



Article NH₄⁺-N and Low Ratios of NH₄⁺-N/NO₃⁻-N Promote the Remediation Efficiency of Salix linearistipularis in Cdand Pb-Contaminated Soil

Dongliu Di ^{1,2,†}, Shaokun Wang ^{1,3,†}, Guangcai Chen ², Qian Wang ¹, Jingwei Zhang ¹, Xiaoyun Niu ^{1,*} and Dazhuang Huang ^{1,*}

- ¹ College of Landscape Architecture and Tourism, Hebei Agricultural University, Baoding 071001, China; dongliudi@caf.ac.cn (D.D.); wangshaokun@caf.ac.cn (S.W.); wangqian_hebau@163.com (Q.W.); jwz_637192@163.com (J.Z.)
- ² Research Institute of Subtropical Forestry, Chinese Academy of Forestry, Hangzhou 311400, China; guangcaichen@sohu.com
- ³ Institute of Wetland Research, Chinese Academy of Forestry, Beijing 100091, China
- * Correspondence: nxy850101@hebau.edu.cn (X.N.); huangdazhuang@126.com (D.H.); Tel.: +86-0312-7520207 (D.H.)
- [†] These authors contributed equally to this work.

Abstract: (1) Background: the utilization of fast-growing trees for phytoremediation in heavymetal-contaminated soil is increasingly recognized as an effective remediation method. Nitrogen (N) fertilizer enhances plant tolerance to heavy metals, yet the impact of various N levels and ammonium $(NH_4^+-N)/nitrate (NO_3^--N)$ ratios on the remediation of cadmium (Cd) and lead (Pb) by trees remains unclear. (2) Methods: the efficiency of Salix linearistipularis in remediating Cd- and Pb-contaminated soil was investigated using a pot experiment with three N levels (60, 120, 200 kg hm⁻¹ year⁻¹) and five NH₄⁺-N/NO₃⁻-N ratios (6/0, 4/2, 3/3, 2/4, 0/6) employed, resulting in 16 treatments including a control. (3) Results: the levels and ratios of NH_4^+ -N/NO₃⁻-N significantly affected the Cd and Pb uptake by S. linearistipularis. The highest increases in Cd and Pb in S. linearistipularis were observed for the N_{120} -6/0 treatment, which increased by 104.36% and 95.23%, respectively. In addition, in the N120-6/0 treatment, the stem and leaf bioconcentration factors of Cd were significantly enhanced by 28.66% and 40.11%, respectively. Structural equation modeling revealed that the uptake of Cd and Pb was predominantly influenced by plant traits (biomass and root traits) rather than soil properties. (4) Conclusions: Our findings highlight the potential of the NH₄⁺-N/NO₃⁻-N ratio to regulate plant traits, thereby improving the phytoremediation efficiency of heavy-metal-contaminated soil.

Keywords: willow; nitrogen; heavy metals; phytoremediation; plant trait; soil contamination

1. Introduction

Heavy metal (HM) contamination in soil has occurred globally due to intensive anthropogenic activities and emissions, such as cadmium (Cd) and lead (Pb) [1,2]. Cd and Pb are non-degradable toxicants that can accumulate in plants, cause chlorosis, affect plant photosynthesis [3], reduce the absorption and translocation of essential elements [4], and pose a threat to human health via the food chain [5]. Therefore, the remediation of Cd-and Pb-contaminated soil is an essential and urgent issue concerning environmental safety, food security, and human health [6,7].

Phytoremediation has garnered recognition for its environmentally friendly nature, aesthetic appeal, efficiency, cost-effectiveness, and high level of acceptance [8,9]. Numerous hyperaccumulators that accumulate high levels of HMs have been screened, such as *Sedum alfredii*, *Thlaspi caerulescens*, and *Amaranthus Hypochondriacus* [10,11]. However, the herbaceous traits of hyperaccumulators, which result in relatively low biomass and high costs for



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Copyright: © 2024 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). annual plants and harvests, restrict their widespread application [12]. Conversely, there has been considerable interest in the use of woody species for phytoremediation in recent years. Among them, willow (genus: *Salix*) was most extensively studied [13]. *Salix* has excellent traits such as rapid growth, ease of cultivation, hypoxia tolerance, a deep root system, large biomass, and high metal accumulation [14–16]. However, the growth of woody plants is mainly limited by their low nitrogen (N) content in urban contaminated soil, which restricts plant growth, especially root development and remediation efficiency [17]. As a component of amino acids, enzymes, and hormones, N plays an important role in stress resistance [3]. In addition, sufficient N increases transpiration and water flow within plants, resulting in a high uptake of mineral elements, including HMs [18–21]. Ammonium (NH₄⁺-N) and nitrate (NO₃⁻-N) are the main forms of N taken up by plants [19]. Therefore, regulating N levels represents an effective strategy to improve the efficiency of HM remediation.

Multiple studies have emphasized species-specific responses to different N sources in plants [22–24]. Research has predominantly focused on herbaceous species, with notable findings such as Cheng's study demonstrating increased Cd accumulation in the herbaceous plant Solanum nigrum when exposed to NH4+-N [25]. Conversely, increased Cd uptake was observed in the cereal crop Oryza sativa and the metal hyperaccumulator Thlaspi caerulescens under NO_3^--N conditions [26,27]. The combined application of NH_4^+-N and NO_3^--N has been shown to enhance plant growth by influencing various nitrogen-related physiological processes [28]. Recent decades have seen a growing consensus that the NH_4^+-N/NO_3^--N ratio significantly impacts the biomass production and phytoremediation capabilities of various species, including agricultural crops like wheat [18,29], Brassica napus, Pakchoi, *Ricinus communis*, and *Panicum maximum* [24,30–32]. Wu's research indicated that higher NH4⁺-N ratios reduce Cd levels in rice shoots and roots, enhance biomass, and diminish total rice Cd accumulation at NH_4^+ - N/NO_3^- -N ratios of 2/1 and 1/0 [33]. Additionally, Tanzania guinea grass grown at NH4⁺-N/NO3⁻-N ratios of 1/1 demonstrated the increased uptake, transport, and accumulation of Cd [24]. However, the specific effects of this ratio on the growth and phytoremediation efficiency of willow, a fast-growing woody plant, have not been adequately explored, suggesting a potential difference in response compared to the commonly studied herbaceous and agricultural species.

Willow demonstrates considerable proficiency in accumulating Cd and Pb [34,35]. This characteristic can be enhanced through suitable amendments, including fertilizer applications. The effects of N fertilizer on HMs in soil and plant accumulation have been extensively studied [21]. It was found that N fertilizer application increased metal accumulation in plants by altering soil properties and promoting plant growth [36]. Moreover, it was established that N regulates soil pH and HMs' bioavailability, thereby altering remediation efficiency [37]. Previous studies have also shown that Cd and Pb are more mobile in acidic soil than in neutral or alkaline soil [38]. Arnaud showed that the application of the N fertilizer enhanced the accumulation of HMs by increasing the biomass of *N. caulescens* [39]. Additionally, Sterckeman demonstrated that roots are the main driver of HM accumulation, with root mass being positively correlated with plant metal uptake [40]. However, the underlying mechanisms by which N levels and ratios regulate Cd and Pb accumulation in *S. linearistipularis* remain unclear.

In this study, we investigated the effects of the levels and ratios of NH_4^+-N/NO_3^--N on the remediation efficiency of *S. linearistipularis* in Cd- and Pb-contaminated soil. The objective of this study was twofold: (1) to investigate the effects of different NH_4^+-N/NO_3^--N levels and ratios on the growth of *S. linearistipularis* and its accumulation of HMs in Cd- and Pb-contaminated soil, and (2) to explore the underlying mechanism by which the N fertilizer application influences the phytoremediation efficiency of willow. The outcomes of this study are anticipated to provide valuable insights into enhancing fertilization methods in order to optimize the phytoremediation efficiency of woody plants in soil contaminated with HMs.

2. Materials and Methods

2.1. Plant Material and Growth Conditions

One-year-old *S. linearistipularis* branches with uniform sizes were collected from the experimental station of Hebei Agricultural University ($38^{\circ}45'21''$ N, $115^{\circ}24'31''$ E). In June 2020, branches measuring 15 cm in length were planted in pots filled with 20 kg of soil. The dimensions of the pots were 34 cm in diameter and 25 cm in height. The soil was classified as meadow cinnamon soil (surface soil, depth 0–20 cm) and spiked with CdCl₂ and Pb(NO₃)₂ for more than 5 years prior to our experiment [41]. The soil Cd and Pb concentrations were 7 mg kg⁻¹ and 210 mg kg⁻¹, respectively. The other soil properties are listed in Table A1. Soil water content was maintained at 60% of the maximum field capacity throughout the experiment.

2.2. Experimental Design

Four N levels (0, 60, 120, and 200 kg hm⁻² year⁻¹) and five NH₄⁺-N/NO₃⁻-N ratios (6/0, 4/2, 3/3, 2/4, and 0/6) were used in this study (Table 1). Nitrogen application levels were established based on the standard nitrogen application rate of 120 kg·hm⁻²·year⁻¹ for one-year seedlings [42]. To minimize the influence of other experimental conditions on plant growth, NH₄ Cl was chosen as the source of NH₄⁺-N, and NaNO₃ was chosen as the source of NO₃⁻-N, considering the relatively minor effects of Cl⁻ and Na⁺ on plant growth. All reagents were of analytical grade.

Table 1. Fertilization treatments.

Treatmonte	Nitrogen Levels (kg⋅hm ⁻² ⋅year ⁻¹)						
freatments	0	60	120	200			
$\begin{array}{c} NH_4^{+}-N/NO_3^{-}-N=6/0\\ NH_4^{+}-N/NO_3^{-}-N=4/2\\ NH_4^{+}-N/NO_3^{-}-N=3/3\\ NH_4^{+}-N/NO_3^{-}-N=2/4\\ NH_4^{+}-N/NO_3^{-}-N=0/6\\ \end{array}$	Control	N ₆₀ -6/0 N ₆₀ -4/2 N ₆₀ -3/3 N ₆₀ -2/4 N ₆₀ -0/6	N ₁₂₀ -6/0 N ₁₂₀ -4/2 N ₁₂₀ -3/3 N ₁₂₀ -2/4 N ₁₂₀ -0/6	N ₂₀₀ -6/0 N ₂₀₀ -4/2 N ₂₀₀ -3/3 N ₂₀₀ -2/4 N ₂₀₀ -0/6			

The experiment was designed with 16 distinct treatments to comprehensively assess the impact of various fertilization treatments on the growth of *Salix*, and each treatment consisted of 4 replicates. One-year-old branches approximately 0.6 cm in diameter were cut into 15 cm in length and were then transplanted into the pot. The nitrogen fertilizer was applied in four equal portions each year, at intervals of 10 days. The fertilizer, once dissolved in water, was evenly sprayed onto the surface of the pots using a small sprayer, followed immediately by watering. To ensure uniform conditions throughout the growing season, all pots were randomly arranged and received identical care, including watering, soil loosening, weeding, and were harvested 150 days after planting.

2.3. Experimental Analysis

2.3.1. Gas Exchange Measurements

Prior to harvest, six leaves of each plant were selected for the measurement of gas exchange parameters such as the photosynthetic rate (Pn, μ mol CO₂·m⁻² s⁻¹), transpiration rate (E, mmol H₂O·m⁻² s⁻¹), stomatal conductance (Gs, mol H₂O·m⁻² s⁻¹), and intercellular CO₂ concentration (CO₂ int, μ mol CO₂·mol⁻¹). They were determined using a photosynthesis analysis system (Li-6400, Lincoln, NE, USA) with 1200 μ mol m⁻² s⁻¹ of photons and 400 μ mol mol⁻¹ of CO₂ concentration [43]. All measurements were performed outdoors between 9:00 and 11:00 a.m. on a sunny day with relatively stable environmental conditions.

2.3.2. Root Morphological Parameters and Biomass Determination

Upon harvest, each plant was carefully removed from the growth substrate, rinsed with tap and deionized water, and separated into roots, stems, and leaves. Roots were collected for the measurement of morphological parameters, including the average root diameter, surface area, total volume, and total root length, using an automatic root scanner (LA-S Plant image analyzer system). Finally, each sample (roots, stems and leaves) was oven-dried at 65 °C to constant weight for dry biomass determination.

2.3.3. Rhizosphere Soil Analyses

Rhizosphere soil was collected via shaking during the harvesting of plants [43]. The soil was subsequently air-dried, pulverized, and sieved using 20- and 100-mesh sieves. Soil pH, soil organic matter (SOM), cation exchange capacity (CEC), soil urease activity (Ure), and soil catalase activity (S-CAT) were determined using soil sieved through a 20-mesh sieve. Soil pH was measured at a volume ratio of 1/2.5 (soil/water) using a pH meter (PHSJ-3F). SOM was determined by oxidation with potassium dichromate and titration with a ferrous sulfate reagent [44]. The CEC was determined using the barium chloride-sulfuric acid exchange method. Urea content was determined via phenol hypochlorite colorimetry using a spectrophotometer (AA-680, Shimadzu, Kyoto, Japan) at 578 nm [45]. The potassium permanganate titration method was used for S-CAT [46]. Ammonium and nitrate concentrations were determined using a COMIN kit [47]. The total concentrations of Cd and Pb were determined by acid digestion (HCL, HNO₃, and HClO₄) [48], and chemical fractions were extracted with the sequential extraction method (BCR) described by Rauret using soil sieved through a 100-mesh sieve [49]. Atomic absorption spectrophotometry (AA-680; SHIMADZU, Kyoto, Japan) was used to determine Cd and Pb concentrations.

2.3.4. Plant Tissue Analysis

Dried plant samples (roots, stems, and leaves) were powdered and sieved through a 60-mesh sieve before digestion with concentrated $HNO_3/HCLO_4$ (v/v, 4:1) at 170 °C for approximately 3.5 h until the solution became clear [50]. The solution was filtered through a 0.45 µm membrane, diluted to 50 mL, and analyzed for Cd and Pb concentrations using flame atomic absorption spectrometry (AA-680, Shimadzu, Kyoto, Japan).

2.4. Statistical Analysis

Data preprocessing included tests for normality, homogeneity of variances, and independence to meet the assumptions required for two-way ANOVA. Following the successful validation of these assumptions, two-way ANOVA was performed at a significance level of 0.05 using IBM SPSS version 20.0. and the analysis primarily focused on examining the effects of varying levels and ratios on N fertilizer as independent variables. Subsequently, post hoc comparisons were carried out employing Duncan's multiple range tests to ascertain the differences in means. A principal component analysis (PCA) was performed applying the "FactoMineR" and "factoextra" packages in R (4.2.1). Additionally, SEMs were developed to investigate the direct and indirect influences of nitrogen fertilizer levels and ratios on the total bioaccumulation of Cd or Pb using the "lavaan" and "semPlot" packages in R (4.2.1).

The bioconcentration factor (BCF) was computed to evaluate the plant's capability to accumulate HMs from its surroundings [51]. Furthermore, the translocation factor (TF) was calculated to assess the efficiency of metal translocation from the roots to the aerial parts of the plant [52]. This provided a comprehensive understanding of the plant's response to different nitrogen treatments in terms of heavy metal accumulation and translocation.

 $BCF = \frac{Cd \text{ or } Pb \text{ concentration in the harvested tissues}}{Cd \text{ or } Pb \text{ concentration in the soil}}$ $TF = \frac{Cd \text{ or } Pb \text{ concentration in the aboveground tissues}}{Cd \text{ or } Pb \text{ concentration in the roots}}$

3. Results

3.1. Chemical Properties of Rhizosphere Soil

The application of N fertilizer significantly decreased soil pH, and this effect was significantly associated with both N levels and ratios (p < 0.01; Table 2). The impact of a high N level or balanced NH₄⁺-N and NO₃⁻-N ratios on pH was pronounced, with pH reaching a minimum of 7.66 for the N₂₀₀-3/3 treatment. However, N fertilizer increased the concentrations of NH₄⁺-N and NO₃⁻-N, as well as the SOM content and urea activity (Table A2). N levels, N ratios, and their interactions had a significant impact on the concentrations of Cd and Pb in rhizosphere soil (p < 0.01; Table 2). The N₂₀₀-0/6 treatment revealed the highest Cd and Pb concentrations of 6.09 mg kg⁻¹ and 201.92 mg kg⁻¹, respectively. In contrast, the N₆₀-6/0 treatment had the lowest Cd concentration (2.65 mg kg⁻¹). The N₁₂₀-6/0 and N₁₂₀-3/3 treatments showed the lowest Pb concentrations, measuring 103.24 mg kg⁻¹ and 102.26 mg kg⁻¹, respectively. Additionally, the NH₄⁺-N-dominated treatment and the N₂₀₀ group increased the acid-soluble concentrations of Cd and Pb. The N₁₂₀-6/0 treatment recorded the highest Cd concentration, measuring 0.83 mg kg⁻¹, whereas the N₂₀₀-6/0 treatment exhibited the highest Pb concentration at 27.62 mg kg⁻¹.

Nitrogen Levels (kg hm ⁻² year ⁻¹)	Ratios	рН	Total Cd (mg kg ⁻¹)	Total Pb (mg kg ⁻¹)	Acid-Soluble Cd (mg kg ⁻¹)	Acid-Soluble Pb (mg kg ⁻¹)
0	Control	7.96 ± 0.05 $^{\rm a}$	$4.09\pm0.56~^{ef}$	$184.17 \pm 11.04 \ ^{\rm abc}$	0.77 ± 0.04 $^{\rm a}$	$20.78\pm2.22~^{\rm de}$
	6/0	$7.88\pm0.06~^{\rm bc}$	$2.65 \pm 0.17~^{g}$	$132.24\pm19.28~^{\mathrm{fgh}}$	0.81 ± 0.13 $^{\rm a}$	$22.82 \pm 2.53 \ ^{bcd}$
	4/2	7.92 ± 0.03 $^{ m ab}$	2.77 ± 0.15 ^g	$121.61 \pm 23.95~^{ m ghi}$	0.81 ± 0.05 ^a	$18.91 \pm 1.61 \ ^{ m e}$
60	3/3	7.78 ± 0.06 ^{de}	2.71 ± 0.24 g	$159.21 \pm 10.82 \; ^{ m de}$	0.82 ± 0.1 a	$22.43\pm3.51^{\rm\ bcde}$
	2/4	$7.77\pm0.05~^{\rm e}$	3.06 ± 0.42 g	$136.74\pm9.19~^{\rm efg}$	0.37 ± 0.03 ^c	22.24 ± 1.87 ^{cde}
	0/6	$7.84\pm0.01~^{\rm cd}$	$4.29\pm0.46~^{\rm de}$	$110.34\pm18.48~^{\rm hi}$	$0.41\pm0.09~^{\rm c}$	$19.42\pm2.48~^{\rm de}$
	6/0	$7.88\pm0.02~^{\rm bc}$	$4.74\pm0.44~^{\rm cd}$	$103.24 \pm 12.79^{\;i}$	$0.83\pm0.05~^{\rm a}$	$19.53\pm2.23~^{\rm de}$
	4/2	$7.78\pm0.08~\mathrm{de}$	$5.14\pm0.16~\mathrm{^{bc}}$	$125.9\pm10.48~^{\rm ghi}$	0.81 ± 0.08 a	$19.22\pm1.19~^{ m de}$
120	3/3	$7.76\pm0.03~^{\rm e}$	4.54 ± 0.45 ^{cde}	$102.26 \pm 18.05\ ^{\rm i}$	$0.41\pm0.03~^{ m c}$	$19.85\pm1.15~\mathrm{de}$
	2/4	7.80 ± 0.03 ^{de}	$4.57\pm0.48~^{ m cde}$	$135.74\pm12.25~^{\mathrm{efg}}$	0.48 ± 0.04 ^{bc}	19.11 ± 1.64 $^{ m e}$
	0/6	$7.80\pm0.03~^{\rm de}$	$3.64\pm0.35~^{\rm f}$	$155.35\pm15~^{def}$	$0.55\pm0.08^{\text{ b}}$	$14.32\pm2.78~^{\rm f}$
	6/0	$7.78\pm0.04~^{\rm de}$	$4.33\pm0.46~^{\rm de}$	$132.97 \pm 14.31 \ {}^{\rm fgh}$	$0.5\pm0.07~^{ m bc}$	$27.62\pm0.69~^{\rm a}$
	4/2	$7.74\pm0.04~\mathrm{^{ef}}$	5.55 ± 0.48 $^{ m ab}$	$194.67 \pm 12.03 \ ^{ m ab}$	0.78 ± 0.07 ^a	25.9 ± 0.59 $^{ m ab}$
200	3/3	$7.66\pm0.03~^{\rm g}$	5.09 ± 0.33 ^{bc}	$173.13 \pm 15.84 \ ^{ m bcd}$	0.77 ± 0.05 $^{\rm a}$	$25.81\pm3.66~^{\rm ab}$
	2/4	$7.69\pm0.04~^{\mathrm{fg}}$	5.62 ± 0.17 $^{ m ab}$	$168.91 \pm 19.3 \ { m cd}$	0.8 ± 0.07 ^a	$25.37\pm2.87~^{ m abc}$
	0/6	7.73 ± 0.05 ef	6.09 ± 0.62 a	$201.92\pm16.63~^{a}$	0.73 ± 0.17 $^{\rm a}$	$25.18\pm0.77~^{\rm abc}$
L		**	**	**	**	**
R		**	**	**	**	**
$L \times R$		ns	**	**	**	ns

Note: All data were analyzed using two-way analysis of variance (ANOVA) followed by Duncan's test for comparison with the controls. N levels are denoted as "L", the N ratio as "R", and the interactive effect of N levels and ratios as "L × R". Different superscript lowercase letters indicate significant differences among the treatments at p < 0.05 in specific tissues. Data represent means \pm SD (n = 4), "ns" denotes non-significance and double asterisks (**) represent significant differences at p < 0.01.

3.2. Salix linearistipularis Traits under Different N Treatments

The N level, N ratio, and their interactions significantly influenced the plant traits of *S. linearistipularis*, including photosynthesis, root morphology, and biomass (p < 0.01, Figures 1 and 2). Specifically, when NH₄⁺-N was added alone, it significantly enhanced the photosynthetic and transpiration rates of *S. linearistipularis*. Furthermore, the NH₄⁺-N treatment markedly increased the root surface area, volume, and length, resulting in a substantial enhancement of root growth (Figures 1 and A2). Significant increases in total biomass were noted, with particular prominence in the N₆₀-6/0 and N₁₂₀-6/0 treatments, where the values reached 33.45 and 36.42 g plant⁻¹, respectively (Figure 2D). Compared

to the control, the biomass increased by 38.91% and 51.25%, respectively. Conversely, the introduction of NO₃⁻-N led to divergent outcomes, resulting in diminished growth (Figure A3), a significant reduction in photosynthetic activity (Figure A1), and a decrease in *S. linearistipularis* biomass (Figure 2). Additionally, higher N levels resulted in an evident decrease in the photosynthetic rate of plants, thereby limiting root expansion and compromising overall plant vitality. Notably, leaves in the N₂₀₀ treatment, characterized by a high N content, exhibited diminished size and yellowing and even displayed signs of toxicity (Figure A3).



Figure 1. Root traits of *S. linearistipularis* under different N treatments. Note: All data were analyzed using two-way analysis of variance (ANOVA) followed by Duncan's test for comparison with the controls. N levels are denoted as "L", the N ratio as "R", and the interactive effect of N levels and ratios as "L × R". Different superscript lowercase letters indicate significant differences among the treatments at p < 0.05 in specific tissues. Data represent means \pm SD (n = 4), "ns" denotes non-significance and double asterisks (**) represent significant differences at p < 0.01.

3.3. Distribution of Cd and Pb in Salix linearistipularis

The concentrations of Cd and Pb within *S. linearistipularis* were significantly influenced by the interactions between N levels and ratios (p < 0.01, Figure 3). In the N₁₂₀ and N₂₀₀ groups, the predominance of NH₄⁺-N (NH₄⁺-N/NO₃⁻-N = 6/0 and 4/2) resulted in a substantial increase in the Cd concentration in the roots, with a range of 60.31%–83.81% (Figure 3A). The highest Cd concentrations in the stems and leaves were observed in the N₁₂₀-6/0 treatment, reaching 26.43 and 64.35 mg kg⁻¹ (Figure 3B,C), representing increases of 28.60% and 40.03%, respectively, compared to the control. Nevertheless, a decrease in the Cd concentration was observed in the stems and leaves, which was primarily induced by NO₃⁻-N treatments (NH₄⁺-N/NO₃⁻-N = 3/3, 2/4, and 0/6). The N₂₀₀-3/3 treatment

displayed the lowest Cd concentration in stems (10.47 mg kg⁻¹), whereas the N₁₂₀-0/6 treatment demonstrated the lowest Cd concentration in leaves (20.23 mg kg⁻¹). The Pb concentration in the roots significantly increased under the NO₃⁻-N and high N treatments (Figure 3D). The N₁₂₀-0/6 treatment recorded the highest concentration at 14.85 mg kg⁻¹, followed by the N₆₀-2/4 treatment at 14.24 mg kg⁻¹, representing increases of 103.59% and 95.15% compared to the control, respectively. Moreover, N fertilization increased Pb concentrations in both the stem and leaf tissues (Figure 3E,F). The N₆₀-2/4 treatment exhibited the highest stem Pb concentration of 10.45 mg kg⁻¹, whereas the N₂₀₀-3/3 treatment displayed the highest leaf Pb concentration of 12.10 mg kg⁻¹.



Figure 2. Biomass of *S. linearistipularis*. Note: All data were analyzed using two-way analysis of variance (ANOVA) followed by Duncan's test for comparison with the controls. N levels are denoted as "L", the N ratio as "R", and the interactive effect of N levels and ratios as "L × R". Different superscript lowercase letters indicate significant differences among the treatments at p < 0.05 in specific tissues. Data represent means \pm SD (n = 4), a single asterisk (*) represents significant differences at p < 0.05, and double asterisks (**) represent significant differences at p < 0.01.

The N level, N ratio, and their interactions significantly influenced the BCF of Cd in *S. linearistipularis* (p < 0.001; Table 3). The use of a N fertilizer significantly increased the BCF of Cd in root tissues, with the N₂₀₀-4/2 treatment attaining the highest value of 3.69, followed by the N₁₂₀-6/0 treatment at 3.33. Furthermore, the N₁₂₀-6/0 treatment significantly increased stem BCF (4.13) and leaf BCF (10.06), representing increases of 28.66% and 40.11%, respectively. Conversely, stem BCF was significantly reduced under treatments such as N₁₂₀-3/3, 0/6, and N₂₀₀-4/2, 3/3, 2/4, 0/6. Similarly, leaf BCF exhibited significant reductions under various treatments, including N₆₀-0/6, N₁₂₀-3/3, 2/4, 0/6, and N₂₀₀-6/0, 4/2, 3/3, 0/6, with reductions ranging from 23.36% to 48.91% for stem BCF and from 7.02% to 55.99% for leaf BCF compared to the control.



Figure 3. Pb and Cd concentration in *S. linearistipularis* under different treatments. Note: All data were analyzed using two-way analysis of variance (ANOVA) followed by Duncan's test for comparison with the controls. N levels are denoted as "L", the N ratio as "R", and the interactive effect of N levels and ratios as "L × R". Different superscript lowercase letters indicate significant differences among the treatments at *p* < 0.05 in specific tissues. Data represent means \pm SD (*n* = 4), "ns" denotes non-significance, and double asterisks (**) represent significant differences at *p* < 0.01.

Nitrogen Levels	Dation	Cd			Pb		
$(kg hm^{-2} year^{-1})$		BCF of Roots	BCF of Stems	BCF of Leaves	BCF of Roots	BCF of Stems	BCF of Leaves
0	Control	$2.01\pm0.19~^{\rm de}$	$3.21\pm0.4~^{bc}$	$7.18\pm0.49~^{bcd}$	$0.04\pm0.01~^{\rm e}$	$0.04\pm0.01~^{\rm b}$	$0.03\pm0.01~^{fg}$
60	6/0 4/2 3/3 2/4 0/6	$\begin{array}{c} 2.19 \pm 0.47 \ ^{de} \\ 1.87 \pm 0.28 \ ^{de} \\ 2.14 \pm 0.35 \ ^{de} \\ 1.99 \pm 0.26 \ ^{de} \\ 2.22 \pm 0.47 \ ^{de} \end{array}$	$\begin{array}{c} 2.84 \pm 0.37 \ \mbox{bcd} \\ 3.15 \pm 0.6 \ \mbox{bcd} \\ 2.52 \pm 0.24 \ \mbox{cdef} \\ 3.52 \pm 0.25 \ \mbox{ab} \\ 2.52 \pm 0.18 \ \mbox{cdef} \end{array}$	$\begin{array}{c} 8.37 \pm 1.28 \ ^{\rm b} \\ 8.55 \pm 0.69 \ ^{\rm b} \\ 5.85 \pm 1.1 \ ^{\rm de} \\ 7.9 \pm 1.51 \ ^{\rm bc} \\ 4.01 \pm 0.79 \ ^{\rm fgh} \end{array}$	$\begin{array}{c} 0.03 \pm 0.01 \ ^{e} \\ 0.04 \pm 0.01 \ ^{e} \\ 0.03 \pm 0.01 \ ^{e} \\ 0.07 \pm 0.01 \ ^{ab} \\ 0.07 \pm 0.01 \ ^{abc} \end{array}$	$\begin{array}{c} 0.04\pm 0.01 \ ^{b} \\ 0.05\pm 0.01 \ ^{ab} \\ 0.05\pm 0.01 \ ^{ab} \\ 0.05\pm 0.01 \ ^{a} \\ 0.05\pm 0.01 \ ^{a} \\ 0.04\pm 0.01 \ ^{b} \end{array}$	$\begin{array}{c} 0.03 \pm 0.01 \; ^{defg} \\ 0.04 \pm 0.01 \; ^{cde} \\ 0.04 \pm 0.01 \; ^{cdef} \\ 0.04 \pm 0.01 \; ^{cdef} \\ 0.04 \pm 0.01 \; ^{efg} \end{array}$
120	6/0 4/2 3/3 2/4 0/6	$\begin{array}{c} 3.33 \pm 0.52 \ ^{a} \\ 3.22 \pm 0.39 \ ^{abc} \\ 2.17 \pm 0.35 \ ^{de} \\ 2.57 \pm 0.34 \ ^{cd} \\ 1.79 \pm 0.49 \ ^{e} \end{array}$	$\begin{array}{c} 4.13 \pm 0.85\ ^{a} \\ 2.84 \pm 0.16\ ^{bcde} \\ 2.39 \pm 0.09\ ^{ef} \\ 2.51 \pm 0.15\ ^{cdef} \\ 1.9 \pm 0.47\ ^{fg} \end{array}$	$\begin{array}{c} 10.06 \pm 0.44 \ ^{a} \\ 6.72 \pm 0.2 \ ^{cd} \\ 4.77 \pm 0.35 \ ^{efg} \\ 3.31 \pm 0.41 \ ^{gh} \\ 3.16 \pm 0.61 \ ^{h} \end{array}$	$\begin{array}{c} 0.05 \pm 0 \; ^{cde} \\ 0.06 \pm 0.01 \; ^{bcd} \\ 0.06 \pm 0.01 \; ^{bcd} \\ 0.05 \pm 0.01 \; ^{de} \\ 0.07 \pm 0.02 \; ^{ab} \end{array}$	$\begin{array}{c} 0.05 \pm 0.01 \ ^{a} \\ 0.04 \pm 0.01 \ ^{b} \end{array}$	$\begin{array}{c} 0.03 \pm 0.01 \ ^{cdefg} \\ 0.03 \pm 0.01 \ ^{g} \\ 0.03 \pm 0.01 \ ^{efg} \\ 0.04 \pm 0.01 \ ^{bcde} \\ 0.04 \pm 0.01 \ ^{bc} \end{array}$

Nitrogen Levels (kg hm $^{-2}$ year $^{-1}$) Ratios	D (1	Cd			Pb			
	Katios	BCF of Roots	BCF of Stems	BCF of Leaves	BCF of Roots	BCF of Stems	BCF of Leaves	
	6/0	3.3 ± 0.38 ^{ab}	3.3 ± 0.38 ^b	5.24 ± 0.24 ef	$0.07\pm0~^{ab}$	0.04 ± 0.01 ^b	$0.04\pm0.01~^{\rm bc}$	
	4/2	3.69 ± 0.66 a	2.46 ± 0.11 def	$3.78\pm0.22~^{\rm fgh}$	0.08 ± 0.01 a	0.04 ± 0.01 ^b	$0.04\pm0.01~^{\mathrm{bcd}}$	
200	3/3	2.52 ± 0.33 ^{cde}	$1.64 \pm 0.22~{ m g}$	$3.98\pm0.45~^{\rm fgh}$	0.08 ± 0.01 $^{\rm a}$	0.05 ± 0.01 ^b	0.06 ± 0.01 $^{\rm a}$	
	2/4	2.12 ± 0.22 de	$1.87\pm0.22~^{\mathrm{fg}}$	$7.51 \pm 1.02 \ ^{ m bc}$	$0.05\pm0.01~^{ m cde}$	0.04 ± 0.01 ^b	0.05 ± 0.01 ^b	
	0/6	$2.6\pm0.36~^{bcd}$	$2.19\pm0.21~^{efg}$	$4.77\pm1.06~^{efg}$	$0.06\pm0.01~^{abcd}$	0.04 ± 0.01 $^{\rm b}$	$0.04\pm0.01~^{bcd}$	
L		**	**	**	**	ns	**	
R		**	**	**	**	ns	ns	
$L \times R$		**	**	**	**	**	**	

Table 3. Cont.

Note: All data were analyzed using two-way analysis of variance (ANOVA) followed by Duncan's test for comparison with the controls. N levels are denoted as "L", the N ratio as "R", and the interactive effect of N levels and ratios as "L × R". Different superscript lowercase letters indicate significant differences among the treatments at p < 0.05 in specific tissues. Data represent means \pm SD (n = 4), "ns" denotes non-significantce, and double asterisks (**) represent significant differences at p < 0.01.

The application of N fertilizer increased the BCF of Pb within root tissues, resulting in increases ranging from 25% to 100%. Notably, the N₂₀₀-4/2 and N₂₀₀-3/3 treatments exhibited the highest BCF values. The increase in stem BCF of Pb was relatively modest, with only the N₆₀-2/4 and N₁₂₀-6/0 treatments showing a 25% increase compared to the control. In terms of the leaf BCF of Pb, significant increases were observed under the N₆₀-4/2, N₁₂₀-2/4, N₁₂₀-0/6, and N₂₀₀ treatments. Among these, the N₂₀₀-3/3 treatment demonstrated the most substantial increase, with a 100% increase compared with the control.

3.4. Cd and Pb Accumulation in Salix linearistipularis

The N level, N ratio, and their interactions significantly increased the accumulation of Cd and Pb in *S. linearistipularis* tissues (p < 0.01, Figure 4). Most Cd and Pb concentrations accumulated in the stems, which were substantially higher than those in the roots and leaves. The total Cd accumulation in *S. linearistipularis* significantly increased in the N₁₂₀-6/0 treatment, reaching 0.11, 0.66, 0.36, and 1.13 mg plant⁻¹ in roots, stems, leaves, and the entire plant, respectively. These values represented increases of 174.67%, 93.05%, 110.06%, and 104.36%, respectively, compared to the control. This was followed by N₆₀-6/0 treatment, with Cd accumulation rates of 0.06, 0.44, 0.23, and 0.74 mg plant⁻¹ in the roots, stems, leaves, and entire plant, respectively, corresponding to increases of 49.44%, 29.03%, 37.83%, and 33.25%. Meanwhile, Cd accumulation was lower in the N₆₀-0/6, N₁₂₀-4/2, 0/6, and N₂₀₀-4/2, 3/3, 2/4, and 0/6 treatments, ranging from 0.16 to 0.39 mg plant⁻¹.

The total Pb accumulation significantly increased in the N_{60} -6/0, 4/2, 2/4, and N_{120} -6/0, 3/3 treatments compared to the control (Figure 4H). In the N_{120} -6/0 treatment, maximum Pb accumulation was observed, with accumulation rates of 0.05, 0.26, 0.04, and 0.35 mg plant⁻¹ in the roots, stems, leaves, and entire plant, respectively, exhibiting increases of 124.79%, 92.45%, 80.76%, and 95.23%, respectively, compared to the control. The N_{60} -2/4 treatment ranked second in Pb accumulation, with accumulation rates of 0.06, 0.21, 0.03, and 0.29 mg plant⁻¹ in the roots, stems, leaves, and whole plant, respectively, signifying increases of 42.31%, 55.01%, 27.14%, and 63.92% compared to the control. However, a decrease in Pb accumulation was observed in the N_{120} -4/2 and N_{200} -4/2, 3/3, and 0/6 treatments, with measurements of 0.11, 0.12, and 0.13 mg of the plant⁻¹, respectively.

3.5. The Direct and Indirect Mechanisms Regulating the Accumulation of Cd and Pb

An SEM was constructed that consisted of the level and ratio of N, soil properties, plant traits, and the total accumulation of Cd and Pb in *S. linearistipularis* (Figure 5). Plant traits were characterized using the first principal component of the PCA. The Cd interpretation rate was 0.92, whereas that of Pb was 0.87. The total accumulation of Cd and Pb was mainly regulated by plant traits. The level and ratio of N indirectly affected plant traits by affecting NH₄⁺-N. In addition, the effect of the N fertilizer ratio on Pb accumulation was more



significant in *S. linearistipularis*. The total absorption of Cd and Pb by *S. linearistipularis* increased with increasing NH₄⁺-N/NO₃⁻-N ratios.

Figure 4. Pb and Cd accumulation in *S. linearistipularis* under different treatments. Note: All data were analyzed using two-way analysis of variance (ANOVA) followed by Duncan's test for comparison with the controls. N levels are denoted as "L", the N ratio as "R", and the interactive effect of N levels and ratios as "L × R". Different superscript lowercase letters indicate significant differences among the treatments at *p* < 0.05 in specific tissues. Data represent means \pm SD (*n* = 4), a single asterisk (*) represents significant differences at *p* < 0.05, and double asterisks (**) represent significant differences at *p* < 0.01.



Figure 5. A structural equation model was employed to illustrate the direct and indirect effects of N application, soil properties (NH₄⁺-N, Ure, and pH), and plant traits on the total accumulation of Cd (**A**) and Pb (**B**) in *S. linearistipularis*, along with the effect value of the total Cd accumulation (**C**) and Pb accumulation (**D**). Solid black and red arrows indicate significantly positive and negative correlations, respectively, while dotted arrows indicate insignificant correlations. The numbers adjacent to the arrows represent standardized path coefficients (λ). Significance levels are denoted as * *p* < 0.05, ** *p* < 0.01, *** *p* < 0.001.

4. Discussion

4.1. Nitrogen Reduced Soil pH and Improved Soil Quality

Appropriate N application can increase soil fertility and improve soil quality. Our study demonstrated that the application of N a fertilizer to meadow cinnamon soil enhanced urea activity while concurrently increasing the contents of NH_4^+ -N, NO_3^- -N, and SOM. This improvement in soil quality can be attributed to the promotion of nutrient cycling due to the addition of N [53]. This observation is consistent with most findings in the field [18]. However, excessive N affects plant health and poses environmental risks, necessitating the control of N fertilizer application [54]. In this study, N application resulted in soil acidification within the rhizosphere (Table 2). However, some studies have shown that NO_3^- -N addition can increase the soil's pH [55]. This variation may be related to soil characteristics because the application of inorganic N in alkaline soil can result in the accumulation of phenolic acids [56]. Furthermore, willows release organic acids during their growth cycle [57], which contribute to soil pH reduction as N fertilization accelerates protein synthesis in soil contaminated with HMs, resulting in the secretion of various organic acids by plants [58]. At a constant N application ratio (NH_4^+ -N/ NO_3^- -N = 3/3),

the pH decreased significantly (Table 1). This was attributed to the nitrification of NH_4^+-N and the absorption of NH_4^+-N by plants, which resulted in the release of H^+ . Moreover, $NH_4^+-N/NO_3^--N = 3/3$ further acidifies the soil through the leaching of acid-soluble alkali ions [59,60].

The bioavailability of HMs is linked to their chemical form [61]. Acid-soluble HMs were identified as the primary forms absorbed and utilized by plants [62]. The NH_4^+ -N treatment increased the acid-soluble Cd concentrations while decreasing the total concentrations of Cd and Pb in the rhizosphere soil (Table 2). In alkaline soil, a decreased pH facilitates the dissolution of acid-soluble Cd and Pb [63]. Additionally, complexes readily form between Cl⁻ and Cd²⁺, thereby promoting an increase in acid-soluble Cd [64]. This ultimately leads to an increase in soluble metal ions, with more metals being taken up by plants, resulting in a decrease in soil Cd and Pb.

4.2. NH4⁺-N Promotes Plant Growth

Photosynthesis and plant traits are important indicators of the tolerance or resistance of plants to various stressors [32]. N fertilization significantly affected the photosynthesis, root morphology, and growth of S. linearistipularis. High N levels and separate $NO_3^{-}-N$ strongly inhibited S. linearistipularis growth, whereas this inhibitory effect was mitigated by the addition of NH_4^+ -N at low and medium N levels (Figures 1 and 2). Comparable outcomes have been documented in other studies of Solanum nigrum, Carpobrotus rossii, *Oryza sativa*, and *Carpobrotus rossii* [25,33,60]. The acceleration of NH_4^+ -N can be attributed to increased cell division and subsequent root branching, which consequently enhances root vitality [33]. Additionally, NH_4^+ -N is directly absorbed by plants as a protein, reducing energy and photosynthetic product consumption and thereby providing more energy for plant growth [28]. Additionally, high N fertilizer application can result in metabolic imbalance, oxidative stress, lipid peroxidation, the disruption of root respiration, cell osmotic balance, and plant hormone homeostasis [65]. High N levels led to a decrease in pH and an increase in the soluble HM content (Table 2), thereby increasing toxicity to S. *linearistipularis* [66]. Additionally, the low biomass of NO_3^{-} -N might have been due to the high BCF of HMs in the roots (Table 2), which generated large amounts of HMs, inhibited nutrient absorption, and impeded root growth, thereby causing adverse damage to plant traits [32].

A low NH₄⁺-N/NO₃⁻-N ratio also promoted *S. linearistipularis* growth, such as NH₄⁺-N/NO₃⁻-N = 2/4. Carbon and N fluxes were stimulated by applying NH₄⁺-N and NO₃⁻-N concurrently and reducing the carbon loss caused by respiration, which can directly contribute to biomass accumulation [23,28]. It has also been observed that stimulated root development promotes enzyme activity, productivity, and fruit quality in peppers through a low NH₄⁺-N/NO₃⁻-N ratio (1/3) [67]. We believe that *S. linearistipularis* prefers NH₄⁺-N in isolation and low NH₄⁺-N/NO₃⁻-N ratios. In summary, at N levels of 60 and 120 kg hm⁻² year⁻¹, applying NH₄⁺-N or low NH₄⁺-N/NO₃⁻-N can alleviate HM stress and promote root growth, thereby affecting the biomass accumulation and uptake of HMs [3,68].

4.3. NH₄⁺-N Increases Cd and Pb Accumulation

N application significantly affected Cd and Pb concentrations in *S. linearistipularis*. NH₄⁺-N increased Cd concentrations and their accumulation in plant tissues, especially in the N₆₀-6/0 and N₁₂₀-6/0 treatments, where *S. linearistipularis* showed the highest Cd accumulation. The total absorption of Cd by *S. linearistipularis* was 0.6 mg pot⁻¹, which is three times and twenty times higher than that of the herbaceous plants *Amaranthus Hypochondriacus* and *Pentas lanceolata*, respectively [11,69]. Furthermore, the TF of HMs ranged from 1.6 to 3.56, approaching that of hyperaccumulator plants. Appropriate NH₄⁺⁻N application increased the transpiration and water flow of plants (Figure A1B), facilitating the absorption of various elements, including HMs, from the soil. Chai found that NH₄⁺⁻N competes with Cd²⁺, expediting Cd transport from the root cell walls to the aboveground

parts, thereby promoting Cd translocation from the root to the stem (TF) [70]. Several studies have concluded that higher protein levels in NH_4^+ -N-fed plant roots, resulting from N assimilation, can upregulate the expression of genes related to Cd loading and transport them in the xylem [25]. Additionally, NH_4^+ -N is believed to stimulate cell division, root branching, and root elongation (Figure 1) and promote nutrient uptake and biomass production, ultimately leading to increased Cd accumulation. *S. linearistipularis*, treated with the N_{120} -6/0 and N_{60} -2/4 treatments, accumulated the highest levels of Pb. Moreover, for *S. linearistipularis*, the N_{60} -2/4 treatment may help *S. linearistipularis* relieve Pb stress and increase metal ion concentrations and plant growth, thereby improving remediation efficiency. Similar results were obtained by *Zerche*, who suggested that the N form shifts hormone levels and carbon flux, which, in turn, affects the plant form [71].

N application may increase plant Cd uptake by promoting root growth and increasing Cd and Pb bioavailability in the soil-plant system. In the present study, plant traits were the main factors promoting HM accumulation, but not the available Cd and Pb in the soil, which directly affected the nutrient absorption and health of the plants (Figure 5). NH₄⁺-N is the main factor affecting plant traits and can be directly absorbed by plants, accelerating cell division, reducing energy and photosynthetic product usage, and ultimately promoting plant growth [28]. In conclusion, S. linearistipularis prefers NH_4^+ -N, which can change root traits, increase S. linearistipularis biomass, and alleviate HM stress, thereby improving plant resistance and increasing total absorption. For S. linearistipularis, we proposed a potential fertilization strategy: N_{120} -6/0 is the most effective at improving the remediation efficiency in Cd- and Pb-contaminated soil. Further validation can be conducted through field-based, in situ experiments. However, our study specifically targeted *S. linearistipularis*, acknowledging that phytoremediation capabilities vary across willow species. The applicability of our results to the phytoremediation effectiveness of other willow types remains uncertain. Additionally, our study was limited to examining Cd and Pb as pollutants despite the fact that contaminated soil frequently comprises a mix of pollutants [1,16,36]. Furthermore, the short-term nature of our experimental design leaves the long-term impacts of N fertilizer's use on soil quality and phytoremediation effectiveness unexplored. Future investigations should delve into its combined remediation efficiency against various pollutants and prioritize longitudinal studies to thoroughly understand the persistent effects of N in phytoremediation and its environmental implications.

5. Conclusions

The levels and ratios of NH_4^+ -N and NO_3^- -N significantly affected the uptake of Cd and Pb by *S. linearistipularis*. The highest accumulation rates of Cd and Pb in plants were observed under the N_{120} -6/0 treatment, which increased by 104.36% and 95.23%, respectively, compared to that of the control. In addition, under the N_{120} -6/0 treatment, the stem and leaf BCF of Cd significantly increased by 28.66% and 40.11%, respectively. Furthermore, the N_{60} -2/4 treatment significantly enhanced the plant growth and remediation efficiency of HMs, resulting in an increase of 63.92% in the total Pb uptake in the whole plant compared to the control. Plant traits, rather than soil properties, were the main factors affecting Cd and Pb uptake. The levels and ratios of NH_4^+ -N and NO_3^- -N affect Cd and Pb accumulation by regulating plant traits. Our findings highlight the potential of the ratio of NH_4^+ -N and NO_3^- -N to regulate plant traits, thereby improving the phytoremediation efficiency of HMs in contaminated soil.

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Conflicts of Interest: The authors declare no conflicts of interest.

Appendix A

Table A1. Soil properties and metal concentrations before planting.

Soil Properties	Value
pH	8.21
Soil organic matter	$15.67 \mathrm{~g~kg^{-1}}$
Total phosphorus	$185.25 \text{ mg kg}^{-1}$
Total potassium	$6.12 \mathrm{g kg^{-1}}$
Pb concentration	210 mg kg^{-1}
Cd concentration	$6.5 \mathrm{mgkg^{-1}}$

Table A2. Chemical properties of rhizosphere soil.

Nitrogen Concentratioin (kg hm $^{-2}$ year $^{-1}$)	Ratios	$_{(g\cdot kg^{-1})}^{SOM}$	CEC (cmol·Kg ⁻¹)	$(mg g^{-1} d^{-1})$	S-CAT (ml·g ⁻¹)	$NO_3^{-}-N$ (mg kg ⁻¹)	NH_4^+-N (mg kg ⁻¹)	Cd (mg kg ⁻¹)	Pb (mg kg ⁻¹)
0	0	$17.71\pm1.81~\mathrm{cd}$	$12.52\pm0.06\ ^a$	$1\pm0.09~\text{cd}$	$0.94\pm0.14~^{efg}$	$0.64\pm0.09~^{fg}$	$0.24\pm0.13\ h$	$4.09\pm0.56~\text{ef}$	$184.17\pm11.04~\mathrm{abc}$
	6/0	$19.6\pm0.4~ab$	$11.86 \pm 0.83 \ a$	$0.91 \pm 0.05 \text{ de}$	1.03 ± 0.13 cdef	$1.47\pm0.1~{\rm efg}$	$0.37\pm0.07~gh$	$2.65\pm0.17~\text{g}$	132.24 ± 19.28 fgh
	4/2	19.42 ± 0.66 ^{ab}	10.46 ± 0.28 ^a	0.73 ± 0.07 ^t	0.77 ± 0.12 g	5.08 ± 0.45 e	$0.31 \pm 0.07 \text{ gh}$	2.77 ± 0.15 g	121.61 ± 23.95 ghi
	3/3	$18.56 \pm 0.8 \text{ bc}$	11.52 ± 1.25 ^a	$0.89 \pm 0.08 \ e$	1.09 ± 0.08 bcde	$2.75 \pm 0.19 \text{ efg}$	$0.56 \pm 0.09 \text{ de}$	2.71 ± 0.24 g	159.21 ± 10.82 de
	2/4	17.35 ± 0.35 de	11.52 ± 0.58 ^a	0.99 ± 0.04 cd	$0.96 \pm 0.1 \text{ defg}$	10.59 ± 1.14 ^d	$0.4 \pm 0.05 \text{ fg}$	$3.06\pm0.42~{ m g}$	136.74 ± 9.19 efg
60	0/6	$19.07\pm0.66\ ab$	$12.19\pm0.69\ a$	$0.97\pm0.05~de$	$1.03\pm0.16\ cdef$	$11.76\pm0.92\ d$	$0.65\pm0.06\ cd$	$4.29\pm0.46~de$	$110.34 \pm 18.48 \ hi$
	6/0	$16.32\pm0.35^{\text{ e}}$	$11.03\pm1.45~^{a}$	$0.91\pm0.08~de$	$0.87\pm0.04~^{fg}$	$0.35\pm0.03~\text{g}$	$0.33\pm0.05\text{gh}$	$4.74\pm0.44~\text{cd}$	$103.24 \pm 12.79 \ i$
	4/2	$18.56 \pm 0.01 \text{ bc}$	11.87 ± 0.79 ^a	$1.19 \pm 0.05 \text{ a}$	$0.96 \pm 0.11 \text{ defg}$	19.91 ± 1.88 ^c	$0.52 \pm 0.05 \text{ ef}$	$5.14 \pm 0.16 \text{ bc}$	125.9 ± 10.48 ghi
	3/3	$18.55 \pm 0.68 \text{ bc}$	12.55 ± 1.1 ^a	$1.16 \pm 0.08 \ a$	0.95 ± 0.07 defg	22.71 ± 0.86 ^c	0.31 ± 0.06 gh	4.54 ± 0.45 cde	102.26 ± 18.05 ⁱ
	2/4	$18.73 \pm 0.66 \text{ bc}$	12.18 ± 0.99 ^a	1.16 ± 0.06 ^a	$1.18 \pm 0.09 \text{ abc}$	36.41 ± 4.92 b	0.77 ± 0.14 b	4.57 ± 0.48 cde	135.74 ± 12.25 efg
120	0/6	$19.41\pm0.35\ ab$	$12.43\pm0.86\ ^a$	$1.15\pm0.04~^{a}$	$1.34\pm0.08~a$	$23.04\pm3.49\ ^{\text{c}}$	$0.98\pm0.08\ a$	$3.64\pm0.35\ f$	$155.35\pm15~def$
	6/0	$19.25 \pm 0.57 \ ab$	$11.55 \pm 0.72 \ ^{\rm a}$	$1.06\pm0.06^{\rm bc}$	$0.91\pm0.07~\mathrm{efg}$	$4.47\pm0.52~{\rm ef}$	$0.33\pm0.06~\text{gh}$	$4.33\pm0.46~\text{de}$	132.97 ± 14.31 fgh
	4/2	20.16 ± 0.4 ^a	11.98 ± 1.75 ^a	$1.13 \pm 0.03 \text{ ab}$	1.11 ± 0.14 bcde	39.46 ± 2.76 b	$0.72 \pm 0.08 \text{ bc}$	5.55 ± 0.48 ab	194.67 ± 12.03 ab
	3/3	18.79 ± 0.79 bc	10.78 ± 0.4 ^a	1.18 ± 0.02 ^a	$1.28 \pm 0.14 \text{ ab}$	14.09 ± 1.68 d	1.35 ± 0.12 ^a	5.09 ± 0.33 bc	173.13 ± 15.84 bcd
	2/4	$18.73 \pm 0.86 \text{ bc}$	12.55 ± 0.83 ^a	1.16 ± 0.04 ^a	1.11 ± 0.1 bcde	58.17 ± 5.38 ^a	0.52 ± 0.04 def	5.62 ± 0.17 ab	168.91 ± 19.3 ^{cd}
200	0/6	$20.15\pm0.39\ a$	$12.74\pm0.79\ ^{a}$	$1.16\pm0.04~^a$	$1.15\pm0.16~abcd$	$38.59 \pm 2.26 \ ^{b}$	$1.05\pm0.04~^a$	$6.09\pm0.62~^a$	$201.92 \pm 16.63 \ a$
L		**	ns	**	**	**	**	**	**
R		**	ns	**	**	**	**	**	**
$L \times R$		**	ns	**	**	**	**	**	**

Note: All data were analyzed using two-way analysis of variance (ANOVA) followed by Duncan's test for comparison with the controls. N levels are denoted as "L", the N ratio as "R", and the interactive effect of N levels and ratios as "L × R". Different superscript lowercase letters indicate significant differences among the treatments at p < 0.05 in specific tissues. Data represent means \pm SD (n = 4), "ns" denotes non-significance, and double asterisks (**) represent significant differences at p < 0.01.

Table A3. TF of Cd and Pb.

Nitrogen Level	Pation	C	d	РЬ		
(kg hm $^{-2}$ year $^{-1}$)	Ratios	TF of Stems	TF of Leaves	TF of Stems	TF of Leaves	
0	0	$1.6\pm0.21~^{\rm abc}$	$3.59\pm0.4~^{\rm abc}$	$1.11\pm0.14~^{bcd}$	$0.78\pm0.04~^{bcdef}$	
	6/0	1.35 ± 0.41 ^{abcd}	3.86 ± 0.24 ^{abc}	1.23 ± 0.24 ^{abc}	1.04 ± 0.24 ^{ab}	
60	4/2 3/3	1.7 ± 0.34 ab 1.21 ± 0.32 ^{cde}	$4.62 \pm 0.75^{\text{ a}}$ $2.86 \pm 1.09^{\text{ cde}}$	1.33 ± 0.01 ab 1.54 ± 0.5 a	1.05 ± 0.18 ab 1.15 ± 0.24 a	
	2/4 0/6	$1.79 \pm 0.34~^{ m a}$ $1.18 \pm 0.32~^{ m cde}$	$4.08 \pm 1.26 \; ^{ m ab} \ 1.86 \pm 0.56 \; ^{ m ef}$	$0.74 \pm 0.06 \ ^{ m ef}$ $0.66 \pm 0.12 \ ^{ m f}$	$0.51 \pm 0.16 ~^{ m fg} \ 0.47 \pm 0.11 ~^{ m g}$	
120	6/0 4/2 3/3 2/4 0/6	$\begin{array}{c} 1.25 \pm 0.27 \ ^{\rm bcde} \\ 0.89 \pm 0.07 \ ^{\rm def} \\ 1.12 \pm 0.14 \ ^{\rm def} \\ 0.99 \pm 0.16 \ ^{\rm def} \\ 1.08 \pm 0.26 \ ^{\rm def} \end{array}$	3.08 ± 0.58 bcd 2.11 ± 0.3 def 2.23 ± 0.33 def 1.29 ± 0.1 f 1.84 ± 0.56 ef	$1.03 \pm 0.11^{\text{ bcde}}$ $0.72 \pm 0.09^{\text{ ef}}$ $0.76 \pm 0.09^{\text{ def}}$ $0.92 \pm 0.08^{\text{ cdef}}$	$0.71 \pm 0.1 { m cdefg} \\ 0.47 \pm 0.08 { m g} \\ 0.54 \pm 0.05 { m efg} \\ 0.81 \pm 0.08 { m bcde} \\ 0.62 \pm 0.27 { m defg} \end{cases}$	

Nitrogen Level (kg hm ⁻² year ⁻¹)	Ratios –	C	Cd	Pb		
		TF of Stems	TF of Leaves	TF of Stems	TF of Leaves	
	6/0	$1\pm0.02~^{ m def}$	1.61 ± 0.19 f	0.63 ± 0.05 f	$0.61\pm0.04~^{ m defg}$	
	4/2	$0.68\pm0.09~{ m f}$	1.03 ± 0.13 $^{ m f}$	0.61 ± 0.06 $^{ m f}$	$0.56\pm0.06~\mathrm{^{efg}}$	
200	3/3	0.65 ± 0.01 $^{ m f}$	1.58 ± 0.03 $^{ m f}$	$0.65\pm0.01~^{ m f}$	$0.88\pm0.06~^{\mathrm{bcd}}$	
	2/4	$0.89\pm0.18~^{ m def}$	$3.58\pm0.78~^{ m abc}$	$0.87\pm0.16~^{ m def}$	$0.91\pm0.05~\mathrm{^{abc}}$	
	0/6	$0.85\pm0.12~{ m ef}$	$1.84\pm0.32~^{ m ef}$	0.69 ± 0.16 ef	$0.65\pm0.05~^{ m cdefg}$	
L	**	**	**	**	**	
R	ns	ns	**	ns	ns	
$L \times R$	**	ns	**	ns	**	

Table A3. Cont.

Note: All data were analyzed using two-way analysis of variance (ANOVA) followed by Duncan's test for comparison with the controls. N levels are denoted as "L", the N ratio as "R", and the interactive effect of N levels and ratios as "L × R". Different superscript lowercase letters indicate significant differences among the treatments at p < 0.05 in specific tissues. Data represent means \pm SD (n = 4), "ns" denotes non-significance, and double asterisks (**) represent significant differences at p < 0.01.



Figure A1. Leaf gas exchange characteristics. Note: All data were analyzed using two-way analysis of variance (ANOVA) followed by Duncan's test for comparison with the controls. N levels are denoted as "L", the N ratio as "R", and the interactive effect of N levels and ratios as "L × R". Different superscript lowercase letters indicate significant differences among the treatments at p < 0.05 in specific tissues. Data represent means \pm SD (n = 4), double asterisks (**) represent significant differences at p < 0.01.



Figure A2. Root phenotype.



Figure A3. Plant growth status.



Figure A4. PCA of plant traits (A) and soil properties (B).

References

- 1. Cheng, H.; Shen, R.L.; Chen, Y.Y.; Wan, Q.J.; Shi, T.Z.; Zhu, T.; Wang, J.J.; Wan, Y.; Hong, Y.S.; Li, X.C. Estimating heavy metal concentrations in suburban soils with reflectance spectroscopy. *Geoderma* **2019**, *336*, 59–67. [CrossRef]
- Luo, X.S.; Yu, S.; Zhu, Y.G.; Li, X.D. Trace metal contamination in urban soils of China. *Sci. Total Environ.* 2012, 421–422, 17–30. [CrossRef]
- 3. Zhang, S.M.; Yang, C.; Chen, M.M.; Chen, J.; Pan, Y.H.; Chen, Y.L.; Rahman, S.U.; Fan, J.F.; Zhang, Y. Influence of nitrogen availability on Cd accumulation and acclimation strategy of *Populus* leaves under Cd exposure. *Ecotoxicol. Environ. Saf.* **2019**, *180*, 439–448. [CrossRef]
- 4. Yang, X.; Wen, F.; Ge, C.J. Iron-modified phosphorus- and silicon-based biochar exhibited various influences on arsenic, cadmium, and lead accumulation in rice and enzyme activities in a paddy soil. *J. Hazard. Mater.* **2023**, *443*, 130203. [CrossRef]
- 5. Tóth, G.; Hermann, T.; Silva, M.R.; Montanarella, L. Heavy metals in agricultural soils of the European Union with implications for food safety. *Environ. Int.* **2016**, *88*, 299–309. [CrossRef]
- 6. Zhang, X.F.; Hu, Z.H.; Yan, T.X.; Lu, R.R.; Peng, C.L.; Li, S.S.; Jing, Y.X. Arbuscular mycorrhizal fungi alleviate Cd phytotoxicity by altering Cd subcellular distribution and chemical forms in *Zea mays. Ecotoxicol. Environ. Saf.* **2019**, *171*, 352–360. [CrossRef]
- Mandal, P. An insight of environmental contamination of arsenic on animal health. *Emerg. Contam.* 2017, *3*, 17–22. [CrossRef]
 Dai, Y.Y.; Liu, R.; Zhou, Y.M.; Li, N.; Hou, L.Q.; Ma, Q.; Gao, B. Fire Phoenix facilitates phytoremediation of PAH-Cd cocontaminated soil through promotion of beneficial rhizosphere bacterial communities. *Environ. Int.* 2020, *136*, 105421. [CrossRef]
- [PubMed]
 9. Tao, Q.; Li, J.X.; Liu, Y.K.; Luo, J.P.; Xu, Q.; Li, B.; Li, Q.Q.; Li, T.Q.; Wang, C.Q. Ochrobactrum intermedium and saponin assisted phytoremediation of Cd and B[a]P co-contaminated soil by Cd-hyperaccumulator Sedum alfredii. Chemosphere 2020, 245, 125547.
 [CrossRef] [PubMed]
- 10. Vijendra, S.; Achlesh, D. Phytoremediation: A multidisciplinary approach to clean up heavy metal contaminated soil. *Environ. Technol. Innov.* **2020**, *18*, 100774. [CrossRef]
- Cui, X.Y.; Mao, P.; Sun, S.; Huang, R.; Fan, Y.X.; Li, Y.X.; Li, Y.W.; Zhuang, P.; Li, Z.A. Phytoremediation of cadmium contaminated soils by *Amaranthus Hypochondriacus* L.: The effects of soil properties highlighting cation exchange capacity. *Chemosphere* 2021, 283, 131067. [CrossRef]
- 12. Dos Santos Utmazian, M.N.; Wenzel, W.W. Cadmium and zinc accumulation in willow and poplar species grown on polluted soils. *J. Plant Nutr. Soil Sci.* 2007, 170, 265–272. [CrossRef]
- 13. Tozser, D.; Magura, T.; Simon, E. Heavy metal uptake by plant parts of willow species: A meta-analysis. *J. Hazard. Mater.* 2017, 336, 101–109. [CrossRef]
- 14. Cao, Y.N.; Ma, C.X.; Chen, G.C.; Zhang, J.F.; Xing, B.S. Physiological and biochemical responses of *Salix integra* Thunb. under copper stress as affected by soil flooding. *Environ. Pollut.* **2017**, 225, 644–653. [CrossRef] [PubMed]
- Yang, W.D.; Liu, D.; Wang, Y.Y.; Hussain, B.L.; Zhao, F.L.; Ding, Z.L.; Yang, X.; Zhu, H.Q.; Dawood, M. Variations in phytoremediation potential and phytoavailability of heavy metals in different *Salix genotypes* subjected to seasonal flooding. *J. Environ. Manag.* 2021, 299, 113632. [CrossRef] [PubMed]
- Andrej, P.; Ronald, S.Z.J.; Srđan, R.; Nataša, N.; Saša, O.; Jelena, B.; Marina, K. Growth, physiology, and phytoextraction potential of poplar and willow established in soils amended with heavy-metal contaminated, dredged river sediments. *J. Environ. Manag.* 2019, 239, 352–365. [CrossRef]
- 17. Sandra, C.G.; María, E.Y.; David, M.C.; Rocío, M.; Luis, E.H. Influence of nitrate fertilization on Hg uptake and oxidative stress parameters in alfalfa plants cultivated in a Hg-polluted soil. *Environ. Exp. Bot.* **2012**, *75*, 16–24. [CrossRef]
- 18. Ata-Ul-Karim, S.T.; Cang, L.; Wang, Y.J.; Zhou, D.M. Effects of soil properties, nitrogen application, plant phenology, and their interactions on plant uptake of cadmium in wheat. *J. Hazard. Mater.* **2020**, *384*, 121452. [CrossRef]
- Carr, N.F.; Boaretto, R.M.; Mattos, D. Coffee seedlings growth under varied NO₃⁻:NH₄⁺ ratio: Consequences for nitrogen metabolism, amino acids profile, and regulation of plasma membrane H⁺-ATPase. *Plant Physiol. Biochem.* 2020, 154, 11–20. [CrossRef]
- 20. Nishida, H.; Suzaki, T. Nitrate-mediated control of root nodule symbiosis. Curr. Opin. Plant Biol. 2018, 44, 129–136. [CrossRef]
- 21. Srivastava, S.; Pathare, V.S.; Sounderajan, S.; Suprasanna, P. Nitrogen supply influences arsenic accumulation and stress responses of rice (*Oryza sativa* L.) seedlings. *J. Hazard. Mater.* **2019**, *367*, 599–606. [CrossRef]
- 22. Boschiero, B.N.; Mariano, E.; Azevedo, R.A.; Paulo, C.O.T. Influence of nitrate—*Ammonium* ratio on the growth, nutrition, and metabolism of sugarcane. *Plant Physiol. Biochem.* **2019**, *139*, 246–255. [CrossRef]
- 23. Li, S.; Zhang, H.; Wang, S.L.; Xu, F.S.; Wang, C.; Cai, H.M.; Ding, G.D. The rapeseed genotypes with contrasting NUE response discrepantly to varied provision of ammonium and nitrate by regulating photosynthesis, root morphology, nutritional status, and oxidative stress response. *Plant Physiol. Biochem.* **2021**, *166*, 348–360. [CrossRef] [PubMed]
- 24. Leite, T.S.; Monteiro, F.A. Nitrogen form regulates cadmium uptake and accumulation in *Tanzania guinea* grass used for phytoex-traction. *Chemosphere* **2019**, 236, 124324. [CrossRef] [PubMed]
- 25. Cheng, M.M.; Wang, P.; Kopittke, P.M.; Wang, A.A.; Tang, C.X. Cadmium accumulation is enhanced by ammonium compared to nitrate in two hyperaccumulators, without affecting speciation. *J. Exp. Bot.* **2016**, *17*, 5041–5050. [CrossRef]
- 26. Jalloh, M.A.; Chen, J.H.; Zhen, F.R. Effect of different N fertilizer forms on antioxidant capacity and grain yield of rice growing under Cd stress. *J. Hazard. Mater.* **2009**, *162*, 1081–1085. [CrossRef]

- 27. Xie, H.L.; Jiang, R.F.; Zhang, F.S.; McGrath, S.P.; Zhao, F.J. Effect of nitrogen form on the rhizosphere dynamics and uptake of cadmium and zinc by the hyperaccumulator *Thlaspi caerulescens*. *Plant Soil* **2009**, *318*, 205–215. [CrossRef]
- 28. Hachiya, T.; Sakakibara, H. Interactions between nitrate and ammonium in their uptake, allocation, assimilation, and signaling in plants. *J. Exp. Bot.* 2017, *68*, 2501–2512. [CrossRef] [PubMed]
- Wang, M.; Chen, S.H.B.; Zheng, H.; Li, S.S.; Chen, L.; Wang, D. The responses of cadmium phytotoxicity in rice and the microbial community in contaminated paddy soils for the application of different long-term N fertilizers. *Chemosphere* 2020, 238, 124700. [CrossRef]
- Zeng, X.Y.; Zou, D.S.; Wang, A.D.; Zhou, A.Y.; Liu, Y.H.; Li, Z.H.; Liu, F.; Wang, H.; Zeng, Q.; Xiao, Z.H. Remediation of cadmium-contaminated soils using Brassica napus: Effect of nitrogen fertilizers. J. Environ. Manag. 2020, 255, 109885. [CrossRef]
- Han, H.; Cai, H.; Wang, X.Y.; Hu, X.M.; Chen, Z.J.; Yao, L.G. Heavy metal-immobilizing bacteria increase the biomass and reduce the Cd and Pb uptake by pakchoi (*Brassica chinensis* L.) in heavy metal-contaminated soil. *Ecotoxicol. Environ. Saf.* 2020, 195, 110375. [CrossRef] [PubMed]
- 32. Zhou, X.P.; Huang, G.Y.; Liang, D.; Liu, Y.H.; Yao, S.Y.; Ali, U.; Hu, H.Q. Influence of nitrogen forms and application rates on the phytoextraction of copper by castor bean (*Ricinus communis* L.). *Environ. Sci. Pollut. Res.* **2019**, *27*, 647–656. [CrossRef] [PubMed]
- Wu, Z.C.; Zhang, W.J.; Xu, S.J.; Shi, H.Z.; Wen, D.; Huang, Y.D.; Peng, L.J.; Deng, T.H.B.; Du, R.Y.; Li, F.R.; et al. Increasing ammonium nutrition as a strategy for inhibition of cadmium uptake and xylem transport in rice (*Oryza sativa* L.) exposed to cadmium stress. *Environ. Exp. Bot.* 2018, 155, 734–741. [CrossRef]
- Cao, Y.N.; Tan, Q.; Zhang, F.; Ma, C.X.; Xiao, J.; Chen, G.C. Phytoremediation potential evaluation of multiple Salix clones for heavy metals (Cd, Zn and Pb) in flooded soils. *Sci. Total Environ.* 2022, *813*, 152482. [CrossRef] [PubMed]
- 35. Wang, S.K.; Niu, X.Y.; Di, D.L.; Huang, D.Z. Nitrogen and sulfur fertilizers promote the absorption of lead and cadmium with *Salix integra* Thunb. by increasing the bioavailability of heavy metals and regulating rhizosphere microbes. *Front. Microbiol.* **2022**, 13, 945847. [CrossRef]
- Taiwo, A.M.; Gbadebo, A.M.; Oyedepo, J.A.; Ojekunle, Z.O.; Oyeniran, A.A.; Onalaja, O.J.; Ogunjimi, D. Bioremediation of industrially contaminated soil using compost and plant technology. J. Hazard. Mater. 2016, 304, 166–172. [CrossRef]
- 37. Meng, J.; Li, W.J.; Diao, C.M. Microplastics drive microbial assembly, their interactions, and metagenomic functions in two soils with distinct pH and heavy metal availability. *J. Hazard. Mater.* **2023**, *458*, 131973. [CrossRef]
- 38. Han, F.; An, S.Y.; Liu, L.; Wang, Y.; Ma, L.Q.; Yang, L. Sulfoaluminate cement-modified straw biochar as a soil amendment to inhibit Pb-Cd mobility in the soil-romaine lettuce system. *Chemosphere* **2023**, *332*, 138891. [CrossRef]
- 39. Arnaud, J.; Léna, D.B.; Thomas, D.; Thibault, S.; Nausicaa, N. Phytoextraction of Cd and Zn with *Noccaea caerulescens* for urban soil remediation: Influence of nitrogen fertilization and planting density. *Ecol. Eng.* **2018**, *116*, 178–187. [CrossRef]
- 40. Sterckeman, T.; Goderniaux, M.; Sirguey, C.; Cornu, J.Y.; Christophe, N. Do roots or shoots control cadmium accumulation in the hyperaccumulator *Noccaea caerulescens? Plant Soil* 2015, 392, 87–99. [CrossRef]
- Niu, X.Y.; Zhou, J.; Wang, X.N.; Su, X.Y.; Du, S.H.; Zhu, Y.F.; Yang, J.Y.; Huang, D.Z. Indigenous bacteria have high potential for promoting *Salix integra* Thunb. remediation of lead-contaminated soil by adjusting soil properties. *Front. Microbiol.* 2020, 11, 924. [CrossRef]
- 42. Cheng, F.Y. Nursery Science for Gardens; China Forestry Publishing House: Beijing, China, 2012.
- 43. Zheng, H.F.; Zhang, X.; Ma, W.J.; Song, J.Y.; Rahman, S.U.; Wang, J.H.; Zhang, Y. Morphological and physiological responses to cyclic drought in two contrasting genotypes of *Catalpa bungei*. *Environ*. *Exp. Bot*. **2017**, *138*, 77–87. [CrossRef]
- 44. Riley, D.; Barber, S.A. Bicarbonate accumulation and pH changes at the soybean (*Glycine max* (L.) Merr.) root-soil interface. *Soil Sci. Soc. Am. J.* **1970**, *34*, 154–155. [CrossRef]
- Lu, K. Analytical Methods of Soil and Agricultural Chemistry; China Agricultural Science and Technology Press: Beijing, China, 1999; pp. 127–332.
- 46. Bao, S.D. Soil Agrochemical Analysis, 3rd ed.; China Agriculture Press: Beijing, China, 2000.
- 47. Gosewinkel, U.; Broadbent, F.E. Conductimetric determination of soil urease activity. *Commun. Soil Sci. Plant Anal.* **1984**, *15*, 1377–1389. [CrossRef]
- 48. Zheng, S.N.; Zhang, M.K. Effect of moisture regime on the redistribution of heavy metals in paddy soil. *J. Environ. Sci.* 2011, 23, 434–443. [CrossRef] [PubMed]
- 49. Rauret, G.; López-Sánchez, J.F.; Sahuquillo, A. Improvement of the BCR three step sequential extraction procedure prior to the certification of new sediment and soil reference materials. *J. Environ. Monit.* **1999**, *1*, 57–61. [CrossRef] [PubMed]
- He, J.L.; Qin, J.J.; Long, L.Y.; Ma, Y.L.; Li, H.; Li, K.; Jiang, X.N.; Liu, T.X.; Polle, A.; Liang, Z.S.; et al. Net cadmium flux and accumulation reveal tissue-specific oxidative stress and detoxification in *Populus × canescens*. *Physiol. Plant.* 2011, 143, 50–63. [CrossRef]
- Kim, I.S.; Kang, K.H.; Johnson, G.P.; Lee, E.J. Investigation of heavy metal accumulation in *Polygonum thunbergii* for phytoextraction. *Environ. Pollut.* 2003, 126, 235–243. [CrossRef]
- 52. Yoon, J.; Cao, X.D.; Zhou, Q.X.; Ma, L.Q. Accumulation of Pb, Cu, and Zn in native plants growing on a contaminated Florida site. *Sci. Total Environ.* **2006**, *368*, 456–464. [CrossRef]
- Ma, Q.; Qian, Y.S.; Yu, Q.Q.; Cao, Y.F.; Tao, R.R.; Zhu, M.; Ding, J.F.; Li, C.Y.; Guo, W.S.; Zhu, X.K. Controlled-release nitrogen fertilizer application mitigated N losses and modified microbial community while improving wheat yield and N use efficiency. *Agric. Ecosyst. Environ.* 2023, 349, 108445. [CrossRef]

- 54. Zhang, Y.J.; Ye, C.; Su, Y.W.; Peng, W.C.; Lu, R.; Liu, Y.X.; Huang, H.C.; He, X.H.; Yang, M.; Zhu, S.S. Soil acidification caused by excessive application of nitrogen fertilizer aggravates soil-borne diseases: Evidence from literature review and field trials. *Agric. Ecosyst. Environ.* **2022**, *340*, 108176. [CrossRef]
- 55. Souza, J.C.; Nogueirol, R.C.; Monteiro, F.A. NO₃⁻/NH₄⁺ ratios affect nutritional homeostasis and production of *Tanzania guinea* grass under Cu toxicity. *Environ. Sci. Pollut. Res.* **2018**, *25*, 14083–14096. [CrossRef]
- Zhang, L.D.; Liu, X.; Wei, M.Y.; Guo, Z.J.; Zhao, Z.Z.; Gao, C.H.; Li, J.; Xu, J.X.; Shen, Z.J.; Zheng, H.L. Ammonium has stronger Cd detoxification ability than nitrate by reducing Cd influx and increasing Cd fixation in *Solanum nigrum* L. *J. Hazard. Mater.* 2022, 425, 127947. [CrossRef] [PubMed]
- Li, X.; Kang, X.F.; Zou, J.Z. Allochthonous arbuscular mycorrhizal fungi promote *Salix viminalis* L.-mediated phytoremediation of polycyclic aromatic hydrocarbons characterized by increasing the release of organic acids and enzymes in soils. *Ecotoxicol. Environ. Saf.* 2023, 249, 114461. [CrossRef] [PubMed]
- 58. Enggrob, K.L.; Jakobsen, C.M.; Pedersen, I.F.; Rasmussen, J. Newly depolymerized large organic N contributes directly to amino acid uptake in young maize plants. *New Phytol.* **2019**, *224*, 689–699. [CrossRef]
- Giansoldati, V.; Tassi, E.; Morelli, E.; Gabellieri, E.; Pedron, F.; Barbafieri, M. Nitrogen fertilizer improves boron phytoextraction by *Brassica juncea* grown in contaminated sediments and alleviates plant stress. *Chemosphere* 2012, *87*, 1119–1125. [CrossRef] [PubMed]
- 60. Cheng, M.M.; Wang, A.; Tang, C.X. Ammonium-based fertilizers enhance Cd accumulation in *Carpobrotus rossii* grown in two soils differing in pH. *Chemosphere* **2017**, *188*, 689–696. [CrossRef]
- 61. Sarker, A.; Masud, M.A.; Deep, D.M.; Kallol, D.; Rakhi, N.; Most, W.R.A.; Reza, A.M.; Islam, T. Biological and green remediation of heavy metal contaminated water and soils: A state-of-the-art review. *Chemosphere* **2023**, *332*, 138861. [CrossRef] [PubMed]
- Gu, S.G.; Zhang, W.; Wang, F.; Meng, Z.H.; Cheng, Y.; Geng, Z.X.; Lian, F. Particle size of biochar significantly regulates the chemical speciation, transformation, and ecotoxicity of cadmium in biochar. *Environ. Pollut.* 2023, 320, 121100. [CrossRef] [PubMed]
- 63. Zhang, P.; Xue, B.; Jiao, L.; Meng, X.Y.; Zhang, L.Y.; Li, B.X.; Sun, H.W. Preparation of ball-milled phosphorus-loaded biochar and its highly effective remediation for Cd- and Pb-contaminated alkaline soil. *Sci. Total Environ.* **2022**, *813*, 152648. [CrossRef]
- 64. Guo, S.H.; Hu, N.; Li, Q.S.; Yang, P.; Wang, L.L.; Xu, Z.M.; Chen, H.J.; He, B.Y.; Zeng, E.Y. Response of edible amaranth cultivar to salt stress led to Cd mobilization in rhizosphere soil: A metabolomic analysis. *Environ. Pollut.* **2018**, 241, 422–431. [CrossRef]
- 65. Esteban, R.; Ariz, I.; Cruz, C.; Jose, F.M. Review: Mechanisms of ammonium toxicity and the quest for tolerance. *Plant Sci.* 2016, 248, 92–101. [CrossRef] [PubMed]
- 66. Su, S.Q.; Zhou, Y.M.; Qin, J.G.; Wang, W.; Yao, W.Z.; Song, L. Physiological responses of *Egeria densa* to high ammonium concentration and nitrogen deficiency. *Chemosphere* **2012**, *86*, 538–545. [CrossRef]
- 67. Bock, B.R. Increasing cereal yields with higher ammonium/nitrate ratios: Review of potentials and limitations. *J. Environ. Sci. Health* **1986**, *21*, 723–758. [CrossRef]
- 68. Babalar, M.; Sokri, S.M.; Lesani, H.; Mohammad, A.; Allen, V.B. Effects of nitrate: Ammonium ratios on vegetative growth and mineral element composition in leaves of apple. *J. Plant Nutr.* **2015**, *38*, 2247–2258. [CrossRef]
- 69. Chang, Y.S.; Chang, Y.J.; Lin, C.T.; Lee, M.C.; Wu, C.W.; Lai, Y.H. Nitrogen fertilization promotes the phytoremediation of cadmium in Pentas lanceolata. *Int. Biodegrad.* 2013, *85*, 709–714. [CrossRef]
- 70. Chai, M.W.; Li, R.Y.; Shen, X.X.; Nora FY, T.; Zan, Q.J.; Li, R.L. Does ammonium nitrogen affect accumulation, subcellular distribution and chemical forms of cadmium in *Kandelia obovata*? *Ecotoxicol. Environ. Saf.* **2018**, *162*, 430–437. [CrossRef] [PubMed]
- 71. Zerche, S.; Druege, U.; Kadner, R. Nitrogen absorption, growth of stock plants, and adventitious rooting of *Pelargonium* × *hortorumcuttings* as affected by the form and dosage of nitrogen. *J. Hortic. Sci. Biotechnol.* **2008**, *83*, 207–217. [CrossRef]

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