

Article

Regional Variation in Forest Canopy Height and Implications for Koala (*Phascolarctos cinereus*) Habitat Mapping and Forest Management

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Abstract: Previous research has shown that the Koala (*Phascolarctos cinereus*) prefers larger trees, potentially making this a key factor influencing koala habitat quality. Generally, tree height is considered at regional scales which may overlook variation at patch or local scales. In this study, we aimed to derive a set of parameters to assist in classifying koala habitat in terms of tree height, which can then be used as an overlay for existing habitat maps. To determine canopy height variation within a specific forest community across a broad area in eastern Australia, we used freely available Airborne Laser Scanning (ALS) data and adopted a straightforward approach by extracting maximum-height ALS returns within a total of 288 30 m × 30 m “virtual” ALS plots. Our findings show that while maximum tree heights generally fall within published regional-scale parameters (mean height 33.2 m), they vary significantly between subregions (mean height 28.8–39.0 m), within subregions (e.g., mean height 21.3–29.4 m), and at local scales, the tree heights vary in response to previous land-use (mean height 28.0–34.2 m). A canopy height dataset useful for habitat management needs to recognise and incorporate these variations. To examine how this information might be synthesised into a usable map, we used a wall-to-wall canopy height map derived from ALS to investigate spatial and nonspatial clustering techniques that capture canopy height variability at both intra-subregional (100s of hectares) and local (60 hectare) scales. We found that nonspatial K-medians clustering with three or four height classes is suited to intra-subregional extents because it allows for simultaneous assessment and comparison of multiple forest community polygons. Spatially constrained clustering algorithms are suited to individual polygons, and we recommend the use of the Redcap algorithm because it delineates contiguous height classes recognisable on a map. For habitat management, an overlay combining these height classification approaches as separate attributes would provide the greatest utility at a range of scales. In addition to koala habitat management, canopy height maps could also assist in managing other fauna; identifying forest disturbance, regenerating forest, and old-growth forest; and identifying errors in existing forest maps.



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1. Introduction

1.1. Koala Habitat

Wildlife managers are required to balance competing demands, firstly, to provide solutions to ensure the sustainability of wildlife populations, and secondly, to mitigate potential conflicts arising from human resource needs. Resolution of these conflicts requires adequate knowledge about how wildlife uses habitats, this information is collected using, for example, presence/absence surveys, dietary studies, and telemetry. In the case of the Koala (*Phascolarctos cinereus*, henceforth koala), use of habitat is primarily influenced by the density of preferred food trees, e.g., [1]. This knowledge is transferred to vegetation maps to generate a habitat map which ranks vegetation communities in terms of their relative



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importance to the koala [2]. These maps are derived from regional-scale vegetation maps covering ~100,000 hectares and might not capture habitat variability at site or landscape scales relevant to the utility needs of both the koala and habitat management stakeholders. We refer to “stakeholders” as a wide group including government agencies, developers, individuals, and community groups; all of whom have an interest in making decisions about koala habitat.

Larger trees have been identified as a factor influencing koala habitat quality by many studies, using methods ranging from; correlations between koala use and tree size at the plot scale; to mixed multi-scale models with additional environmental factors (Table 1). Mixed models have provided varying results, but no studies published have shown that the koala prefers smaller trees.

Table 1. Previous studies investigating use of larger trees by the koala. DBH = diameter at breast height.

Method	Scale	Finding	Reference
Use vs. Availability by DBH class	plot	positive correlation with food tree, negative with non-food trees	[3]
tree girth and koala use	koala radio tracking	positive correlation with food tree size	[4]
generalised linear mixed model	koala radio tracking	taller non-eucalypts on hot days	[5]
mixed-effects logistic model in 3 regions with 9 habitat variables	tree	tree size importance varies regionally. Noosa: Ranked 1 of 9. Port Stephens: 7 of 9. Ballarat: 6 of 9	[6]
koalas observed in trees, generalised additive models, generalised additive mixed models	site (7.6 ha)	increasing DBH highly positively significant	[7]
generalised linear mixed effects model, 3 seasonal surveys	transect	positively associated with DBH, negative with tree height	[8]
Use vs. Availability by DBH class	transect	positively associated with large trees	[9]
2-way ANOVA	koala radio tracking	positively associated with larger trees within plots	[10]
DBH, <i>t</i> -test	plot	positively associated with larger trees within plots	[11]
crown size, <i>t</i> -test	plot	positively associated with crown size within plots	[11]
Logistic regression and hierarchical partitioning of 12 variables	plot	larger DBH had strongest independent effect, but negative	[12]
9 variables, mixed-effects resource selection probability functions	plot/landscape	tree species and tree height most important factors at both scales	[13]
Observed and expected tree visitation based on DBH	tree	koalas use trees significantly larger than expected	[14]
one-way ANOVA	koala radio tracking	large trees selected disproportionately to availability	[15]

The preference for larger trees has been attributed to several factors; (a) there may be a threshold tree size that guides the selection of foraging behaviour [4]; (b) larger trees have larger canopies providing increased food resources [7,9] and shelter [5,13]; and (c) tree health and vigour, as determined by environmental gradients [13], in more challenging environments trees are likely to be smaller, with less nutritional quality and greater levels of deterrent foliar chemicals [16], and site competition also affects tree size [17]. In addition, other factors have been recognised, for example, soil fertility influences habitat occupancy at the landscape scale [3,18] and surrogate measures of fertility (slope and elevation) have been incorporated into some koala habitat models (e.g., [19]). A study by Callaghan et al. [3] suggested that, where detailed structural mapping is available, ‘an overlay map could indicate areas where koala habitat quality is enhanced by the predominance of large, older-growth trees.’ Identifying these areas might therefore provide useful information for habitat management at a range of scales.

1.2. The Use of LiDAR (Light Detection and Ranging) for Canopy Height Information

Canopy height, defined as the statistical aggregation of individual tree heights over a nominated area [20] is one of the most important forest metrics [21]. It can be a categorical or continuous variable, and is scale-independent whether used at plot, landscape, regional or even global scales [22]. Canopy heights are traditionally measured from the ground using a clinometer and measuring tape, which can be a slow and laborious technique. In dense forests, observing the exact canopy top is difficult because of visual obstruction, and tree heights can be underestimated by up to 2 m [23]. Since 2000, remote sensing of forests using LiDAR (Light Detection and Ranging) has developed apace, and LiDAR can be used to assess forest structure.

LiDAR data collection methods include terrestrial, unmanned aerial vehicle (UAV), airborne (ALS), and, more recently, satellite platforms. LiDAR offers several useful products: digital terrain models, tree canopy or height models, tree mapping, and plant projective cover maps [24]; and provides opportunities to explore wildlife interactions with their habitat in the third dimension, e.g., canopy height and forest structure [25,26]. Typically, ALS data is pre-classified into discrete-return classes, for example, ground, buildings, roads, shrubs, and trees. ALS has been used to develop canopy height models, and, with a sufficient density of within-canopy returns, internal structure of individual trees can be modelled [27]. ALS provides broad-scale wall-to-wall coverage at individual tree resolution and is freely available over large areas of the koala's geographic range in eastern Australia [28].

ALS research generally focusses on two different methods to obtain canopy height, firstly, area-based methods, and secondly, individual tree detection (ITD) [27] which focusses on the delineation of individual tree crowns, or individual tree vertical profiles [29]. Both methods have different uses, for example, area-based methods provide maximum canopy height while ITD can be used to obtain average canopy height. ITD methods begin with the assumption that local elevation maxima determine the top of the tree, and then, for example, use pixel-based region-growing methods to identify "watersheds" or "valley bottoms" to determine the base of the canopy [27] or Delaunay triangulation [30]. Stopping the algorithm when the base of the tree crown is sufficiently identified can be a complicated, multi-step process [31]. Over broader regional and landscape areas, ITD, because it can detect smaller trees, reduces errors of omission, and helps maintain consistent analysis between structurally variable forests [32].

Area-based approaches, e.g., within a nominated plot or grid cell, are computationally less intensive and provide similar results as ITD for certain metrics, including the height of the largest trees [32]. Area-based approaches include Canopy Height Models (CHMs), a raster-based product with the height value determined by ALS maximum return height within the grid cell [33]. Other definitions of a dominant-height metric have been formulated, e.g., the 50 highest trees/ha and the 95th percentile of all returns [22]; the predominant height based on means of local maxima [34]; the 99th percentile [35]; modal height, mean height, and 95th percentile height [36]; the arithmetic mean of returns above the 80th percentile [35]; and the median height of all returns [37]. A recent development is "cloth simulation", where a virtual cloth is draped over a fine-scale CHM to extract large trees across different scales by varying cloth tension [33].

Virtual ALS plots have been proposed as a proxy for field-based data collection and offer efficient and cost-effective data collection over large or inaccessible areas with provision of information, including canopy height, equivalent to the accuracy of field-based plots [38]. Furthermore, ALS plots can be matched to field-based plots when necessary, for example, when data calibration is required [39]. With efficient sampling design, virtual plots can be used to analyse forest structure in wall-to-wall ALS coverage [38].

In Australia, forestry research using LiDAR began in the early 2000s, and is commonly used by forestry agencies, for example, in biomass assessment of plantations and harvest planning [40] as well as by land management agencies [41]. However, most ITD and CHM methods have been developed in northern-hemisphere boreal forests with *Pinus*

species and are not readily applicable to the more structurally-heterogenous *Eucalyptus* species forests found in Australia. These forests typically have sparse, highly clumped canopies and erectophile leaf angle distributions, allowing ALS pulses to penetrate deeper into the canopy; therefore, they may underestimate dominant or maximum height [42]. Conversely, denser canopies in some tree species affect the accuracy of the underlying terrain model from which canopy heights are obtained [43]. Considering these two factors, point densities above 0.5 pts/m² provide forest metrics, including height, that did not differ significantly from those obtained with a density of 10 pts/m² [43]. ALS metadata typically specifies a target pulse density of 1 pt/m². At broader scales, ALS has been used to calibrate space-borne LiDAR data to produce vegetation community structural classifications in Queensland [36], and across Australia [44].

Koalas prefer vegetation communities with a high proportion of food trees [2,3], and they also prefer habitats with larger trees (Table 1). However, no previous research has attempted to combine these factors in a habitat management map. Given extensive ALS coverage is available in the southeast Queensland and the New South Wales north coast regions where koalas are widely distributed, our aim is to evaluate the utility of ALS in developing high resolution canopy height models which can complement existing low resolution regional-scale habitat maps and further assist wildlife managers to better manage koalas. To do this, we ask the following research questions: (1) do published herbarium descriptions adequately capture height variability at regional, sub-regional, and intra-subregional scales, and; (2) how can canopy height variability best be incorporated into a koala habitat map that can be useful to wildlife managers across different spatial scales.

2. Materials and Methods

Using herbarium descriptions as a baseline, we first examine height variability at different scales, and then, how this information could be incorporated into existing koala habitat management maps to provide greater utility at a range of scales. We make the following hierarchical scale definitions:

- Regional: The whole study region encompassing southeast Queensland and the adjoining Tweed Local Government Area in New South Wales (2,500,000 hectares),
- Subregional Group: Any identified area which is geographically separate from other areas within the study region (30,000–230,000 ha),
- Intra-subregional Group: Any identified area which is geographically separate from other areas within a subregional group (400–1600 ha), or our subregional study area,
- Polygon Group: For map production, we used contiguous groups of 25 m × 25 m cells derived from a vector polygon within intra-subregional groups (<100 ha).

2.1. Variation in Forest Canopy Height

We chose a floristically variable vegetation community with tree species known to be used by the koala [2,45] within the Southeast Queensland (SEQ) Interim Bioregion [46], which covers SEQ and part of the New South Wales north coast. In SEQ, the community is mapped as Regional Ecosystem (RE) 12.11.3 [47]. Its floristic composition is described by the Regional Ecosystem Descriptions Database [48] as mainly composed of *Eucalyptus siderophloia* (Grey Ironbark) and *E. propinqua* (Small-fruited Grey Gum), with other patchily distributed species including *E. microcorys* (Tallowwood), *Lophostemon confertus* (Brush Box), *Corymbia intermedia* (Pink Bloodwood), *E. biturbinata* (a Grey Gum), *E. acmenoides* (White Mahogany), and *E. tereticornis* (Queensland Blue Gum). Vegetation communities similar to RE 12.11.3 extend into New South Wales [49,50].

RE 12.11.3 forest structure is described in a Technical Description [51], and varies from Tall Open Forest to Woodland, with overstorey heights ranging between 15–37 m (average 28.1 m). The Description uses 66 Herbarium CORVEG plots, data collected at these plots includes floristic composition and crown cover from all forest structural layers at relatively undisturbed and representative plots, CORVEG data was collected between 1991–2011, 30 of these plots were collected within 10 years of ALS acquisition which began in 2009.

We aimed to use all 38 plots within the ALS footprint but discarded five sites which had been disturbed by development post-CORVEG data collection.

RE 12.11.3 in southeast Queensland and the associated NSW community covers 76,390 ha. Our study was limited to those areas containing ALS data (49,680 ha) (Figure 1).

We developed a workflow (Figure 2) providing a multiscale approach to assess the forest height variability within the vegetation community (henceforth referred to as RE 12.11.3) at the regional, subregional, and intra-subregional scales.

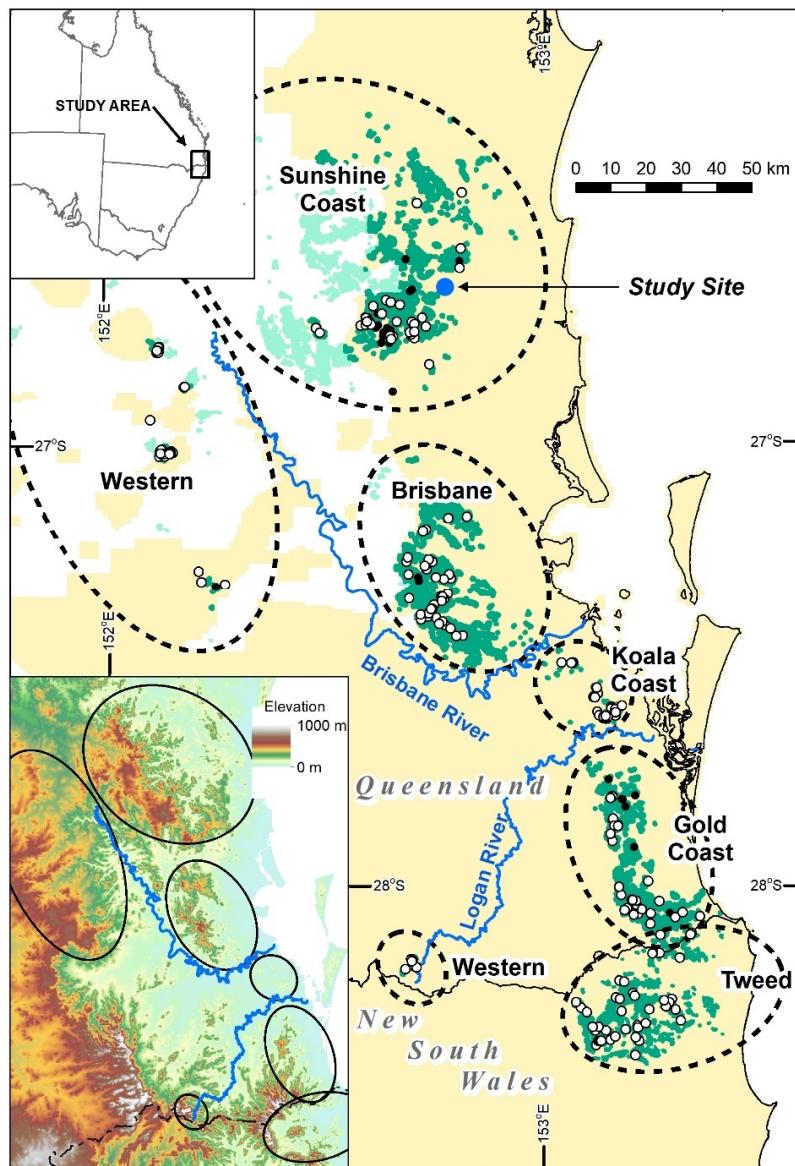


Figure 1. We examined a forest community used by koalas (Regional Ecosystem 12.11.3) occurring across Southeast Queensland and the New South Wales north coast (Tweed). Within this study area, six distinct subregions (dotted black ellipses) of RE 12.11.3 can be recognised by spatial proximity or separation by rivers or state borders. We selected the Koala Coast to examine the intra-subregion canopy height variability (Figure 3) and an additional site on the Sunshine Coast to examine the height variability resulting from historical forest disturbances (blue dot; see Figure 4). ALS coverage is shown (pale yellow background); vegetation community RE 12.11.3 is covered by ALS (dark green, with similar areas in NSW), and RE 12.11.3 not covered by ALS (light green). Random plots (205) within ALS-covered RE 12.11.3 (white circles) and CORVEG plots (33) within ALS-covered RE 12.11.3 (black circles) are shown.

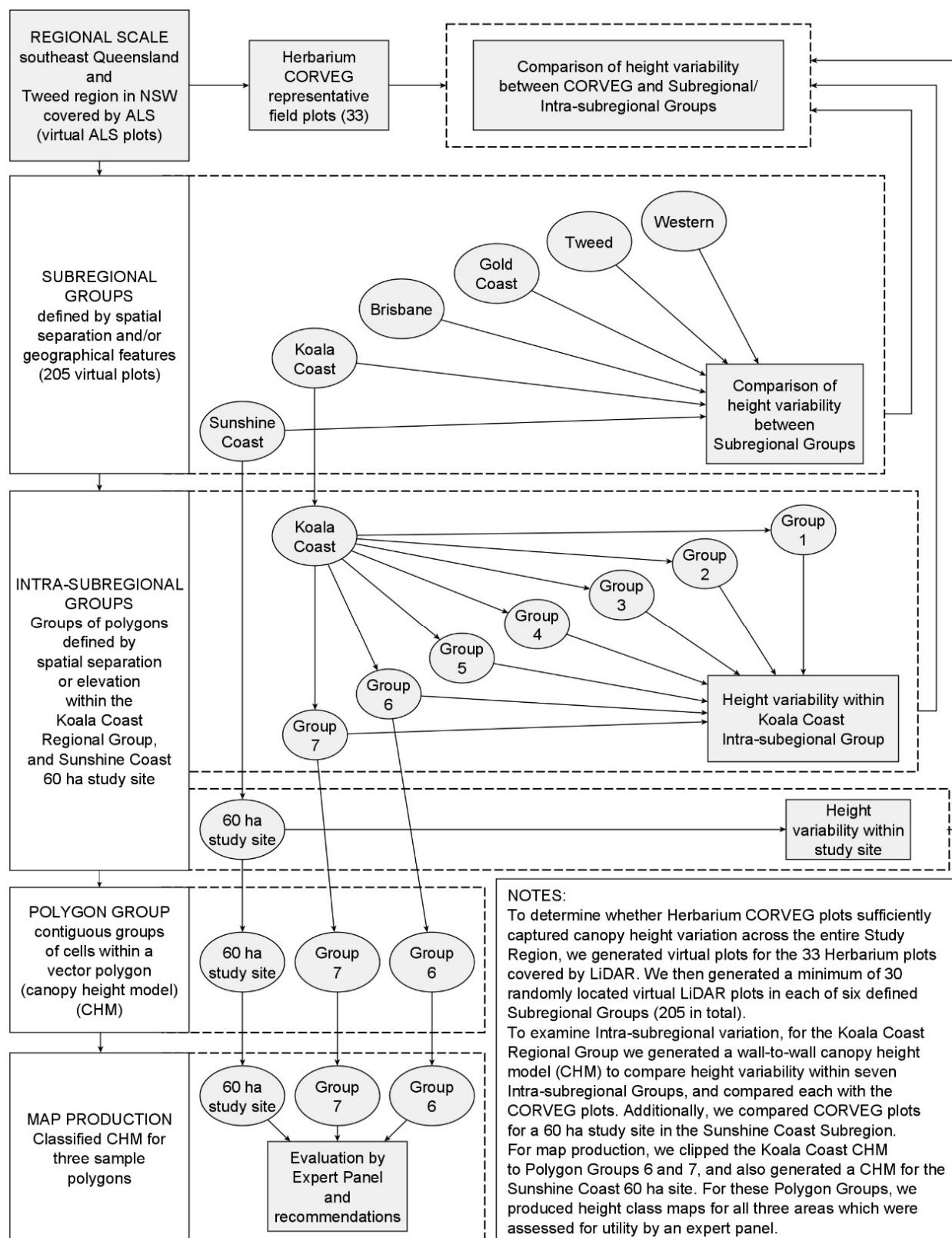


Figure 2. Workflow implemented to examine the forest height variability at the regional, subregional, and intra-subregional scales for the nominated forest community (Regional Ecosystem 12.11.3). We assessed Herbarium field plots (CORVEG) to examine how well the published information captures the height variabilities within RE 12.11.3 at these scales.

2.1.1. Regional Scale

Within the study area, we nominated six sub-regions (Figure 1). The Sunshine Coast and Western subregions have 38% and 29% ALS coverage, respectively, but the other subregions are entirely covered. Within the region covered by ALS, in addition to the 33 CORVEG plots, we used ArcMap [52] to generate 205 random 30 m × 30 m “virtual” plots within RE 12.11.3, which were at least 100 m apart.

The technical description for RE 12.11.3 provides the average canopy height for the RE by estimating the heights of at least three trees representing the median canopy height in each of the 66 [53] representative 20 m × 50 m plots [54]. Owing to the absence of individual plot orientation, we could not replicate this procedure, so we used 30 m × 30 m plots centred on the CORVEG plots and extracted the maximum canopy height in each of these plots using ALS.

2.1.2. Subregional-Scale

Within the wider region, we nominated six spatially separate subregional groups (Figure 1) and ensured that each subregion had at least 30 of the 205 virtual plots to enable an assessment of the cross-region height variation and compare it against the CORVEG plot heights.

2.1.3. Intra-Subregional-Scale

To examine the variability at finer scales, we selected the geographically distinct Koala Coast subregion (Figure 1), which also has the highest density of koalas in the wider region. Within the Koala Coast, we nominated seven distinct intra-subregional groups based on spatial separation, with two close groups demarcated by differences in elevation (Figure 3).

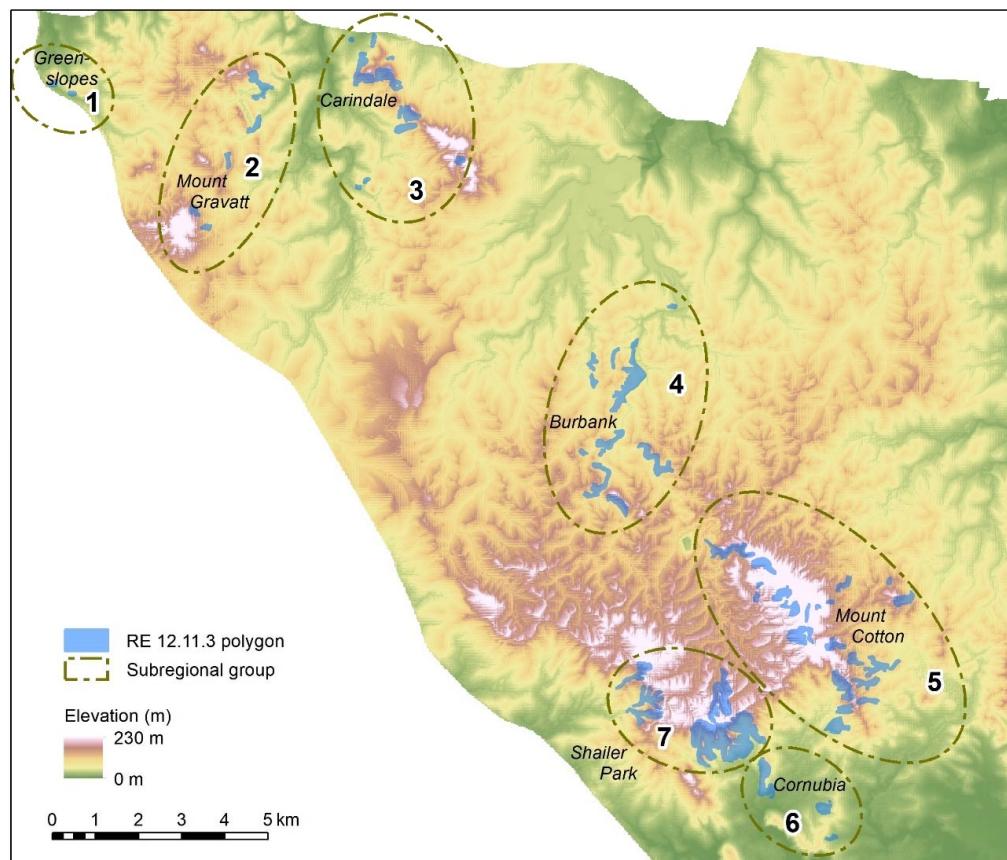


Figure 3. RE 12.11.3 Koala Coast intra-subregional groups, numbered 1–7. We used spatial separation to define most groups and elevation to separate Groups 6 and 7.

For the Koala Coast intra-subregional groups, we generated a CHM covering all RE 12.11.3 polygons and populated an aligned rectangular 25 m × 25 m polygon grid with CHM height values.

To further investigate the intra-subregional variation, and to examine the influence of disturbance on the forest height, we used data from a 60-hectare study site in the Sunshine Coast subregion. This focus study site encompassed two tenures with different land use histories—a private property “Quinlans” and Mapleton National Park. Quinlans was a dairy farm up until the 1960s, with the vegetation mostly cleared; the subsequent regrowth vegetation is now classified as Remnant Vegetation [47]. Mapleton NP was declared a forest reserve in 1923, and its vegetation is also mapped as Remnant. Figure 4 shows the vegetation cover in 1940; in some areas of Quinlans, the vegetation was almost completely removed, and the remaining forest areas were extensively thinned to an open woodland. A 1958 aerial photo (not shown) indicates that tree regrowth was actively occurring in Quinlans at that time. Areas of possible disturbance in Mapleton NP 1940 are also shown and might be because of previous logging, fire, and/or storm damage or topographic influences (drier northerly aspect).

Within this site, we used 50 30 m × 30 m virtual plots spaced on a 100-m map grid (Map Grid of Australia MGA94 Zone 56). In May 2019, we collected data, including tree species, diameter at breast height (DBH), and MGA coordinates, for every tree over 10-cm DBH at 44 sites. Data collection was aimed at obtaining detailed knowledge of both tree species variability and structural variability across the site and to provide a context to assist the interpretation of the ALS data.

2.1.4. Height Variability—Extracting Maximum Canopy Height from ALS

We obtained all ALS data from ELVIS (Elevation Information System) [28] at <https://elevation.fsdf.org.au/>, accessed on 24 June 2021. This site collects elevation data from many Australian government agencies and includes classified ALS point clouds. Our regional study area encompassed multiple ALS acquisition projects (Table 2); all available metadata showed that the minimum spacing requirement of 0.5 returns/m², appropriate for extracting forest metrics, such as maximum canopy height, was met [43].

Table 2. Available metadata for ALS data used in this paper. Point density was not included, but most datasets include ground point spacing. Following Reference [43], we regard all datasets to be suitable for our study. All data was downloaded from ELVIS (Elevation Information System) at <https://elevation.fsdf.org.au/>, accessed on 24 June 2021.

Project and Year	Height (m)	Swath Width (m)	Ground pt Spacing (m)	Pulse Rate (kHz)	Sensor
Lismore 2017	3340	1959	1.62		ALS80
Murwillumbah 2013	2120	1024	1.51		ALS50
Tweed Heads 2013	2120	1024	1.51		ALS50
Fraser Coast 2009	1500	1084	2	145.3	ALS60
Inland Towns Stage 4 2012			1.74	125–150	Gemini/Pegasus
Inland Towns Stage 6 2014				300	varies
Noosa-Gympie 2015				300	Q1560
Sunshine Coast LGA 2008	1000	924	1.74	125–150	Optech ALTM
Sunshine Coast LGA 2014	700	1155	5.08	300	Trimble Harrier 68i INS
Logan LGA 2009			0.7	126.2	ALS50-II
Logan LGA 2013	1400	992	1	100	ALS50-II
Moreton Bay LGA 2014		738	3.2	144	ALS60
Scenic Rim 2011	30–1160		2.03		Riegl LMS-Q56
Southeast Queensland 2009			0.7	126.2	ALS50-II
Southeast Queensland 2014	2060	240			ALS50, SN069, Riegl 680i
Toowoomba LGA 2010	1200	1000	1		Optech ALTM Gemini

For all virtual plots, we clipped each ALS tile to the 30 m × 30 m plot footprints using the lasclip tool [55] in ArcMap [52] and obtained the maximum heights using lasheight [55]. We then manually extracted the maximum height from the Arc Map’s layer display while visually checking all plots for ALS misclassifications (e.g., birds) and correcting when necessary.

2.1.5. Statistical Analysis

We first examined the datasets for sample size, normality, and variance (Table 3). The sample sizes varied considerably; the majority were not distributed normally, and uneven variances occurred at all the scales. For two-group comparisons, we used Welch's *t*-test, and for more than two groups, we determined that ANOVA with Games-Howell post hoc tests (with bootstrapping) were suitable for analysis, since these tests do not assume normality, equal variances, or sample sizes [56].

Table 3. Virtual ALS plot datasets are characterised by uneven sample sizes, non-normal data distribution, and unequal variances. We used both the Kolmogorov–Smirnov (with Lilliefors Significance Correction) and Anderson–Darling normality tests. Non-normal datasets are shown in blue, and highly unequal variances are shown in orange. * denotes maximum significance level reported by software package.

Regional Groups	Samples	Kolmogorov–Smirnov		Anderson–Darling		
		Statistic	Sig.	Normality A	Normality <i>p</i>	Variance
CORVEG virtual plots	33	0.109	0.200 *	0.476	0.23	42.2
Regional virtual plots (VPs)	205	0.071	0.02	1.064	0.00	55.6
Subregional Groups						
Brisbane VPs	35	0.072	0.200 *	0.176	0.91	33.4
Gold Coast VPs	31	0.148	0.08	1.100	0.00	30.0
Koala Coast VPs	30	0.111	0.200 *	0.465	0.24	15.1
Sunshine Coast VPs	32	0.121	0.200 *	0.441	0.27	50.9
Tweed VPs	40	0.127	0.1	0.696	0.07	46.4
Western VPs	37	0.171	0.01	1.249	0.00	51.9
Intra-subregional Groups—Koala Coast						
Koala Coast Group 1 VPs	38	0.114	0.200 *	0.335	0.49	13.4
Koala Coast Group 2 VPs	241	0.032	0.200 *	0.471	0.24	13.2
Koala Coast Group 3 VPs	625	0.082	<0.001	8.960	0.00	25.9
Koala Coast Group 4 VPs	1061	0.041	<0.001	2.645	0.00	13.9
Koala Coast Group 5 VPs	1667	0.062	<0.001	9.297	0.00	23.6
Koala Coast Group 6 VPs	309	0.09	<0.001	2.716	0.00	37.2
Koala Coast Group 7 VPs	2066	0.022	0.02	1.431	0.00	15.3
Intra-subregional Groups—Sunshine Coast						
Sunshine Coast VPs	32	0.121	0.200 *	0.441	0.27	50.9
Study site all VPs	50	0.131	0.03	0.753	0.05	22.8
Mapleton National Park VPs	22	0.134	0.200 *	0.397	0.34	21.6
Quinlans VPs	28	0.09	0.200 *	0.295	0.57	7.9

2.2. Map Production—Canopy Height Model (CHM)

This part of our research examined the implications of the regional variations in tree height in the context of providing a forest height map that could be used for habitat or forest management at the regional, subregional, and intra-subregional scales. We considered using individual tree detection (ITD), but since we only required the maximum canopy height, we decided that a canopy height model (CHM) would suffice for our purposes, with the added advantage of this approach being much easier to implement over broad areas. For the Koala Coast subregion and the Sunshine Coast study site, we extracted all ALS data overlapping RE 12.11.3 polygons and generated a 25-m resolution CHM using the *canopymodel* batch function in Fusion software [57] to extract the maximum height for each 25 m × 25 m cell. We again visually checked all tiles and removed misclassifications (birds, telecommunications towers, and a high-voltage powerline). We generated a 25 m × 25 m “fishnet” aligned with the CHM and extracted the CHM height values to the fishnet for analysis. We chose this cell size (0.0625 hectares) as a compromise between the Regional Ecosystem minimum mapping unit (0.5–1 ha) in Southeast Queensland [54] and some smaller units in the Tweed dataset (e.g., 0.025 ha). This cell size also provides sufficient resolution for decision-making at the habitat management and development scales.

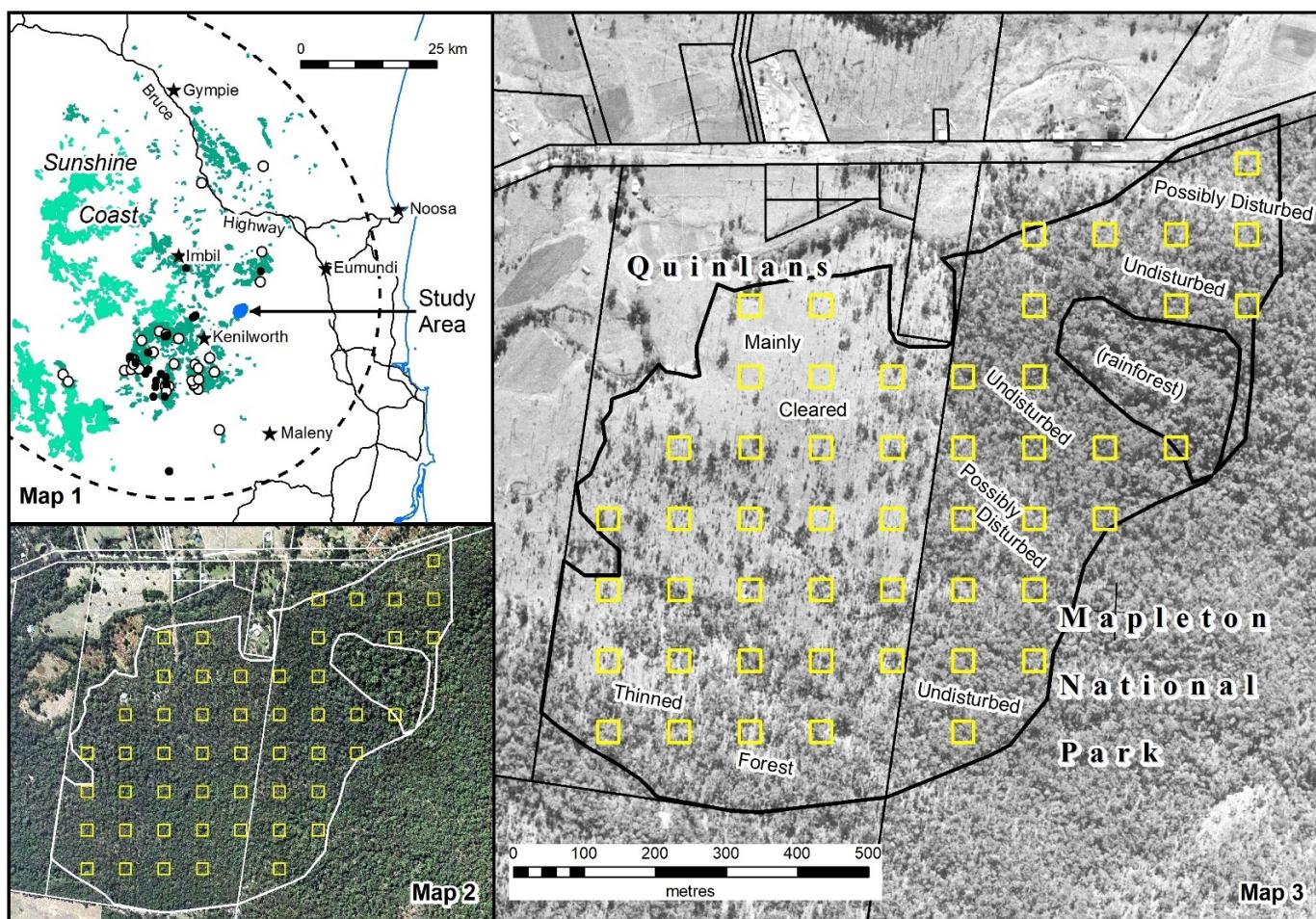


Figure 4. **Map 1:** Location of the study site in the Sunshine Coast subregion (see, also, Figure 1). **Map 2:** 2006 aerial photo (Photo: QAP6183058.jpg, Source: <https://qimagery.information.qld.gov.au/>, accessed on 27 June 2018) showing the forest cover as near-contemporaneous with the ALS acquisition in 2009. **Map 3:** Focus study area (thick black outline, rainforest excluded from study) in a 1940 aerial photo (Photo: 1940jul08_MAP648_76044_2011.tif, Source: <https://qimagery.information.qld.gov.au/>, accessed on 19 August 2018). Thinner black lines outline the tenure boundaries. Labels show disturbance types from the interpretations of the 1940 photo. Areas in Mapleton National Park marked “Possibly Disturbed Forest” might reflect the topographic position (drier northerly aspect) and not disturbance per se. Yellow squares in Maps 2 and 3 show locations of 50 ALS virtual plots.

Producing a Classified Height Map

For those involved in koala habitat or forest management decision-making, classified maps are a valuable tool that enable a feature of interest to be interpreted, compared, and evaluated in a spatial context. We now address our second research aim, “How can canopy height variability best be incorporated into a map?” Our approach was to use statistical clustering methods. Spatial clusters can be defined as a grouping of similar objects in a geographic space that are more homogenous than those from different clusters and that occupy contiguous regions [58,59] a cluster analysis uses analytical pattern recognition to group similar data points [60]. There are many different clustering algorithms, and a selection of the most suitable method requires some consideration of the data structure, e.g., data density and arrangement, and purpose, i.e., whether we sought out global or more localised cluster identification. Our dataset had distinctive characteristics: (a) height was the only variable, (b) the sample size (i.e., CHM) was 100%, (c) the data was in the form of unconnected polygons, each composed of connected groups of cells, and (d) the polygon groups had different sizes and shapes, i.e., they may be compact, linear, or complex. Unconnected polygon groups required a nonspatial global approach, whereas individual

polygon groups (i.e., with contiguous cells) were also suited to spatially constrained algorithms, since they met the contiguity requirement. Our goal was to evaluate several clustering algorithms and recommend those that (a) displayed canopy height information in a user-friendly map format useful for habitat management, (b) were reproducible using readily available software, and (c) were amenable to a high degree of automation. For map production, we used two polygon groups drawn from the Koala Coast and our Sunshine Coast study site.

For these areas, we firstly derived maps showing interquartile height ranges (i.e., four classes with equal numbers of data points) and visually identified six apparent clusters of higher canopy heights. Cluster maps were generated using algorithms implemented in *Geoda* [61], chosen because (a) it is freely available and regularly updated software with a graphical user interface (<https://geodacenter.github.io/download.html>, accessed on 1 October 2021), (b) the data formats are compatible with most GIS software, and (c) a wide selection of clustering algorithms is available. We chose eight candidate algorithms suited to our cellular data and used default values for each algorithm. For each algorithm we produced maps with two, three, and four classes, and to assess the map utility and interpretability, a three-member expert panel subjectively evaluated each output using the criteria shown in Table 4.

Table 4. Procedure used by an expert panel to rate each cluster map. Binary scoring (yes or no) for each identified feature was designed to discourage any assessment subjectivity.

Criterion	Evaluation	Score
Is the feature identified?	yes	1
	no	0
Is the areal extent similar?	Within $\pm 20\%$ of the IQR * map	1
	More than 25% different to the IQR map	0
Is the feature internally contiguous?	all cells joined (edges and/or corners)	1
	no better than the IQR map	0

* IQR Interquartile range.

3. Results

3.1. Regional Variation

The RE 12.11.3 technical description [53] uses canopy heights obtained from an average of three to five trees at each CORVEG plot [54] and summarises the data from 66 plots as the mean, maximum, and minimum heights (Table 5). We were not able to replicate this procedure, so we recorded the maximum height in each of our 33-plot ALS subsamples of the CORVEG plots, which we used to compare with all the other ALS virtual plots. Since we measured the maximum height at each CORVEG plot (Table 4), the ALS plots were substantially higher (maximum height 51.4 m) than the technical description (maximum average height 37 m), and the average maximum ALS height (33.2 m) was higher than the published mean height (28 m) [53].

Table 5. RE 12.11.3 mean, maximum, and minimum canopy heights from 66 CORVEG plots [53] compared with maximum heights obtained from the 33 CORVEG plots available to ALS, and these 33 plots were compared to the 205 ALS virtual plots within the Southeast Queensland Bioregion (SEQ), randomly sampled from six identified subregional areas within SEQ. Height differences were not significantly different (Welch's *t*-test, $df = 46.66$, $F = 1.205$, $p = 0.057$). All measurements are in metres. Max. = maximum, Min. = minimum, Std = standard, Std Dev. = standard deviation.

n Plots	Mean height	Max. Height	Min. Height					
66	28	37	15					
	Mean max. height	Std Error	Max. max. height	Min. max. height	Median height	Range	Variance	Std Dev.
33	33.15	1.13	51.36	22.36	32.36	29.00	42.24	6.50
205	37.03	1.03	57.02	18.47	36.21	27.18	35.09	5.92

These results indicate that the CORVEG plots were not significantly different in height to the random plots ($p = 0.057$) and so, for RE 12.11.3, the Herbarium technical description adequately captures the height variations across the wider Southeast Queensland region and the adjoining Tweed Coast in New South Wales.

3.2. Subregional Variation

The 205 randomly selected virtual plots across the Southeast Queensland Bioregion were assigned to six identified subregional groups to, firstly, examine the variations between the subregions and, secondly, to identify those subregions outside the CORVEG height ranges. The average maximum ALS canopy heights measured at the CORVEG plots were substantially below the average of all the subregional group heights, except the Koala Coast and Western. The Sunshine Coast and Tweed subregional groups had higher canopy heights with a larger range of heights; in contrast, the Western group had the lowest canopy. The Koala Coast canopies were also lower, with the smallest variations in height range (Figure 5).

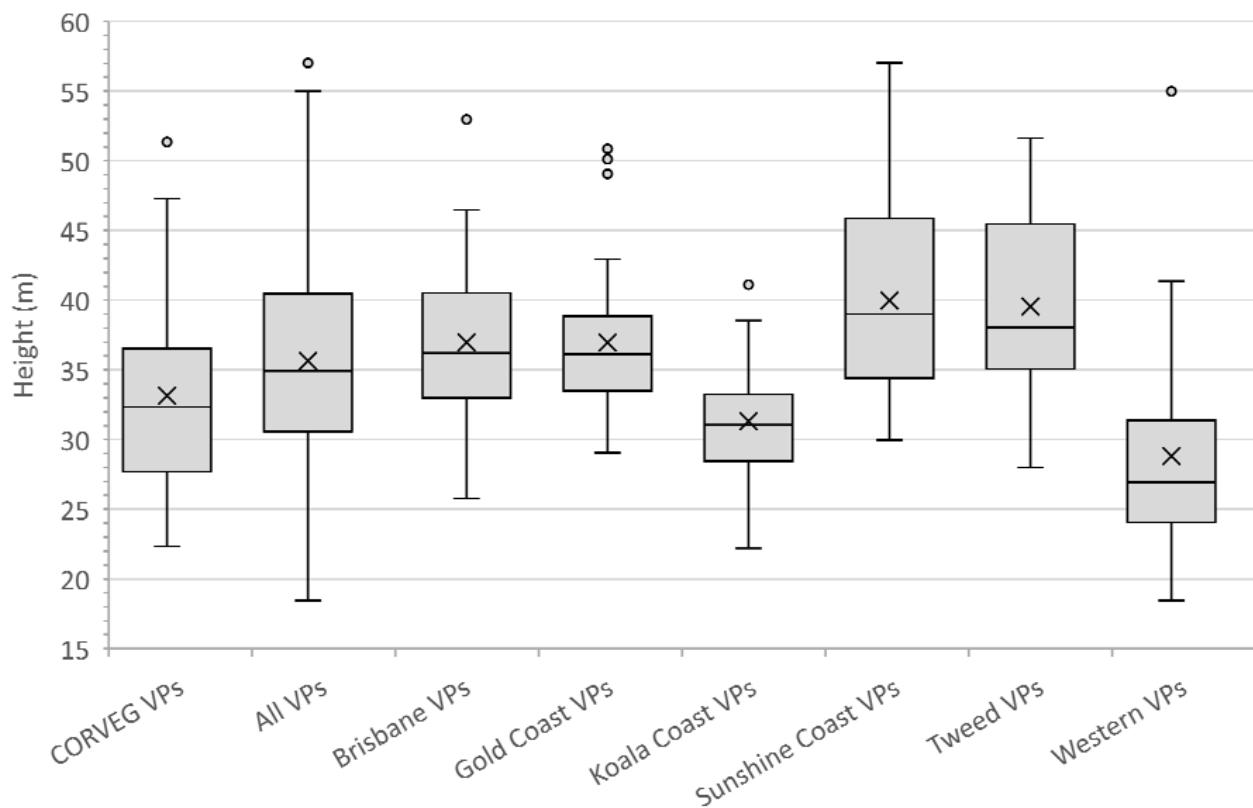


Figure 5. ALS virtual plot (VP) maximum tree heights for 33 Qld Herbarium CORVEG plots and 205 VPs from identified subregional areas in the Southeast Queensland Bioregion (SEQ).

The Koala Coast had substantially lower mean maximum heights (31.32 m) than the average of the other near-coastal regional groups and a lower standard deviation in maximum height (3.89 m) than the average of these groups (6.5 m). The Koala Coast and Western subregions had lower average canopy heights than CORVEG; all the other subregions had higher average heights than CORVEG. The height distribution was positively skewed, particularly for the Gold Coast and Western Regions (Table 6).

Table 6. Variations in the canopy height between CORVEG and six geographically distinct subregional groups within the study area (Figure 1).

Measure	CORVEG	Brisbane	Gold Coast	Koala Coast	Sunshine Coast	Tweed	Western
Mean	33.15	36.97	36.97	31.32	39.98	39.55	28.82
Standard Error	1.13	0.98	0.98	0.71	1.26	1.08	1.18
Maximum	51.36	52.97	50.86	41.12	57.02	51.64	54.99
Minimum	22.36	25.79	29.05	22.21	29.95	28.01	18.47
Median	32.36	36.21	36.12	31.055	39.01	38.05	26.95
Range	29.0	27.2	21.8	18.9	27.1	23.6	36.5
Standard Deviation	6.50	5.78	5.47	3.89	7.13	6.81	7.21
Variance	42.24	33.39	29.97	15.10	50.89	46.35	51.94
Skewness	0.87	0.38	1.21	0.43	0.42	0.21	1.59
Count	33	35	31	30	32	40	37

The differences between CORVEG and the subregional plots were not significant for the Sunshine Coast and Tweed subregions, and between the subregions, the Koala Coast and Western subregions were not significantly different to each other, but both were significantly different to the other four subregions. These four near-coastal subregions (Brisbane, Gold Coast, Sunshine Coast, and Tweed) were not significantly different from each other (Table 7).

Table 7. Differences in the maximum heights between the CORVEG virtual ALS plots and sub-regional virtual plots; significant *p*-values (ANOVA with post hoc tests, df = 231, F = 16.065) are shown in bold. The CORVEG plots were significantly different to the Sunshine Coast and Tweed subregional groups, and the Koala Coast and Western groups were significantly different to all the other subregional groups.

Subregional Group	(C)	B	GC	KC	SC	T	W
(CORVEG)	1	0.158	0.161	0.815	0.003	0.002	0.128
Brisbane		1	1.000	<0.001	0.498	0.570	<0.001
Gold Coast			1	<0.001	0.500	0.573	<0.001
Koala Coast				1	<0.001	<0.001	0.543
Sunshine Coast					1	1.000	<0.001
Tweed						1	<0.001
Western							1

3.3. Intra-Subregional Variation

3.3.1. Intra-Subregional Variation (Koala Coast)

We previously identified (Figure 3) seven intra-subregional groups within the Koala Coast subregional group, and for this part of our analysis, we used a 100% sample rate, i.e., all 25 m × 25 m cells within each group. Figure 6 shows similar within-group variations, except for Group 6, where there was a greater range of height values.

Table 8 shows height variations between these groups; compared to CORVEG, these groups showed a more compact canopy (standard error, standard deviation, and variance were all lower), and the CORVEG distribution was strongly positively skewed. The statistical comparisons are shown in Table 9.

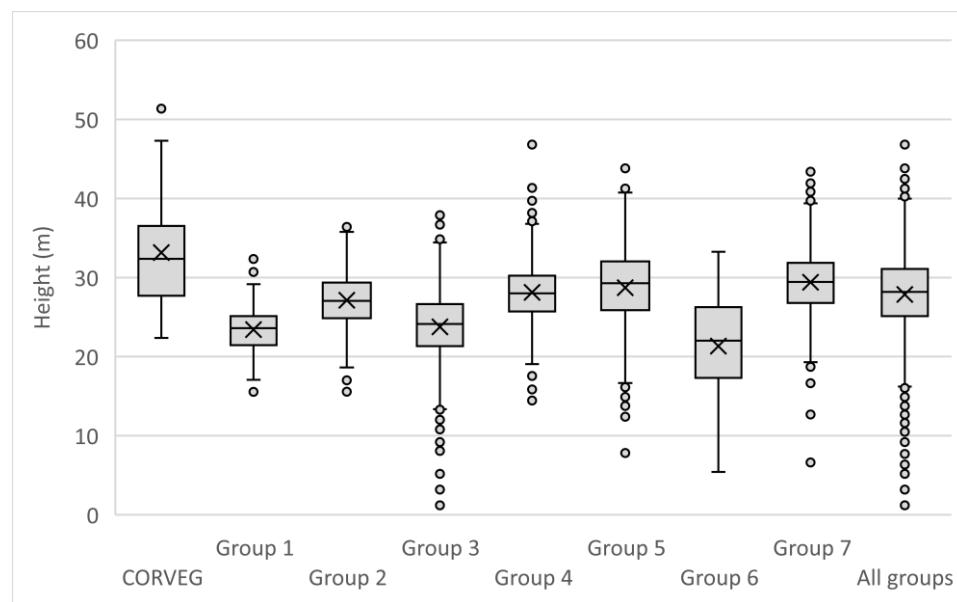


Figure 6. Canopy height variations within the seven Koala Coast intra-subregional groups. Groups 2, 4, 5, and 7 are generally taller than Groups 1, 3, and 6. Heights below 15 m are typically near the polygon group edges and can be attributed to mapping errors or edge disturbances.

Table 8. Variations in the canopy heights for seven geographically distinct intra-subregional groups within the Koala Coast regional group. Compared to CORVEG, several groups have negatively skewed distributions and lower variances.

Measure	CORVEG	Group 1	Group 2	Group 3	Group 4	Group 5	Group 6	Group 7	All Groups
Mean	33.15	23.43	27.13	23.75	28.13	28.72	21.33	29.41	27.86
Standard Error	1.13	0.59	0.23	0.20	0.11	0.12	0.35	0.09	0.06
Maximum	51.36	32.35	36.96	37.89	46.85	43.80	33.25	44.26	46.85
Minimum	22.36	15.54	15.56	1.20	14.45	7.82	5.41	6.62	1.20
Median	32.36	23.60	27.05	24.12	28.01	29.29	22.02	29.45	28.18
Range	29.00	16.81	21.40	36.69	32.40	35.98	27.84	37.64	45.65
Standard Deviation	6.50	3.66	3.63	5.08	3.73	4.86	6.10	3.91	4.96
Variance	42.24	13.41	13.18	25.85	13.90	23.62	37.23	15.26	24.59
Skewness	0.87	0.18	-0.07	-0.97	0.27	-0.49	-0.50	0.12	-0.65
Count	33	38	241	625	1061	1667	309	2066	6607

Table 9. Differences in the maximum heights between the CORVEG and Koala Coast intra-subregional virtual plots; significant *p*-values (ANOVA with post hoc tests, df 1 = 7, df2 = 6032, F = 231.483) are shown in bold. We also compared the CORVEG plots to the pooled group plots (Welch's *t*-test, df = 6038, F = 4.434). Except for Group 7, all the groups were significantly different to CORVEG, and Group 1 was not significantly different to Groups 3 and 6. n.a. = not applicable.

Koala Coast Group	(CORVEG)	Group 1	Group 2	Group 3	Group 4	Group 5	Group 6	Group 7	All Groups
(CORVEG)	1	<0.001	<0.001	<0.001	0.002	0.010	<0.001	0.043	0.002
Group 1		1	<0.001	0.999	<0.001	<0.001	0.063	<0.001	n.a.
Group 2			1	<0.001	0.004	<0.001	<0.001	<0.001	n.a.
Group 3				1	<0.001	<0.001	<0.001	<0.001	n.a.
Group 4					1	0.009	<0.001	<0.001	n.a.
Group 5						1	<0.001	<0.001	n.a.
Group 6							1	<0.001	n.a.
Group 7								1	n.a.
All Groups									1

3.3.2. Intra-Subregional Variation (Sunshine Coast Study Site)

The effects of forest disturbances are readily apparent at the Sunshine Coast intra-subregional study site, which encompasses two land tenures with different land use histories. The canopy heights fall within the CORVEG height ranges but are substantially less than Sunshine Coast virtual plots, and the canopy height in Quinlans averages 6.2 m below the adjoining Mapleton National Park (Figure 7 and Table 10) but still falls within

the CORVEG parameters (e.g., range and standard deviation from the mean). Pairwise differences between CORVEG and the entire study site, and CORVEG and Mapleton National Park, were not significantly different, but all other pairwise comparisons were significantly different (Table 11).

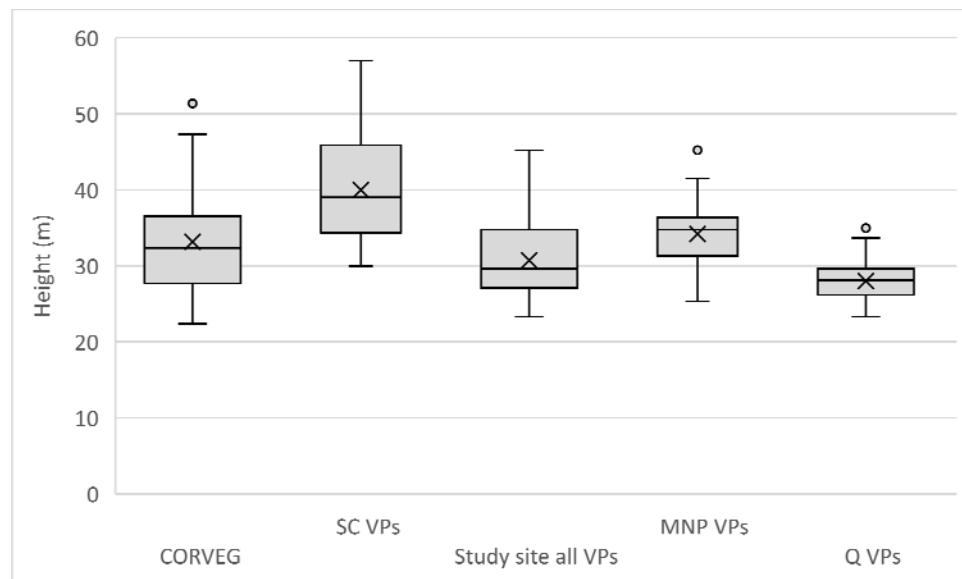


Figure 7. Canopy height variations for the Sunshine Coast intra-subregional study area. The study site, with 50 virtual plots, shows similar distribution to CORVEG as a whole, but Quinlans (Q) is substantially lower than CORVEG, and all the datasets are lower than the 32 Sunshine Coast random virtual plots.

Table 10. Regional CORVEG virtual plots (VPs) and subregional Sunshine Coast VP canopy heights compared to the Sunshine Coast intra-subregional study site. Quinlans (Q) has a lower canopy characterised by a lower standard deviation and low variance.

Measure	CORVEG	SC VPs	Study Site VPs	MNP VPs	Q VPs
Mean	33.15	39.98	30.71	34.19	28.00
Standard Error	1.13	1.26	0.67	0.99	0.53
Maximum	51.36	57.02	45.21	45.21	34.97
Minimum	22.36	29.95	23.29	25.31	23.29
Median	32.36	39.01	29.62	34.75	28.09
Range	29.00	27.07	21.92	19.90	11.68
Standard Deviation	6.50	7.13	4.78	4.65	2.82
Variance	42.24	50.89	22.82	21.65	7.95
Skewness	0.87	0.42	0.78	0.25	0.43
Count	33	32	51	22	28

Table 11. Differences in the maximum heights between the CORVEG and Sunshine Coast intra-subregional study sites; significant *p*-values (Welch's *t*-test) are shown in bold.

Pairwise Combination	F	df	<i>p</i>
CORVEG-Study site	2.163	54.06	0.088
CORVEG-Mapleton NP	2.189	52.68	0.503
CORVEG-Quinlans	10.692	45.13	0.001
Sunshine Coast VPs-Study site	9.816	48.52	0.001
Sunshine Coast VPs-Mapleton NP	7.768	51.90	0.001
Sunshine Coast VPs-Quinlans	26.365	41.53	0.001
Mapleton NP-Quinlans	3.143	32.74	0.001

While it may technically fall within the definition of Remnant Forest (Neldner et al. 2019; Table 4), compared to CORVEG and all the other Sunshine Coast datasets, Quinlans is

lower in height, and we can conclude that Quinlans has a regenerating forest. The difference in canopy heights between the Mapleton National Park and Quinlans is illustrated in Figure 8 using the ALS profile.

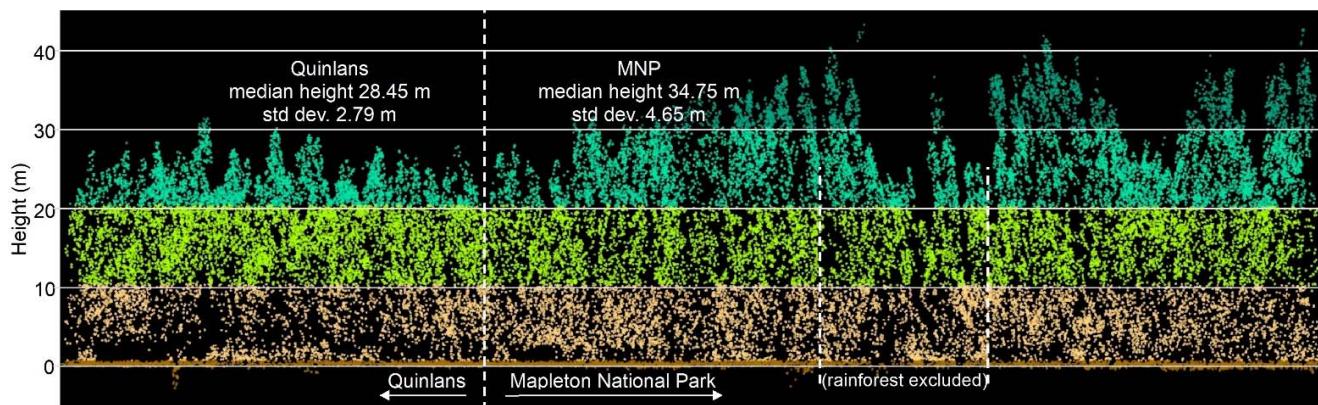


Figure 8. ALS profile (southwest to northeast) across the Sunshine Coast local study site illustrating the forest canopy height differences between the regenerating forest (Quinlans) and mature/old growth forest (Mapleton National Park). Rainforest area is excluded from our research. Height information from Table 8. Transect length 1300 m, vertical exaggeration 7×.

3.4. Clustering and Sample Map Production

The Sunshine Coast study site and two polygon groups in the Koala Coast were selected to evaluate clustering algorithms. We chose these because; (a) we have detailed disturbance history and field data for the study site; (b); the two Koala Coast groups represent higher and lower canopy cohorts; and (c) the three polygon groups are representative of compact, linear, and complex shapes which might, potentially, influence clustering algorithms in different ways. The three groups are shown in Figure 9, coloured by the interquartile ranges (IQR) for each group of cells, with median heights for each IQR shown. By visual inspection we identified six distinct height canopy areas, marked A–E (higher canopy), and an ecotone identified during fieldwork marked F (moderate canopy height) in Figure 9.

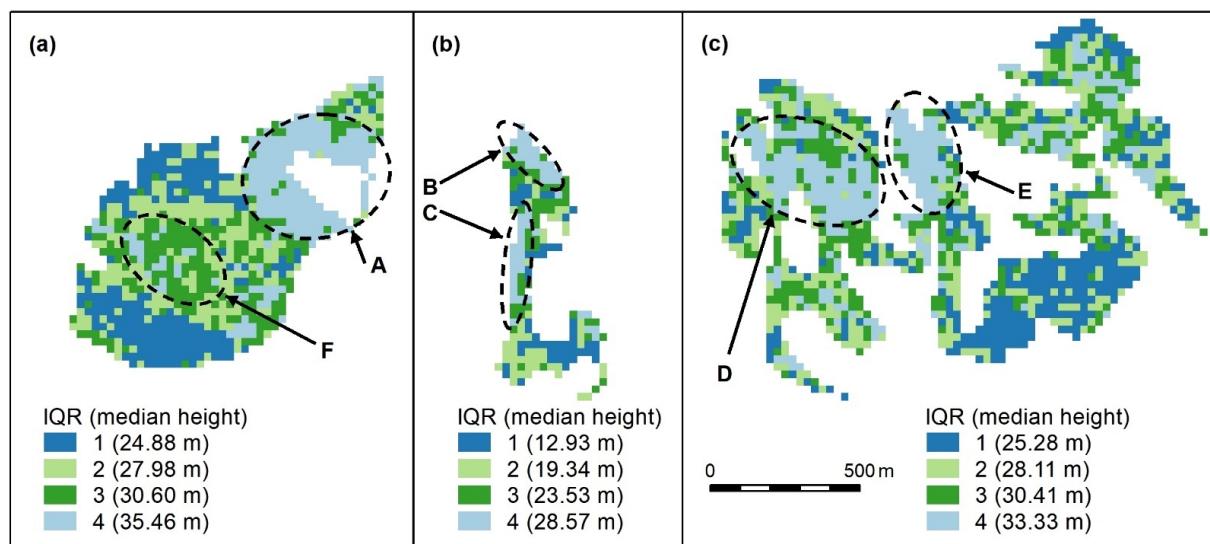


Figure 9. Polygon groups representing compact (a), linear (b), and complex (c) shapes, with interquartile ranges (IQR) and interquartile median height. We used these maps to assess clustering algorithms. Higher-canopy spatial patterns are evident, for e.g., patches of higher canopy (A–E, light blue), and a canopy area of moderate height (F, dark green, some light green and light blue), surrounded by canopy with slightly lower height (light green). Heights appear more randomly distributed in other areas (light green, dark green) except for areas of lowest canopy height (dark blue). All maps use the same scale.

For 90 output cluster maps, an expert panel of three visually evaluated each map to assess how well the algorithm identified the six nominated features for individual polygons, results are summarised in Table 12. Elbow plots (see Appendix A) show the class variance explained by each algorithm, indicate the appropriate number of clusters, and algorithm efficiency. The complete evaluation process is detailed in Appendix A.

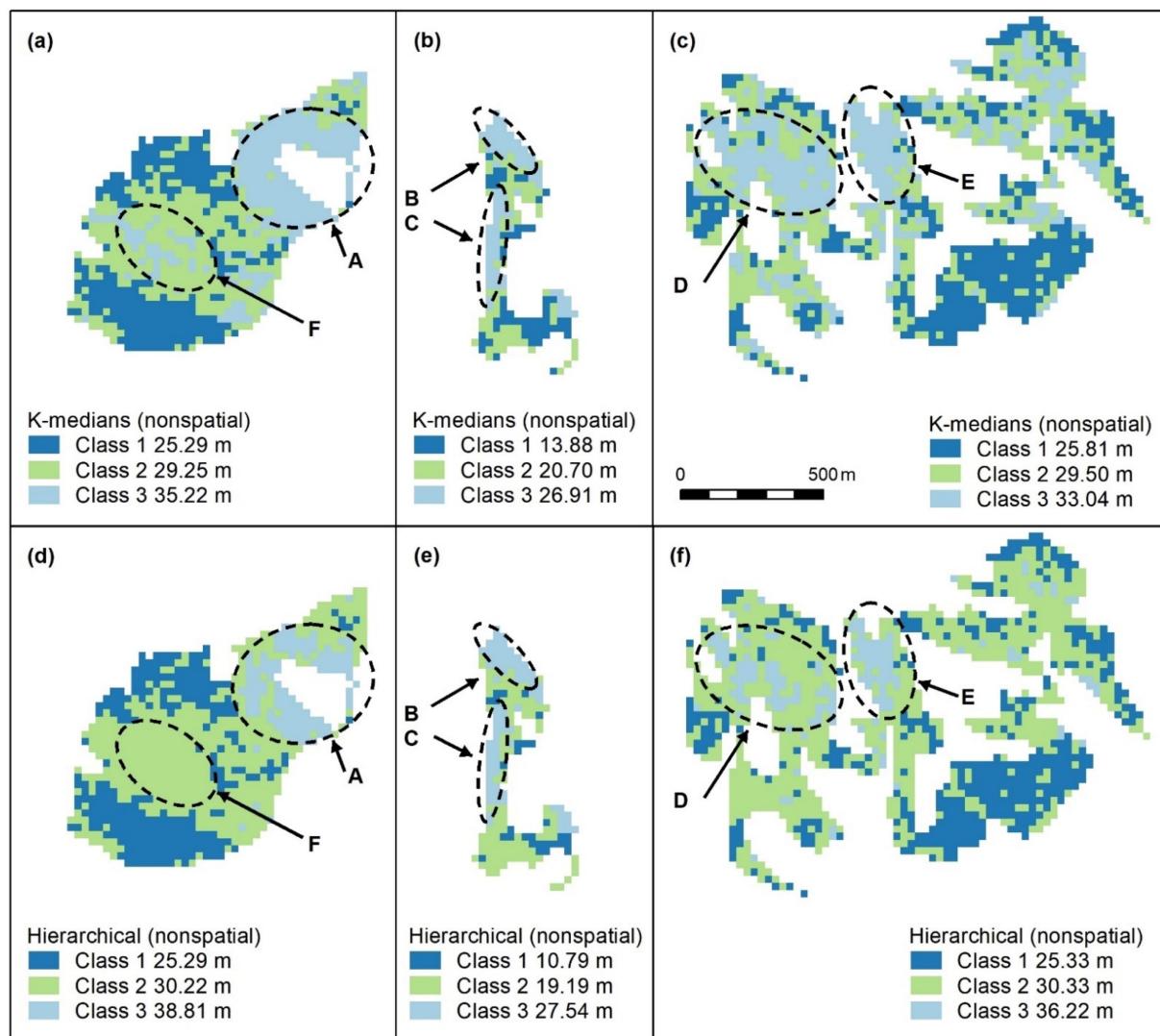


Figure 10. Nonspatial K-medians (a–c) and nonspatial hierarchical cluster maps (d–f) for individual polygons with 3 height classes. Letters A–F indicate areas of higher canopy as shown in Figure 9. Five of six features were identified, and cluster areas are similar except for feature D (c,f). Feature F was partially captured by K-medians (d).

Three-class cluster maps with the highest evaluation scores are shown in Figure 10. Nonspatial K-means and hierarchical clustering successfully identified five of six features, but cluster areas varied substantially between algorithms for Feature D. Feature F, the forest ecotone community, was only partially captured by K-medians.

Four-class cluster maps are shown in Figure 12. K-medians again ranked highest within the nonspatial algorithms, but several features have less well-defined clusters than three-class maps. Redcap [62] had the highest score from the expert panel, with tightly defined contiguous clusters for all features. Redcap extended the area of feature A, and effectively shows the height difference between Mapleton National Park and regenerating forest in Quinlans (see Figure 4). Redcap also accurately captures the extent of the forest ecotone in Quinlans (Feature F). However, compared to Figure 9 and other cluster

maps (Figure 10), Redcap overestimated the extent of one feature (D) and underestimated another (E).

Table 12. Visual evaluation by a panel composed of one author (DM), a GIS analyst (DS) and an ecologist (DJ); for maps produced for individual polygons by clustering algorithms. Maps were ranked according to how well six nominated features (A–F in Figure 10) were captured by each algorithm, each with 2, 3, and 4 classes. Panellists first independently allocated scores, disagreements were highlighted, and the panellists re-evaluated their scores. Key: P—feature identified (1), not identified (0); A—feature has similar area (1), smaller or larger (0); C—feature contiguous (1), not contiguous (0), ‘s’ prefix denotes spatially constrained algorithm. **1 1 1** = panellists agree that feature recognised by algorithm using all criteria- **0 0 0** feature not recognised using any criteria- Dark grey cells indicate panellists agree with individual criteria, **0/1** one panellist disagrees. For all cells, there was 86.3% agreement by all panellists. Panellists most often disagreed on whether feature D was captured. Nonspatial K-medians (all classes) has the highest overall score, followed by nonspatial K-means. For 4 classes, Redcap ranked slightly higher than nonspatial K-medians.

Algorithm	Feature Classes	P	A	C	P	A	C	P	A	C	P	A	C	P	A	C	P	A	C	Sum	Rank
K-means	2	1	1	1	1	1	1	1	1	1	1	0	1	1	1	1	1	1	0	16	2
K-medians	2	1	1	1	1	1	1	1	1	1	1	0	1	1	1	1	1	1	1	17	1
Hier. Clust.	2	1	0	1	1	1	1	1	1	1	1	0	1	1	0	1	1	1	1	12	3
s K-means	2	1	0	1	1	0	1	1	0	1	0	0	0	0	0	0	0	0	0	6	6
s K-medians	2	1	0	1	1	0	1	0	0	0	1	0	1	1	0	1	0	0	0	8	5
s Hier. Clust.	2	1	0	1	1	1	1	0	0	0	1	0	1	1	0	1	1	0	0	9	4
s Skater	2	1	0	1	1	1	1	0	0	0	1	0	1	1	0	1	0	0	0	9	4
s Redcap	2	1	0	1	1	1	1	0	0	0	1	0	1	1	0	1	0	0	0	9	4
K-means	3	1	0	1	1	1	1	1	1	1	1	0	0	1	1	1	0	0	0	12	2
K-medians	3	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	0	0	16	1
Hier. Clust.	3	1	0	1	1	1	1	1	1	1	1	0	0	1	0	1	1	0	0	11	3
s K-means	3	1	0	1	1	0	1	0	0	0	1	0	1	1	1	1	0	0	0	9	4
s K-medians	3	1	0	1	1	0	1	0	0	0	1	0	1	1	1	1	0	0	0	9	4
s Hier. Clust.	3	1	0	1	1	1	1	0	0	0	0	0	0	1	0	1	0	0	0	7	5
s Skater	3	1	0	1	1	1	1	0	0	0	1	0	1	1	0	1	0	0	0	9	4
s Redcap	3	1	0	1	1	1	1	0	0	0	0	0	0	1	0	1	1	1	0	9	4
K-means	4	1	0	1	1	1	1	1	1	1	0	0	0	1	0	1	1	0	0	11	5
K-medians	4	1	1	1	1	1	1	1	1	1	1	0	0	1	1	1	1	0	1	15	2
Hier. Clust.	4	1	1	0	1	1	1	1	1	1	1	0	0	0	1	0	1	1	0	12	4
s K-means	4	1	0	1	0	0	0	0	0	0	1	0	1	1	1	1	0	0	0	7	6
s K-medians	4	1	0	1	0	0	0	0	0	0	1	0	1	1	1	1	0	0	0	7	6
s Hier. Clust.	4	1	0	1	1	1	1	1	1	1	1	0	1	1	0	1	1	0	1	14	3
s Skater	4	1	0	1	1	1	1	1	1	1	1	0	1	1	1	1	1	0	1	15	2
s Redcap	4	1	1	1	1	1	1	1	1	1	1	0	1	1	0	1	1	1	1	16	1
Disagreements	(Sum)	2	4	2	2		3		13	2	11	3	8	3	4		2				

The panel evaluation for multiple polygon cluster maps produced using all Koala Coast subregional data is summarised in Table 13.

Table 13. Visual evaluation of the clustering algorithms using all height data within the Koala Coast subregional group (6007 cells), scored using the same methods as Table 12. Maps were ranked according to how well four nominated features (B–E in Figure A11) were captured by each algorithm, each with 2, 3, and 4 classes. For all cells, there was 84.3% agreement by all panellists, and the panellists disagreed most often on whether the feature area (column A) was captured. K-means and K-medians are ranked equally for 3 classes, but K-medians is ranked higher for 4 classes.

Algorithm	Feature Classes	P	B	A	C	P	C	A	C	P	D	A	C	P	E	A	C	Sum	Rank
K-means	2	1	0	0	1	0	0	0	1	1	0	1	1	0	1	1	1	6	2
K-medians	2	1	0	0	1	0	0	0	1	1	0	1	1	0	1	0	1	6	2
Hier. Clust.	2	1	1	0	1	0	0	1	0	1	0	1	1	0	1	0	1	7	1
K-means	3	1	1	0	1	1	0	0	1	1	1	1	1	1	1	1	1	10	1
K-medians	3	1	1	0	1	1	0	0	1	1	1	1	1	1	1	1	1	10	1
Hier. Clust.	3	1	1	0	1	1	0	0	1	0	0	1	1	0	1	0	1	8	2
K-means	4	1	0	0	0	1	0	0	0	1	0	0	0	1	1	1	1	6	3
K-medians	4	1	1	0	1	1	0	0	1	1	1	1	1	1	1	1	1	10	1
Hier. Clust.	4	1	1	0	1	1	0	0	1	0	0	0	1	1	1	1	1	8	2
Disagreements	(Sum)		3		1	4			2			1	5	1					

K-means and K-medians were ranked equally highly for three classes (Figure 11), but K-medians was ranked higher for four classes. Features D and E were identified, but because features B and C were not high compared to D and E, were assigned to lower canopy height classes. In addition, areas of High Value Regrowth Forest [63] were identified by both algorithms.

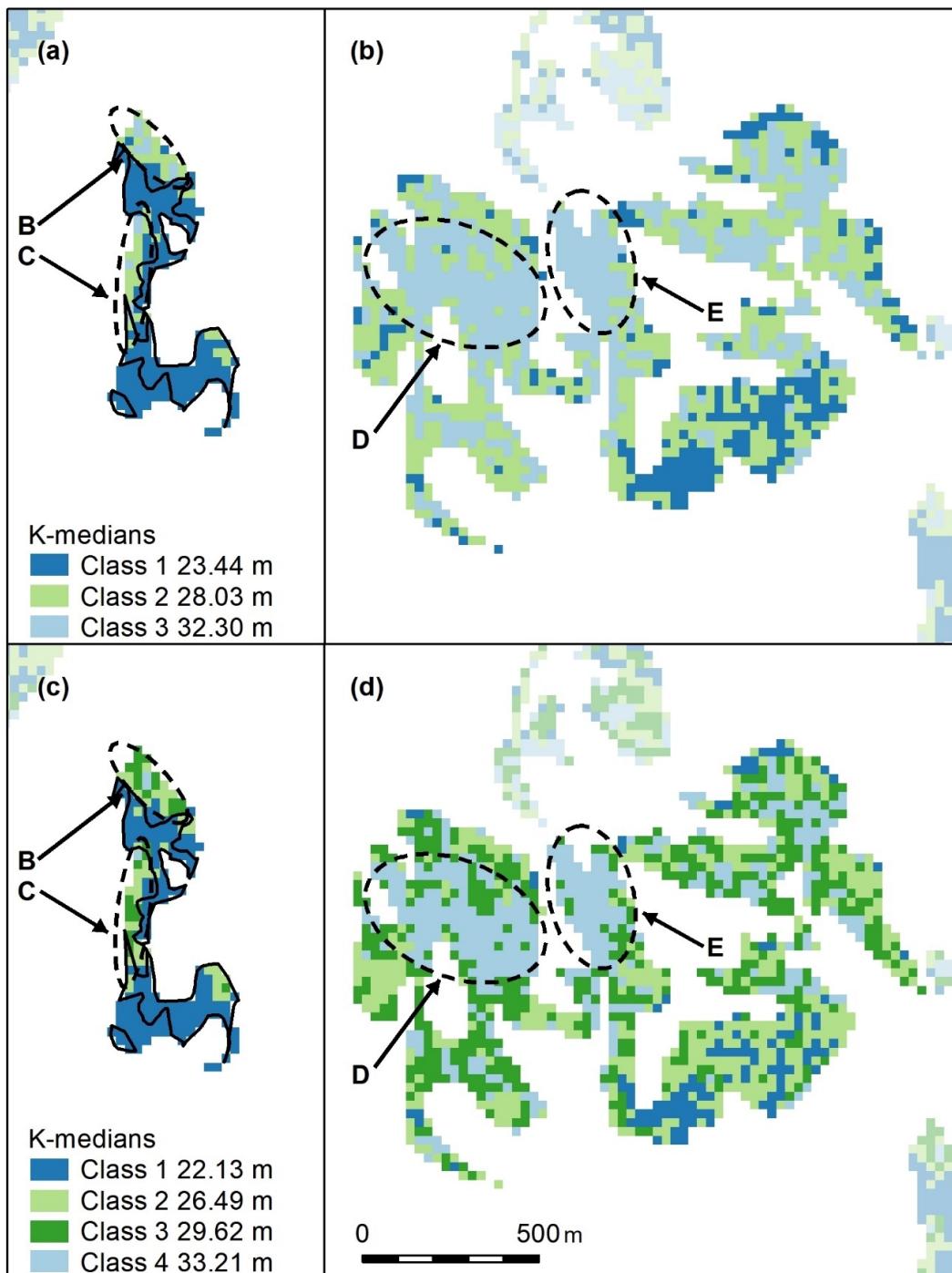


Figure 11. Nonspatial K-medians cluster maps for the Group 6 (low canopy) polygon (a,c) and Group 7 (high canopy) polygon (b,d) using all the height data from Koala Coast subregional Groups 1–7 (Figure 3). Letters B–E indicate areas of higher canopy as shown in Figure 9. Features D and E are clearly identified but B and C are not, because they belong to the second- and third-highest height classes across the wider Koala Coast subregion. B and C are still significantly higher than the lowest height class (a,c, dark blue). Dark blue areas in (a,c) accurately capture areas of Mature Regrowth vegetation [63], outlined in black.

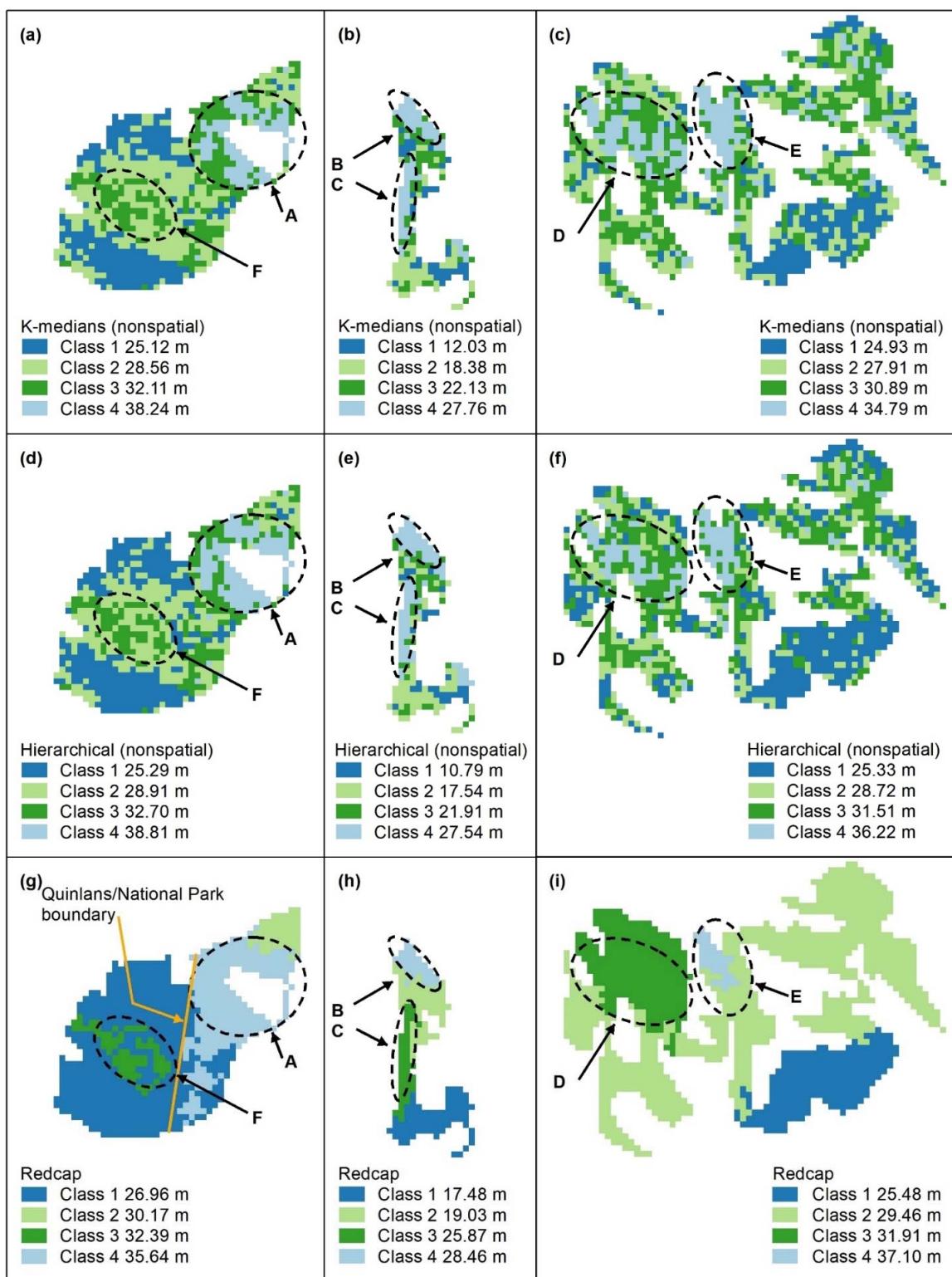


Figure 12. Nonspatial K-medians (a–c), nonspatial hierarchical (d–f) and Redcap [64] spatial hierarchical (g–i) cluster maps for individual polygons (refer Figure 9) with 4 height classes. Letters A–F indicate areas of higher canopy as shown in Figure 9. All six features are identified, but features A, D, and F are not contiguous. In comparison, the spatially constrained algorithm Redcap (g–i) identifies all the features as contiguous but assigns C, D, and F to a lower height class. Compared to Figure 9, Redcap overestimates D and underestimates E; A is also larger, because it also captures height differences reflecting the land use history of Quinlans compared to Mapleton National Park. Redcap also accurately captures the extent of an ecotone identified during fieldwork (F); this ecotone has a higher proportion of a preferred koala food tree than the surrounding area (dark blue), and which is absent from the national park.

4. Discussion

Current koala habitat mapping uses regional-scale vegetation maps incorporating a high degree of floristic variability, which may limit their application at the local scale (1–300 hectares), where tree species variability, i.e., food resource availability, becomes more apparent. Since koalas prefer larger trees [13,16], incorporating this variable into a map overlay might be useful for habitat management at this scale [3]. Our study was aimed at how best to incorporate larger trees into the current habitat maps using heights obtained from ALS. To derive the parameters for an overlay map, we used published information as a baseline and found that, at the regional scale, forest canopy heights are within published parameters [53] but, at the subregional and intra-subregional scales, the height variability is generally outside these parameters. Additionally, the variability is both significantly different between subregions and within subregions. Two height parameters demonstrated the effects of previous land use. In regenerating forests, not only is the canopy height lower, but the canopy has a lower standard deviation in height, i.e., the canopy is “flatter”, with shade-tolerant and shade-intolerant species in active competition [17]. The range in the standard deviation therefore has potential for identifying similar areas of regenerating forests.

To incorporate ALS canopy height information into a habitat map overlay, we examined several clustering algorithms and found that nonspatial K-medians best characterised the nominated high-canopy areas for multiple polygons at the intra-subregional scale. This approach might be useful for height comparisons between different areas for broader management purposes but produced clusters suffering from a lack of contiguity, limiting their use at finer scales. At the site scale, i.e., individual habitat map polygons, the spatially constrained hierarchical clustering algorithm Redcap [61,62] was able to: (a) provide greater map utility by identifying contiguous clusters; (b) separate historically disturbed, lower canopy height forests from relatively undisturbed higher canopy forests, and (c) capture a forest ecotone with a higher proportion of preferred koala food resources. Redcap further showed its utility in another polygon by capturing an area independently mapped as Mature Regrowth Forest [63]. We used algorithms requiring no user input so that our approach could be automated, allowing any number of forest map polygons to be clustered at any desired scale. The number of algorithms we used, and the consistent results between many of them, confirmed that most of our visually nominated features were, indeed, real clusters of similar heights and that our algorithm recommendations were valid. This multifaceted approach was recommended by Ketchen and Shook [64], who warned against adopting any single clustering algorithm. However, we noted that two of our nominated features were not recognised as being statistically significant by any algorithm, which shows that arbitrary classifications of canopy height (a) may not be meaningful and (b) may not be repeatable.

The previous use of multivariate modelling that included ALS-derived canopy height as a variable has shown that this metric features prominently within these analyses, for example, by using multivariate K-means clustering to portray forest structures, one study found that most landscape indices were rendered redundant by the median and standard deviation of the canopy height [65], i.e., ALS provides directly measured predictor variables and improves the accuracy of maps compared to inferred variables. In Alberta, Canada, a cluster analysis with ALS variables (canopy height, cover, and height variation) was used to identify areas of high conservation values within mapped forest polygons at the regional scale [66]. In another example, a Northern Spotted Owl (*Strix occidentalis caurina*) nesting habitat in Oregon, USA, mapped using traditional approaches (aerial photo interpretation, satellite imagery, and environmental predictor variables), was substantially improved with the addition of the ALS-derived large tree density and average stand height [67], and the importance of height was confirmed by another study in Northern California [68]. The maximum canopy height was found to be the best predictor of Spotted Owl habitats in Northern California (*Strix occidentalis*) [69]. In The Netherlands, a study found that, for species distribution models, the ALS-derived canopy height was a better predictor of bird

species presence/absence than a suite of land cover variables [70] and had advantages at multiple scales where species habitat requirements are either not known or vary according to particular needs (e.g., feeding, breeding, and dispersal) [71].

We have recommended the use of Redcap with four classes for individual polygons, but to further maximise the overlay map utility at broader scales, for example, in conservation planning, we recommend that a height overlay also includes height classes derived from K-medians. For at-risk species dependent on large hollow-bearing trees, such as the greater glider (*Petauroides volans*) [72], this multiscale approach might be more appropriate. Koala researchers have previously stratified field sites based on low-resolution vegetation community types and/or geology maps [18]; sampling strategies using height clusters could reduce field efforts by effectively sampling areas of mature forest, regrowth, or ecotones within these broader map units. A similar approach using LiDAR has been used to increase the efficiency of survey plot sampling designs in pine plantations [73].

In Queensland, the distinction between Remnant and Regrowth vegetation is important, because, e.g., Remnant Vegetation [54] is subject to stricter legislation governing land use compared to Regrowth or other undefined forest; therefore, it is necessary to accurately differentiate between these forest types. We identified potential mapping errors within our sample forest communities by identifying areas below the minimum height required for classification as Remnant Vegetation [54]; the application of our methods could highlight these mapping errors so they can be corrected. During our research, we noted gaps and overlaps between the Remnant and Mature Regrowth datasets, and because we successfully captured mapped areas of Mature Regrowth [63], our methods could also be used to rectify these mapping errors.

We acknowledge that our study was limited in scope, with the Sunshine Coast and Western areas having only approximately 30% LiDAR coverage. As time goes on, we can expect greater LiDAR coverage, not only from ALS but from newer LiDAR platforms, e.g., the recently commissioned Global Ecosystem Dynamics Investigation (GEDI) satellite [74]. Our study only examined one regional vegetation community and assessed three community polygons with clustering, so our results may not be applicable to all forest communities. However, by choosing a floristically variable community with a likely corresponding variability in canopy height, and with disturbance information for one polygon and partial disturbance information for another, we regard our study as a valuable contribution to this topic.

5. Conclusions

In this study we showed that, because forest canopy heights vary at the subregional and intra-subregional scales, height overlay maps need to incorporate this variability. A regional framework provides an overall picture of the forest canopy height, but at the subregional scale and intra-regional scale relevant to habitat management, the overlay can incorporate attributes derived from the canopy height at these scales. This approach can be implemented over broad areas to produce a multiscale map that can be used to capture an important koala habitat attribute, i.e., that koalas prefer larger trees at any given location.

For koalas, which constantly face development pressure, habitat maps depicting areas of large trees could provide valuable assistance in identifying valuable habitats for protection. Our clustering approach could provide unambiguous and independent tools to assess development at both the site and local government authority scales, reducing the reliance of development approval authorities on information supplied by development proponents. Other potential uses include the identification of candidate old-growth forests, mapping areas of larger hollow-bearing trees required for fauna, conservation reserve planning, and improvement in the spatial accuracy of forest community mapping.

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Appendix A

Appendix A.1. Introduction

We regarded simple height classification using arbitrary height classes unsatisfactory for two reasons: (a) class definitions are arbitrary and (b) may not identify actual areas of interest. Clustering algorithms are capable of separating heights in a statistically robust and repeatable manner and require little or no user input. With the aim of producing classified canopy height maps suitable for Koala habitat management and other forest management purposes, we firstly used clustering algorithms implemented in *Geoda* software (<https://geodacenter.github.io/cheatsheet.html>, accessed on 24 June 2021) to produce three separate cluster maps for individual polygon groups representing compact, linear, and complex shapes. Then, for the two of these polygons within the Koala Coast, we ran those algorithms capable of clustering multiple polygons to (a) enable comparisons between any polygon groups and (b) compare these with maps generated with individual polygon groups.

We evaluated the maps in three ways. Firstly, we constructed elbow plots, which (a) indicate the appropriate number of clusters and, (b) by displaying the variance explained by the number of clusters for each algorithm, provide comparisons between algorithms. Secondly, an expert panel composed of one of the authors (DM) and two analysts scored each map visually using three simple criteria. Then, the maps were assessed by the magnitude of separation between the highest canopy class and the next-highest class. The three methods were individually ranked, and finally, a total rank score was used to choose the best map.

The cluster maps, all with two, three, or four classes, are shown below (Appendix A.2), together with the panel score for each map (see Appendix A.5 for the score allocation).

Appendix A.2. Individual Polygon Clustering

Our aim was to evaluate which *Geoda* algorithm best detects higher canopy areas (features A–F) in each of the three polygons depicted in Figure 11, reproduced below (Figure A1). The algorithms for the individual polygon groups used only the height data within that group.

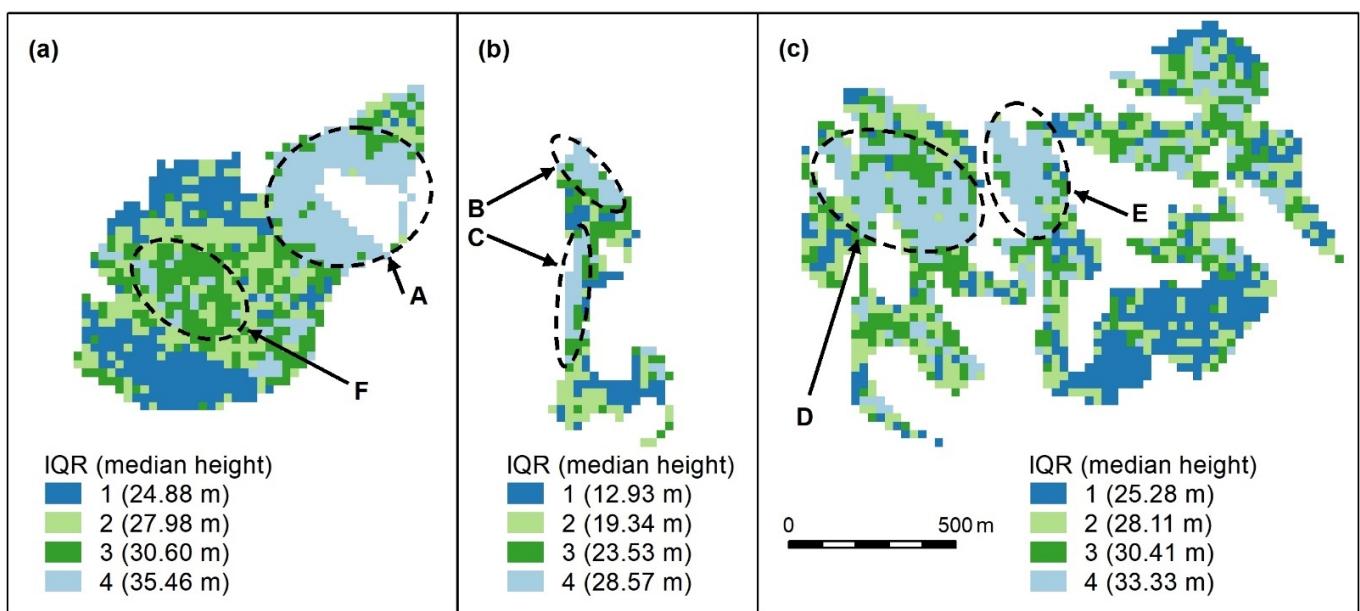


Figure A1. Compact (a), linear (b), and complex (c) polygon groups, showing the interquartile ranges (IQR) and interquartile median canopy heights within each group. We used this map to assess and compare the clustering algorithm outputs which follow. Spatial patterns are evident, for e.g., patches of the highest canopy (A–E, light blue) and an area of moderate height (F, dark green), surrounded by a canopy with a slightly lower height (light green). The height appears more randomly distributed in other areas, except for lower canopy height areas (dark blue). All maps use the same scale.

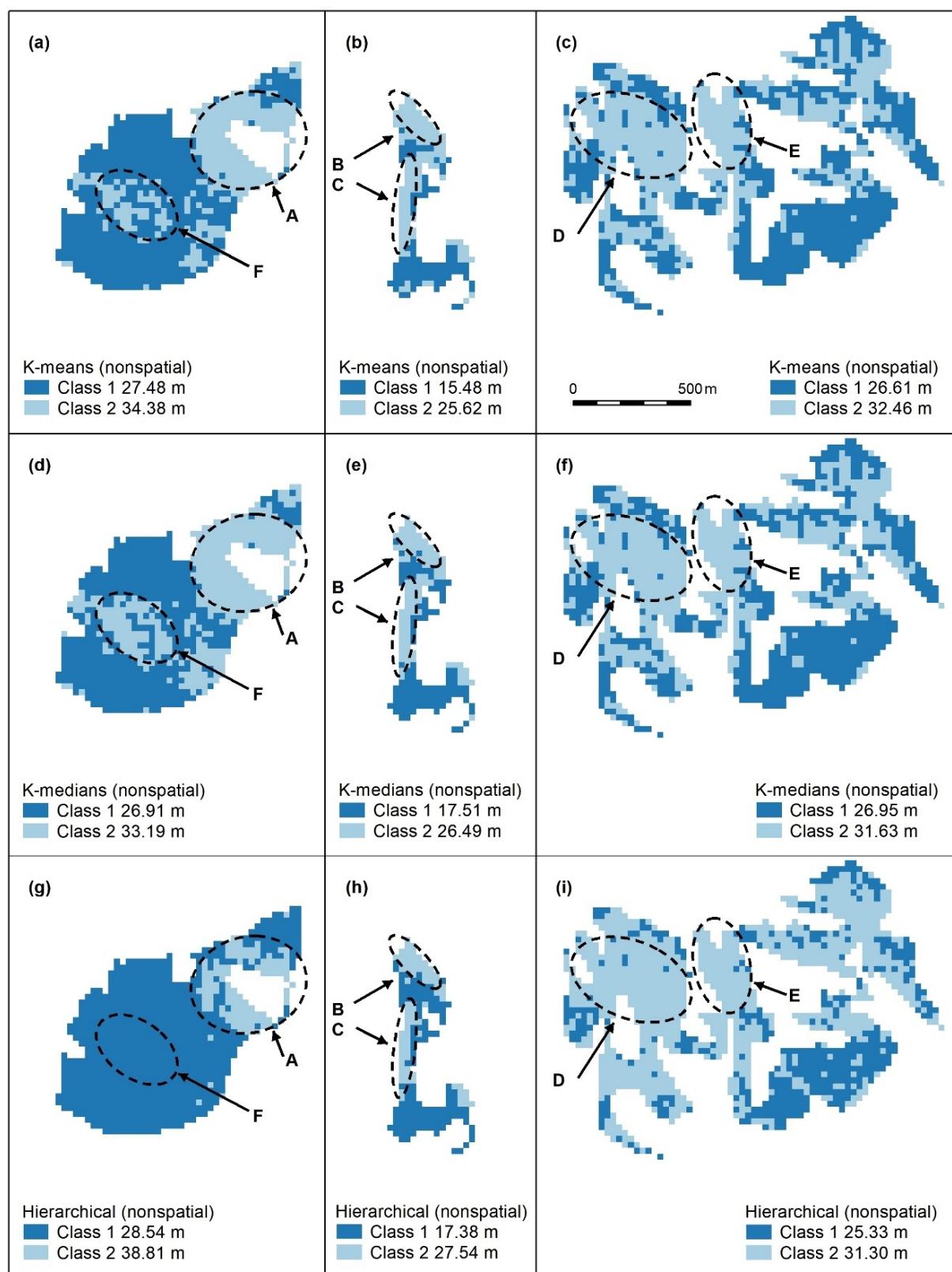


Figure A2. Non-spatial algorithms with 2 classes, using K-means (a–c), K-medians (d–f) and non-spatial hierarchical clustering (g–i). Letters A–F indicate areas of higher canopy. Features A–E are captured. K-medians best captures the areal extent of these features and best captures F, an unmapped area of the koala food tree *Eucalyptus tereticornis* (d). Panel scores: K-means = 16, K-medians = 17, and Hierarchical = 12.

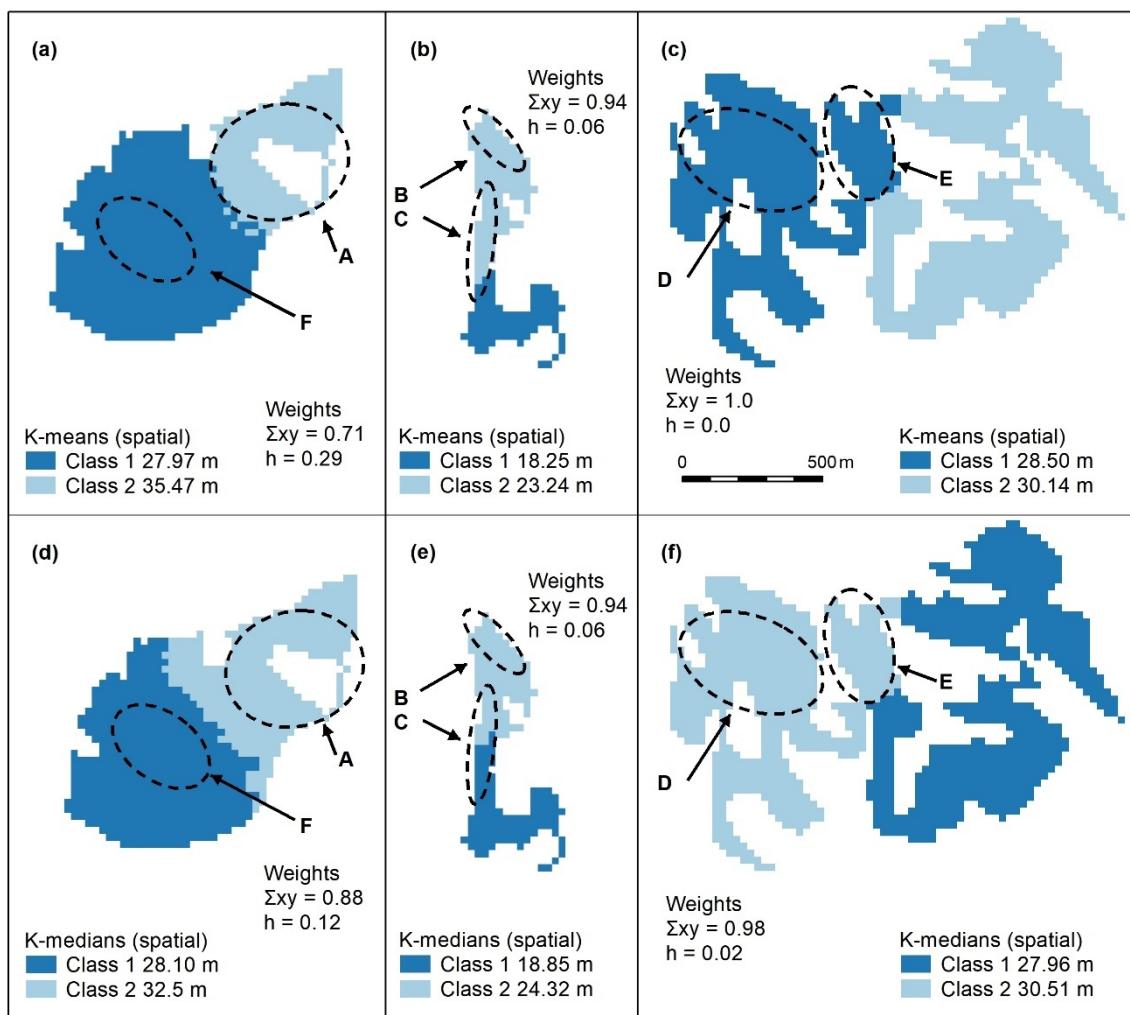


Figure A3. Spatially constrained K-means (**a–c**) and K-medians (**d–f**) with 2 classes. These algorithms use spatial weights (x, y), which, in all cases, outweigh those of the object (h = height). Letters A–F indicate areas of higher canopy, only one of the designated features (E) is adequately captured using these algorithms (**f**). Panel scores: K-means = 6 and K-medians = 8.

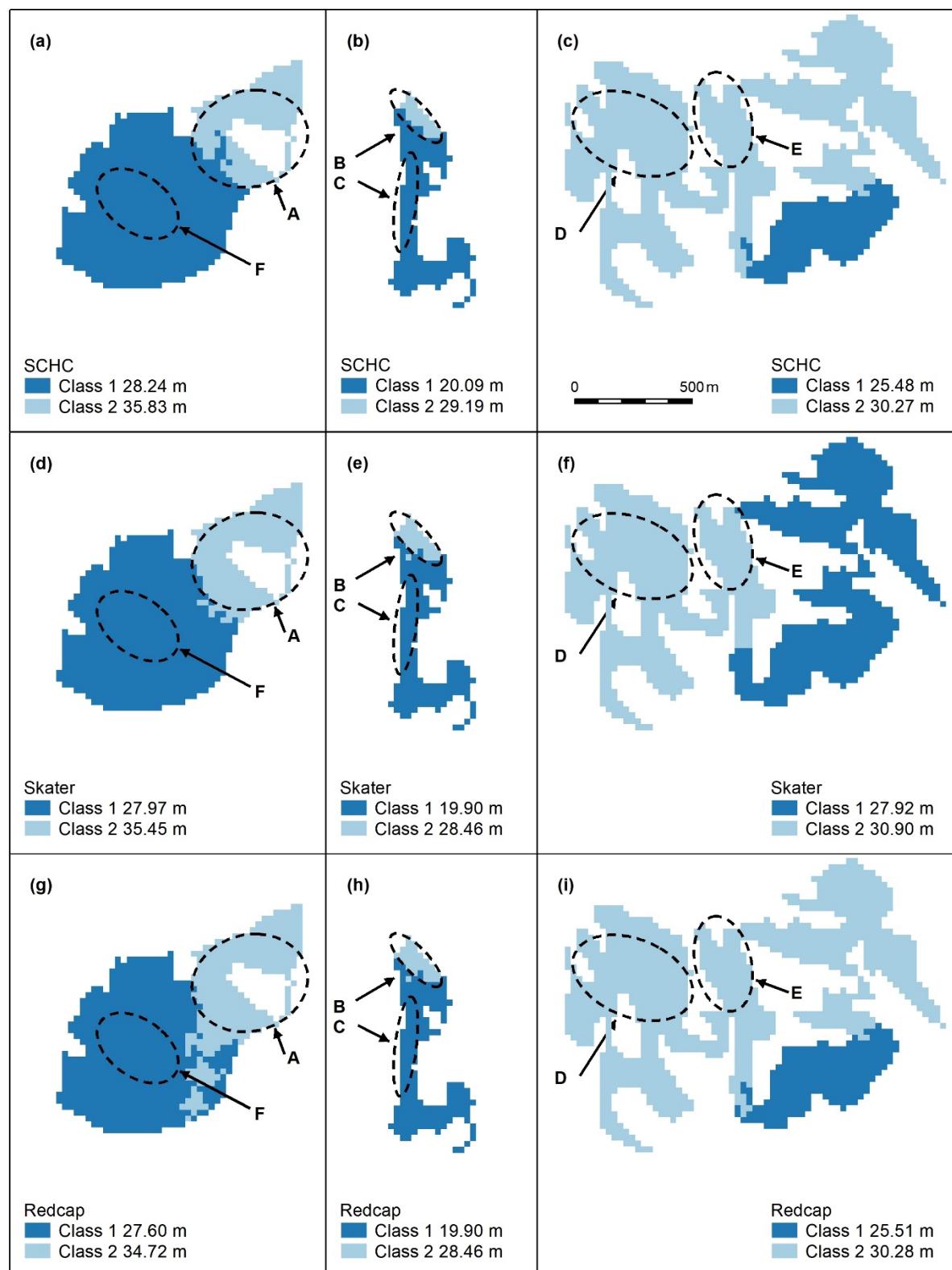


Figure A4. Spatially constrained hierarchical (SCHC, (a–c), Skater (d–f), and Redcap (g–i)) clustering maps with 2 classes. Letters A–F indicate areas of higher canopy, features A and B are adequately captured by these algorithms (a,b,d,e,g,h), but the areal extents of D and E (e,f,i) are too large. C and F are not recognised. Redcap separates an area of regrowth forest on land previously used for farming (g, dark blue) adjoining a relatively undisturbed national park (light blue), which encompasses Feature A. Panel scores: Hierarchical clustering = 9, Skater = 9, and Redcap = 9.

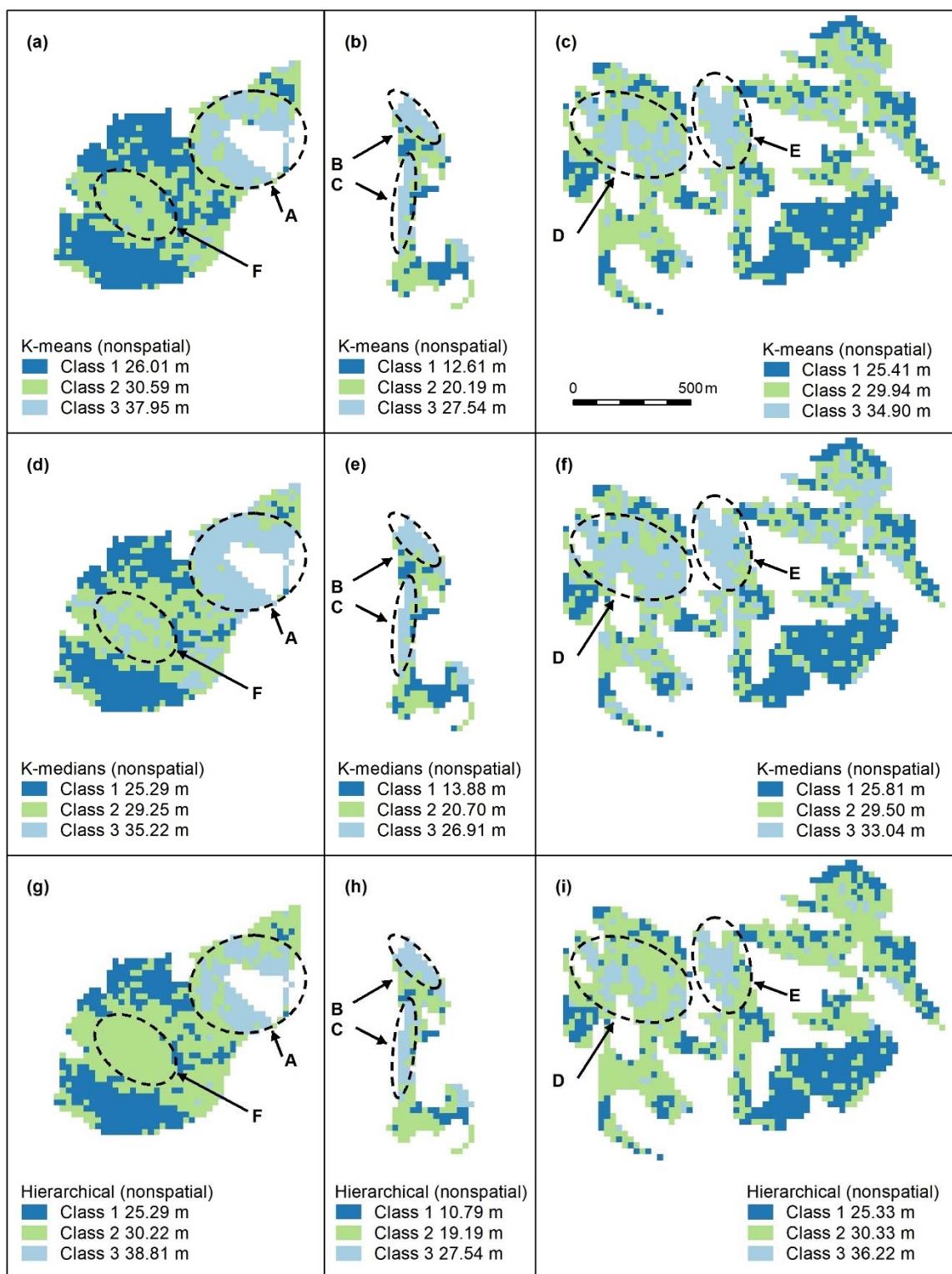


Figure A5. Nonspatial algorithm maps (K-means, a–c; K-medians, d–f; nonspatial hierarchical, g–i) with 3 classes. Letters A–F indicate areas of higher canopy. K-medians (d–f) best captures features A, D, C, all algorithms capture B and C. No algorithm adequately captures F, the unmapped area of *Eucalyptus tereticornis*. Panel scores: K-means = 12, K-medians = 16, and Hierarchical = 11.

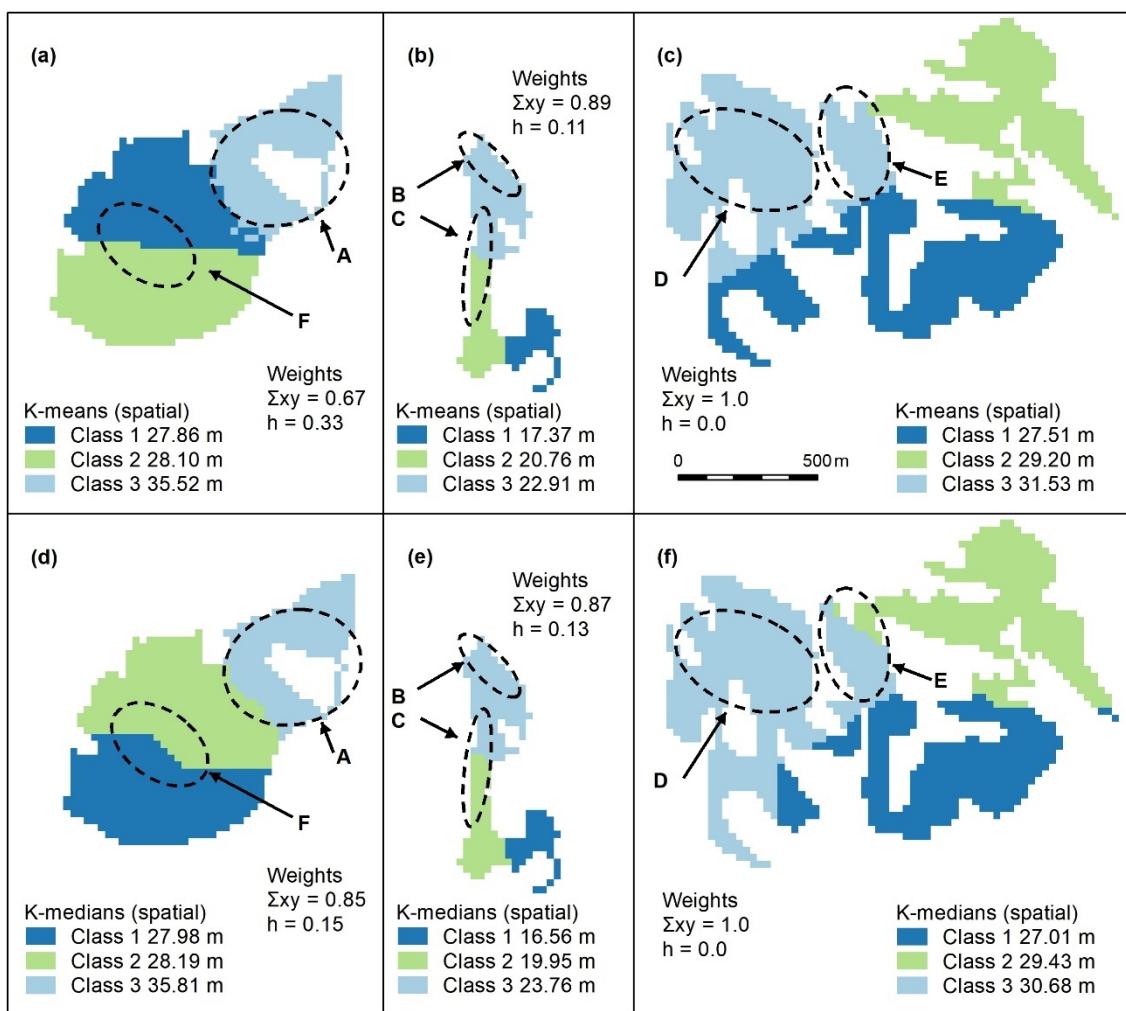


Figure A6. Spatially constrained K-means (**a–c**) and K-medians (**d–f**) cluster maps with 3 classes. Letters A–F indicate areas of higher canopy. Again, these algorithms use spatial weights (x, y) which outweigh those of the object (h = height). The only features adequately captured by these algorithms are A (**a,d**) and E (**c,f**). Panel scores: K-means = 9 and K-medians = 9.

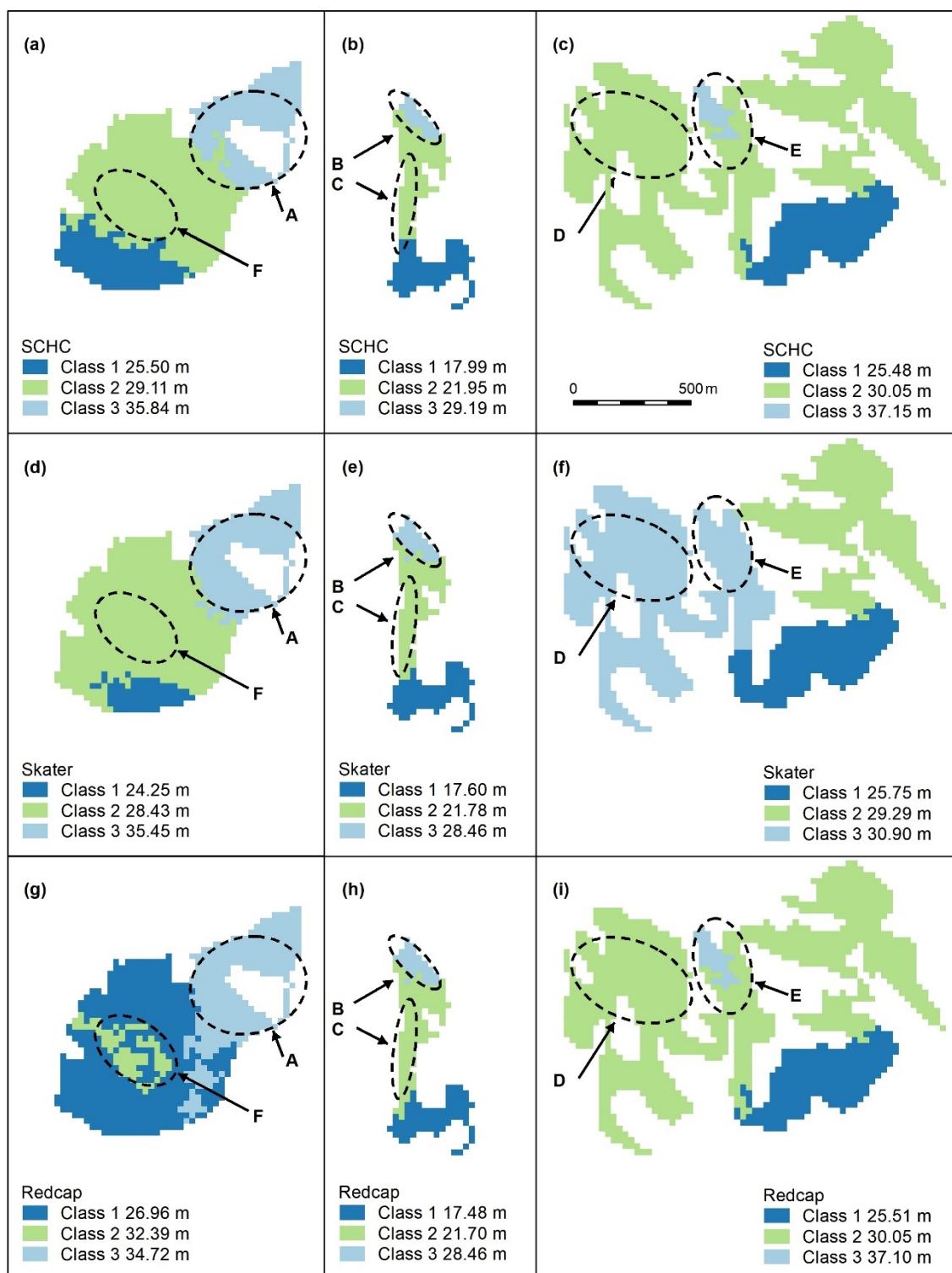


Figure A7. Spatially constrained hierarchical (SCHC, (a–c), Skater (d–f), and Redcap (g–i)) clustering maps with 3 classes. Letters A–F indicate areas of higher canopy, features A and B are adequately captured by all algorithms, but the areal extent of E is too small (c,i) or too large (f). C and D are not recognised. Redcap (g) separates an area of regrowth forest on land previously used for farming (dark blue) adjoining a relatively undisturbed national park (light blue), which encompasses Feature A and recognises F, the unmapped area of *Eucalyptus tereticornis*. Panel scores: Hierarchical clustering = 7, Skater = 9, and Redcap = 9.

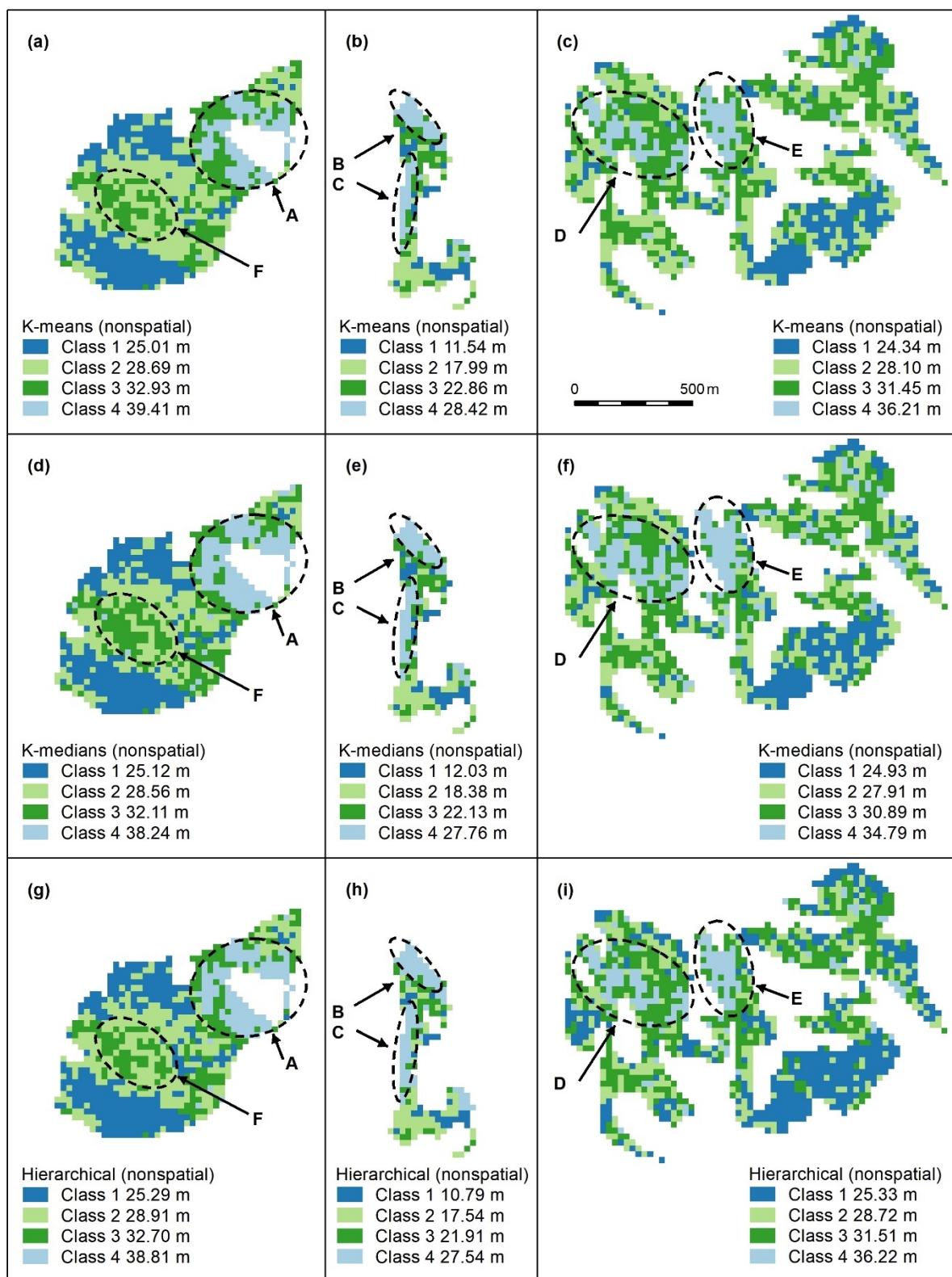


Figure A8. Nonspatial algorithm cluster maps, K-means (a–c), K-medians (d–f) and hierarchical (g–i) with 4 classes. Letters A–F indicate areas of higher canopy. K-medians (d–f) best captures all the features, but D and F are not contiguous. B and C are captured in all maps, with little difference between algorithms. Panel scores: K-means = 11, K-medians = 15, and Hierarchical = 12.

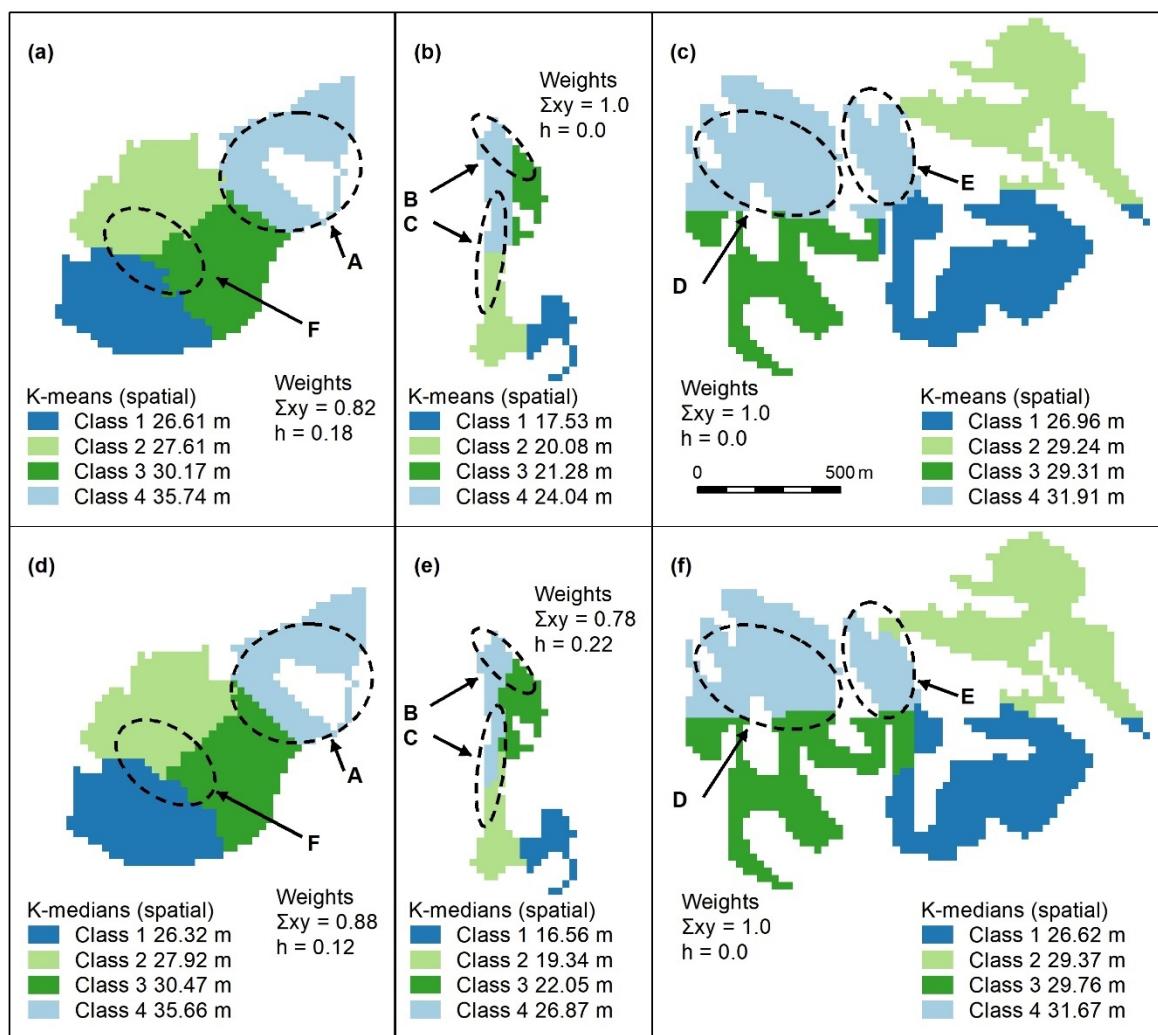


Figure A9. Spatially constrained K-means (a–c) and K-medians (d–f) cluster maps with 4 classes. Letters A–F indicate areas of higher canopy. Again, these algorithms use spatial weights (x, y) that far outweigh those of the object (h = height). As with 3 classes for these algorithms, the only features adequately captured are A (a,d) and E (c,f). Panel scores: K-means = 7 and K-medians = 7.

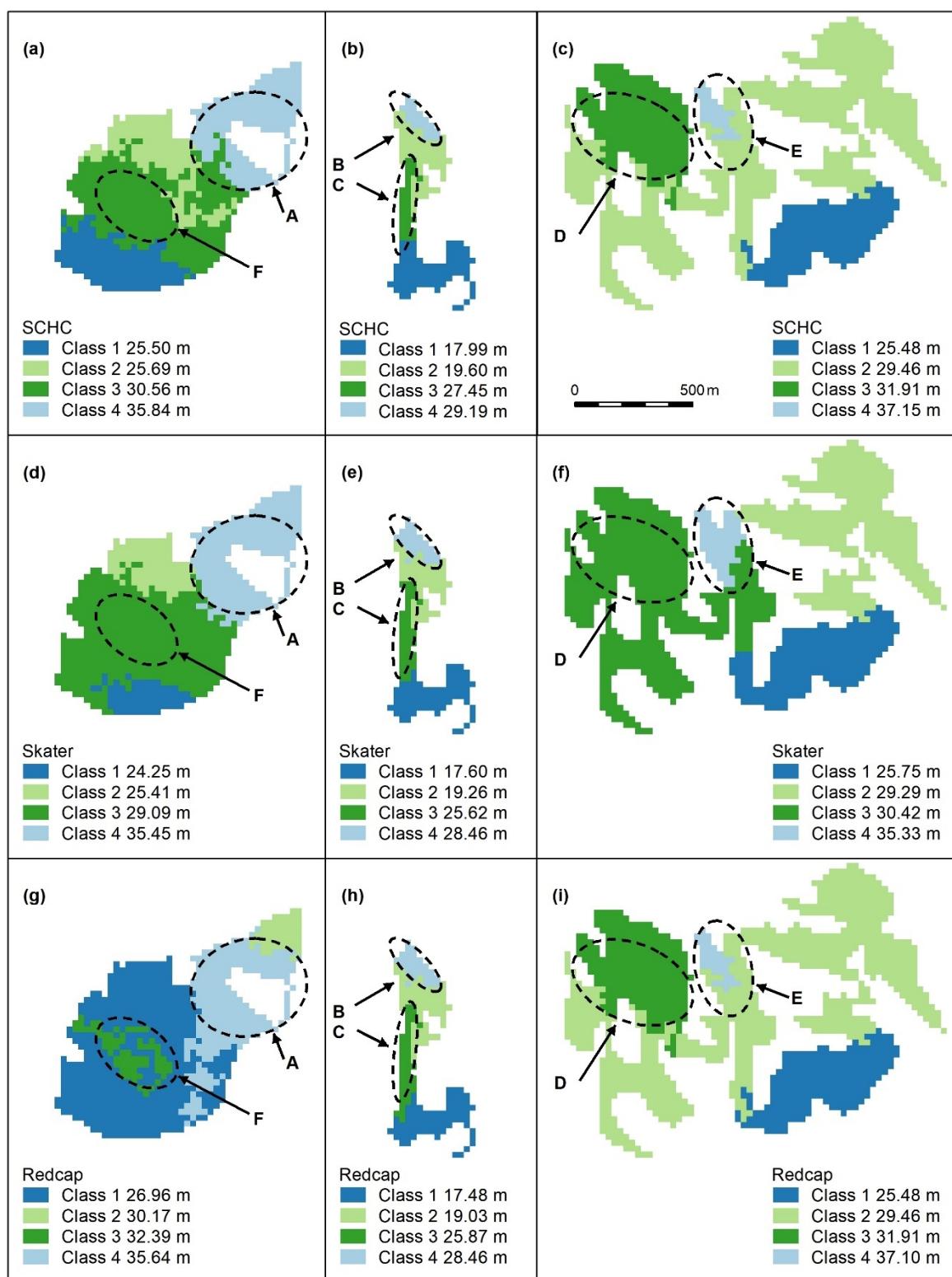


Figure A10. Spatially constrained hierarchical (SCHC, (a–c), Skater (d–f), and Redcap (g–i)) clustering maps with 4 classes. Letters A–F indicate areas of higher canopy. Redcap (g) separates an area of regrowth forest on land previously used for farming (A, dark blue) adjoining a relatively undisturbed national park (light blue). Redcap also recognises F, the unmapped area of *Eucalyptus tereticornis* (dark green), which is slightly lower in height than the national park canopy but higher than the surrounding forest (dark blue). All algorithms portray C (b,e,h) as slightly lower than B, but the areal extent of D is exaggerated by Skater (i). Skater is the algorithm that best captures the areal extent of E (i). Panel scores: Hierarchical clustering = 14, Skater = 15, and Redcap = 16.

Appendix A.3. Multiple Polygon Clustering

Non-contiguous polygons can only be clustered by nonspatial algorithms, i.e., K-means, K-medians, and hierarchical clustering. For the Koala Coast area, we selected two polygon groups, one from intra-subregional Group 6 (212 cells, 13.3 ha) and one from Group 7 (1357 cells, 84.8 ha), i.e., a polygon from each of the high- and low-canopy cohorts. We again produced a canopy height map classified by interquartile height ranges (IQR) as a reference point but with height data from 6007 cells from all seven Koala Coast intra-subregional groups (Figure A11). Some clustering is apparent in the maps—in particular, in Map (b), two areas of higher canopy (D and E), and a section of lower canopy in the southeast. In Map (a), the higher canopy areas (B and C) are differentiated from the low-canopy height (dark blue, IQR 1). Figure A11 thus identifies the same features as Figure A1 but provides a subregional context, i.e., whilst B and C are higher than the remainder of their polygon group, they are significantly lower than D and E and similar areas across the subregion.

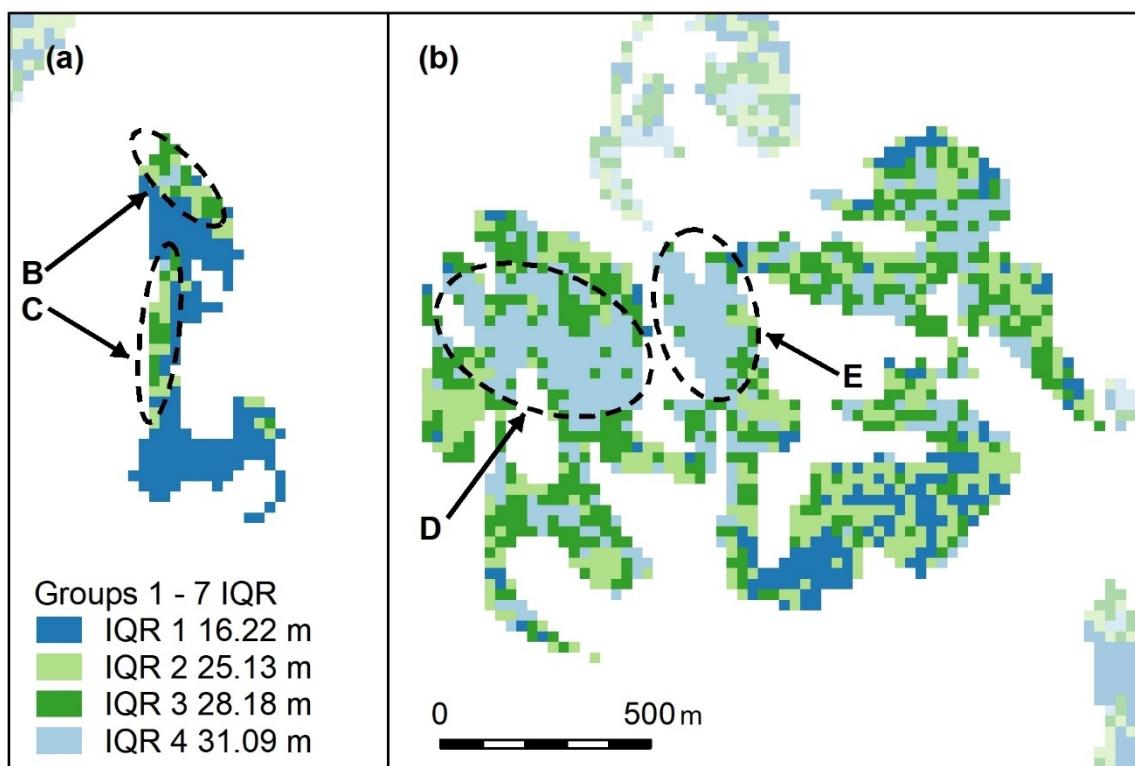


Figure A11. Selected polygon groups from Koala Coast intra-subregional Groups 6 (a, low-canopy cohort) and 7 (b, high-canopy cohort). Letters B–E indicate areas of higher canopy. These maps are similar to Figure A1, with features B–E recognisable, but, because they encompass the entire canopy height range within the Koala Coast, B and C (a) are in middle IQRs (interquartile ranges), D and E (b) are still in the upper quartile. Lower canopy heights (b, dark blue) adjoining B and C are pronounced. Cell colours represent the IQRs of all height data within the Koala Coast, i.e., intra-subregional Groups 1–7. IQR median heights are also shown.

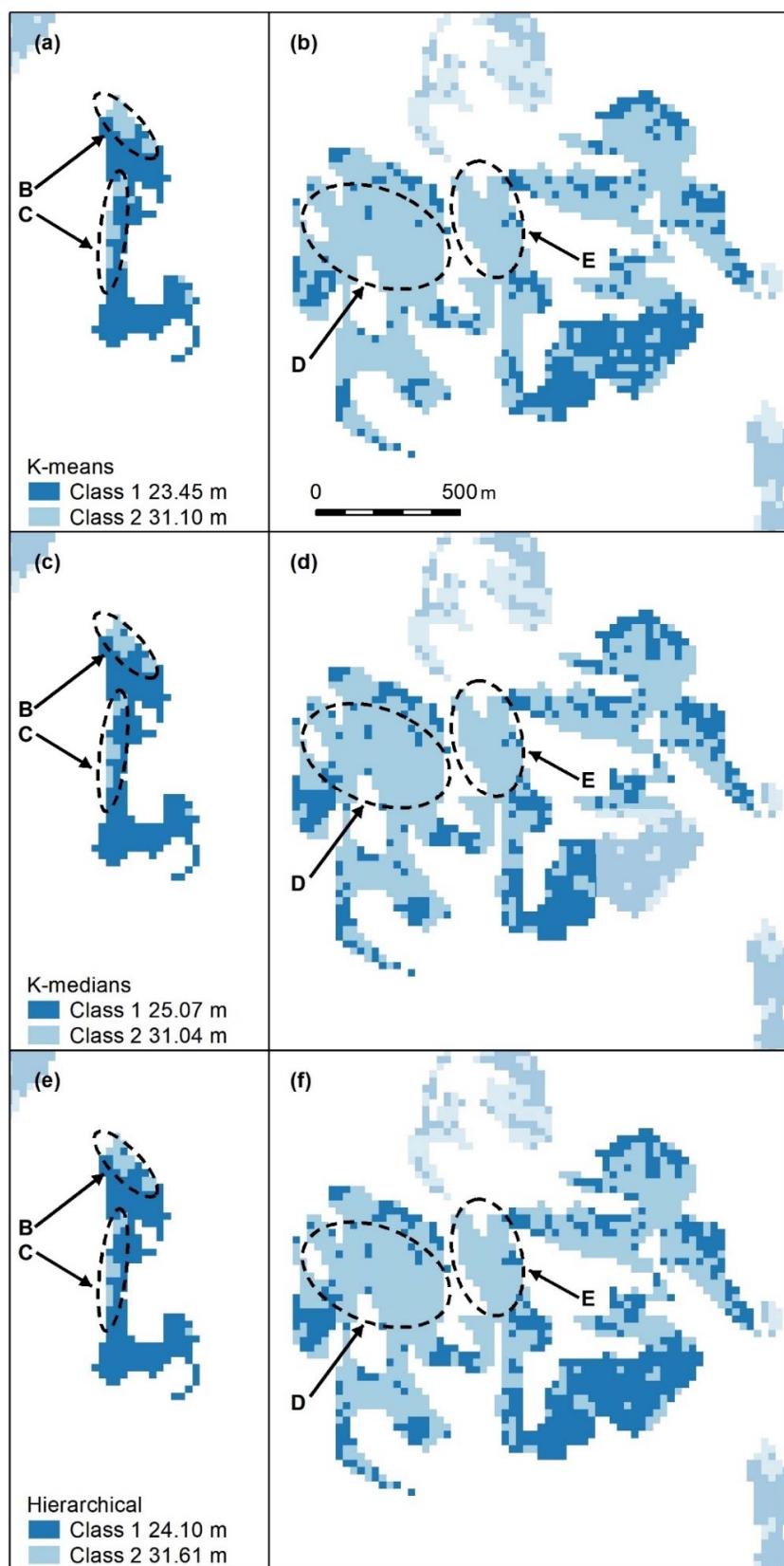


Figure A12. Nonspatial algorithm (K-means, (a,b), K-medians (c,d), hierarchical (e,f)) cluster maps with 2 classes. Letters B–E indicate areas of higher canopy. Features B and C (a,c,e) are captured, but the areal extent of D and E (b,d,f) is not as well-defined as in Figure A11 Panel scores: K-means = 6, K-medians = 6, and Hierarchical = 7.

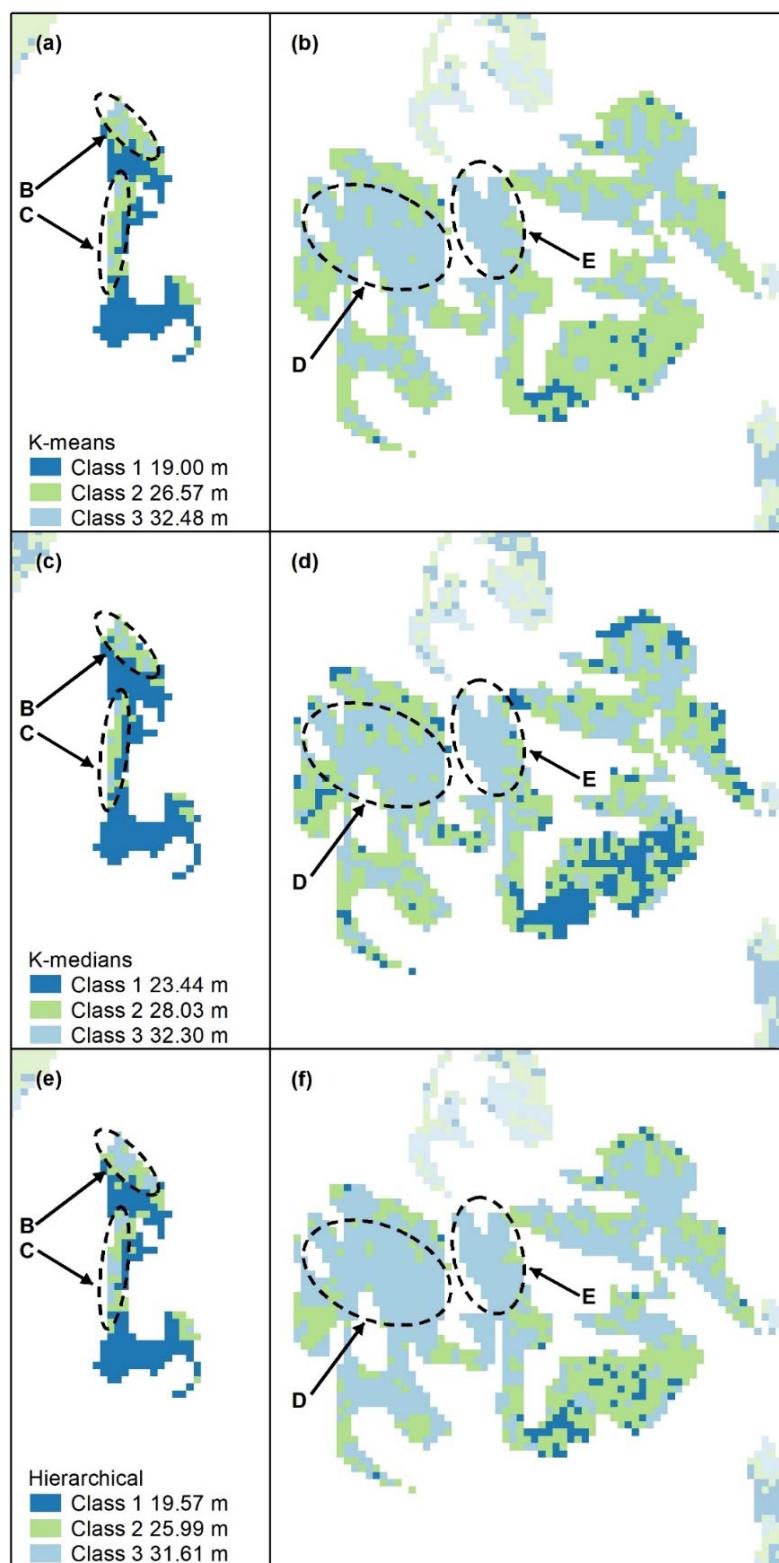


Figure A13. Nonspatial algorithm (K-means, **a,b**; K-medians **c,d**; hierarchical, **e,f**) cluster maps with 3 classes. Letters B–E indicate areas of higher canopy. Features B and C are captured by all algorithms, but D is best captured by K-means (**b**). Panel scores: K-means = 10, K-medians = 10, and Hierarchical = 8.

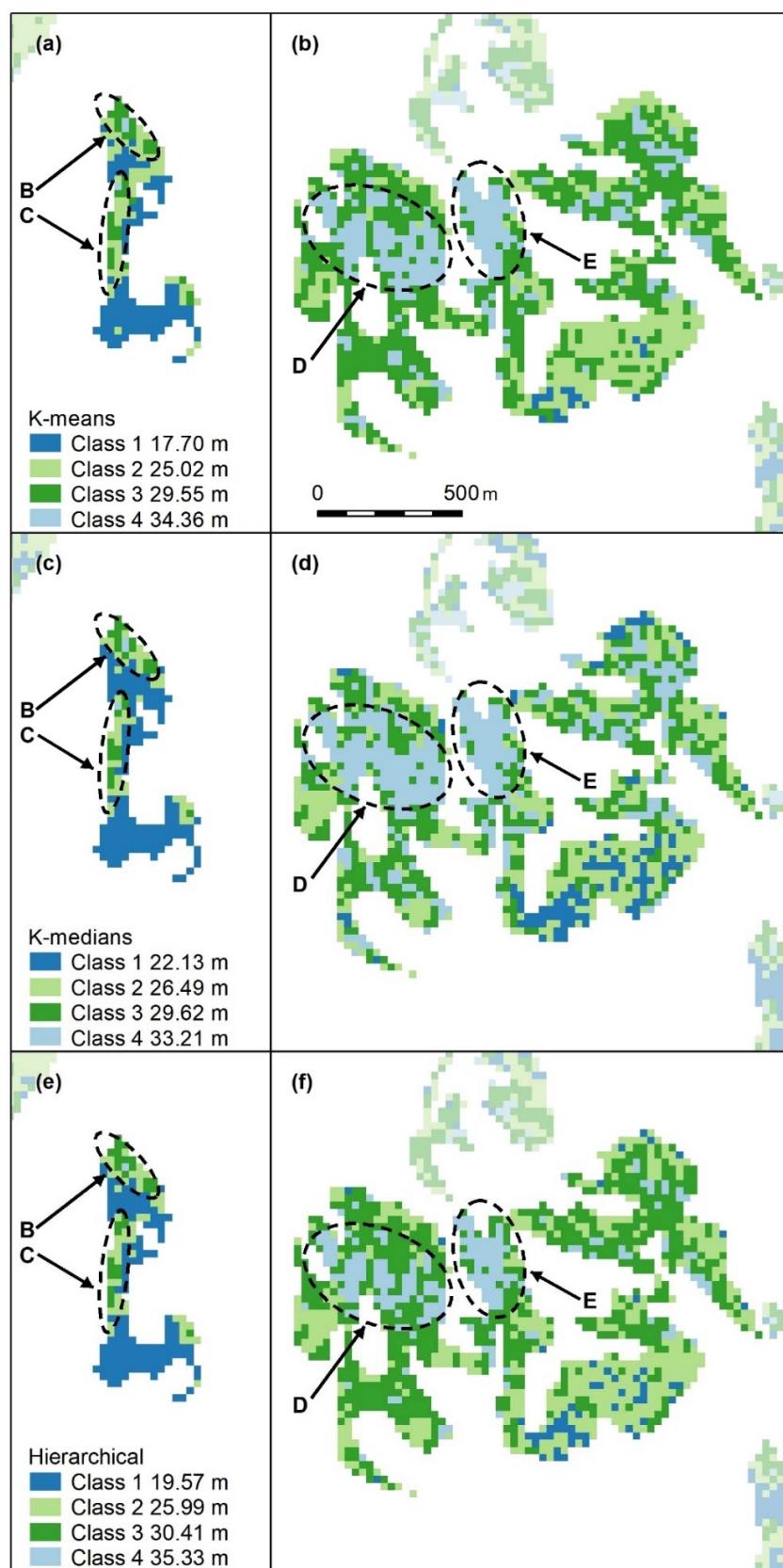


Figure A14. Nonspatial algorithm (K-means, **a,b**; K-medians **c,d**; hierarchical, **e,f**) cluster maps with 4 classes. Letters B–E indicate areas of higher canopy. Features B and C are captured by all algorithms, but D and E are best captured by K-medians (**c,d**). Panel scores: K-means = 6, K-medians = 10, and Hierarchical = 8.

Appendix A.4. Elbow Plots

Our assumption was that two, three, or four cluster class maps would have the greatest utility in terms of user practicality, but we generated maps with 2–10 clusters to allow use of the elbow method (Figures A15–A17) to estimate the optimal number of clusters [64]. The plots indicate that the elbow position is not clearly demarcated but occurs between two and four clusters; after four clusters, additional clusters explain less and less variance. Nonspatial methods show the greatest consistency between algorithms, and differences between spatial algorithms may be due to different polygon group shapes, i.e., compact, linear, and complex.

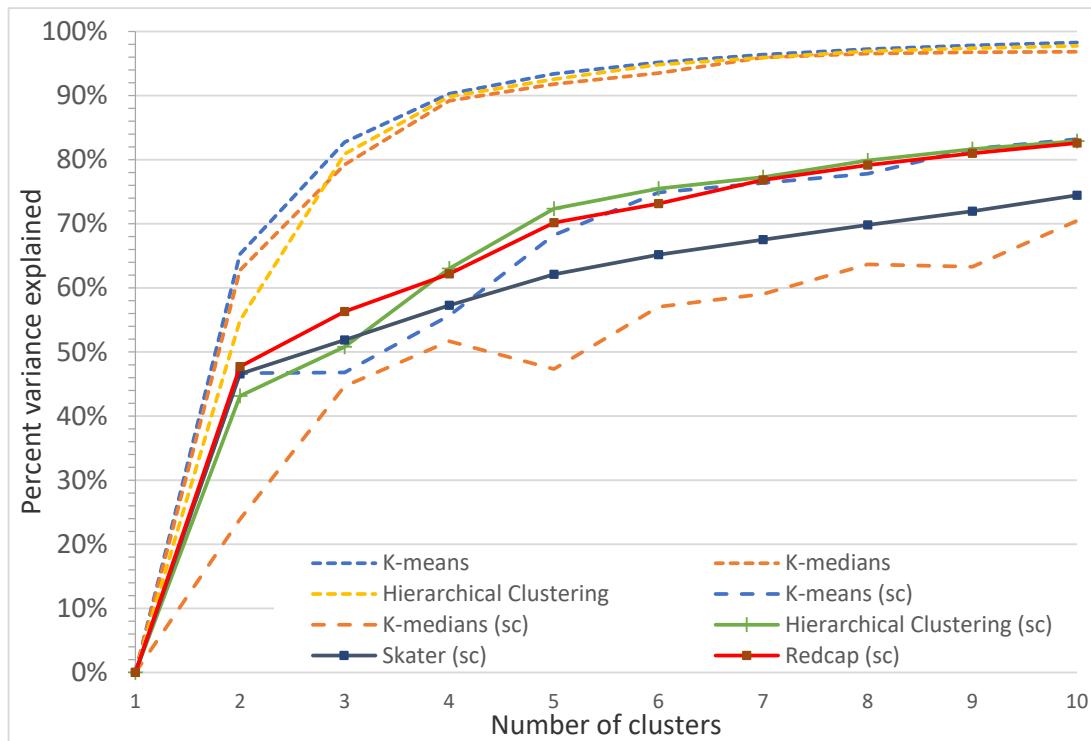


Figure A15. Plot of the percentage of variance explained as a function of the number of clusters generated by eight algorithms for the Sunshine Coast study site. This elbow plot indicates that 2–4 classes is the optimum, e.g., for nonspatial algorithms, 3 classes explain 82.7% (K-means), 79.2% (K-medians), and 80.9% (Hierarchical clustering) of the variance, and 4 classes explain 90.3% (K-means), 89.2% (K-medians), and 90.0% (Hierarchical clustering) of the variance. In comparison, spatially constrained algorithms (sc) explain a maximum of 56% (3 classes) and 63% (4 classes) of the variance.

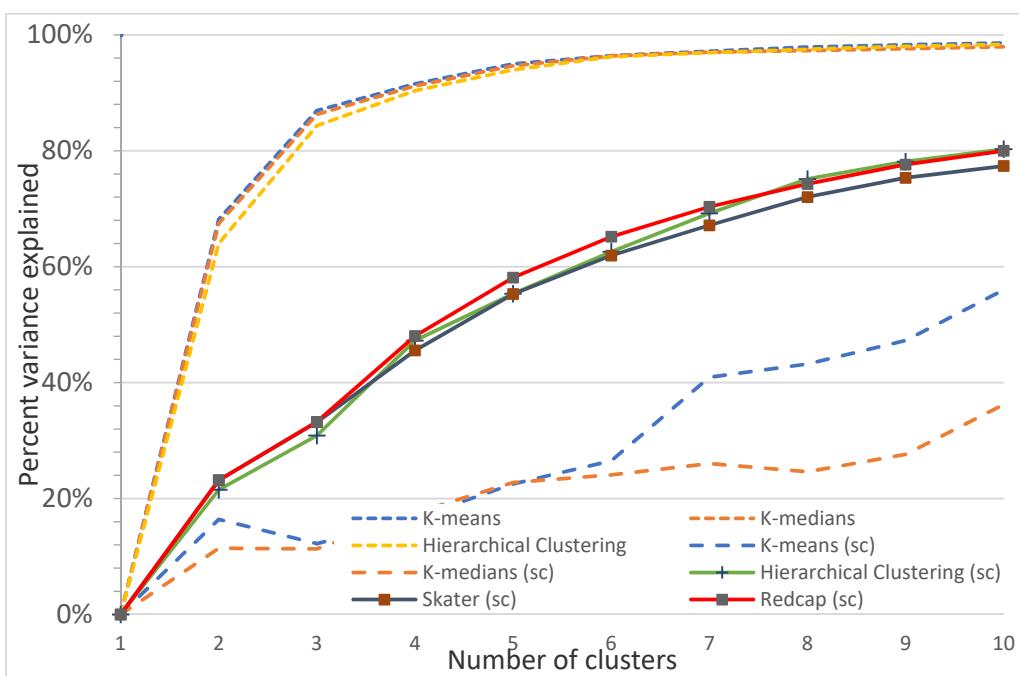


Figure A16. Plot of percentage of the variance explained as a function of the number of clusters generated by eight algorithms for the Koala Coast intra-subregional Group 6 polygon group. This elbow plot indicates that 2 to 3 classes is the optimum, e.g., for nonspatial algorithms, 2 classes explain 86.9% (K-means), 86.3% (K-medians), and 84.3% (Hierarchical clustering) of the variance. For spatial algorithms (sc), there is no clear elbow, with the Redcap algorithm explaining a maximum of 33.2% of the variances for 3 classes.

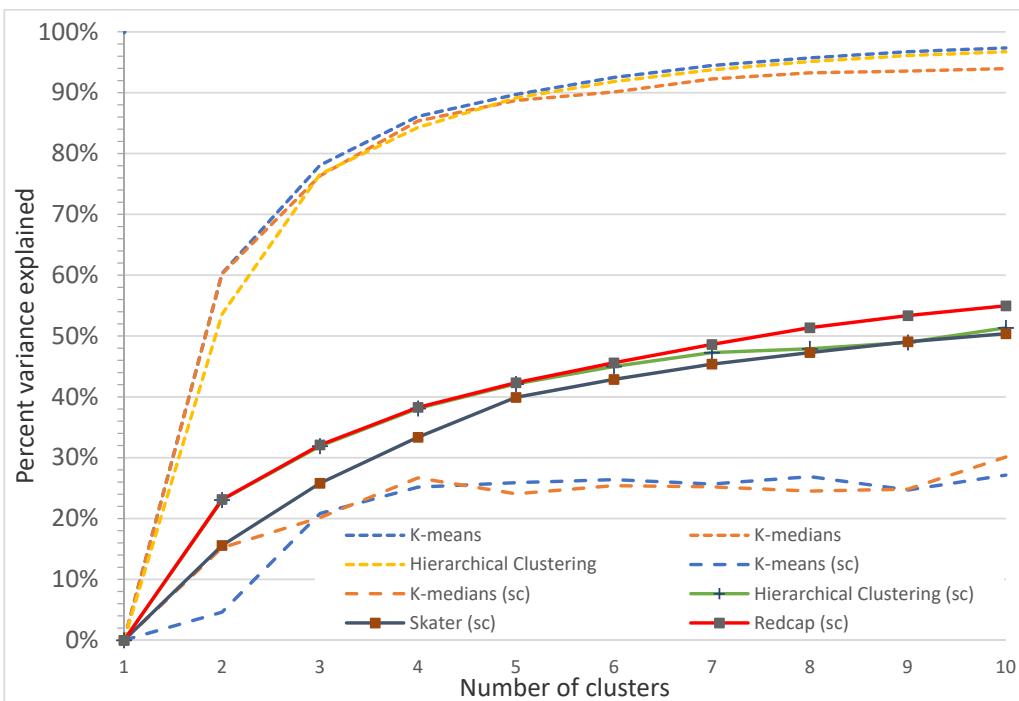


Figure A17. Plot of percentage of the variance explained as a function of the number of clusters generated by eight algorithms for the Koala Coast intra-subregional Group 7 polygon group. This elbow plot indicates that 2–4 classes is the optimum, e.g., for nonspatial algorithms, 3 classes explain 78.1% (K-means), 76.3% (K-medians), and 76.6% (Hierarchical clustering) of the variance, and 4 classes explain 86.1% (K-means), 85.4% (K-medians), and 84.3% (Hierarchical clustering) of the variance. For spatial algorithms (sc), Redcap explains a maximum of 32.1% (3 classes) and 38.3% (4 classes) of the variance.

Appendix A.5. Visual Evaluation

To evaluate the utility of each algorithm, we used three criteria: (a) Does the algorithm identify the six features shown in Figure A1? (b) Is the spatial extent of each feature similar? (c) Is the feature internally contiguous, i.e., are all the cells connected? The maps (Figures A2–A10 and A12–A14) were evaluated by an expert panel composed of one of the authors (DM), an ecologist specialising in forest mapping (DJ), and a GIS analyst (DS), i.e., potential height cluster map users. To provide a “hard” rule and reduce the subjectivity, we scored these criteria using a yes/no binary (Table A1).

Table A1. Procedure used to rate each cluster map. Binary scoring (yes or no) for each identified feature is designed to discourage subjectivity. Letters P, A, and C refer to codes in Tables 12 and 13.

Criterion	Evaluation	Score
Is the feature identified? P = present	yes no	1 0
Is the areal extent similar? A = area	Within $\pm 20\%$ of the IQR * map More than 25% different to the IQR map	1 0
Is the feature internally contiguous? C = contiguity	all cells joined (edges and/or corners) no better than the IQR map	1 0

* IQR Interquartile range (Figures A1 and A11).

The panellists first independently allocated scores, their disagreements were highlighted, and the panellists then re-evaluated their scores. These scores are shown with each of the above figures. The results for all algorithms are shown in Table 12; for nonspatial clustering, K-medians was assigned the highest rank for the two, three, and four classes. In comparison, the spatially constrained algorithms (Hierarchical, Skater, Redcap) performed poorly for the two and three classes but slightly better than the nonspatial algorithms for the four class. No single algorithm was judged to capture all the features, and the panellists had the most disagreement with feature D. The spatially constrained K-means and K-medians performed particularly poorly, identifying only one feature.

For two or more non-contiguous polygon groups, only the nonspatial algorithms (K-means, K-medians, and hierarchical clustering) are suitable clustering methods. For two sample polygon groups (Figure A11) selected from the highest (Group 7) and lowest (Group 6) canopy height cohorts, we used the entire Koala Coast canopy height model to evaluate these algorithms. Comparing multiple polygon groups within a subregion allows for the generation of multiscale clustering maps. Algorithms with two classes did not capture any features. K-medians captured features D and E for three and four classes and was assigned the highest rank (Table 13).

Appendix A.6. Cluster Height Differentiation

The mapped highest-height clusters may be significantly higher than the next-highest height, but these differences need to have practical significance, for example, a 4-m spread would be apparent to field observers and is well outside the error range of the LiDAR sensor. The spread between classes is directly influenced by the number of height classes, for example, for K-means, the average spread varies from 8.13 m (two classes) to 6.42 m (three classes) and 5.48 m (four classes) (Table A2). For the spatial algorithms, there is a trade-off between class separability and location, for example, spatial K-means and K-medians and Redcap have well-defined map clusters, but at the expense of height differentiation, which is less than 4 m.

Table A2. Height differences (metres) between the highest cluster median height and next-lowest cluster median height. Spatially constrained K-means and K-medians have the smallest spread between these height classes, while the other spatial algorithms have a larger spread but are still less than the nonspatial methods. n.a. = algorithm not applicable.

Algorithm	Classes	Figure A1 Map (a)	Figure A1 Map (b)	Figure A1 Map (c)	Figure A11 Maps (a), (b)	Average	Rank
K-means (ns)	2	8.29	10.14	5.86	8.25	8.13	2
K-medians (ns)	2	6.28	8.98	4.68	5.42	6.34	5
Hierarchical Clustering (ns)	2	10.66	10.16	5.96	8.44	8.81	1
K-means (sc)	2	7.50	4.99	1.64	n.a.	4.71	7
K-medians (sc)	2	4.40	5.47	2.55	n.a.	4.14	8
Hierarchical Clustering (s)	2	7.59	9.11	4.79	n.a.	7.16	3
Skater (s)	2	7.48	8.55	2.98	n.a.	6.34	6
Redcap (s)	2	7.12	8.55	4.77	n.a.	6.82	4
K-means (ns)	3	7.61	7.35	4.95	5.76	6.42	3
K-medians (ns)	3	5.92	6.22	3.54	3.93	4.90	6
Hierarchical Clustering (ns)	3	8.69	8.36	5.89	4.66	6.90	2
K-means (sc)	3	7.43	2.14	2.33	n.a.	3.97	8
K-medians (sc)	3	7.62	3.81	1.25	n.a.	4.23	7
Hierarchical Clustering (s)	3	6.73	7.24	7.10	n.a.	7.02	1
Skater (s)	3	7.02	6.68	1.60	n.a.	5.10	5
Redcap (s)	3	2.34	6.75	7.05	n.a.	5.38	4
K-means (ns)	4	6.62	5.56	4.76	4.97	5.48	1
K-medians (ns)	4	6.07	5.63	3.90	3.30	4.72	3
Hierarchical Clustering (ns)	4	6.27	5.63	4.72	4.66	5.32	2
K-means (sc)	4	5.56	2.76	2.61	n.a.	3.64	8
K-medians (sc)	4	5.19	4.82	1.91	n.a.	3.97	6
Hierarchical Clustering (s)	4	5.28	1.74	5.24	n.a.	4.09	5
Skater (s)	4	6.36	2.83	4.91	n.a.	4.70	4
Redcap (s)	4	3.25	2.58	5.19	n.a.	3.67	7

References

- McAlpine, C.A.; Bowen, M.E.; Callaghan, J.G.; Lunney, D.; Rhodes, J.R.; Mitchell, D.L.; Pullar, D.V.; Possingham, H.P. Testing Alternative Models for the Conservation of Koalas in Fragmented Rural-Urban Landscapes. *Austral Ecol.* **2006**, *31*, 529–544. [[CrossRef](#)]
- Department of Environment and Science (DES). *Spatial Modelling for Koalas in South East Queensland v1.0*; Biodiversity Assessment Team, Queensland Herbarium, Science and Technology Division, Department of Environment and Science: Brisbane, QLD, Australia, 2020. Available online: <https://environment.des.qld.gov.au/wildlife/animals/living-with/koalas/mapping/koalamaps#toc-2> (accessed on 3 February 2021).
- Callaghan, J.; McAlpine, C.; Thompson, J.; Mitchell, D.; Bowen, M.; Rhodes, J.; De Jong, C.; Sternberg, R.; Scott, A. Ranking and Mapping Koala Habitat Quality for Conservation Planning on the Basis of Indirect Evidence of Tree-Species Use: A Case Study of Noosa Shire, South-Eastern Queensland. *Wildl. Res.* **2011**, *38*, 89–102. [[CrossRef](#)]
- Ellis, W.A.H.; Melzer, A.; Bercovitch, F.B. Spatiotemporal Dynamics of Habitat Use by Koalas: The Checkerboard Model. *Behav. Ecol. Sociobiol.* **2009**, *63*, 1181–1188. [[CrossRef](#)]
- Crowther, M.S.; Lunney, D.; Lemon, J.; Stalenberg, E.; Wheeler, R.; Madani, G.; Ross, K.A.; Ellis, M. Climate-Mediated Habitat Selection in an Arboreal Folivore. *Ecography* **2014**, *37*, 336–343. [[CrossRef](#)]
- McAlpine, C.A.; Rhodes, J.R.; Bowen, M.E.; Lunney, D.; Callaghan, J.G.; Mitchell, D.L.; Possingham, H.P. Can Multiscale Models of Species’ Distribution Be Generalized from Region to Region? A Case Study of the Koala. *J. Appl. Ecol.* **2008**, *45*, 558–567. [[CrossRef](#)]
- Moore, B.D.; Lawler, I.R.; Wallis, I.R.; Beale, C.M.; Foley, W.J. Palatability Mapping: A Koala’s Eye View of Spatial Variation in Habitat Quality. *Ecology* **2010**, *91*, 3165–3176. [[CrossRef](#)] [[PubMed](#)]

8. Dargan, J.R.; Moriyama, M.; Mella, V.; Lunney, D.; Crowther, M.S. The Challenge for Koala Conservation on Private Land: Koala Habitat Use Varies with Season on a Fragmented Rural Landscape. *Anim. Conserv.* **2019**, *22*, 543–555. [CrossRef]
9. Hindell, M.A.; Lee, A.K. Habitat Use and Tree Preferences of Koalas in a Mixed Eucalypt Forest. *Wildl. Res.* **1987**, *14*, 349–360. [CrossRef]
10. Santamaria, F.; Keatley, M.; Schlagloth, R. Does Size Matter: Tree Use by Translocated Koalas. *Vic. Nat.* **2005**, *122*, 4–13.
11. Thompson, J. The Comparative Ecology and Population Dynamics of Koalas in the Koala Coast Region of South-East Queensland. Ph.D. Thesis, School of Integrative Biology, University of Queensland, Brisbane, QLD, Australia, 2006. [CrossRef]
12. Januchowski, S.R.; McAlpine, C.A.; Callaghan, J.G.; Griffin, C.B.; Bowen, M.; Mitchell, D.; Lunney, D. Identifying Multiscale Habitat Factors Influencing Koala (*Phascolarctos cinereus*) Occurrence and Management in Ballarat, Victoria, Australia. *Ecol. Manag. Restor.* **2008**, *9*, 134–142. [CrossRef]
13. Smith, A.G.; McAlpine, C.; Rhodes, J.R.; Seabrook, L.; Baxter, G.; Lunney, D.; Bradley, A. At What Spatial Scales Does Resource Selection Vary? A Case Study of Koalas in a Semi-Arid Region. *Austral Ecol.* **2013**, *38*, 230–240. [CrossRef]
14. Moore, B.; Foley, W. Tree Use by Koalas in a Chemically Complex Landscape. *Nature* **2005**, *435*, 488–490. [CrossRef] [PubMed]
15. Matthews, A.; Lunney, D.; Gresser, S.; Maitz, W. Tree Use by Koalas (*Phascolarctos cinereus*) after Fire in Remnant Coastal Forest. *Wildl. Res.* **2007**, *34*, 84–93. [CrossRef]
16. Moore, B.D.; Foley, W.J.; Wallis, I.R.; Cowling, A.; Handasyde, K.A. Eucalyptus Foliar Chemistry Explains Selective Feeding by Koalas. *Biol. Lett.* **2005**, *1*, 64–67. [CrossRef]
17. Florence, R.G. *Ecology and Silviculture of Eucalypt Forests*, 1st ed.; CSIRO: Collingwood, VIC, Australia, 1996. [CrossRef]
18. Phillips, S.; Callaghan, J.G. Tree Species Preferences of Koalas (*Phascolarctos cinereus*) in the Campbelltown Area South-West of Sydney, New South Wales. *Wildl. Res.* **2000**, *27*, 509–516. [CrossRef]
19. Gutteridge Haskins & Davey (GHD). *South East Queensland Koala Habitat Assessment and Mapping Project*; GHD: Brisbane, QLD, Australia, 2009. Available online: <https://qldgov.softlinkhosting.com.au/liberty/opac/search.do?queryTerm=GHDReport%3ASouthEastQueenslandkoalaHabitatAssessmentandMappingProjectandmode=ADVANCEDand=undefinedandmodeRadio=KEYWORDandoperator=ANDandactiveMenuItem=false> (accessed on 24 June 2021).
20. Parker, G.G.; Brown, M.J. Forest Canopy Stratification—Is It Useful? *Am. Nat.* **2000**, *155*, 473–484. [CrossRef]
21. Jones, S.; Haywood, A.; Suárez, L.; Wilkes, P.; Woodgate, W.; Soto-Berelov, M.; Mellor, A.; Axelsson, C. Literature Review for Determining Optimal Data Primitives for Characterising Australian Woody Vegetation and Scalable for Landscape-Level Woody Vegetation Feature Generation, 2013. CRC Project Report. Available online: <http://goo.gl/knst81> (accessed on 3 February 2021).
22. Suarez, L.; Jones, S.D.; Haywood, A.; Wilkes, P.; Woodgate, W.; Soto-Berelov, M.; Mellor, A. Woody Vegetation Landscape Feature Generation from Multispectral and LiDAR Data (A CRC 2.07 Woody Attribution Paper). In Proceedings of the International Geoscience and Remote Sensing Symposium (IGARSS), Melbourne, VIC, Australia, 21–26 July 2013; pp. 262–265. [CrossRef]
23. Coops, N.C.; Wulder, M.A.; Culvenor, D.S.; St-Onge, B. Comparison of Forest Attributes Extracted from Fine Spatial Resolution Multispectral and Lidar Data. *Can. J. Remote Sens.* **2004**, *30*, 855–866. [CrossRef]
24. Arroyo, L.A.; Johansen, K.; Armston, J.; Phinn, S. Integration of LiDAR and QuickBird Imagery for Mapping Riparian Biophysical Parameters and Land Cover Types in Australian Tropical Savannas. *For. Ecol. Manag.* **2010**, *259*, 598–606. [CrossRef]
25. Vogeler, J.C.; Cohen, W.B. A Review of the Role of Active Remote Sensing and Data Fusion for Characterizing Forest in Wildlife Habitat Models. *Rev. Teledetección* **2016**, *45*, 1–14. [CrossRef]
26. Carrasco, L.; Giam, X.; Papés, M.; Sheldon, K.S. Metrics of Lidar-Derived 3D Vegetation Structure Reveal Contrasting Effects of Horizontal and Vertical Forest Heterogeneity on Bird Species Richness. *Remote Sens.* **2019**, *11*, 743. [CrossRef]
27. Wallace, L.; Lucieer, A.; Watson, C.W. Evaluating Tree Detection and Segmentation Routines on Very High Resolution UAV LiDAR Data. *IEEE Trans. Geosci. Remote Sens.* **2014**, *52*, 7619–7628. [CrossRef]
28. Geosciences Australia. ELVIS—Elevation and Depth—Foundation Spatial Data Version: 0.7.0. 2020. Available online: <https://elevation.fsdf.org.au/> (accessed on 24 June 2021).
29. Coops, N.C.; Hilker, T.; Wulder, M.A.; St-Onge, B.; Newnham, G.; Siggins, A.; Trofymow, J.A. Estimating Canopy Structure of Douglas-Fir Forest Stands from Discrete-Return LiDAR. *Trees—Struct. Funct.* **2007**, *21*, 295–310. [CrossRef]
30. Alexander, C. Delineating Tree Crowns from Airborne Laser Scanning Point Cloud Data Using Delaunay Triangulation. *Int. J. Remote Sens.* **2015**, *30*, 3843–3848. [CrossRef]
31. Zhen, Z.; Quackenbush, L.J.; Zhang, L. Impact of Tree-Oriented Growth Order in Marker-Controlled Region Growing for Individual Tree Crown Delineation Using Airborne Laser Scanner (ALS) Data. *Remote Sens.* **2014**, *6*, 555–579. [CrossRef]
32. Jeronimo, S.M.; Kane, V.R.; Churchill, D.J.; McGaughey, R.J.; Franklin, J.F. Applying LiDAR Individual Tree Detection to Management of Structurally Diverse Forest Landscapes. *J. For.* **2018**, *116*, 336–346. [CrossRef]
33. Zhang, W.; Cai, S.; Liang, X.; Shao, J.; Hu, R.; Yu, S.; Yan, G. Cloth Simulation-Based Construction of Pit-Free Canopy Height Models from Airborne LiDAR Data. *For. Ecosyst.* **2020**, *7*, 1–14. [CrossRef]
34. Lovell, J.L.; Jupp, D.L.; Culvenor, D.S.; Coops, N.C. Using Airborne and Ground-Based Ranging Lidar to Measure Canopy Structure in Australian Forests. *Can. J. Remote Sens.* **2003**, *29*, 607–622. [CrossRef]
35. Jenkins, R.B. Airborne Laser Scanning for Vegetation Structure Quantification in a South East Australian Scrubby Forest-Woodland. *Austral Ecol.* **2012**, *37*, 44–55. [CrossRef]

36. Scarth, P.; Armston, J.; Goodwin, N. If You Climb Up A Tree, You Must Climb Down The Same Tree. But How High Was It? In Proceedings of the 15th Australasian Remote Sensing and Photogrammetry Conference, Alice Springs, NT, Australia, 13–19 September 2010. Available online: https://www.researchgate.net/publication/309211791>If_You_Climb_Up_A_Tree_You_Must_Climb_Down_The_Same_Tree_But_How_High_Was_It (accessed on 24 June 2021).
37. Liao, Z.; Van Dijk, A.I.; He, B.; Larraondo, P.R.; Scarth, P.F. Woody vegetation cover, height and biomass at 25-m resolution across Australia derived from multiple site, airborne and satellite observations. *Int. J. Appl. Earth Obs. Geoinf.* **2020**, *93*, 102209, ISSN 0303-2434. [CrossRef]
38. Wulder, M.A.; White, J.C.; Bater, C.W.; Coops, N.C.; Hopkinson, C.; Chen, G.; Wulder, M.A. Lidar Plots—A New Large-Area Data Collection Option: Context, Concepts, and Case Study. *Can. J. Remote Sens.* **2014**, *38*, 600–618. [CrossRef]
39. Wilkes, P.; Jones, S.D.; Suarez, L.; Haywood, A.; Soto-Berelov, M.; Mellor, A.; Axelsson, C.; Woodgate, W. Deriving Metrics of Vertical Structure at the Plot Level for Use in Regional Characterisation of S.E. Australian Forests. *CEUR Workshop Proc.* **2012**, *1328*. Available online: <https://researchrepository.rmit.edu.au/esploro/outputs/conferenceProceeding/Deriving-metrics-of-vertical-structure-at-the-plot-level-for-use-in-regional-characterisation-of-SE-Australian-forests/9921862129601341> (accessed on 24 June 2021).
40. Saremi, H.; Kumar, L.; Turner, R.; Stone, C. Airborne LiDAR Derived Canopy Height Model Reveals a Significant Difference in Radiata Pine (*Pinus Radiata*, D. Don) Heights Based on Slope and Aspect of Sites. *Trees-Struct. Funct.* **2014**, *28*, 733–744. [CrossRef]
41. Nguyen, T.H.; Jones, S.D.; Soto-berelov, M.; Haywood, A.; Hislop, S. Monitoring Aboveground Forest Biomass Dynamics over Three Decades Using Landsat Time-Series and Single-Date Inventory Data. *Int. J. Appl. Earth Obs. Geoinf.* **2020**, *84*, 101952. [CrossRef]
42. Lee, A.C.; Lucas, R.M. A LiDAR-Derived Canopy Density Model for Tree Stem and Crown Mapping in Australian Forests. *Remote Sens. Environ.* **2007**, *111*, 493–518. [CrossRef]
43. Wilkes, P.; Jones, S.D.; Suarez, L.; Haywood, A.; Woodgate, W.; Soto-berelov, M.; Mellor, A.; Skidmore, A.K. Understanding the Effects of ALS Pulse Density for Metric Retrieval across Diverse Forest Types. *Photogramm. Eng. Remote Sens.* **2015**, *81*, 625–635. [CrossRef]
44. Scarth, P.; Armston, J.; Lucas, R.; Bunting, P. A Structural Classification of Australian Vegetation Using ICESat/GLAS, ALOS PALSAR, and Landsat Sensor Data. *Remote Sens.* **2019**, *11*, 147. [CrossRef]
45. Office of Environment and Heritage (OEH). *A Review of Koala Tree Use across New South Wales*; Office of Environment and Heritage: Sydney, NSW, Australia, 2018. Available online: <https://www.environment.nsw.gov.au/research-and-publications/publications-search/a-review-of-koala-tree-use-across-new-south-wales> (accessed on 20 November 2020).
46. Department of Agriculture, Water and the Environment (DAWE). Interim Biogeographic Regionalisation for Australia (IBRA). 2012. Available online: <https://www.environment.gov.au/land/nrs/science/ibra#ibra> (accessed on 24 February 2021).
47. Department of Environment and Science (DES). Biodiversity Status of Remnant Regional Ecosystems—Queensland, Version 11. State of Queensland, Department of Environment and Science. 2018. Available online: <http://qldspatial.information.qld.gov.au/catalogue/custom/detail.page?fid=%7B8FDF54D2-654C-4822-8295-1D8E8E772373%7D> (accessed on 21 April 2020).
48. Department of Environment and Science (DES). Regional Ecosystem Description Database (REDD). Version 11. State of Queensland, Department of Environment and Science. 2018. Available online: https://www.qld.gov.au/_data/assets/excel_doc/0021/91551/redd-v11-1-2019.xlsx (accessed on 21 April 2020).
49. Kingston, M.; Turnbill, J.; Hall, P. *Tweed Vegetation Management Strategy 2004*; Ecograph: Limpinwood, NSW, Australia. Available online: https://www.google.com/Url?Sa=T&Rct=j&Q=&Es=1&S=1&Source=Web&Cd=1&Cad=Rja&Uact=8&Ved=2ahukewiawuyv6_Xahxy4nmbhw6qdrwqfjaaegqibhad&Url=https%3a%2f%2fwww.Tweed.Nsw.Gov.Au%2fdownload.aspx%3fpath%3d~%2fdocuments%2fenvironment%2fflora%2520and%2520fauna%2ftsc00219_Tweed_Vegetation_Management_Strategy_Volume_3_Of_3_Appendices.Pdf&Usg=Aovvaw3y4xb16tpotzq6ljkzxg (accessed on 24 June 2021).
50. Fisher, M.; Body, M.; Gill, J. The Vegetation of the Coffs Harbour City Council LGA. A Report to Coffs Harbour City Council. Vegetation Community Profiles—Volume 2. 1996. Available online: <https://www.coffsharbour.nsw.gov.au/environment/Plants-and-Animals/Pages/Class5VegetationMapforCoffsHarbour.aspx> (accessed on 24 June 2021).
51. Ryan, T.S. (Ed.) *Technical Descriptions of Regional Ecosystems of Southeast Queensland*; Queensland Herbarium, Science and Technology Division, Department of Environment and Science: Brisbane, QLD, Australia, 2012. Available online: <https://www.publications.qld.gov.au/dataset/re-technical-descriptions/resource/1771f391-44b9-4d25-8315-f92033610a9b> (accessed on 24 June 2021).
52. ESRI. *ArcMap (Version 10.7.1)*; Environmental Systems Research Institute (ESRI): Redlands, CA, USA, 2019. Available online: <https://www.esri.com/en-us/arcgis/products/arcgis-desktop/overview> (accessed on 12 May 2020).
53. Ryan, T.S. (Ed.) *Draft Technical Description of 12.11.3*; Queensland Herbarium, Science and Technology Division, Department of Environment and Science: Brisbane, QLD, Australia, 2018.
54. Neldner, V.J.; Wilson, B.A.; Dilleward, H.A.; Ryan, T.S.; Butler, D.W.; McDonald, W.J.F.; Addicot, E.P.; Appelman, C.N. Methodology for Surveying and Mapping Regional Ecosystems and Vegetation Communities in Queensland. Version 5.1; Queensland Herbarium, Queensland Department of Environment and Science, Brisbane, Australia. 2019. Available online: <https://www.publications.qld.gov.au/dataset/redd/resource/6dee78ab-c12c-4692-9842-b7257c2511e4> (accessed on 3 March 2020).

55. Isenburg, M. ArcGIS LAStools Toolbox for LiDAR Processing; Rapidlasso GmbH, Gilching, Germany. 2013. Available online: <https://rapidlasso.com/2013/07/02/arcgis-lastools-toolbox-for-lidar-processing/> (accessed on 24 June 2021).
56. Ruxton, G.D.; Beauchamp, G. Time for some a priori thinking about post hoc testing. *Behav. Ecol.* **2008**, *19*, 690–693. [CrossRef]
57. McGaughey, R.J. FUSION/LDV: Software for LIDAR Data Analysis and Visualization, FUSION Version 4.0. *US Dep. Agric.* **2020**, *170*. Available online: <http://scholar.google.com/scholar?hl=en&btnG=Search&q=intitle:FUSION+/+LDV+:+Software+for+LIDAR+Data+Analysis+and+Visualization#0> (accessed on 20 February 2021).
58. Assunção, R.M.; Neves, M.C.; Câmara, G.; Da Costa Freitas, C. Efficient Regionalization Techniques for Socio-Economic Geographical Units Using Minimum Spanning Trees. *Int. J. Geogr. Inf. Sci.* **2006**, *20*, 797–811. [CrossRef]
59. Waller, L.A. Detection of Clustering in Spatial Data. In *The SAGE Handbook of Spatial Analysis*; Sage Publications: London, UK, 2012; pp. 298–320. Available online: <https://uk.sagepub.com/en-gb/eur/the-sage-handbook-of-spatial-analysis/book227940> (accessed on 24 June 2021). [CrossRef]
60. Jacquez, G.M. Spatial Cluster Analysis. *Handb. Geogr. Inf. Sci.* **2008**, *2008*, 395–416. [CrossRef]
61. Anselin, L.; Syabri, I.; Kho, K. GeoDa: An Introduction to Spatial Data Analysis. *Geogr. Anal.* **2006**, *38*, 5–22. [CrossRef]
62. Guo, D. Regionalization with Dynamically Constrained Agglomerative Clustering and Partitioning (REDCAP). *Int. J. Geogr. Inf. Sci.* **2008**, *22*, 801–823. [CrossRef]
63. Department of Environment and Science (DES). Environmental Protection Act 1994—High Value Regrowth. Department of Environment and Science (DES). 2012. Available online: <http://qldspatialinformation.qld.gov.au/catalogue/custom/search.page> (accessed on 3 February 2021).
64. Ketchen, D.J.; Shook, C.J. The Application of Cluster Analysis in Strategic Management Research: An Analysis and Critique. *Strateg. Manag. J.* **1996**, *17*, 441–458. [CrossRef]
65. Pascual, C.; García-Abril, L.G.; García-Montero, L.; Martín-Fernández, S.; Cohen, W.B. Object-Based Semi-Automatic Approach for Forest Structure Characterization Using Lidar Data in Heterogeneous Pinus Sylvestris Stands. *For. Ecol. Manag.* **2008**, *255*, 3677–3685. [CrossRef]
66. Guo, X.; Coops, N.C.; Tompalski, P.; Nielsen, S.E.; Bater, C.W.; John Stadt, J. Regional Mapping of Vegetation Structure for Biodiversity Monitoring Using Airborne Lidar Data. *Ecol. Inform.* **2017**, *38*, 50–61. [CrossRef]
67. Ackers, S.H.; Davis, R.J.; Olsen, K.A.; Dugger, K.M. The Evolution of Mapping Habitat for Northern Spotted Owls (*Strix Occidentalis Caurina*): A Comparison of Photo-Interpreted, Landsat-Based, and Lidar-Based Habitat Maps. *Remote Sens. Environ.* **2015**, *156*, 361–373. [CrossRef]
68. Hagar, J.C.; Yost, A.; Haggerty, P.K. Incorporating LiDAR Metrics into a Structure-Based Habitat Model for a Canopy-Dwelling Species. *Remote Sens. Environ.* **2020**, *236*, 111499. [CrossRef]
69. North, M.P.; Kane, J.T.; Kane, V.R.; Asner, G.P.; Berigan, W.; Churchill, D.J.; Conway, S.; Gutiérrez, R.; Jeronimo, S.; Keane, J.; et al. Cover of Tall Trees Best Predicts California Spotted Owl Habitat. *For. Ecol. Manag.* **2017**, *405*, 166–178. [CrossRef]
70. Ficetola, G.F.; Bonardi, A.; Mücher, C.A.; Gilissen, N.L.; Padoa-Schioppa, E. How Many Predictors in Species Distribution Models at the Landscape Scale? Land Use versus LiDAR-Derived Canopy Height. *Int. J. Geogr. Inf. Sci.* **2014**, *28*, 1723–1739. [CrossRef]
71. Seavy, N.E.; Viers, J.H.; Wood, J.K.; Seavy, N.E.; Viers, J.H.; Julian, K.W. Riparian Bird Response to Vegetation Structure: A Multiscale Analysis Using LiDAR Measurements of Canopy Height. *Ecol. Appl.* **2016**, *19*, 1848–1857. [CrossRef]
72. Eyre, T.J. Regional Habitat Selection of Large Gliding Possums at Forest Stand and Landscape Scales in Southern Queensland, Australia. I. Greater Glider (*Petauroides volans*). *For. Ecol. Manag.* **2006**, *235*, 270–282. [CrossRef]
73. Melville, G.; Stone, C.; Turner, R. Application of LiDAR Data to Maximise the Efficiency of Inventory Plots in Softwood Plantations. *N. Z. J. For. Sci.* **2015**, *45*, 9. [CrossRef]
74. Dubayah, R.; Blair, J.B.; Goetz, S.; Fatoyinbo, L.; Hansen, M.; Healey, S.; Hofton, M.; Hurtt, G.; Kellner, J.; Luthcke, S.; et al. The Global Ecosystem Dynamics Investigation: High-Resolution Laser Ranging of the Earth’s Forests and Topography. *Sci. Remote Sens.* **2020**, *1*, 100002. [CrossRef]