. The vibration peaks corresponding to the triazine unit vibration, stretching vibration of CN, and stretching vibration of N-H in free amino groups were located at 810 cm⁻¹, 1200–1700 cm⁻¹, and 3000–3400 cm⁻¹, respectively [24]. g-C₃N₄ and FCNBs showed similar characteristic peaks, indicating that trace amounts of Fe did not significantly alter the basic structure of g-C₃N₄. The XRD patterns depicted in Figure 3b reveal the spectra of g-C₃N₄ and FCNB. Both catalysts show two distinct peaks at around 13.20° and 27.50°, associated with the stacking structure of the conjugated aromatic system and the in-plane structure of the tris-s-triazine unit, respectively [25]. Compared with g-C₃N₄, the diffraction peaks of FCNBs are generally weaker, which can be attributed to the interaction between

Figure 2. TEM image (**a**), the magnified TEM image of a single structural unit (**b**), and elemental mapping (**c**–**e**) of FCNBs.

Figure 3. FT-IR spectra (a) and XRD spectra (b) of g-C₃N₄ and FCNBs.

The XPS analysis of FCNBs revealed a diverse surface elemental composition, as depicted in Figure 4a. Elements including C, N, O, and Fe were identified. The presence of O may be attributed to the adsorption of O_2 or H_2O on the composite surface [26]. Despite the low doping content of iron, distinct peaks corresponding to the iron species were not

evident. The Fe 2p high-resolution spectrum (Figure 4b) exhibited peaks at 725.44 and 710.63 eV, indicative of the $2p_{1/2}$ and $2p_{3/2}$ states of Fe²⁺, respectively. Similarly, peaks at 730.01 and 717.05 eV corresponded to the $2p_{1/2}$ and $2p_{3/2}$ states of Fe³⁺ [23]. This observation confirms the coexistence of Fe^{2+} and Fe^{3+} in FCNBs. Figure 4c shows the high-resolution C 1s spectrum, which showed three peaks at 284.67, 286.11, and 288.05 eV. The peak at 284.67 eV represents amorphous carbon, while the peak at 286.11 eV indicates the presence of C-O bonds. The peak at 288.05 eV was associated with sp²-hybridised carbon in the N-C=N structure of the triazine ring [27]. In addition, Figure 4d shows the high-resolution N 1s spectrum with three distinct peaks. The peaks at 398.5, 399.9, and 400.92 eV correspond to sp²-bonded N atoms (C-N=C), tertiary nitrogen atoms in N-(C)₃ groups, and N-H groups, respectively [28]. The tabulated form of elemental analysis from XPS is shown in Table S2 to confirm the element ratio of FCNB. We also conducted nitrogen adsorption and desorption tests on FCNB and g-C₃N₄. As shown in Table S3, the specific surface area of FCNB is slightly smaller than that of g-C₃N₄, which can be attributed to the introduction of Fe nanoparticles, leading to the destruction of the layered structure of g-C₃N₄ [26].

Figure 4. XPS survey spectra of FCNBs (**a**) and high-resolution XPS spectra of Fe 2p (**b**), C 1s (**c**), and N 1s (**d**).

Figure 5a depicts the UV-Vis absorption spectra of $g-C_3N_4$ and FCNBs. Although $g-C_3N_4$ absorbs light across the ultraviolet and visible-light ranges, its absorption in the visible region is relatively weak. In contrast, FCNBs exhibited significantly improved absorption properties in the visible-light range due to the incorporation of Fe. Furthermore, there is a noticeable shift towards longer wavelengths in the absorption edge of FCNBs compared to that of $g-C_3N_4$, indicating that the addition of Fe can greatly enhance the capacity to absorb visible light. The improvement in the efficiency of solar energy utilization and boosting of the visible-light catalytic activity of the catalyst are the benefits of this enhancement. The evaluation of the photocatalytic performance of the photocatalysts was primarily determined by their efficiency in separating photogenerated electron–hole pairs. Therefore, we carried out photoluminescence (PL) testing. Figure 5b shows that $g-C_3N_4$ has a significant fluorescence peak at 450 nm, indicating a high rate of electron–hole

pair recombinations after irradiation. In contrast, the fluorescence intensity of FCNBs notably decreased, suggesting that the incorporation of Fe can effectively suppress carrier recombination and enhance photocatalytic activity.

Figure 5. UV-vis diffuse reflectance spectra (**a**) and steady-state photoluminescence spectra (**b**) of g-C₃N₄ and FCNBs.

We carried out an in-depth study of the light-induced electron transfer process and enhanced photocatalytic activity using electrochemical impedance spectroscopy (EIS) and photocurrent testing. The objective of this study is to explore the charge transfer resistance and efficiency in photogenerated electron–hole pairs' separation. A comparison of the Nyquist plots of $g-C_3N_4$ and FCNBs in Figure 6a reveals that the arc radius of FCNBs is significantly smaller, the charge transfer resistance is reduced, and the electron transfer rate is accelerated. In addition, Figure 6b illustrates that the photocurrent intensity of FCNBs surpasses that of $g-C_3N_4$, indicating a substantial enhancement in the separation of the photogenerated electron–hole pairs.

Figure 6. EIS Nyquist plots (**a**) and transient photocurrent density response (**b**) of nanosized and $g-C_3N_4$ and FCNBs.

3.2. Photocatalytic Activity Measurement

The study investigated the degradation of OTC in the presence of visible light using the prepared materials. Figure 7a illustrates the OTC photodegradation process for differ-

ent photocatalyst systems. Prior to visible-light irradiation (300 W xenon lamp), a dark reaction was conducted for 30 min to ensure catalyst adsorption saturation. The figure clearly demonstrates that, in the absence of any photocatalyst, OTC photodegradation is minimal, indicating the high stability of OTC under visible light. After 60 min of visible-light irradiation, g-C₃N₄ only degraded 37% of the OTC. In contrast, FCNBs achieved a photocatalytic efficiency of 82.5% within the same timeframe, which is superior to the similarly reported g-C₃N₄ catalyst at present (Table S4). This indicates that the introduction of iron notably augmented the photocatalytic properties of the composites. We increased the amount of FCNB from 30 to 50 mg to conduct a photodegradation experiment, and compared the effects of different amounts of the catalyst on the photodegradation performance. As shown in Figure S2a, the OTC degradation efficiency of 50 mg FCNB within 60 min is 84.9%, indicating that the degradation efficiency increases with an increase in the amount of catalyst. In addition, we carried out an in-depth study of the prepared samples to analyze the reaction kinetic properties of the photocatalytic degradation of OTC and applied Equation (1) to calculate the apparent rate constant (*k*) [29]:

$$\ln\left(C_0/C\right) = kt \tag{1}$$

where *k* is the apparent rate constant, *t* is the light irradiation time, *C* is the concentration at irradiation time *t*, and C_0 is the concentration at adsorption equilibrium.

Figure 7. Photocatalytic activities (**a**) and kinetic fit (**b**) of the $g-C_3N_4$ and FCNBs for OTC photodegradation; (**c**) cycle runs for OTC photodegradation on FCNBs; (**d**) effect of different scavengers on the photodegradation of OTC compared to FCNBs.

The plot of $ln(C_0/C)$ versus time (*t*) demonstrates linearity in the kinetic analysis (Figure 7b), affirming that its degradation process adheres to first-order kinetics. The *k* for g- C_3N_4 is calculated as 0.0055 min⁻¹. In contrast, the *k* value for FCNBs was determined to be 0.0205 min⁻¹, which is 3.7 times greater than that of g- C_3N_4 , underscoring the significantly enhanced degradation rate of the composite photocatalyst. In practical applications, the reusability and stability of composite materials are paramount. Consequently, cycling tests for OTC photocatalytic degradation were conducted on FCNBs to assess their reusability. In addition, we tested the photocatalytic degradation of OTC under the full solar spectrum with other conditions remaining unchanged. The results show that the photodegradation efficiency of the sample is 85.0%, as shown in Figure S2b. As depicted in Figure 7c, after four cycles, the degradation efficiency of the catalyst exhibited no notable decline, highlighting the excellent stability of FCNBs during photocatalytic degradation. We carried out SEM and XRD tests on the materials after the degradation test, as shown in Figure S3a,b. The considerable photodegradation efficiency of OTC makes FCNB an excellent material for the photocatalytic degradation of OTC.

3.3. Photocatalytic Mechanism

To delve deeper into the photocatalytic degradation mechanism, active radical trapping experiments were performed by using hydroquinone (BQ) and isopropanol (IPA) as a trap for the superoxide anion radical (\bullet O²⁻) and hydroxyl radical (\bullet OH), and disodium ethylenediamine tetraacetate (EDTA-2Na) as a hole-trapping agent [30]. We conducted an in-depth study on the role of these actives in the photocatalytic degradation process. As seen in Figure 7d, the introduction of IPA had no discernible impact on the degradation rate of OTC, suggesting that \bullet OH had an insignificant effect on the degradation of OTC. However, the addition of BQ (10 mM) and EDTA-2Na (10 mM) led to a significant decrease in the degradation rate of OTC, highlighting the key role of \bullet O²⁻ and h⁺ in the degradation process.

Furthermore, the Tauc diagram (Figure S4a) of $g-C_3N_4$ and FCNBs was obtained through UV-vis spectrum calculations, revealing band gaps of 2.92 eV and 2.48 eV for $g-C_3N_4$ and FCNBs, respectively. By combining this with the Mott–Schottky curves (Figure S4b,c), the band structure of $g-C_3N_4$ and FCNBs can be deduced (Figure S4d). The conduction band potential (E_{CB}) of FCNBs is -0.55 V vs. NHE, which is lower than the $O_2/\bullet O^{2-}$ potential (-0.33 V vs. NHE). This indicates that the photoelectrons on the CB of FCNBs are conducive to the reduction in $\bullet O^{2-}$. The valence band potential (E_{VB}) of FCNBs is 1.93 V vs. NHE, which is lower than that of $H_2O/\bullet OH$ (2.69 V vs. NHE). It can be inferred that holes (h^+) cannot oxidize OH⁻ to $\bullet OH$ but directly oxidize OTC instead, based on these findings. This suggests that the influence of $\bullet OH$ on photocatalytic performance is negligible, consistent with the results of the capture experiment [31].

To explain how FCNBs degrade OTC, a tenable photocatalytic process is put forth (Figure 8). Under the given light conditions, electrons in the $g-C_3N_4$ material move to the conduction band through the valence band and then migrate to the iron state, improving the separation efficiency of electron–hole pairs. The h⁺ remaining in the valence band can oxidize contaminants. Subsequently, the photogenerated electrons may be captured by oxygen molecules to generate a potent oxidizing species $\bullet O_2^-$, which decomposes antibiotics into smaller molecular species. Additionally, the incorporation of iron considerably boosts the absorption of visible light and enhances solar utilization. In conclusion, the improved photocatalytic degradation performance of FCNBs can be attributed to the increased light harvesting and higher electron–hole pair separation efficiency.

Figure 8. The possible photocatalytic mechanism of FCNBs for oxytetracycline degradation.

3.4. Possible Degradation Pathway of OTC by FCNBs

The degradation intermediates of OTC were detected using LC-MS in order to elucidate its degradation pathway. Figure S5 and Table S5 present the potential intermediates, while Figure 9 illustrates two plausible main degradation pathways. In the first pathway, OTC undergoes dehydroxylation and deamidation, resulting in the formation of P2 (m/z = 382.9), which is subsequently further degraded to P3 (m/z = 279.1) through fragmentation. Further dehydroxylation leads to the generation of P4 (m/z = 262.9). In the second pathway, dehydration converts OTC into P1 (m/z = 443.1), which then decomposes and fragments into P5 (m/z = 244.2) via deamidation, demethylation, and dehydration processes. Subsequently, P5 is converted to P6 (m/z = 232.2) through deoxidation reactions. Finally, both P4 and P6 undergo additional decomposition steps, yielding lower-molecularweight substances such as P7 (m/z = 110.1), P8 (m/z = 142.0), P9 (m/z = 178.0), and P10 (m/z = 192.0). These small molecules eventually degrade into CO₂ and H₂O after undergoing further degradation steps.

Figure 9. The possible degradation pathways of OTC.

4. Conclusions

Graphene carbon nitride materials modified with iron were successfully synthesized using a homogeneous mixture of urea and chlorophyll iron sodium salt via chemical vaporphase co-precipitation. Fe is uniformly dispersed within the internal structure of $g-C_3N_4$ through chemical coordination with the $g-C_3N_4$ matrix. Furthermore, Fe functions as an electron trap, thereby improving the separation efficiency of the generation of electronhole pairs through light. The composite demonstrates efficient and stable photocatalytic degradation activity during the decomposition of OTC, making it a promising material for mitigating antibiotic contamination in the environment.

Supplementary Materials: The following supporting information can be downloaded at: https://www.mdpi.com/article/10.3390/ma17112488/s1, Figure S1: SEM image of FCNBs; Figure S2: Photocatalytic activities of (a) 50 mg FCNBs, (b) 30 mg FCNBs in full solar spectrum of light; Figure S3: (a) XRD image and (b) SEM image of FCNBs after photocatalytic activity measurement; Figure S4: (a) Tauc curves of FCNBs and g-C₃N₄, Mott-Schottky curves of (b) g-C₃N₄ and (c) FCNBs, (d) the band structure of g-C₃N₄ and FCNBs; Figure S5: MS spectra of OTC solution after 60 min degradation; Table S1: Iron content of FCNBs; Table S2: Element content (wt.%) from XPS; Table S3: Specific surface area and pore parameter information of g-C₃N₄ and FCNBs; Table S4: Comparison of photocatalytic performance of photocatalysts based on g-C₃N₄ for degradation of OTC; Table S5: Structural information of the possible intermediates. Refs. [32–35] can be found in Supplementary Materials

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