

Article

Sustainability Assessment of Public Transport, Part I—A Multi-Criteria Assessment Method to Compare Different Bus Technologies

Jonas Ammenberg  and Sofia Dahlgren * 

Environmental Technology and Management and Biogas Research Center (BRC), Department of Management and Engineering, Linköping University, SE-581 83 Linköping, Sweden; jonas.ammenberg@liu.se

* Correspondence: sofia.dahlgren@liu.se

Abstract: This article departs from the perspective of Swedish regional transport authorities and focuses on the public procurement of bus transports. Many of these public organizations on the county level have the ambition to contribute to a transition involving the continued marginalization of fossil fuels and improved sustainability performance. However, there are several renewable bus technologies to choose between and it can be difficult to know what alternative (or combination) is preferable. Prior research and the authors' experiences indicate a need for improved knowledge and supportive methods on how sustainability assessments can support public procurement processes. The purpose of this article is to develop a multi-criteria assessment (MCA) method to support assessments of public bus technologies' sustainability. The method, which was established in an iterative and participatory process, consists of four key areas and 12 indicators. The article introduces the problem context and reviews selected prior research of relevance dealing with green or sustainable public procurement and sustainability assessments. Further on, the process and MCA method are presented and discussed based on advice for effective and efficient sustainability assessments. In the companion article (Part II), the MCA method is applied to assess several bus technologies involving biodiesel, biomethane, diesel, electricity, ethanol and natural gas.

Keywords: bus technologies; multi-criteria assessment; MCA; MCDA; public transport; sustainability assessment; sustainable or green public procurement



Citation: Ammenberg, J.; Dahlgren, S. Sustainability Assessment of Public Transport, Part I—A Multi-Criteria Assessment Method to Compare Different Bus Technologies. *Sustainability* **2021**, *13*, 825. <https://doi.org/10.3390/su13020825>

Received: 17 December 2020

Accepted: 11 January 2021

Published: 15 January 2021

Publisher's Note: MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Copyright: © 2021 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (<https://creativecommons.org/licenses/by/4.0/>).

1. Introduction

Transportation systems are linked to essential environmental, health and resource challenges [1] and many organizations take actions for improved sustainability performance, for example, References [2–4]. A rapid transition is needed to reach objectives such as those regarding renewable energy, climate impact, air quality, for example, in the EU and countries like Sweden [5–8]. For many different types of transportation, there is a wide array of possible alternative technologies involving, for example, biodiesel, biomethane, electricity and ethanol [9–11].

This article deals with the generic problem of how to assess transport technologies' sustainability performance (including technical and short-term economic aspects). It departs from the perspective of Swedish regions, which are authorities on the county level that, among other duties, are responsible for the procurement of public transport as regional public transport authorities. The focus is on buses, as they have a dominating position among the public transport modes in Sweden [12], as in many parts of the world [13].

Regarding bus fuels in public transport in Sweden, a major transition has occurred during the last two decades. At the beginning of the 21st century, the bus fleet was almost totally driven on fossil fuels, whereas in 2017, more than 60% of the buses used renewables [14]. The regions, specifically the regional public transport authorities, have been key actors in this transition [11]. Via green public procurement (GPP) or sustainable

public procurement (SPP), they have importantly contributed to this shift. However, several different factors seem to influence how public procurements are arranged and what technical solutions are preferred [11,15]. For example, the shift in Sweden to a large extent started with bioethanol, which reached a peak in 2011 but has since decreased by 50% (see Figure 1 in Reference [14]). The use of biomethane for transport started in the 1990s and has grown relatively rapidly and steadily since 2004. Biodiesel became popular around 2010—first, the use of FAME (Fatty Acid Methyl Ester) grew fast and since 2015 there has been a strong focus on HVO (Hydrotreated Vegetable Oil). Lately, electric buses have entered the stage. Their numbers are still relatively low but this alternative gets much attention and is generally expected to play an important role in the near future, at least in city centers [16–18].

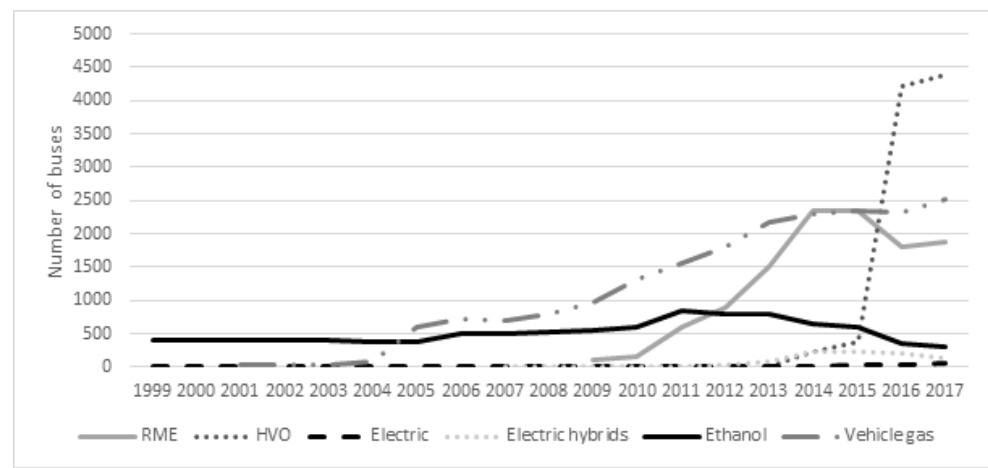


Figure 1. The number of buses in Sweden in the last decades based on fuel use based on estimations of the number of buses using each type of fuel, done by Reference [19].

The described palette with renewable technologies provides great opportunities for the continued marginalization of fossil fuels. However, when the regions (or other actors) ask what alternative or what combination of alternatives, is best suited for them, it may be challenging to answer as the question encompasses many sub-questions. Generally, there seems to be a great need for improved knowledge and supportive methods when it comes to including environmental/sustainability assessments in public procurement processes [20] and in other types of related appraisals [21,22]. This is not least relevant in connection with transportation [15,23,24]. These observations are in line with the authors' experience of working within the Swedish Biogas Research Center (BRC) and in related projects, where we have been approached by regions, municipalities and biogas sector organizations requesting support on how to compare different bus transport technologies. In addition, biofuel producers and distributors and other stakeholders, for example, see Figure 1 in Reference [11] have shown great interest in developing these kinds of assessments.

Aim and Scope

The purpose of this article is to develop a multi-criteria assessment (MCA) method to support assessments of public bus technologies' sustainability. The method shall:

- Be based on existing knowledge of practitioners and scientists
 - Be adapted to the context and input from stakeholders, particularly Swedish regions' challenges related to the procurement of bus services and their views on sustainability
- Include common indicators such as monetary costs and function/quality [20] but broaden the scope in relation to many existing methods

- Cover essential areas regarding sustainability and the most relevant aspects for the different technologies to be assessed
- Be relatively simple to use with a reasonable number of indicators to facilitate data collection and overview
 - When used, provide results for a wide range of indicators without weighting, thus leaving the users to decide based on their own preferences if any indicators are more important, such as local conditions and prioritized objectives.

This article is the first (Part I) of two associated articles and presents the MCA method establishment process. Following the introduction, Section 2 provides a literature review on green or sustainable public procurement (GoSPP) and environmental or sustainability assessments (EoSA), focusing on MCA and transport systems. Section 3 gives an overview of the working process, while Section 4 presents the outcome—the established MCA method. Finally, there is a concluding discussion in Section 5. In Part II, the method is applied for the comparison of different bus technologies (Dahlgren and Ammenberg, 2020 (“Part II” is hereafter used to refer to this second article.)).

2. Prior Research and Methodological Approach

2.1. Green or Sustainable Public Procurement

Cheng et al. [20] conducted an extensive literature review on GoSPP and found an increased focus on it in recent years, both in practice and research (cf. [25], focusing on the private sector). Key procurement actors, like local authorities, struggle in the implementation and improvements are required regarding the follow-up of requirements.

Budget constraints are commonly emphasized as the main barrier to GoSPP [20,26,27], due to additional costs of products or services with superior sustainability performance (cf. [22]) and a need for additional staff resources [15,28]. Conflicting objectives were also found to be a common barrier in several studies [28]. A broader and more long-term perspective is needed to account for sustainability impacts and shift the direction toward more comprehensive socio-economic outcomes [28–30]. Accordingly, based on life cycle costing (LCC), it has been established that GPP within the EU has led to decreased costs for the purchasing organizations. A short-term cost focus makes it, in general, very difficult to address sustainability issues [15].

A lack of sufficient competency is another frequently mentioned barrier [20,28]. Michelsen and de Boer [31] found, in a large study among Norwegian municipalities and counties, that only about 5 percent of the municipalities “... felt they had sufficient competence to formulate demands on environmental performance and evaluate the information they received in the suppliers’ offers” (p. 163). Testa et al. [32] even concluded knowledge and training to be more important than economic resources.

Perera et al. [33] refer to a study from 2007, where hundreds of tools, including SPP tools, were identified, including guidelines, handbooks, databases and software. This report does not include the characteristics of these tools but indicates that a large share consists of unique tools developed by individual organizations to be applied in their particular context. Several studies highlight the need for simple methods and guidance [20,28,32,34]. In line with this, Cheng et al. [20] noted the lack of official guidance on green public procurement. Thomson and Jackson [35] emphasize the need to develop new models to quantify the wider impacts of procurement decisions. About 70% of the Norwegian municipalities in the study by Michelsen and de Boer [31] emphasized a need for templates, including standardized environmental requirements. The EU has provided a training toolkit on GPP, with criteria for ten key sectors including transport but Palmujoki et al. [36] found that this has the potential for further improvements, for example, by including more detailed criteria.

Some articles deal with EU policy and practical implications. Luttenberger and Luttenberger [29] state that the 2014 EU Directive on Public Procurement implies that technical specifications can relate to sustainability impacts at any stage of the life cycle of a product or service, that the considered costs can regard acquisition, use, maintenance, end

of life and environmental externalities. Aldenius and Khan [15] mention the two options with the EU to use minimum compliance criteria and award criteria, where the latter can give additional points to the bidder in the tender. They state that minimum requirements are dominating regarding renewable fuels in the Swedish bus sector and found two types of such requirements to be used: functional (e.g., a maximum amount of CO₂ that can be released) and specific (e.g., that a specific fuel has to be used) (cf. Reference [37]).

In the reviewed literature, there is limited detailed information on methods or tools for comparison and evaluation of different types of products and services, including transport. Therefore, the literature review was extended to encompass publications on sustainability assessments that could provide more relevant information considering this article's purpose.

2.2. Environmental or Sustainability Assessments

There has been a focus on sustainable development for several decades [38] and much has been written about its meaning and implementation [39–42]. Despite abundant “sustainability initiatives” within organizations of different types, Waas et al. [43] emphasize an implementation gap (as do [44] and other researchers) and urgent need to focus more on implementation in essential decision-making processes, stating, “. . . sustainable development must be considered as a decision-making strategy” ([43], p. 5513). In line with many others, they argue for the need for environmental or sustainability assessment (EoSA) methods that can support and influence decision-making towards sustainability.

Sustainability assessment methods are intended to provide decision-makers with knowledge on sustainability implications of different actions or alternatives [45]. There are several different methods for the integration of sustainability into decision making [46]. As we see the sustainability assessment of bus technologies as a complex problem or wicked problem [47,48] and acknowledge value pluralism [49], it seemed reasonable to use a multi-criteria approach. In practice, the process of establishing an MCA method and applying it often incorporates several of the other methods (or techniques) described by Bueno et al. [46], such as combining an MCA with cost-benefit analysis [50,51]. For example, our application of the MCA method has been influenced by sustainability indicators used in other sustainability evaluation tools and results from LCC and LCA (Life Cycle Assessment/Analysis) studies were used. There are many different types of MCA methods in the literature, for example, References [24,51–55]. Belton and Stewart [56] define multi-criteria decision analysis as “an umbrella term to describe a collection of formal approaches which seek to take explicit account of multiple criteria in helping individuals or groups explore decisions that matter”. A strength of MCA methods is that they allow for the inclusion of several different types of indicators [21], both of a quantitative and qualitative character [57–59] and can involve expert assessments and participatory processes [52,60,61]. They can be used to handle large amounts of information and to select and structure the most relevant information, thereby providing more holistic evaluations [51,52] and simultaneously facilitating overview and communication [54,62]. In addition to the mentioned strengths, there are also drawbacks cited in the literature regarding MCA, for example:

- The use of several indicators with different directions or units, can imply that some effects are “double”-counted [63,64], that is, in case indicators are overlapping, which can be hard to avoid
- Critique regarding arbitrariness and subjectivity [63].

The use of MCA in connection with decisions involving sustainability considerations seems to have increased in recent years/decades [65–73]. Bueno et al. [46] report on several MCA studies in the field of transport. Commonly, an MCA process involves the following steps, adapted based on, for example, References [21,51,54]:

1. Definition of problem, purpose and the alternatives to be assessed;
 - This can depart from generic definitions and principles regarding sustainability (e.g., Waas et al., 2014)

- It is important to make a context specific definition or adaption, for example, Reference [55]
- 2. Selection/Definition of criteria/indicators, deciding what to consider and how it shall be assessed. Many articles deal with criteria/indicators, for example, Reference [74]
- 3. Data collection for each alternative
- 4. Weighting, meaning that different indicators/criteria can be assigned a weight in relation to their relative importance
- 5. Assessment and analysis
- 6. Presentation and interpretation of results, recommendations.

These steps are also similar to the MCA-based decision-making process illustrated by Foxon et al. [74] and Oltean-Dumbrava et al. [75]. There are several specified methods for MCA, for example, to assign weights and calculate scores that can be relatively advanced, such as the analytic hierarchy process developed by Saaty [76]. Hüging et al. [21] have reviewed assessment methods and conclude that there is a lack of suitable methods for sustainability assessments and specific demand for simple but broad approaches (cf. [22,75]). This has influenced our aim and recommendations from the mentioned studies (and others, see below) has influenced the MCA method development process.

Effective and Efficient Sustainability Assessments

The literature on sustainability assessments, including MCA, contains much advice regarding methods and processes. Based on their review, Waas et al. [43] (based on original references, such as Baker and McLelland [77]) present four categories of effectiveness: substantive (e.g., to achieve the intended outcome), normative (e.g., social learning), procedural (e.g., that the process is open, fair and objective) and transactive (efficient use of resources, including time). Furthermore, it is emphasized that the ideal-typical sustainability assessment needs a “top-down/expert-driven” and “bottom-up/stakeholder-driven” integration and be able to make use of different kinds of knowledge and create opportunities for learning. Oltean-Dumbrava et al. [75] present practical requirements for good multi-criteria decision-making tools, such as consistency and logical soundness, user-friendliness and good visual presentation of results. Through combining advice from several articles [43,74,78–82], criteria for suitability and usability are listed below:

1. Comprehensiveness and relevance: the indicators should cover economic, environmental, social and technical aspects in order to ensure that account is being taken of progress towards sustainability objectives (cf. [44]). The indicators should be relevant in relation to the studied problem and the context of the study (democratic, good stakeholder participation). The indicators should allow grading in relation to sustainability, that is, provide results on the sustainability performance.
2. Practicability: a reasonable number of indicators that are straightforward and possible to use, considering the time frames and resources available for the assessment and which form a practicable set for the purposes of the decision.
3. Applicability: the indicators should be applicable for every alternative under consideration and interpretable. Reference values can facilitate.
4. Tractability: there should be sufficient reliable data (numerical or qualitative data should be available to enable the estimation).
5. Transparency: the indicators (including criteria/scales) should be easy to understand and chosen in a transparent way, not least to enable stakeholders to clearly identify what is being considered, to understand the criteria/scales used and to propose other criteria for consideration.
6. The indicators should be predictable in response, sensitive and responding to relevant changes or differences of performance.
7. The indicators should not be (strongly) correlated.
8. The indicators should be acceptable from an ethical perspective.

The advice and these criteria have been considered in the MCA method development process. They are related to in the coming sections and discussed at the end of this article and in Part II. It can be challenging to establish a method with a strong set of indicators in relation to all criteria. The selection process commonly involves different tradeoffs [43,83].

3. Methods

3.1. Process Description

The process of establishing the MCA method was participatory (or collaborative, [84]) and iterative, with different actors being involved throughout the process. It was initiated in 2017 by the authors, who created a first version of the MCA method in the form of a presentation dealing with key areas, key questions and indicators to possibly include. This was done based on previous MCA method establishment experiences [54,62,85], an initial literature review of other studies dealing with assessments of bus technologies and discussions. From that point, the method has been developed iteratively in cooperation with several stakeholders and experts, as presented in Table 1, in parallel with extended literature reviews.

Table 1. Actors that have been involved in the multi-criteria assessment (MCA) method development process, more actively as participants (P) in research projects and stakeholders (S) that have been given the opportunity to provide input at meetings and conferences.

Project Participants (P, p ¹) and Involved Stakeholders (S)	Relevance, Competences	Comment
Region Östergötland (P), part in BRC	Environmental strategist with long-term experience regarding bus technologies, sustainability issues and public procurement processes. In later stages, an energy and climate strategist was also involved	Has participated in the whole MCA establishment process and has been part of several workshops dealing with indicators, scales and results
Other regions (p)	Long-term experience regarding public bus transports and other relevant issues	The regions of Gotland, Kalmar and Jönköping participated in later stages of the process (the last two years). They, for example, provided input at a dedicated workshop
Gasum (P), part in BRC	Represented by a business development specialist, civil engineer specialized in environmental and energy management. Also represented by a business development manager	Participated in the whole MCA establishment process and has been part of several workshops dealing with indicators, scales and results
Linköping University (P, p), part in BRC	Experts in: - environmental systems analysis and biofuels (P) - business administration (p) - sociotechnical systems (p)	Four researchers, namely the authors and two other colleagues, participated through the whole MCA establishment process. Researchers with expertise in business administration and sociotechnical systems provided input in later stages
Municipalities (p), part in BRC	Long-term experience regarding public bus transports and several other relevant issues	The municipalities of Linköping and Norrköping participated in later stages of the MCA establishment process (the last two years). They, for example, provided input at a dedicated workshop
Tekniska Verken (p), part in BRC	A municipally owned utility company, for example, with expertise in energy and waste management. Long-term experience of biogas and electricity production and use for public bus transports	Provided input at the later stages of the MCA establishment process (the last two years)

Table 1. Cont.

Project Participants (P, p ¹) and Involved Stakeholders (S)	Relevance, Competences	Comment
Scania (p), part in BRC	This company provides transport solutions. Manufacturer of buses and trucks, with expertise regarding all the studied vehicle types and fuels	Provided input at the later stages of the MCA establishment process (the last two years). They, for example, provided input at a dedicated workshop
Borlänge energi (p), part in BRC	A municipally owned utility company, for example, with expertise in energy and waste management. Experience with biogas and electricity production and use for public bus transports	Provided input at a dedicated workshop
JES (S)	A management consultancy firm that has been working with, for example, biogas and public transport	Provided input at a dedicated workshop
Vattenfall (S)	A state-owned energy utility company operating in Europe with expertise in, for example, electrification and other relevant issues	Took part in a research project on green buses in which the authors participated. Provided input to two presentations of the MCA method
Other partners of BRC (S)	Other than the already-listed organizations taking part in this transdisciplinary research center. Expertise within many areas related to socio-technical systems, fuels, transport, etc.	The project and MCA method were presented at large BRC meetings during poster sessions and other events, which resulted in relevant input from those with different backgrounds and competences

¹ A capital letter indicates participation through the whole MCA method establishment process or a large contribution in any stage, while a lowercase letter indicates a somewhat smaller contribution, often in later stages.

During 2018, the then-existing version of the MCA method was further developed by students supervised by the authors, first, by one master's thesis student (30 ECTS) and then by a group of four students taking a project course (12 ECTS per student) (further information in the acknowledgements). The students provided further knowledge on relevant indicators and data and their work involved interactions with several of the actors shown in Table 1 and others within BRC. After the student contributions, the MCA establishment process continued with further literature studies and finetuning of the set of indicators. The project participants listed in Table 1 provided input during a dedicated workshop, where the indicators were discussed and they have regularly been involved in the process via BRC project meetings. In the final stages of the process, when the MCA method was established and had been applied, the pre-final drafts of both articles (Part I & II) were reviewed by a group of 10 selected experts. They were selected to include competency on sustainability systems analysis, specifically, multi-criteria analysis, LCA, LCC and energy analysis; transportation, especially regarding other fuels than biomethane to complement the competency profile of the authors; sociotechnical systems; and environmental innovations.

3.2. The Assessed Bus Technologies

Seven different kinds of technologies are used in Swedish public buses: biomethane, diesel, electricity, ethanol, FAME, HVO and natural gas [86]. They formed the basis for the development of the MCA method and are the alternatives assessed in Part II.

The FAME used is almost exclusively produced from rapeseed [87], meaning that the focus was on RME (Rapeseed Methyl Ester). Regarding electric buses, both slow-charging options, which use less infrastructure but larger batteries and fast-charging options, which use more infrastructure but smaller batteries, have been considered. These technologies/fuels are further specified in Part II. The assessment has focused on 12-m-long Euro VI buses.

3.3. Selection of Key Areas, Indicators and Scales for Assessment

Several key areas were selected for inclusion in the MCA method, considering what issues were focused on in the relevant literature regarding transport and sustainability, the specific technologies and other relevant contexts [83,88]. The participants and stakeholders influenced the outcome in different ways. The participating regions were central, providing input on what is characteristic of a “sustainable” or resource-efficient bus technology. To be able to make assessments for each key area, several indicators were selected and scales were defined for assessment. This process was iterative, where indicators have been, for example, suggested and discussed, then kept, revised, combined or discarded; the process was similar for the scales. In some cases, already-existing indicators and scales for assessment have been used, while other indicators and scales have been defined more from scratch.

For each indicator, we defined five-step scales using quantitative intervals or qualitative descriptions, ranging from very poor to very good. In some cases, when it was not seen as reasonable to use five steps, a three-step scale was used where the scales for poor and good were removed. In addition, a simple three-step scale was used to indicate the uncertainty of the assessor: “*” referred to high uncertainty (not certain), “**” referred to some uncertainties and “***” referred to low uncertainty (rather certain).

3.4. Additional Steps in the MCA Process

The previously described general MCA method (see Section 2) has been followed to a large extent, although we have not conducted any weighting (Step 4 is dealt with in the concluding discussion). The steps concerning data collection (3), assessment and analysis (5) and presentation and interpretation of results, recommendations (6) are mainly described in the associated Part II article, where the MCA method is applied.

4. The MCA Method for Sustainability Assessment of Bus Technologies

An overview of the established MCA method, including key areas, key questions and indicators, is provided in Table 2. The four key areas are based on the dimensions of sustainable development and are in line with Oltean-Dumbrava [75], Foxon et al. [74] and others. In the following sub-sections, the different areas are presented with their indicators and scales for assessment, which are introduced and motivated. Generally, the indicators and scales have been selected and defined to focus on aspects where the bus technologies perform differently, that is, excluding issues where they are performing similarly, which is clarified in the following descriptions and in the concluding discussion.

Table 2. Key areas, key questions and indicators of the MCA methodology for assessment of the sustainability of different bus technologies.

Key Areas and Key Questions	Indicators
Technical performance <i>Is the technology cost-efficient with a stable cost development?</i>	- Technical maturity - Daily operational availability
Economic performance <i>Is the technology cost-efficient with a stable cost development?</i>	- Total cost of ownership - Need for investments in infrastructure - Cost stability
Environmental performance <i>Is the technology favorable concerning environmental impacts and management of natural resources?</i>	- Non-renewable primary energy efficiency - Greenhouse gas emission savings - Local/regional impact on land and aquatic environments - Air pollution - Noise
Social performance <i>Is the technology favorable concerning societal and social issues?</i>	- Energy security - Sociotechnical system services

4.1. Technical Performance

For any bus technology, it is a key requirement that it works well from a technical perspective [88], being able to provide the desirable transport function and additional services during the contracted period (commonly around 10 years in Sweden). In cases where public organizations take part in technical development projects, this could differ and technical indicators can then be reformulated, deprioritized or removed. Regarding comfort, accessibility and security, the standard Bus Nordic (see Bus Nordic, Common Nordic bus procurement requirements, version 2018) is commonly used, which includes predefined requirements that each service provider must fulfill cf. [89]. As these requirements are the same for any technology, they have not been included in the MCA method.

Two technical indicators were selected. The first indicator concerns *technological maturity* (or technological reliability or readiness, for example, see Reference [90]). This indicator was incorporated to assess the stage of development and implementation of bus technologies. Generally, a technology can be more or less developed, ranging from early research stages to mature solutions that have been used in society for many years. Newer technologies are prone to come with technological challenges, both directly related to their function and related to less established support networks [91]. It is thus considered favorable if a technology is well-established, both nationally and internationally. Table 3 presents the qualitative scale for the indicator technological maturity. It has been assumed that the technology is the same or similar, during the contract period. In the context of long-term assessments, however, a new technology can have larger development potentials [91].

Table 3. Scale for the indicator “technical maturity.”

Value	Scale Definition
Very good	Well-established technology on the national and international market. No relevant doubts regarding the technical performance. High operational availability is expected.
Good	Well-established technology on the national market. No relevant doubts regarding the technical performance. High operational availability is expected.
Satisfactory	Relatively new technology, commercially implemented and proven to work well in some cases, in conditions similar to the national/regional context. More limited support networks compared to the levels of good and very good. Some uncertainties regarding the performance, for example, regarding operational availability, energy use, replacement of critical components or needs of maintenance.
Poor	New technology, tested in several cases or commercially implemented in some cases with different conditions from the national/regional context. Very limited support networks. Large uncertainties regarding the performance, for example, regarding operational availability, energy use, replacement of critical components or needs of maintenance.
Very poor	Possibly coming technology but not developed enough to be seen as a reasonable alternative from a technical perspective.

The second technology indicator is oriented towards the daily operational duties. For efficient use, it is relevant to consider if and to what extent necessary stops influence the ability to perform the duties. Therefore, we included the indicator *daily operational availability* considering the range and time for refueling or recharging; see the scale in Table 4.

Table 4. Scale for the indicator “daily operational availability.”

Value	Scale Definition
Very Good	Refueling or recharging is conducted during the night (or during another period with low demand) and results in a vehicle range that is sufficient to carry out the daily duties without any additional stops for refueling/recharging.
Satisfactory	Refueling or recharging is conducted during the day (or during another period with relatively high demand) but without significant negative impact on the wanted timetables or any need for additional vehicles due to refueling/recharging.
Very Poor	Refueling or recharging is conducted during the day (or during another period with high demand), significantly influencing the wanted timetables negatively or leading to needs of additional vehicles due to refueling/recharging.

4.2. Economic Performance

All procurers of public transport need to consider the costs involved; they should try to be cost-efficient and maximize the transport service level within set budget frames [92]. This key area focuses on costs directly linked to the vehicles and their use and costs related to infrastructure. Thus, the perspective is relatively narrow and short term, not broader socio-economic and long term. However, the broader environmental performance assessment compensates for this to some extent.

The first economic indicator is the *total cost of ownership* (TCO), including costs for purchasing the bus, fuel and maintenance/repair and considering the vehicles’ residual value. Many sustainability assessments consider the costs; some use TCO or life-cycle costs (LCC) that can have similarities (when focused on direct costs, LCCs differ, considering indirect costs and externalities), while others focus on some of these costs or are not that transparent regarding what has been included [30,83,88,93]. Even if the focused TCO is most often mainly taken by the service provider (depending on the type of contract - there might for example be indexed costs (i.e., not fixed costs) where the regions must pay more in case of increasing fuel prices), it strongly influences the price offered by the providers and thus influences the procurer’s budget. Table 5 shows the quantitative scale for the indicator total cost of ownership, grading each bus technology’s cost against the average/median cost (of all the studied technologies in our case, of all offerings in a practical case). The percentage levels are loosely based on the costs according to Ecotrafic [94] that compared Euro VI buses of all the relevant technologies apart from electric buses. However, the scale could be adapted and based on offered price differences in a procurement rather than using these percentages. There are also other options for such an indicator/scale, for example, to specify absolute cost intervals (e.g., in SEK/km) for each level.

Table 5. Scale for the indicator “total cost of ownership.”

Value	Scale Definition
Very good	The costs are at least 15% lower than the average/median cost.
Good	The costs are at least 5% lower than the average/median cost but not lower than 15%.
Satisfactory	The costs are average, within a range of 5% from the average/median cost.
Poor	The costs are at least 5% higher than the average/median cost but not higher than 15%.
Very poor	The costs are at least 15% higher than the average/median cost.

The indicator *need for investments in infrastructure* concerns the level of investment a technology requires. These investments can be related to new infrastructure but also maintenance or expansion of existing infrastructure. It is an indicator that also has been part of several similar studies [88,93]. These costs have been separated from the total cost of

ownership, as they normally (at least in a Swedish context) are not taken by the bus service provider as a part of the public procurement of bus service contract. In this study, these investments may concern, for example, infrastructure for storage and transport of fuels, for refueling and for recharging. However, we have not included infrastructure or facilities for the production of fuel/electricity or any parts upstream from the production. Table 6 shows the qualitative three-step scale. In the assessment, it can be the logic to compare the costs for infrastructure with the total cost of ownership, when deciding what is minor and significant (or in-between).

Table 6. Scale for the indicator need for “investments in infrastructure.”

Value	Scale Definition
Very good	No investments in infrastructure are needed for this technology.
Satisfactory	Minor but acceptable investments in infrastructure are needed for this technology.
Very poor	Significant investments in infrastructure are needed for this technology.

Cost stability concerns risks for significantly increased or unexpected costs during the contract period but also chances of lower costs. Such increased costs can, for example, be due to scarce resources and increased demand for some fuels or for unplanned repairs. The costs of public bus transport in Sweden have increased significantly since 2010 [95]. This is partly explained by costs that we have not considered, like salaries, as they are assumed to be independent of bus technology but costs related to vehicles and fuels have also contributed to this development [16,95]. Depending on the contract between the public transport authority and the service providers, changed costs have different implications. The focus of the indicator is on changed costs taken by the public transport procurer, the scale of which is shown in Table 7. In many cases, it may be seen as non-controversial to avoid or lower the economic risks, in line with the suggested indicator and in any case wise to learn about existing risks. Nevertheless, there may be negative implications if such an assessment results in excessively low marginals for the service providers or leads them to take an unreasonable share of the risks (in cases where it is difficult to avoid/lower risks and still provide the services wanted) [96]. In addition, chances of high profits for the service providers may be disliked by the taxpayers (seen as non-efficient management of public funds but this depends on how profits are used). The scale could be complemented to cover such cases as well.

Table 7. Scale for the indicator “cost stability.”

Value	Scale Definition
	<i>Focusing on Costs Taken by the Procurer/Region:</i>
Very good	The costs related to vehicles or fuels are expected to significantly decrease during the time period of the service contract. There is a good chance of costs significantly below the expected budget level.
Good	The costs related to vehicles or fuels are expected to slightly decrease during the time period of the service contract. There is a good chance of costs below the expected budget level.
Satisfactory	The costs related to vehicles or fuels are expected to remain stable during the time period of the service contract. There is a good chance of costs in line with the expected budget level.
Poor	The costs related to vehicles or fuels are expected to slightly increase during the time period of the service contract. There is a risk of costs above the expected budget level.
Very poor	The costs related to vehicles or fuels are expected to significantly increase during the time period of the service contract. There is a risk of costs significantly above the expected budget level.

4.3. Environmental Performance

In line with the introduction, Swedish regional transport authorities commonly want bus technologies with a favorable environmental performance. There are many possible impact categories [97] or indicators [79,98] that could be considered, of which a few were chosen to reach a reasonable total number.

The first indicator deals with *non-renewable primary energy efficiency* from a well-to-wheel perspective, in line with the reasoning by Marcus Gustafsson et al. [99]. The Swedish Association of Local Authorities and Regions provides yearly statistics for the comparison of different regions' performance [100], where they regarding energy focuses on vehicle energy use. However, this narrow systems perspective does not account for important energy and environmental issues, as 'efficient buses' may be associated with high energy use and environmental impact (cf. [30]). The scale is defined in Table 8, based on the findings of Gustafsson et al. [99] regarding the energy use of buses in Swedish regions.

Table 8. Scale for the indicator "non-renewable primary energy efficiency."

Value	Scale Definition
Very good	The bus technology uses less than 1 kWh of non-renewable primary energy/vehicle kilometer.
Good	The bus technology uses between 1 and 1.5 kWh of non-renewable primary energy/vehicle kilometer.
Satisfactory	The bus technology uses between 1.5 and 2 kWh of non-renewable primary energy/vehicle kilometer.
Poor	The bus technology uses between 2 and 2.5 kWh of non-renewable primary energy/vehicle kilometer.
Very poor	The bus technology uses more than 2.5 kWh of non-renewable primary energy/vehicle kilometer.

Greenhouse gas emission (GHG) savings is an issue of great environmental importance [101] that receives much attention from decision makers [102] and is commonly included in similar MCA methods, for example, References [83,88,93] and other relevant studies. The indicator deals with the amount of carbon dioxide equivalent emissions (gCO₂eq) that can be reduced in comparison with a baseline diesel bus technology reference. As for energy, a well-to-wheel perspective should be used, including all emissions related to the transport service (fuels, vehicles and infrastructure). The scale (Table 9) was formulated with even steps in the range of 0 to 100%. It can be noted that the EU's renewable energy directive (Directive (EU) 2018/2001 of the European Parliament and of the Council of 11 December 2018 on the promotion of the use of energy from renewable sources.) requires 60% GHG emission savings for biofuels used from 2015 to 2020, matching the lower level of the interval for good. It is recommended to use results from life-cycle assessments based on ISO 14040 and ISO 14044 [103] that is, with system expansion to include a broad range of relevant climate issues. The LCA methods used for the calculations in the directive above do not account for essential issues, and system expansion, as recommended by the ISO standards, gives more accurate results. Information about GHG emissions is also of relevance to get a rough understanding of other global/regional environmental impact categories, indirectly providing information about emissions of NO_x and SO₂ being linked to fossil fuel use [54].

Table 9. Scale for the indicator “greenhouse gas emission savings.”

Value	Scale Definition <i>Compared to the Diesel Bus Technology Reference (of 1241 g CO₂-Eq/Vehicle Kilometer¹), the GHG Emissions Savings Are:</i>
Very good	80% or higher ($x \geq 80\%$).
Good	At least 60% or higher but not higher than 80% ($80 > x \geq 60$).
Satisfactory	At least 40% or higher but not higher than 60% ($60 > x \geq 40$).
Poor	At least 20% or higher but not higher than 40% ($40 > x \geq 20$).
Very poor	Less than 20% ($x < 20$).

¹ Based on the calculations made by Prussi et al. [104].

Air pollution consists of pollutants that commonly cause health problems in city areas and environmental problems, like carbon monoxide (CO), hydrocarbons (HC), nitrogen oxides (NO_x) and particles (PM). The transport sector is a significant contributor to local air pollution [105], causing serious negative health effects in many cities worldwide [106]. For simplicity, we have focused on tailpipe emissions, in line with EU emission standards for buses. However, there are also some limited considerations of the broader lifecycle emissions for electricity [107], as the very good level requires ‘clean and safe’ renewable energy sources, meaning that highly polluting electricity production and production linked to risks of nuclear radiation get lower grading [108–110]. Further on, it is assumed that all bus technologies contribute to similar amounts of road wear particles and are therefore not included. The scale is based on the EU emission standards (Euro V and VI), also considering the latest Swedish legislation concerning low emission zones (SFS 2018:1562). The scale for the indicator air pollution is shown in Table 10.

Table 10. Scale for the indicator “air pollution.”

Value	Scale Definition
Very good	The buses have no tailpipe emissions AND The electricity is to 100% produced from renewable sources with very low health impacting emissions, like electricity produced from water, wind or solar power. AND The electricity is NOT at all produced from nuclear power (i.e., associated with risks of nuclear radiation).
Good	The buses fulfil the requirements for Low Emission Zone 3 in Sweden, meaning that they: are driven by 100% electricity or fuel cells, OR are driven by gas engines fulfilling the Euro VI requirements, OR are chargeable hybrid vehicles fulfilling the Euro VI requirements.
Satisfactory	The buses fulfil the requirements for Euro VI.
Poor	The buses fulfil the requirements for Euro V.
Very poor	The buses do NOT fulfil the requirements for Euro V.

Noise is a common problem area for many cities, where transports are among the most prevalent sources [111,112], causing disturbances for people and wildlife [113]. There is an abundance of literature dealing with transport-related noise, for example, References [111,114,115], with a relatively large focus on modeled noise or measurements from test environments, that may provide significantly different results compared to real-life situations. Noise levels are important to understand health impacts but also psychological or psychophysiological factors for people concerned [116]. Nevertheless, it was still found reasonable to include noise levels in the MCA method. So-called A-weighted sound levels are commonly used (denoted as dBA units) to adjust the measurements to the sensitivity

of the human ear, with a focus on the range between 1 to 4 kHz. The dBA unit is widely used in assessments of transportation-related noise, commonly assumed to be the default unit [117]. Larsson and Holmes [118] studied noise related to different bus technologies and chose to use the dBA scale, with the motivation that it is most commonly used in socio-economic noise studies. However, although not using it (but providing relevant dBC data), they acknowledge that the dBC scale is also very relevant to pay more attention to low-frequency noise, which can be of relevance for transportation and several other sources (the difference between the dBC and dBA sound levels is used for information about low frequencies). It seems to be widely supported that transport noise in the range from about 10 Hz to 200 Hz can cause indoor noise annoyance [119,120], while higher frequencies contribute more to outdoor problems [121]. It should also be noticed that differently defined maximum (peak) and equivalent (average) levels are used in regulations and noise studies, such as socio-economic studies of transport noise [118,122]. According to Andersson et al. [123], based on a study of some Swedish regions, data on 24-h average sound levels are important for socio-economic noise estimations, while maximum sound levels are not as relevant. Relevant EU noise level requirements are established by EU regulation 540/2014, with maximum levels (as dBA from the vehicle, to be measured in accordance with the latest amendments (see EU regulations/amendments from 2016 and onwards.)) set for buses (focusing on vehicle type M3, which holds more than eight passengers and has engine power greater than 250 kW) for the years 2016–2020 (80 dBA), 2020–2024 (78 dBA) and after 2024 (77 dBA). Braun et al. [124] have studied different vehicle noise sources and conclude that the four major ones are the engine, intake system, exhaust system and tire/road system. Assuming similar weight and speed for the assessed bus technologies, we have focused on noise from the vehicle, that is, excluded noise from tires or roads in the MCA method. However, it is important to notice that for speeds exceeding 50 km/h, noise from the tires/road dominates [118]. Thus, in areas where buses commonly have a speed around 50 km/h or higher, it is less important to focus on noise from the engines, meaning that what bus technology is used is not of great importance regarding noise annoyance. The scale was formulated, considering all this information, as shown in Table 11, with levels chosen in relation to existing regulations.

Table 11. Scale for the indicator “noise.”

Value	Scale Definition (Engine Power 135 < 250 kW)	Scale Definition (Engine Power > 250 kW)	Comments and References
	Sound level as dBA units, measured in accordance with EU regulations	Sound level as dBA units, measured in accordance with EU regulations	
Very good	noise < 76	noise < 77	The limit from year 2024 according to EU regulation 540/2014
Good	$76 \leq \text{noise} < 77$	$77 \leq \text{noise} < 78$	The limit from year 2020 according to EU regulation 540/2014
Satisfactory	$77 \leq \text{noise} < 78$	$78 \leq \text{noise} < 80$	The limit from year 2016 according to EU regulation 540/2014
Poor	$78 \leq \text{noise} < 79$	$80 \leq \text{noise} < 81$	
Very poor	$79 < \text{noise}$	$81 < \text{noise}$	

In addition to the previously mentioned environmental indicators, bus technologies can have *other local/regional impacts on land and aquatic environments*, thus included as an indicator. Production of fuel, electricity, vehicles and infrastructure may involve both

positive and negative (local) effects, which can vary significantly depending on the choice of technology. For example:

- Fossil fuels may involve a wide range of negative, local environmental impacts [125–127].
- Biofuels or electricity produced from food waste, aquatic biomass or other relevant feedstocks may involve recycling of nutrients and reducing eutrophication [128].
- Biofuels produced from straw may lead to too low soil organic carbon levels, being negative regarding soil fertility, while there are also several examples leading to improved soil fertility [62,129].
- Feedstock production may be linked to a positive and/or negative impact on species and ecosystems nearby, for example, ecological farming favorable for biodiversity in contrast to farming involving pesticides [130].

A qualitative scale was chosen for the indicator local/regional impact on land and aquatic environments, shown in Table 12.

Table 12. Scale for the indicator “local/regional impact on land and aquatic environments.”

Value	Scale Definition <i>Focusing on Local/Regional Impact on Land/Soil, Water Resources and Aquatic Environments, Biodiversity/Ecosystems and Other Relevant Local/Regional Impacts that Are Not Clearly Covered by Any Other Indicator:</i>
Very good	The bus technology is found to be very beneficial from a local/regional environmental perspective: - There are significant positive environmental effects AND - There are no significant negative environmental effects
Good	The bus technology is found to be beneficial from a local/regional environmental perspective: - There are relevant positive environmental effects, together judged to be clearly more important than the negative effects AND - There are some negative (but still acceptable) environmental effects
Satisfactory	The bus technology is found to have no or neutral effects from a local/regional environmental perspective: - There are no significant environmental effects OR the negative and positive effects are judged to be of similar importance (where the negative are acceptable)
Poor	The bus technology is found to be negative from a local/regional environmental perspective: - There are relevant negative environmental effects, together judged to be clearly more important than the positive effects AND - There are some positive environmental effects
Very poor	The bus technology is found to be very negative from a local/regional environmental perspective: - There are significant negative environmental effects AND - There are no significant positive environmental effects

4.4. Social Performance

In addition to the economic and environmental performance, social aspects have been considered.

Energy security is defined by the International Energy Agency (IEA) as “the uninterrupted physical availability at a price which is affordable” (cf. [131,132]). In recent decades there has been a re-emerged interest in energy security, driven by rising demand, disrupted supplies and the push towards de-carbonization [133] and in both Europe [134] and Sweden [135], issues regarding energy security receive significant attention. The focus of this indicator is on the physical availability of primary resources cf. [136] and on where the

production of fuels and electricity takes place since fuel and electricity prices are covered by the economic performance indicators. A simple geographical perspective has thus been used to consider to what extent different bus solutions contribute to energy security on local/regional, national and international levels. The scale is presented in Table 13. As Sweden is an EU member country influenced by the EU energy security strategy [137], the EU level was set as the satisfactory level also including the closely connected Schengen area in order to include countries like Norway, which Sweden has close links and long-term good relations with. The resource considerations also include resources for electricity production (such as coal, oil, energy crops, etc.) where renewable sources such as wind, water or solar power are considered to be local to the electricity production site.

Table 13. Scale for the indicator “energy security.”

Value	Scale Definition
Very good	More than 90% of the used fuel or electricity is produced within the actual region ¹ , based on resources from this region.
Good	More than 90% of the used fuel or electricity is produced within the nation, based on resources of national origin.
Satisfactory	More than 90% of the used fuel or electricity is produced within countries that are geographically close to the nation, that the nation has long-term and stable business relations with, based on resources from those countries.
Poor	More than 90% of the used fuel or electricity is produced in countries that are not geographically close to the nation but which the nation has long-term and stable business relations with, based on resources from those countries.
Very poor	More than 90% of the used fuel or electricity is produced within countries that are not geographically close to the nation and which the nation does not have long-term and stable business relations with, based on resources from those countries.

¹ Referring to the term region as used in Sweden, corresponding to county. Other areas could be used.

This geographical orientation is also related to employment, even if we decided not to explicitly include it. For example, regional resource management (or production) and regional production of fuels/electricity will likely be linked to regional employment. In addition, Ekener-Petersen et al. [138] found that both fossil and biofuels can cause significant negative social impacts and emphasized the need for social performance requirements in procurements of fuels. In this context, it is important to consider the country or origin since there can be important differences concerning human and labor rights and work health. There are databases such as the Social Hot Spots Database that can be used for these kinds of assessments [138]. However, we did not choose to explicitly include such considerations as the performance may vary significantly within countries and sectors and between different production sites. Nevertheless, a recent report by The International Trade Union Confederation (ITUC), [139] indicates that EU countries, in general, are performing relatively well regarding human rights. Thus, indirectly, the above-suggested scale takes social performance into account within additional areas than those mainly targeted.

Public organizations like regions and municipalities are important actors for several sociotechnical systems, like the systems for the management of energy, transportation, waste and water [140]. As there are commonly important links between such systems involving public transportation [141], a second indicator was included to consider if and how the choice of bus technology influences the mentioned sociotechnical systems—the indicator *sociotechnical system services* (see Table 14). This assessment can include links to sociotechnical systems outside the specific area studied (such as a certain region). Links to agricultural systems are not included here as they are covered in the indicator *local/regional impact on land and aquatic environments* via involving nutrient management and soil impact.

Table 14. Scale for the indicator “sociotechnical system services.”

Value	Scale Definition
Very good	The bus technology is linked to regional/municipal sociotechnical systems of waste wastewater management and/or energy and significantly facilitates their function and/or economic viability
Satisfactory	The bus technology is not linked to regional/municipal sociotechnical systems of waste wastewater management and/or energy or does not significantly influence their function and/or economic viability
Very poor	The bus technology is linked to regional/municipal sociotechnical systems of waste wastewater management and/or energy and is significantly problematic regarding their function and/or economic viability

5. Concluding Discussion

There are several technologies for transportation, ranging from fossil systems to biofuels, electricity and other future options. There is a demand for modern, well-functioning and cost-effective technologies with superior environmental or sustainability performance to reduce climate impact, air pollution, resource depletion and other urgent challenges. On the one hand, the now-existing palette of renewable technologies brings great opportunities for the continued marginalization of fossil fuels and improved sustainability performance [104]. On the other hand, it is difficult to know what technology is preferable—that is, to systematically compare the different alternatives. This article deals with the assessment of transport technologies, focusing on the public procurement of bus transports by Swedish regional authorities, that have used ‘green or sustainable public procurement’ (GoSPP) to achieve a major transition of the bus fleet but also, to some extent, other types of vehicles [142].

Within the field of sustainability assessments, researchers emphasize the existence of implementation gaps and argue for a need to focus more on implementation in central decision-making processes [43,44]. Similarly, prior research on GoSPP and the authors’ practical experiences, point out a need for improved knowledge and supportive methods on how environmental/sustainability assessments can support public procurement processes. Our study addresses these challenges via the establishment of a multi-criteria assessment (MCA) method for assessments of public bus technologies’ sustainability. The multi-criteria approach was chosen as we see sustainability assessment of bus technologies as a complex problem and acknowledge value pluralism.

To guide the MCA method development process and for reviewing the outcome, a literature review on advice for effective and efficient sustainability assessments was conducted (see Section 2.2). Below, our method development process and the resulting MCA method are discussed in relation to keywords and criteria (in *italic*) from this review. The process of establishing the MCA method has been described in detail for *transparency*, providing more information on management, actors and indicators than most studies reviewed. The process has been iterative and *participatory*, as it engaged staff from a wide range of actors. The method results from a process governed by this study’s researchers based on literature reviews and input from project participants, other stakeholders and a few students. This mix of actors and sources has influenced the key areas, indicators and scales. Thus, it can be difficult to link a certain part of the method to a specific actor, source or part of the process, which limits the *transparency*. All the involved actors have importantly contributed to the development. Of course, an enlarged group would have brought a broader competence base, which could have improved the resulting MCA method [43]. A different composition of the group would presumably also have led to a different method [62,63]. We have tried to establish an MCA method that works well for all the assessed fuels and electricity to cover all technologies’ strengths and weaknesses. However, it may have caused bias that the project has been based within the Biogas Research Center and involved people with special competence and interests in biogas solutions. Still, several participating organizations, such as energy utility companies,

regions and municipalities, have broader interests. The truck and bus manufacturer can be described as a more neutral actor in this respect, as this company sells trucks and buses designed for all the covered fuels and electricity. In addition, the final review round, for example, involving experts on other transport technologies, was intended to reduce potential biogas/biomethane bias. In relation to the four categories of effectiveness presented by Waas et al. [43], this section has dealt with several of the *procedural* components and *normative* components since the participants learnt from the process. *Substantive and transactive effectiveness* is discussed in Part II.

For the method to be useful and efficient in relation to the aim, it is important to consider the choice of indicators and scales from several different angles. Intentionally, the method consists of both quantitative and qualitative indicators, as the purpose has been to focus on the essential issues rather than on what can be easily measured or quantified [43,143]. In this respect, the method is different from many other assessment methods for similar transport contexts, which are more quantitatively oriented [24,83,144]. As researchers, we wanted to avoid weighting, otherwise commonly applied in MCA/MCDM (Multi-Criteria Decision Making) projects, as this is a more ‘political step’ than other parts of the process and as the importance of the indicators can vary between different regions. For example, noise may be highly relevant in large city contexts but less important in the countryside. We also wanted to contribute to broadened requirements in public procurement processes [145] without including too many indicators (*practicability*). This is a balancing act, as more *comprehensive* methods will cover additional sustainability aspects but can involve overcomplicated assessments and difficulties concerning interpretation and decision making. The method consists of 4 key areas and 12 indicators. The relatively low number of indicators, without any decision trees or weighting, should make it easier to understand the *logic* and facilitate the *practical assessment*, interpretation and *visualization*. However, it is certainly possible to further simplify the use, for example, by providing a tool as exemplified by Lindfors and Ammenberg [23]. We believe that the method is broad enough to cover many essential areas, thereby helping in avoiding unintended problem shifting [146,147]. We have tried to reduce the *correlation* between the indicators but there are (certainly) overlaps, for example, environmental impacts related to total costs via taxes and similar issues. This is difficult to avoid. Several of the criteria for effective and efficient assessments are discussed in Part II, as they are closely linked to the actual assessment process and results—for example, the time needed for assessments, data availability and quality and clarity of results. One should also add that good management of uncertainties is required (part II in Reference [148]).

Reviewing the method, it is also relevant to focus on essential issues that may not be well covered by the chosen indicators. An overarching comparison with other MCA methods or sustainability assessments related to transportation shows many similarities. This is quite natural, as they influenced our choices. However, even if the same or similar indicators are included, the level of detail may vary, particularly related to scales for assessment and data collection. For example, our method includes a relatively simple and qualitative scale for infrastructure investments (cf. [93]), while others have a more quantitative and detailed approach (cf. [149]). Several studies with a similar focus have a broader consideration of social issues (cf. [26,73,88]), for example, dealing with availability/mobility and comfort. However, for the buses procured by the regions, there will not be any important differences in this respect, no matter the technology/fuel, why it was not seen as relevant to focus on (the procurement is set to match a pre-decided function or level of service that is not supposed to be influenced by choice of service provider or transport technology). The method could have been complemented with a larger focus on safety (cf. *ibid.*), considering risks related to, for example, vehicles, infrastructure, the production chain and energy systems. However, it may be problematic to assess risk levels, especially for new technologies, due to a lack of data. In the current version, risks related to nuclear power are indirectly included via the indicator “air pollution,” due to the requirements for

“very good,” which is maybe not fully *logical*. During the course of the project, discussions also focused on whether and how to include the following:

- the use of scarce natural resources and use of primary or secondary resources and implications
- flexibility related to the existence of back-up fuels (e.g., biodiesel—diesel; biomethane—natural gas)
- more detailed health effects and costs.

However, for various reasons, we decided not to directly include such indicators, which can be added in other cases or in improved versions, although some of these areas are indirectly covered, such as health costs via air pollution. Public acceptance was also discussed but left out as it was indicated by the involved actors that most customers do not care about the actual technology as long as it is renewable, which was based on previous customer surveys.

Although the method has been developed in a Swedish context, it addresses a challenge of general relevance—how to conduct sustainability assessments of transport technologies. The 4 key areas and 12 indicators would probably be relevant for assessment of buses across the globe. We find the indicators dealing with technical and economic performance to be generally applicable, while some adjustment can be reasonable for a few of those dealing with environmental and social performance: for example, to adapt the levels of the scales to fit local conditions and available technology (such as levels of energy efficiency, air pollution and noise) and adjust the scale on energy security considering relevant political and trade alliances and so forth.

Finally, we would like to stress that methods such as the one proposed can be used to find the best future transport technology or fuel but due to limited raw material and production potentials and other issues, we need a smart combination of several renewable fuels. Thus, it may be wise to also study and focus on what combination of technologies is most efficient.

Author Contributions: Conceptualization, J.A. and S.D.; methodology, J.A. and S.D.; validation, J.A. and S.D.; formal analysis, J.A. and S.D.; investigation, J.A. and S.D.; writing—original draft preparation, J.A.; writing—review and editing, J.A. and S.D.; visualization, J.A. and S.D.; supervision, J.A. and S.D.; project administration, J.A. and S.D.; funding acquisition, J.A. All authors have read and agreed to the published version of the manuscript.

Funding: This research was performed within the Biogas Research Center (BRC), which is funded by the Energy Agency of Sweden, Linköping University and the Swedish University of Agriculture and more than 20 private and public partners.

Institutional Review Board Statement: Not applicable.

Informed Consent Statement: Not applicable.

Acknowledgments: We want to thank the five students who, during 2018, contributed to this project—Jacob Dahlstedt, Agnes Lundgren, Éamon Magorrian, Linnea Orsholm and Anders Wilzén. We are also very grateful for the input from participants, stakeholders and reviewers.

Conflicts of Interest: The authors declare no conflict of interest.

References

1. Black, W.R. *Sustainable Transportation: Problems and Solutions*; Guilford Press: New York, NY, USA, 2010; ISBN 978-1-60623-905-6.
2. De Besi, M.; McCormick, K. Towards a Bioeconomy in Europe: National, Regional and Industrial Strategies. *Sustainability* **2015**, *7*, 10461–10478. [[CrossRef](#)]
3. Engert, S.; Rauter, R.; Baumgartner, R.J. Exploring the Integration of Corporate Sustainability into Strategic Management: A Literature Review. *J. Clean. Prod.* **2016**, *112*, 2833–2850. [[CrossRef](#)]
4. Nijkamp, P.; Perrels, A.; Perrels, A. *Sustainable Cities in Europe*; Routledge: Abingdon, UK, 2018; ISBN 978-1-315-06645-5.
5. Creutzig, F.; Jochem, P.; Edelenbosch, O.Y.; Mattauch, L.; van Vuuren, D.P.; McCollum, D.; Minx, J. Transport: A Roadblock to Climate Change Mitigation? *Science* **2015**, *350*, 911–912. [[CrossRef](#)] [[PubMed](#)]

6. European Commission 2030 Climate & Energy Framework. Available online: https://ec.europa.eu/clima/policies/strategies/2030_en (accessed on 3 November 2019).
7. Jackson, R.B.; Quéré, C.L.; Andrew, R.M.; Canadell, J.G.; Korsbakken, J.I.; Liu, Z.; Peters, G.P.; Zheng, B. Global Energy Growth Is Outpacing Decarbonization. *Environ. Res. Lett.* **2018**, *13*, 120401. [[CrossRef](#)]
8. Swedish Environmental Protection Agency The National Environmental Quality Objectives. Available online: <http://www.swedishepa.se/Environmental-objectives-and-cooperation/Swedens-environmental-objectives/The-national-environmental-objectives/> (accessed on 3 November 2019).
9. Börjesson, M.; Ahlgren, E.O.; Lundmark, R.; Athanassiadis, D. Biofuel Futures in Road Transport—A Modeling Analysis for Sweden. *Transp. Res. Part. Transp. Environ.* **2014**, *32*, 239–252. [[CrossRef](#)]
10. Guo, M.; Song, W.; Buhain, J. Bioenergy and Biofuels: History, Status, and Perspective. *Renew. Sustain. Energy Rev.* **2015**, *42*, 712–725. [[CrossRef](#)]
11. Xylia, M.; Silveira, S. On the Road to Fossil-Free Public Transport: The Case of Swedish Bus Fleets. *Energy Policy* **2017**, *100*, 397–412. [[CrossRef](#)]
12. Transport Analysis. Transport Work 2000–2019 (Translated). Available online: <https://www.trafa.se/globalassets/statistik/transportarbete/transportarbete-2019.pdf>? (accessed on 24 June 2019).
13. International Association of Public Transport. STATISTICS BRIEF—Urban. Public Transport. In *the 21st Century*; International Association of Public Transport: Brussels, Belgium, 2017; p. 8.
14. The Swedish Confederation of Transport Enterprises. *Statistics about the Bus Sector 2017 (Translated)*; The Swedish Confederation of Transport Enterprises: Stockholm, Sweden, 2018.
15. Aldenius, M.; Khan, J. Strategic Use of Green Public Procurement in the Bus Sector: Challenges and Opportunities. *J. Clean. Prod.* **2017**, *164*, 250–257. [[CrossRef](#)]
16. Ammenberg, J.; Anderberg, S.; Lönnqvist, T.; Grönkvist, S.; Sandberg, T. Biogas in the Transport Sector—Actor and Policy Analysis Focusing on the Demand Side in the Stockholm Region. *Resour. Conserv. Recycl.* **2018**, *129*, 70–80. [[CrossRef](#)]
17. Official Inquiry on Fossil Fuel-Free Road Transportation. *Fossil Fuel Free Road Transportation (Translated)*; Official Reports of the Swedish Government; Swedish Government: Stockholm, Sweden, 2013.
18. Xylia, M.; Leduc, S.; Patrizio, P.; Kazraxner, F.; Silveira, S. Locating Charging Infrastructure for Electric Buses in Stockholm. *Transp. Res. Part. C Emerg. Technol.* **2017**, *78*, 183–200. [[CrossRef](#)]
19. Swedish Confederation of Transport Enterprises. *Statistik Om Bussbranschen 2018*; Transportföretagen: Stockholm, Sweden, 2018; p. 134.
20. Cheng, W.; Appolloni, A.; D’Amato, A.; Zhu, Q. Green Public Procurement, Missing Concepts and Future Trends—A Critical Review. *J. Clean. Prod.* **2018**, *176*, 770–784. [[CrossRef](#)]
21. Hüging, H.; Glensor, K.; Lah, O. Need for a Holistic Assessment of Urban Mobility Measures—Review of Existing Methods and Design of a Simplified Approach. *Transp. Res. Procedia* **2014**, *4*, 3–13. [[CrossRef](#)]
22. Johansson, E.; Winslott Hiselius, L.; Koglin, T.; Wretstrand, A. Evaluation of Public Transport: Regional Policies and Planning Practices in Sweden. *Urban. Plan. Transp. Res.* **2017**, *5*, 59–77. [[CrossRef](#)]
23. Lindfors, A.; Ammenberg, J. Using National Environmental Objectives in Green Public Procurement: Method Development and Application on Transport Procurement in Sweden. *J. Clean. Prod.* **2020**. [[CrossRef](#)]
24. Tzeng, G.-H.; Lin, C.-W.; Opricovic, S. Multi-Criteria Analysis of Alternative-Fuel Buses for Public Transportation. *Energy Policy* **2005**, *33*, 1373–1383. [[CrossRef](#)]
25. Appolloni, A.; Sun, H.; Jia, F.; Li, X. Green Procurement in the Private Sector: A State of the Art Review between 1996 and 2013. *J. Clean. Prod.* **2014**, *85*, 122–133. [[CrossRef](#)]
26. Brammer, S.; Walker, H. Sustainable Procurement in the Public Sector: An International Comparative Study. *Int. J. Oper. Prod. Manag.* **2011**, *31*, 452–476. [[CrossRef](#)]
27. Preuss, L. Addressing Sustainable Development through Public Procurement: The Case of Local Government. *Supply Chain Manag. Int. J.* **2009**, *14*, 213–223. [[CrossRef](#)]
28. Günther, E.; Hueske, A.-K.; Stechemesser, K.; Buscher, L. The ‘Why Not’—Perspective of Green Purchasing: A Multilevel Case Study Analysis. *J. Chang. Manag.* **2013**, *13*, 407–423. [[CrossRef](#)]
29. Luttenberger, A.; Luttenberger, L.R. Sustainable Procurement and Environmental Life-Cycle Costing in Maritime Transport. *WMU J. Marit. Aff.* **2017**, *16*, 219–231. [[CrossRef](#)]
30. Nurhadi, L.; Borén, S.; Ny, H. Advancing from Efficiency to Sustainability in Swedish Medium-Sized Cities: An Approach for Recommending Powertrains and Energy Carriers for Public Bus Transport Systems. *Procedia Soc. Behav. Sci.* **2014**, *111*, 1218–1225. [[CrossRef](#)]
31. Michelsen, O.; de Boer, L. Green Procurement in Norway; a Survey of Practices at the Municipal and County Level. *J. Environ. Manag.* **2009**, *91*, 160–167. [[CrossRef](#)] [[PubMed](#)]
32. Testa, F.; Annunziata, E.; Iraldo, F.; Frey, M. Drawbacks and Opportunities of Green Public Procurement: An Effective Tool for Sustainable Production. *J. Clean. Prod.* **2016**, *112*, 1893–1900. [[CrossRef](#)]
33. Perera, O.; Chowdhury, N.; Goswami, A. *State of Play in Sustainable Public Procurement*; International Institute for Sustainable Development: Winnipeg, MB, Canada, 2007.

34. von Oelreich, K.; Philp, M. *Green Procurement: A Tool for Achieving National Environmental Objectives (Translated)*; Swedish EPA: Stockholm, Sweden, 2013.
35. Thomson, J.; Jackson, T. Sustainable Procurement in Practice: Lessons from Local Government. *J. Environ. Plan. Manag.* **2007**, *50*, 421–444. [[CrossRef](#)]
36. Palmujoki, A.; Parikka-Alhola, K.; Ekroos, A. Green Public Procurement: Analysis on the Use of Environmental Criteria in Contracts. *Rev. Eur. Community Int. Environ. Law* **2010**, *19*, 250–262. [[CrossRef](#)]
37. Arvidsson, A.; Stage, J. Technology-Neutral Green Procurement in Practice—An Example from Swedish Waste Management. *Waste Manag. Res.* **2012**, *30*, 519–523. [[CrossRef](#)]
38. Michelsen, G.; Adomßent, M.; Martens, P.; von Hauff, M. Sustainable Development—Background and Context. In *Sustainability Science: An Introduction*; Heinrichs, H., Martens, P., Michelsen, G., Wiek, A., Eds.; Springer: Dordrecht, The Netherlands, 2016; pp. 5–29. ISBN 978-94-017-7242-6.
39. Barkemeyer, R.; Holt, D.; Preuss, L.; Tsang, S. What Happened to the ‘Development’ in Sustainable Development? *Bus. Guidel. Two Decad. Brundtland. Sustain. Dev.* **2014**, *22*, 15–32. [[CrossRef](#)]
40. David, S.-S.; Owen, G.; Farooq, G.; Norichika, R.; Bjorn, K.; Paul, S. Integration: The Key to Implementing the Sustainable Development Goals. *Sustain. Sci.* **2017**, *12*, 911–919.
41. Holliday, C.O.J.; Schmidheiny, S.; Watts, P.; Schmidheiny, S.; Watts, P. *Walking the Talk: The Business Case for Sustainable Development*; Routledge: Abingdon, UK, 2017; ISBN 978-1-351-28195-9.
42. Ramos, T.B.; Caeiro, S.; van Hoof, B.; Lozano, R.; Huisingh, D.; Ceulemans, K. Experiences from the Implementation of Sustainable Development in Higher Education Institutions: Environmental Management for Sustainable Universities. *J. Clean. Prod.* **2015**, *106*, 3–10. [[CrossRef](#)]
43. Waas, T.; Hugé, J.; Block, T.; Wright, T.; Benitez-Capistros, F.; Verbruggen, A. Sustainability Assessment and Indicators: Tools in a Decision-Making Strategy for Sustainable Development. *Sustainability* **2014**, *6*, 5512–5534. [[CrossRef](#)]
44. Gibson, R.B. Sustainability Assessment: Basic Components of a Practical Approach. *Impact Assess. Proj. Apprais.* **2006**, *24*, 170–182. [[CrossRef](#)]
45. Ness, B.; Urbel-Piirsalu, E.; Anderberg, S.; Olsson, L. Categorising Tools for Sustainability Assessment. *Ecol. Econ.* **2007**, *60*, 498–508. [[CrossRef](#)]
46. Bueno, P.C.; Vassallo, J.M.; Cheung, K. Sustainability Assessment of Transport Infrastructure Projects: A Review of Existing Tools and Methods. *Transp. Rev.* **2015**, *35*, 622–649. [[CrossRef](#)]
47. Churchman, C.W. Guest Editorial: Wicked Problems. *Manag. Sci.* **1967**, *14*, B141–B142.
48. Rittel, H.W.J.; Webber, M.M. Dilemmas in a General Theory of Planning. *Policy Sci.* **1973**, *4*, 155–169. [[CrossRef](#)]
49. Martinez-Alier, J.; Munda, G.; O’Neill, J. Weak Comparability of Values as a Foundation for Ecological Economics. *Ecol. Econ.* **1998**, *26*, 277–286. [[CrossRef](#)]
50. Barfod, M.B.; Salling, K.B.; Leleur, S. Composite Decision Support by Combining Cost-Benefit and Multi-Criteria Decision Analysis. *Decis. Support. Syst.* **2011**, *51*, 167–175. [[CrossRef](#)]
51. Beria, P.; Maltese, I.; Mariotti, I. Multicriteria versus Cost Benefit Analysis: A Comparative Perspective in the Assessment of Sustainable Mobility. *Eur. Transp. Res. Rev.* **2012**, *4*, 137. [[CrossRef](#)]
52. Browne, D.; Ryan, L. Comparative Analysis of Evaluation Techniques for Transport Policies. *Environ. Impact Assess. Rev.* **2011**, *31*, 226–233. [[CrossRef](#)]
53. Department for Communities and Local Government. *Multi—Criteria Analysis: A Manual*; Department for Communities and Local Government: London, UK, 2009.
54. Feiz, R.; Ammenberg, J. Assessment of Feedstocks for Biogas Production, Part I—A Multi-Criteria Approach. *Resour. Conserv. Recycl.* **2017**, *122*, 373–987. [[CrossRef](#)]
55. Oltean-Dumbrava, C.; Watts, G.; Miah, A. Towards a More Sustainable Surface Transport Infrastructure: A Case Study of Applying Multi Criteria Analysis Techniques to Assess the Sustainability of Transport Noise Reducing Devices. *J. Clean. Prod.* **2016**, *112*, 2922–2934. [[CrossRef](#)]
56. Belton, V.; Stewart, T. *Multiple Criteria Decision Analysis: An Integrated Approach*; Springer: Berlin/Heidelberg, Germany, 2002.
57. Dixit, A.; McGray, H. *Analyzing Climate Change Adaption Options Useing Multi-Criteria Analysis*; World Resources Institute (WRI) and United States Agency for International Development: Washington, DC, USA, 2013.
58. Mendoza, G.A.; Macoun, P.; Prabhu, R.; Sukadri, D.; Purnomo, H.; Hartanto, H. *Guidelines for Applying Multi-Criteria Analysis to the Assessment of Criteria and Indicators*; CIFOR: Bogor, Indonesia, 1999.
59. Wedley, W.C. Combining Qualitative and Quantitative Factors—An Analytic Hierarchy Approach. *Socioecon. Plann. Sci.* **1990**, *24*, 57–64. [[CrossRef](#)]
60. Loucks, D.P.; Gladwell, J.S.; Programme, I.H. *Sustainability Criteria for Water Resource Systems*; Cambridge University Press: Cambridge, UK, 1999; ISBN 978-0-521-56044-3.
61. Tudela, A.; Akiki, N.; Cisternas, R. Comparing the Output of Cost Benefit and Multi-Criteria Analysis: An Application to Urban Transport Investments. *Transp. Res. Part. Policy Pract.* **2006**, *40*, 414–423. [[CrossRef](#)]
62. Ammenberg, J.; Feiz, R. Assessment of Feedstocks for Biogas Production, Part II—Results for Strategic Decision Making. *Resour. Conserv. Recycl.* **2017**, *122*, 388–404. [[CrossRef](#)]

63. Annema, J.A.; Mouter, N.; Razaeei, J. Cost-Benefit Analysis (CBA), or Multi-Criteria Decision-Making (MCDM) or Both: Politicians' Perspective in Transport Policy Appraisal. *Transp. Res. Procedia* **2015**, *10*, 788–797. [CrossRef]
64. Thomopoulos, N.; Grant-Muller, S. Incorporating Equity as Part of the Wider Impacts in Transport Infrastructure Assessment: An Application of the SUMINI Approach. *Transportation* **2013**, *40*, 315–345. [CrossRef]
65. Achillas, C.; Moussiopoulos, N.; Karagiannidis, A.; Baniyas, G.; Perkoulidis, G. The Use of Multi-Criteria Decision Analysis to Tackle Waste Management Problems: A Literature Review. *Waste Manag. Res.* **2013**, *31*, 115–129. [CrossRef]
66. Ananda, J.; Herath, G. A Critical Review of Multi-Criteria Decision Making Methods with Special Reference to Forest Management and Planning. *Ecol. Econ.* **2009**, *68*, 2535–2548. [CrossRef]
67. Buchholz, T.; Rametsteiner, E.; Volk, T.A.; Luzadis, V.A. Multi Criteria Analysis for Bioenergy Systems Assessments. *Energy Policy* **2009**, *37*, 484–495. [CrossRef]
68. Diaz-Balteiro, L.; González-Pachón, J.; Romero, C. Measuring Systems Sustainability with Multi-Criteria Methods: A Critical Review. *Eur. J. Oper. Res.* **2017**, *258*, 607–616. [CrossRef]
69. Govindan, K.; Rajendran, S.; Sarkis, J.; Murugesan, P. Multi Criteria Decision Making Approaches for Green Supplier Evaluation and Selection: A Literature Review. *J. Clean. Prod.* **2015**, *98*, 66–83. [CrossRef]
70. Herva, M.; Roca, E. Review of Combined Approaches and Multi-Criteria Analysis for Corporate Environmental Evaluation. *J. Clean. Prod.* **2013**, *39*, 355–371. [CrossRef]
71. Macharis, C.; Bernardini, A. Reviewing the Use of Multi-Criteria Decision Analysis for the Evaluation of Transport Projects: Time for a Multi-Actor Approach. *Transp. Policy* **2015**, *37*, 177–186. [CrossRef]
72. Taha, R.A.; Daim, T. Multi-Criteria Applications in Renewable Energy Analysis, a Literature Review. In *Research and Technology Management in the Electricity Industry*; Daim, T., Oliver, T., Kim, J., Eds.; Green Energy and Technology; Springer: London, UK, 2013; pp. 17–30, ISBN 978-1-4471-5096-1.
73. Wang, J.-J.; Jing, Y.-Y.; Zhang, C.-F.; Zhao, J.-H. Review on Multi-Criteria Decision Analysis Aid in Sustainable Energy Decision-Making. *Renew. Sustain. Energy Rev.* **2009**, *13*, 2263–2278. [CrossRef]
74. Foxon, T.J.; McIlkenny, G.; Gilmour, D.; Oltean-Dumbrava, C.; Souter, N.; Ashley, R.; Butler, D.; Pearson, P.; Jowitt, P.; Moir, J. Sustainability Criteria for Decision Support in the UK Water Industry. *J. Environ. Plan. Manag.* **2002**, *45*, 285–301. [CrossRef]
75. Oltean-Dumbrava, C.; Watts, G.; Miah, A. Transport Infrastructure: Making More Sustainable Decisions for Noise Reduction. *J. Clean. Prod.* **2013**, *42*, 58–68. [CrossRef]
76. Saaty, T.L. A Scaling Method for Priorities in Hierarchical Structures. *J. Math. Psychol.* **1977**, *15*, 234–281. [CrossRef]
77. Baker, D.C.; McLelland, J.N. Evaluating the Effectiveness of British Columbia's Environmental Assessment Process for First Nations' Participation in Mining Development. *Environ. Impact Assess. Rev.* **2003**, *23*, 581–603. [CrossRef]
78. Buzási, A.; Csete, M. Sustainability Indicators in Assessing Urban Transport Systems. *Period. Polytech. Transp. Eng.* **2015**, *43*, 138–145. [CrossRef]
79. Efroymson, R.A.; Dale, V.H. Environmental Indicators for Sustainable Production of Algal Biofuels. *Ecol. Indic.* **2015**, *49*, 1–13. [CrossRef]
80. Haghshenas, H.; Vaziri, M. Urban Sustainable Transportation Indicators for Global Comparison. *Ecol. Indic.* **2012**, *15*, 115–121. [CrossRef]
81. Joumard, R.; Gudmundsson, H.; Folkesson, L. Framework for Assessing Indicators of Environmental Impacts in the Transport Sector. *Transp. Res. Rec.* **2011**, *2242*, 55–63. [CrossRef]
82. Zito, P.; Salvo, G. Toward an Urban Transport Sustainability Index: An European Comparison. *Eur. Transp. Res. Rev.* **2011**, *3*, 179–195. [CrossRef]
83. Reisi, M.; Aye, L.; Rajabifard, A.; Ngo, T. Transport Sustainability Index: Melbourne Case Study. *Ecol. Indic.* **2014**, *43*, 288–296. [CrossRef]
84. Roberts, N. Wicked Problems and Network Approaches to Resolution. *Int. Public Manag. Rev.* **2000**, *1*, 1–19.
85. Feiz, R.; Ammenberg, J.; Baas, L.; Eklund, M.; Helgstrand, A.; Marshall, R. Improving the CO₂ Performance of Cement, Part II: Framework for Assessing CO₂ Improvement Measures in the Cement Industry. *J. Clean. Prod.* **2015**, *98*, 282–291. [CrossRef]
86. Fridas Användarförening FRIDA Miljö- Och Fordonsdatabas. Available online: <http://www.frida.port.se/hemsidan/default.cfm> (accessed on 14 January 2021).
87. Swedish Energy Agency. *Drivmedel 2018*; Swedish Energy Agency: Bromma, Sweden, 2019.
88. Osorio-Tejada, J.L.; Llera-Sastresa, E.; Scarpellini, S. A Multi-Criteria Sustainability Assessment for Biodiesel and Liquefied Natural Gas as Alternative Fuels in Transport Systems. *J. Nat. Gas. Sci. Eng.* **2017**, *42*, 169–186. [CrossRef]
89. Barbosa, S.B.; Ferreira, M.G.G.; Nickel, E.M.; Cruz, J.A.; Forcellini, F.A.; Garcia, J.; Guerra, J.B.S.O. Multi-Criteria Analysis Model to Evaluate Transport Systems: An Application in Florianópolis, Brazil. *Transp. Res. Part. Policy Pract.* **2017**, *96*, 1–13. [CrossRef]
90. Mankins, J.C. Technology Readiness Assessments: A Retrospective. *Acta Astronaut.* **2009**, *65*, 1216–1223. [CrossRef]
91. Kemp, R.; Schot, J.; Hoogma, R. Regime Shifts to Sustainability through Processes of Niche Formation: The Approach of Strategic Niche Management. *Technol. Anal. Strateg. Manag.* **1998**, *10*, 175–198. [CrossRef]
92. Vigren, A. Costs in Swedish Public Transport: An Analysis of Cost Drivers and Cost Efficiency in Public Transport Contracts. Ph.D. Thesis, KTH Royal Institute of Technology, Stockholm, Sweden, 2015.
93. Sehatpour, M.-H.; Kazemi, A.; Sehatpour, H. Evaluation of Alternative Fuels for Light-Duty Vehicles in Iran Using a Multi-Criteria Approach. *Renew. Sustain. Energy Rev.* **2017**, *72*, 295–310. [CrossRef]

94. Ecotrafic. *Kunskapssammanställning—Stadsbussar Euro VI*; Ecotrafic: Stockholm, Sweden, 2015.
95. Sundström, B.; Legerius, B. *Kollektivtrafikens Kostnadsutveckling—En Överblick. Vad Förklarar Utvecklingen 2011–2015?* Sveriges Kommuner och Landsting: Stockholm, Sweden, 2017; p. 41.
96. Bloomfield, P. The Challenging Business of Long-Term Public–Private Partnerships: Reflections on Local Experience. *Public Adm. Rev.* **2006**, *66*, 400–411. [[CrossRef](#)]
97. Guinée, J.B. Selection of Impact Categories and Classification of LCI Results to Impact Categories. In *Life Cycle Impact Assessment*; Hauschild, M.Z., Huijbregts, M.A.J., Eds.; LCA Compendium—The Complete World of Life Cycle Assessment; Springer: Dordrecht, The Netherlands, 2015; pp. 17–37. ISBN 978-94-017-9744-3.
98. McBride, A.C.; Dale, V.H.; Baskaran, L.M.; Downing, M.E.; Eaton, L.M.; Efroymson, R.A.; Garten, C.T.; Kline, K.L.; Jager, H.I.; Mulholland, P.J.; et al. Indicators to Support Environmental Sustainability of Bioenergy Systems. *Ecol. Indic.* **2011**, *11*, 1277–1289. [[CrossRef](#)]
99. Gustafsson, M.; Svensson, N.; Anderberg, S. Energy Performance Indicators as Policy Support for Public Bus Transport—The Case of Sweden. *Transp. Res. Part. Transp. Environ.* **2018**, *65*, 697–709. [[CrossRef](#)]
100. Sara, R.; Bo, L.; Ringqvist, S. *Öppna Jämförelser—Kollektivtrafik 2017*; Sveriges Kommuner och Landsting: Stockholm, Sweden, 2017.
101. Wolff, E.; Arnell, N.; Friedlingstein, P.; Gregory, J.; Haigh, J.; Haines, A.; Hawkins, E.; Hegerl, G.; Hoskins, B.; Mace, G.; et al. The Royal Society Climate Updates: What Have We Learnt since the IPCC 5th Assessment Report? Available online: <https://royalsociety.org/-/media/policy/Publications/2017/27-11-2017-Climate-change-updates-report.pdf> (accessed on 24 June 2019).
102. Bulkeley, H.; Newell, P.; Newell, P. *Governing Climate Change*; Routledge: Abingdon, UK, 2015; ISBN 978-1-315-75823-7.
103. Ekvall, T.; Finnveden, G. Allocation in ISO 14041—A Critical Review. *J. Clean. Prod.* **2001**, *9*, 197–208. [[CrossRef](#)]
104. Prussi, C.M.; Yugo, M.; Prada, L.D.; Padella, M.; Edwards, R.; Lonza, L. *JRC Sciency for Policy Report. JEC Well-to-Tank Report v5. Well-to-Wheels Analysis of Future Automotive Fuels and Powertrains in the European Context*; European Commission: Luxembourg, 2020.
105. Anenberg, S.; Miller, J.; Henze, D.; Minjares, R. *A Global Snapshot of the Air Pollution-Related Health Impacts of Transportation Sector Emissions in 2010 and 2015*; ICCT: Washington, DC, USA, 2019.
106. Lelieveld, J.; Evans, J.S.; Fnais, M.; Giannadaki, D.; Pozzer, A. The Contribution of Outdoor Air Pollution Sources to Premature Mortality on a Global Scale. *Nature* **2015**, *525*, 367–371. [[CrossRef](#)]
107. Gagnon, L.; Bélanger, C.; Uchiyama, Y. Life-Cycle Assessment of Electricity Generation Options: The Status of Research in Year 2001. *Energy Policy* **2002**, *30*, 1267–1278. [[CrossRef](#)]
108. Behling, N.; Williams, M.C.; Behling, T.G.; Managi, S. Aftermath of Fukushima: Avoiding Another Major Nuclear Disaster. *Energy Policy* **2019**, *126*, 411–420. [[CrossRef](#)]
109. Monson, P.C. Radioactive Air Pollution from Uranium Mining: Regulatory Abdication in the Face of Scientific Uncertainty Comment. *Environ. Law* **1982**, *13*, 545–588.
110. Przystupa, K.; Vasylykivskyi, I.; Ishchenko, V.; Pohrebennyk, V.; Kochan, O.; Su, J. Assessing Air Pollution from Nuclear Power Plants. In Proceedings of the 2019 12th International Conference on Measurement, Smolenice, Slovakia, 27–29 May 2019; IEEE: New York, NY, USA, 2019; pp. 232–235.
111. Braubach, M.; Tobollik, M.; Mudu, P.; Hiscock, R.; Chapizanis, D.; Sarigiannis, D.A.; Keuken, M.; Perez, L.; Martuzzi, M. Development of a Quantitative Methodology to Assess the Impacts of Urban Transport Interventions and Related Noise on Well-Being. *Int. J. Environ. Res. Public Health* **2015**, *12*, 5792–5814. [[CrossRef](#)]
112. Lercher, P. Noise in Cities: Urban and Transport Planning Determinants and Health in Cities. In *Integrating Human Health into Urban and Transport Planning: A Framework*; Nieuwenhuijsen, M., Khreis, H., Eds.; Springer International Publishing: Cham, Switzerland, 2019; pp. 443–481. ISBN 978-3-319-74983-9.
113. Shannon, G.; McKenna, M.F.; Angeloni, L.M.; Crooks, K.R.; Fristrup, K.M.; Brown, E.; Warner, K.A.; Nelson, M.D.; White, C.; Briggs, J.; et al. A Synthesis of Two Decades of Research Documenting the Effects of Noise on Wildlife. *Biol. Rev.* **2016**, *91*, 982–1005. [[CrossRef](#)] [[PubMed](#)]
114. Brown, A.L.; Van Kamp, I. WHO Environmental Noise Guidelines for the European Region: A Systematic Review of Transport Noise Interventions and Their Impacts on Health. *Int. J. Environ. Res. Public Health* **2017**, *14*, 873. [[CrossRef](#)] [[PubMed](#)]
115. Kim, U.; Maunder, M.; Grant, P.; Mawdsley, D. *Developing a Car to Meet New Pass-By Noise Requirements Using Simulation and Testing*; SAE Technical Paper; SAE: Warrendale, PA, USA, 2015.
116. Shepherd, D.; Welch, D.; Dirks, K.N.; Mathews, R. Exploring the Relationship between Noise Sensitivity, Annoyance and Health-Related Quality of Life in a Sample of Adults Exposed to Environmental Noise. *Int. J. Environ. Res. Public Health* **2010**, *7*, 3579–3594. [[CrossRef](#)]
117. US Federal Highway Administration. *Noise Measurement Handbook*; US Federal Highway Administration: Washington, DC, USA, 2018.
118. Larsson, K.; Holmes, M. *Nyttöberäkningar av Minskat Buller från Elbusstrafik I Göteborg*; RISE—Research Institutes of Sweden: Gothenburg, Sweden, 2016.
119. Leventhall, H.G. Low Frequency Noise and Annoyance. *Noise Health* **2004**, *6*, 59. [[PubMed](#)]

120. Waye, K.P. Effects of Low Frequency Noise on Sleep. *Noise Health* **2004**, *6*, 87.
121. Höstmad, P.; Bergman, P.; Fredriksson, K. Off-Peak Low Noise Heavy-Duty Vehicles, Façade Insulation and Indoor Noise Disturbance. Available online: <https://www.ingentaconnect.com/contentone/incc/inccp/2016/00000253/00000003/art00059> (accessed on 28 June 2019).
122. Van Essen, H.H.; Boon, B.B.; Mitchell, S.S.; Yates, D.D.; Greenwood, D.D.; Porter, N.N. *Sound Noise Limits*; Commission Européenne: Delft, The Netherlands, 2005.
123. Andersson, H.; Swärdh, J.-E.; Ögren, M. *Traffic Noise Effects of Property Prices: Hedonic Estimates Based on Multiple Noise Indicators*; Centre for Transport Studies: Stockholm, Sweden, 2015.
124. Braun, M.E.; Walsh, S.J.; Horner, J.L.; Chuter, R. Noise Source Characteristics in the ISO 362 Vehicle Pass-by Noise Test: Literature Review. *Appl. Acoust.* **2013**, *74*, 1241–1265. [[CrossRef](#)]
125. Allen, L.; Cohen, M.J.; Abelson, D.; Miller, B. Fossil Fuels and Water Quality. In *The World's Water: The Biennial Report on Freshwater Resources*; Gleick, P.H., Ed.; The World's Water; Island Press/Center for Resource Economics: Washington, DC, USA, 2011; pp. 73–96, ISBN 978-1-61091-048-4.
126. Burton, G.A.; Basu, N.; Ellis, B.R.; Kapo, K.E.; Entrekin, S.; Nadelhoffer, K. Hydraulic “Fracking”: Are Surface Water Impacts an Ecological Concern? *Environ. Toxicol. Chem.* **2014**, *33*, 1679–1689. [[CrossRef](#)]
127. Mendelsohn, I.A.; Andersen, G.L.; Baltz, D.M.; Caffey, R.H.; Carman, K.R.; Fleegeer, J.W.; Joye, S.B.; Lin, Q.; Maltby, E.; Overton, E.B.; et al. Oil Impacts on Coastal Wetlands: Implications for the Mississippi River Delta Ecosystem after the Deepwater Horizon Oil Spill. *BioScience* **2012**, *62*, 562–574. [[CrossRef](#)]
128. Hagman, L.; Eklund, M. *The Role of Biogas Solutions in the Circular and Bio-Based Economy*; Biogas Research Center (BRC), Linköping University: Linköping, Sweden, 2016.
129. Prade, T.; Svensson, S.-E.; Björnsson, L. Introduction of Grass-Clover Crops as Biogas Feedstock in Cereal-Dominated Crop Rotations. In Part I: Effects on Soil Organic Carbon and Food Production. In Proceedings of the 9th International Conference on Life Cycle Assessment in the Agri-Food Sector (LCA Food 2014), San Francisco, CA, USA, 8–10 October 2014; pp. 8–10.
130. Bengtsson, J.; Ahnström, J.; Weibull, A.-C. The Effects of Organic Agriculture on Biodiversity and Abundance: A Meta-Analysis. *J. Appl. Ecol.* **2005**, *42*, 261–269. [[CrossRef](#)]
131. Jewell, J. *The IEA Model of Short-Term Energy Security (MOSES)*; IEA Energy Paper: Paris, France, 2011.
132. Winzer, C. Conceptualizing Energy Security. *Energy Policy* **2012**, *46*, 36–48. [[CrossRef](#)]
133. Cherp, A.; Jewell, J. The Concept of Energy Security: Beyond the Four As. *Energy Policy* **2014**, *75*, 415–421. [[CrossRef](#)]
134. Jonsson, D.K.; Johansson, B.; Månsson, A.; Nilsson, L.J.; Nilsson, M.; Sonnsjö, H. Energy Security Matters in the EU Energy Roadmap. *Energy Strategy Rev.* **2015**, *6*, 48–56. [[CrossRef](#)]
135. Johansson, B.; Jonsson, D.K.; Veibäck, E.; Sonnsjö, H. Assessing the Capabilities to Manage Risks in Energy Systems—Analytical Perspectives and Frameworks with a Starting Point in Swedish Experiences. *Energy* **2016**, *116*, 429–435. [[CrossRef](#)]
136. Månsson, A.; Johansson, B.; Nilsson, L.J. Assessing Energy Security: An Overview of Commonly Used Methodologies. *Energy* **2014**, *73*, 1–14. [[CrossRef](#)]
137. Matsumoto, K.; Doumpos, M.; Andriopoulos, K. Historical Energy Security Performance in EU Countries. *Renew. Sustain. Energy Rev.* **2018**, *82*, 1737–1748. [[CrossRef](#)]
138. Ekener-Petersen, E.; Höglund, J.; Finnveden, G. Screening Potential Social Impacts of Fossil Fuels and Biofuels for Vehicles. *Energy Policy* **2014**, *73*, 416–426. [[CrossRef](#)]
139. The International Trade Union Confederation (ITUC). *2018 ITUC GLOBAL RIGHTS INDEX The World's Worst Countries for Workers*; The International Trade Union Confederation: Brussels, Belgium, 2019.
140. Genon, G. Tasks of Local Public Services for Environmental Protection. In Proceedings of the 6th International Conference on Urban Regeneration and Sustainability, La Coruna, Spain, 14–16 April 2010; pp. 191–201.
141. Fallde, M.; Eklund, M. Towards a Sustainable Socio-Technical System of Biogas for Transport: The Case of the City of Linköping in Sweden. *J. Clean. Prod.* **2015**, *98*, 17–28. [[CrossRef](#)]
142. Miljöbarometern Andel Förnybara Drivmedel. Available online: <http://2030.miljobarometern.se/kommun/verksamhet/branslet/andel-fornybara-drivmedel-b2a-kv/compare> (accessed on 3 November 2019).
143. Cameron, W.B. *Informal Sociology: A Casual Introduction to Sociological Thinking*; Random House: New York, NY, USA, 1963; Volume 21.
144. Harris, M.A.; Soban, D.; Smyth, B. Recommendations for a Whole Life Cycle Economic and Environmental Impact Technology Assessment Tool for Alternative Driveline Bus Fleets. In Proceedings of the Irish Transport Research Network Conference, Dublin, Ireland, 28–29 August 2017.
145. Bratt, C.; Hallstedt, S.; Robèrt, K.-H.; Broman, G.; Oldmark, J. Assessment of Criteria Development for Public Procurement from a Strategic Sustainability Perspective. *J. Clean. Prod.* **2013**, *52*, 309–316. [[CrossRef](#)]
146. De Haes, H.U.; Van Rooijen, M. Life Cycle Approaches—The Road from Analysis to Practice. *UNEP/SETAC Life Cycle Initiat.* **2005**, 89.
147. Lindfors, A.; Feiz, R.; Eklund, M.; Ammenberg, J. Assessing the Potential, Performance and Feasibility of Urban Solutions: Methodological Considerations and Learnings from Biogas Solutions. *Sustainability* **2019**, *11*, 3756. [[CrossRef](#)]

-
148. Feiz, R. Systems Analysis for Eco-Industrial Development: Applied on Cement and Biogas Production Systems. Ph.D. Thesis, Linköping University, Linköping, Sweden, 2016.
 149. Barfod, M.B.; Salling, K.B. A New Composite Decision Support Framework for Strategic and Sustainable Transport Appraisals. *Transp. Res. Part. Policy Pract.* **2015**, *72*, 1–15. [[CrossRef](#)]