



Article Estimating the Carbon Emissions of Remotely Sensed Energy-Intensive Industries Using VIIRS Thermal Anomaly-Derived Industrial Heat Sources and Auxiliary Data

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Abstract: The energy-intensive industrial sector (EIIS) occupies a majority of global CO₂ emissions, but spatially monitoring the spatiotemporal dynamics of these emissions remains challenging. In this study, we used the Chinese province with the largest carbon emissions, Shandong Province, as an example to investigate the capacity of remotely sensed thermal anomaly products to identify annual industrial heat source (IHS) patterns at a 1 km resolution and estimated the carbon emissions of these sources using auxiliary datasets and the boosting regression tree (BRT) model. The IHS identification accuracy was evaluated based on two IHS references and further attributed according to corporate inventory data. We followed a bottom-up approach to estimate carbon emissions for each IHS object and conducted model fitting using the explanatory strength of the annual population density, nighttime light (NTL), and relevant thermal characteristic information derived from the Visible Infrared Imaging Radiometer Suite (VIIRS). We generated a time series of IHS distributions from 2012 to 2020 containing a total of over 3700 IHS pixels exhibiting better alignment with the reference data than that obtained in previous work. The results indicated that the identified IHSs mostly belonged to the EIIS, such as energy-related industries (e.g., thermal power plants) and heavy manufacturing industries (e.g., chemistry and cement plants), that primarily use coal and coke as fuel sources. The BRT model exhibited a good performance, explaining 61.9% of the variance in the inventory-based carbon emissions and possessing an index of agreement (IOA) of 0.83, suggesting a feasible goodness of fit of the model when simulating carbon emissions. Explanatory variables such as the population density, thermal power radiation, NTL, and remotely sensed thermal anomaly durations were found to be important factors for improving carbon emissions modeling. The method proposed in this study is useful to aid management agencies and policymakers in tracking the carbon footprint of the EIIS and regulating high-emission corporations to achieve carbon neutrality.

Keywords: thermal anomalies; boosting regression tree; industrial carbon emissions; DBSCAN; Shandong Province

1. Introduction

Fossil fuel consumption is the dominant source of atmospheric greenhouse gas (GHG) concentration increases, which are believed to be responsible for global climate warming and its resulting negative impacts on human well-being, ecosystem health, biodiversity, sea-level rise, and other public concerns [1–6]. According to an assessment report released by the Intergovernmental Panel on Climate Change (IPCC), carbon emissions are sourced from the energy-intensive industrial sector (EIIS), such as the energy (14.0 Gt, ~41.8%) and industry (6.16 Gt, 18.4%) sectors, which together accounted for approximately 60.4% of the total CO_2 emissions in 2019 [7]. In China, these two sectors together accounted for approximately 83% of the total CO_2 emissions of 15.0 Gt [8]. Quantifying the spatial patterns of the carbon emissions from these industries and tracking changes in the EIIS



Citation: Kong, X.; Wang, X.; Jia, M.; Li, Q. Estimating the Carbon Emissions of Remotely Sensed Energy-Intensive Industries Using VIIRS Thermal Anomaly-Derived Industrial Heat Sources and Auxiliary Data. *Remote Sens.* 2022, *14*, 2901. https://doi.org/10.3390/ rs14122901

Academic Editors: Jian Yang and Le Yu

Received: 24 May 2022 Accepted: 14 June 2022 Published: 17 June 2022

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Copyright: © 2022 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). thus provides critical benchmark information for policymakers and management agencies aiming to regulate the climate crisis [9,10]. However, regular inventory methods cannot be applied to these industries due to the vast number of corporations present in this sector as well as many of these corporations being located remotely and running irregularly. On the other hand, there is an urgent need to improve the spatial and temporal resolutions of GHG emissions data to support the achievement of carbon neutrality.

Energy-intensive industries (e.g., thermal power plants, the cement industry, and the metal-smelting industry) are usually high polluters due to the fact of their burning of fossil fuels, as this process can produce waste gas, wastewater, and heat anomalies in contrast with the relatively cool surrounding environment. New opportunities for observing and monitoring EIIS activities have arisen from recent advances and the inherent advantages in thermal infrared remote sensing techniques, such as the high sensitivity of these methods to surface heat anomalies [11,12], the capacity of continuous and up-to-date observations obtained at relatively high resolutions, and the critically important objectiveness and consistency of the resulting data. An increasing number of available thermal infrared sensors, such as the Moderate Resolution Imaging Spectroradiometer (MODIS), the Advanced Spaceborne Thermal Emission and Reflection Radiometer (ASTER), the Visible Infrared Imaging Radiometer Suite (VIIRS), and instruments onboard the Landsat-8 and -9 satellites, have enabled improved surface heat anomaly detection methods related to active fires [13–17] and urban heat islands [18–22].

The application of space-borne thermal imagery or products for identifying and monitoring industrial heat sources (IHSs) through various approaches, such as spatial clustering analyses [23–25], thermal anomaly indices [26,27], and object-oriented approaches [28], has also gained increasing interest. Although the IHS identification accuracy differed among these past studies, the distinct temperature differences observed between industrial processing operation areas (e.g., boilers, blast furnaces, and chimneys) and the cool background can enable the precise detection of IHS objects [26]. The spatial distributions, amounts, and temporal variations in IHSs derived from these past efforts form a critical foundation on which our understanding of how anthropogenic industrial activities influence the environment at various scales can be improved. However, research regarding the attribution of remotely sensed IHS objects remains problematic and lacks evaluations of which types of industrial sectors can be detected [28]. Though some studies have determined that the industrial sectors, which correspond to the remotely sensed IHS objects that can be observed through visual interpretations [23,27,28], may have overlooked the fact that several industrial sectors may coexist within a given remotely sensed IHS polygon.

Furthermore, the spatially explicated tracking of carbon emission footprints is a pragmatic and urgent requirement for environmental governing agencies, but this work remains challenging. State-of-the-art remote sensing techniques have been effectively applied to retrieve CO₂ concentrations using CO₂-observing satellites (e.g., the GHG Observing Satellite (GOSAT), Orbiting Carbon Observatory 2 (OCO-2), and China's TanSat) [29–34] and indirectly estimate carbon emissions by combining nightlight observations with indicators that can reflect socioeconomic activities [35–37]. Remotely sensed thermal anomalies have also been applied to estimate carbon emissions from global wildfires and volcanic eruption events [38–42], but the linkages between these anomalies and industrial-induced carbon emissions have rarely been explored. Integrating remotely sensed IHSs with fine-scale governmental inventory data may provide new insights that can improve the existing GHG emission estimation methodologies.

Therefore, the objectives of our study were two-fold: (1) to determine the explicit types of IHSs that can be detected using VIIRS thermal products and auxiliary datasets and (2) to evaluate the feasibility of estimating carbon emissions by combining remotely sensed IHSs, nightlight observations, and machine learning approaches. We used heat anomaly data derived from the VIIRS active fire product to recognize IHSs based on a spatial cluster analysis approach and auxiliary dataset. We used the corporation-level position of interest (POI) data features with thoroughly official attributive information

to verify the potential strength of the remotely sensed heat anomalies to depict the EIIS. The boosting regression tree (BRT) model was employed to quantify the predictive power of the utilized combination of remotely sensed variables and the official carbon emission footprint inventory to obtain a wall-to-wall map of carbon emissions in Shandong Province, China. The spatially explicit and straightforward modeling proposed in this study is the first attempt to link the remotely sensed IHS objects with their in situ carbon emissions. The combination of physically and empirically based approaches is novel and easy to generalize with other study areas. This can play a vital role in refining the estimation of industrial carbon emissions and forming a consistent approach for tracking carbon footprint dynamics.

2. Materials and Methods

2.1. Study Area

We chose Shandong Province of China as our study area, because it has the largest amount of carbon emissions among all Chinese provinces reported in recent years and has the highest availability of corporation-level inventory data regarding carbon emissions and industrial sector attributes for use in this study (Figure 1). The gross domestic product (GDP) of Shandong Province was over USD 1 trillion (RMB ~8.31 trillion) in 2021, ranking within the top 3 provinces with the highest GDPs in China. The population of this province is over 100 million people, and the land area is 1.558×10^4 km². This province is also well known for its rapid precision farming development, especially regarding vegetable and grain cultivation. At the same time, the province represents a typical Chinese industrial structure dominated by the manufacturing industry mixed with heavy energy-intensive industries such as the power generation industry, steel industry, chemical industry, and mineral and energy extraction industry. The widely distributed farmlands within this province can provide a cool background with which IHSs can be recognized according to thermal anomaly and nightlight datasets.



Figure 1. Spatial location of Shandong Province (**A**) in China (**B**). The background map represents the land cover pattern of the study area in 2020 based on the European Space Agency (ESA) Climate Change Initiative Land Cover (CCI-LC) product.

2.2. Data Sources and Materials

2.2.1. VIIRS-Derived Thermal Anomaly Product

We obtained the Suomi National Polar-Orbiting Partnership (NPP) satellite-derived VIIRS active fire product representing the 2012–2020 period from the Fire Information for Resource Management System (FIRMS) data center under the administration of the National Aeronautics and Space Administration (NASA). This product can provide the global spatial distribution of thermally anomalous pixels (TAPs) at a 375 m spatial resolution, synthesized and distributed annually [14]. The dataset also provides a series of detailed thermal characteristics and attributes including the spatial locations of the observations, the brightness temperatures of the VIIRS I-4 and I-5 channels, the acquisition dates and times, the fire radiative power (FRP), and other attributive characteristics related to the confidence level and inferred fire type. In this study, we obtained a total of 0.214 million TAPs from the VIIRS active fire product (Figure 2). We retained all TAPs for the subsequent analysis considering that IHSs may be characterized by relatively low brightness temperatures and may have low confidence for active fire classification in the product.



Figure 2. Spatial distribution of raw thermally anomalous pixels (TAPs) in Shandong Province. The total number of TAPs that fell in each 1×1 km pixel was logarithmically transformed to improve the visual effect. Two snapshots represent the TAP distribution in Zibo (**A**) and Linyi (**B**) cities, which feature the petrochemical industry and manufacturing industry, respectively.

2.2.2. ESA_CCI Land Cover Dataset

To determine the land cover type corresponding to each TAP, we used a global annual land cover time series dataset with a 300 m spatial resolution named the European Space Agency (ESA) Climate Change Initiative Land Cover (CCI-LC) product. The land cover types included in this dataset follow a classification scheme including 33 land cover types with an overall global accuracy of 71.1% [43]. We extracted the bare area and settlement

categories as the mask to filter out TAPs that were related to forest fires and agricultural residual burning (see Figure 1). To effectively align these data with the temporal information characterizing the thermal anomaly time series, we obtained land cover maps from two versions of the ESA_CCI_LC product. The EAS_CCI_LC V2.0.7 product was used to cover the period from 2012 to 2015, and the EAS_CCI_LC V2.1.1 product was used for the 2016–2020 period. The annual mask area, among which the settlement class accounted for over 99.8%, tended to increase year by year from 1.28×10^4 km² in 2012 to 1.79×10^4 km² in 2020 (see Appendix A Figure A1).

2.2.3. Remotely Sensed Nighttime Light (NTL) and Gridded Population Data

It is well established that nighttime light (NTL) is a useful indicator that reflects the intensity of socioeconomic activities [44]. Thus, we assumed that NTL information, the spatiotemporal pattern of which can be quantified explicitly thanks to recent advances in satellite sensors, may be very closely linked with carbon emissions due to the fact of its strong explanatory power for many industrial-related activities such as energy consumption [45–47]. We used a harmonized version of the global 1 km annual NTL time series (1992–2020) product generated by Li et al. [48] to quantify the accumulated radiance of the recognized IHSs. This NTL product was sourced from the satellite sensors of the Operational Linescan System (OLS) of the Defense Meteorological Satellite Program (DMSP) and the VIIRS on the Suomi NPP. The NTL time series data representing the 1992–2013 period were derived from the intercalibrated observations of the DMSP-OLS sensor, while the 2014–2020 NTL time series data were simulated from the VIIRS observations using a harmonization process. In this study, we used the 2012–2020 NTL time series to support our analysis. We applied a nearest-neighbor interpolation approach to extract the NTL values corresponding to the recognized IHS pixels.

Population density is another ancillary datum that has frequently been used to predict carbon emissions in previous studies [49,50]. The distribution of the human footprint is assumed to be correlated with industrial energy consumption and can thus be applied to better describe carbon emissions statistically. To obtain a spatiotemporally consistent population census for the 2012–2020 period, we used a gridded world population product to estimate the human footprint intensity [51]. This product was a high-resolution (~90 m) annual global population distribution time series covering the period from 2000 to 2020. A random forest model, a type of machine learning algorithm, was combined with a collection of harmonized covariates related to factors such as topography, climate, land cover, and traffic to generate wall-to-wall population density predictions [51]. For each 90 \times 90 m grid, the prediction was subsequently applied as a weighting surface and integrated with the dasymetric function to redistribute the census population counts. We then applied a spatial aggregation approach to calculate the total population count corresponding to each recognized HIS pixel in the corresponding year.

2.2.4. Reference IHS Patches

To verify the accuracy of the identified IHS objects, we randomly sampled 864 IHS polygons as a type of reference data (Figure 3). We used the high-resolution imageries embedded in Google Earth Pro software as the base maps to depict the polygon outlines according to our empirical knowledge and understanding of the local industrial sectors and their characteristics reflected in remote sensing imageries. For example, we used blast furnaces and chimneys as important targets when determining the locations of IHS objects. In addition, we also obtained the global IHS product published by Liu et al. [28] covering the period from 2012 to 2016 (denoted as Liu_2018 hereafter). The IHS objects in this product are composited in spatial polygon format (see Appendix A Figure A1). Although this product was developed using VIIRS-derived active fire time series, the polygons were organized together with no temporal information provided. We noticed that the Liu_2018 dataset omits many HIS objects in our study area, but this product was still the most reliable open-access IHS dataset available in the present case. The overall accuracy of this product

is rather high (~77%) when determining subcategorical industry sectors such as cement plants and steel plants. Here, we applied these two datasets as benchmarks to validate the accuracy of the IHSs identified in our study. Because these datasets cannot be effectively aligned with the corporation-level inventory period, we did not use them in the carbon emission analysis performed in the present study.



Figure 3. The spatial distribution of the reference industrial heat source (IHS) patches (i.e., sample data and Liu_2018 data) applied in this study. The snapshots show the boundaries of our sampled data and corresponding high-resolution images.

2.2.5. Corporation-Level Inventory Data

To verify the attributes of the remotely sensed IHS objects and to establish a carbon emissions estimation model, we obtained a portion of the annual corporate inventory data credited by the official authorities on environmental pollution monitoring. These data consisted of information characterizing the annual consumption of fuels (i.e., coal, coke, and natural gas) and electric power from 2016 to 2020 by each corporation that fell in the spatial extent of the identified IHS objects focused on in this study. The corporations could be classified into three main sectors according to the International Standard Industrial Classification of All Economic Activities (UNSD: 2006) including sectors related to the mining industry (denoted as Sector_B), the manufacturing industry (denoted as Sector_C), and the electricity, heat, gas, and water production and supply industry (denoted as Sector_D). These three main sectors could be further divided into 41 subcategories. Summarized descriptions of these subcategories are provided in Appendix A Table A1.

For each corporation, we calculated the carbon emission amount (in terms of CO_2) following the IPCC method. The equation can be expressed as follows:

$$A_{\rm CE} = \frac{44}{12} \sum_{i=1}^{4} K_i E_i \tag{1}$$

where A_{CE} is the amount of carbon emissions; *i* represents the fuel or energy type; *K* is the carbon emission coefficient of a given type of energy consumption; *E* is the amount of fuel or energy consumed. The 44/12 term is the molar ratio by which CO₂ is converted to C. The conversion factor of different fuel types to standard coal equivalent and emission factors can be found in Appendix A Table A2.

For a given IHS object, we obtained the corporations that were spatially located within the object according to the corporation coordinates. We then calculated the total annual carbon emissions from these corporations and assigned the resulting value to the IHS object. We noted that the actual size of some corporations may be spatially larger than the spatial size of the corresponding IHS objects, but we could not decompose the carbon emission values to match the entire areas of these corporations due to the lack of such information; thus, we roughly assigned the total carbon emission values to the nearest IHS objects.

2.3. Workflow

We graphically summarized our workflow into two main steps (Figure 4). The first step involved identifying and characterizing the IHSs from the VIIRS thermal anomaly and auxiliary datasets by combining a spatial cluster algorithm and a series of spatial processing methods. We evaluated the accuracy of the identification results by comparing them with the reference data. Then, in the second step, we established a statistical model for estimating the carbon emissions of the identified IHSs. After the model training and validation steps, we finally interpreted the model outputs and analyzed the explanatory powers of the applied environmental variables.



Figure 4. Flowchart of the data processing and modeling methodology utilized in this study.

2.4. Data Processing and Analysis

2.4.1. Identification of IHSs

The VIIRS thermal anomalies were observed in the form of discrete points. We reasonably assumed that these anomaly points should collocate with the core of the heat source and be spatially clustered in a small area when using long time series observations, given that industrial heat releases are stable and continuous. Such patterns can support the use of the spatial cluster analysis approach when grouping the spatially adjacent anomaly points into an object. Here, we utilized an unsupervised learning algorithm (i.e., the density-based spatial cluster applications with noise (DBSCAN) approach [52])

to distinguish whether a given point belonged to a cluster defined as the maximum set of densely connected points. This approach can be used to discover clusters of different shapes and sizes from a large amount of data containing noise and outliers.

The DBSCAN approach uses the minimum number of points (Min_Pts) and the maximum radius of the neighborhood (EPS) to justify the density reachability and density connectivity of a given point to a cluster [53]. The density reachability establishes that a point is reachable from another point if it lies within a particular distance (i.e., EPS) from the first point. Connectivity is a transitivity-based chaining approach that is used to determine whether points are located within a particular cluster. This approach randomly selects a point and estimates the reachability of all points from the selected point under the constraints imposed by the EPS and Min_Pts values. If the number of adjacent points is more than Min_Pts, the selected point is defined as a core point of a cluster; otherwise, it is labeled as a noise point, and the process moves to the next point. This process continues until all points have been classified. Because the spatial sizes of individual IHSs usually do not reach the kilometer level or the point density of annual observations, we set the Min_Pts value to four and the EPS value to 1.5 km. The DBSCAN approach is robust to noise, insensitive to the cluster shape, and has a rather high computing speed [54,55].

2.4.2. Characterization of IHSs

We used the annual VIIRS thermal anomaly product as the input data and generated a series of raw clusters that may have contained other thermal sources such as forest fires, agricultural straw burning, or fire accidents. The annual ESA_CCI_LC product was used to filter out the thermal anomaly points that fell within regions of vegetation-dominated land cover types. Due to the coarse resolution of the VIIRS data, we carried out a filtering process following the cluster analysis to prevent the potential omission of points located along edge regions. We converted the spatial points to spatial pixel objects at a 1 km resolution using the spatial overlay approach. We then characterized these pixels by spatially aggregating the corresponding thermal points and calculating the point number, mean brightness temperature, mean FRP, and duration. Some pixels contained very sparse thermal points but featured long durations due to the presence of uncertain factors such as incorrect land cover mapping or seasonal agricultural straw burning. We thus set decision rules based on the point number (>10) and duration (>30 days) to further mitigate these uncertainties. We acknowledge that such a hard classifier may remove some IHSs from the analyzed data, but it should be noted that the retained data had high confidence and should be focused on in practical applications.

2.4.3. Accuracy Assessment of IHSs

We used the two aforementioned reference IHS datasets to validate the accuracy of our results. We further processed our results to solve the issues regarding temporal and spatial inconsistencies between the datasets. To ensure that the temporal periods matched the Liu_2018 dataset, we chose our IHS results from 2012 to 2016 and generated a mosaic to improve the temporal consistency between the two datasets. For the sample data, we generated a full 2012–2020 composite of our IHS results to promote temporal consistency between the two datasets both represented partial samples compared to the true IHS information, we did not think they could be used to fully evaluate the accuracies of the IHS results in terms of the commission and omission errors. Therefore, we used two measures to evaluate the accuracy of our identifications. At the patch level, we calculated the rate of the overlapping area (ROA) for each referenced IHS object, obtained the basic relevant statistics, such as the mean values and standard deviations, and constructed a histogram to evaluate the patch-level accuracy. For the whole dataset, we calculated the percentage of referenced IHSs that overlapped with our identified results.

To determine the industrial sectors corresponding to the identified IHS objects, we did not assign categories or subcategories to the IHS objects as has been conducted previously. As we noticed that one remotely sensed IHS object can cover multiple corporations that belong to different industrial sectors, we used all corporations covered by each IHS and obtained statistics related to the corporation abundances from 2016 to 2020. We subsequently analyzed the composition of the validated IHSs to address our first objective.

2.5. Estimation of Industrial Carbon Emissions

To ensure consistency with the IHS analysis, we also converted the corporate carbon emissions inventory into a raster dataset at a 1 km resolution. We followed the bottom-up approach to generate a carbon emission raster in which the carbon emissions of a given pixel were assigned using the sum of all corporations within the corresponding area. The overlapping pixels between the remotely sensed IHSs and carbon emission grids were then extracted and utilized as the training data in the subsequent modeling work. The summary statistics and descriptions of explanatory variables for the carbon emission modeling can be seen in Table 1.

 Table 1. Summary statistics and descriptions of explanatory variables for the carbon emission modeling.

Variable Abbreviations	Description	Mean \pm SD	Data Source
Рор	Population density per unit area.	1706.238 ± 157.819	[51]
NTL	The maximum nighttime light radiance.	42.755 ± 2.689	[48]
Num_TAPs	Number of TAPs that fell in the IHS pixel.	26.474 ± 2.776	[14]
Duration	The period length in days between the start and end dates of all TAPs that fell in the IHS pixel.	201.106 ± 6.933	[14]
FRP	Accumulated fire radiative power per unit area derived from all TAPs the fell in the IHS pixel.	42.062 ± 4.822	[14]
BT4	Accumulated brightness temperature of the I-4 Channel derived from all TAPs in the IHS pixel.	8150.194 ± 851.2	[14]
BT5	Accumulated brightness temperature of the I-5 Channel derived from all TAPs in the IHS pixel.	7586.415 ± 799.994	[14]

2.5.1. BRT Modeling

Since the remotely sensed features can only provide parameters that are related to carbon emissions, the carbon emission estimation via remote sensing is overwhelmingly empirical. Here, we applied the BRT algorithm to establish the empirical relationships. We did not choose deep learning algorithms, because their "black box" nature would limit us to interpreting those empirical relationships and could be supported in the BRT algorithm. The BRT model is a tree-based ensemble learning algorithm proposed by Elith et al. [56]. This algorithm is robust for processing nonlinear relationships between explanatory and response variables. In the model, the boosting algorithm is integrated with the decision tree model to enhance the predictive ability of the model by combining weak models into a relatively strong model. The boosting algorithm can help the BRT model create weighted trees to improve the subsequently constructed trees, meaning that the model takes into account the error in the prediction of the previous tree when fitting the next tree. By using this sequential procedure, the accuracy of the model results is continuously improved. In addition, the model has a relatively strong tolerance to unbalanced data with skewness, missing values, and complex interactions among predictors. It is widely reported that the BRT model has a stronger regression power than most other machine learning algorithms, such as the well-known random forest model and the support vector machine algorithm, in the same application scenarios.

The BRT model has two important parameters that must be defined by the user: the so-called tree complexity and the learning rate. The tree complexity controls the number of branches in each tree, and the learning rate determines the contribution of each tree to the growing model. Together, these parameters determine the number of trees required to obtain the optimal prediction. We followed the parameter setting recommendations

proposed by Elith, Leathwick, and Hastie [56], setting the tree complexity to 3 and the learning rate to 0.01 to allow the model to grow a sufficient number of trees. We used the relative importance as an indicator to quantify the contribution of predictors aiming to improve the model explanatory strength to the variance in carbon emissions.

2.5.2. Accuracy Assessment

The BRT model's performance was evaluated by comparing the model-fitted data with the gridded corporate inventory data. We calculated the variance explained (R^2) and the index of agreement (IOA) to evaluate the goodness of fit of the model results. In addition, we also calculated the root mean squared error (RMSE) and the mean absolute percentage error (MAPE) as criteria to assess the errors in the results. These metrics can be calculated using the following equations:

$$R^{2} = \left[\frac{\sum_{i=1}^{N} (Y_{o,i} - \overline{Y}_{o}) (Y_{m,i} - \overline{Y}_{m})}{\sqrt{\sum_{i=1}^{N} (Y_{o,i} - \overline{Y}_{o})^{2} \sum_{i=1}^{N} (Y_{m,i} - \overline{Y}_{m})^{2}}}\right]^{2}$$
(2)

$$IOA = 1 - \frac{\sum_{i=1}^{N} (Y_{o,i} - Y_{m,i})^2}{\sum_{i=1}^{N} (|Y_{o,i} - \overline{Y}_o| + |Y_{m,i} - \overline{Y}_o|)^2}$$
(3)

$$RMSE = \sqrt{\frac{1}{N} \sum_{i=1}^{N} (Y_{o,i} - Y_{m,i})^2}$$
(4)

$$MAE = \frac{1}{N} \sum_{i=1}^{N} |Y_{o,i} - Y_{m,i}|$$
(5)

where Y_o is the observation value and Y_m is the predicted value; Y_o is the average observation values and \overline{Y}_m is the average predicted value; N is the number of observation data.

3. Results

3.1. Identification of IHSs

We identified a total of 4031 IHS pixels (1 × 1 km) containing approximately 0.101 million TAPs identified from 2012 to 2020 (Table 2). The annual number of IHS pixels was significantly correlated with the corresponding number of raw TAPs (r = 0.64, p < 0.02). As a result of government control on agricultural straw burning, the raw TAPs showed a remarkable decreasing trend. However, we found that the retained TAPs used to identify IHSs did not represent a statistically significant trend. In contrast, the fraction of retained TAPs exhibited a significantly increasing trend as the number of raw TAPs decreased (r = -0.88, p < 0.02), suggesting that the control of agricultural straw burning helped reduce the uncertainties associated with thermal anomalies.

The DBSCAN clusters, from which the IHS pixels were obtained, also decreased along with the dynamics of the annual raw TAPs (r = 0.98, p < 0.02), ultimately causing a reduction in the number of raw IHS pixels. This procedure excluded approximately 52% of the raw TAPs and generated an average of 541 clusters each year. The further thresholding procedure filtered approximately 42% of the raw IHS pixels, comprising pixels that could not meet the decision rules. We generated accumulated TAPs for each IHS pixel and found that over 20% contained more than 1000 points between 2012 and 2020 (Figure 5A). The IHS pixels were sparsely distributed in general but exhibited a clustered pattern in the south and central parts of the province. Approximately 30% of the IHS pixels were identified more than three times, and approximately 9.4% of the total number of IHS pixels were identified every year (Figure 5B).

Year Raw	The	hermal Anomaly Pixels			Industrial Heat Source Pixels		
	Raw	Retained	Fraction	Cluster	Raw	Retained	Fraction
2012	28,797	10,583	0.37	684	775	389	0.502
2013	28,110	11,886	0.42	699	821	504	0.614
2014	29,346	13,138	0.45	728	876	483	0.551
2015	26,498	11,172	0.42	699	908	460	0.507
2016	21,070	9976	0.47	452	681	425	0.624
2017	21,196	10,605	0.50	459	595	350	0.588
2018	20,185	11,560	0.57	357	637	399	0.626
2019	20,249	10,836	0.54	413	592	351	0.593
2020	18,242	10,087	0.55	377	553	363	0.656
Mean	23,744	11,201	0.48	541	715	414	0.58
(SD)	(4364)	(1001)	(0.07)	(157)	(133)	(58)	(0.05)

Table 2. Statistical variations in thermal anomaly pixels (TAPs), clusters, and industrial heat source (IHS) objects during processing.



Figure 5. The spatial patterns of the thermal anomaly pixels (TAPs) within industrial heat source pixels (**A**) and the observation times (**B**).

3.2. Comparison with the Referenced IHSs

Regarding the relatively coarse resolution of our analysis, the IHS pixel boundaries did not strictly align with the spatial extent of each industrial corporation; however, we still found that the IHS pixels could cover most areas corresponding to these corporations, especially the locations from which heat was released (see Appendix A Figure A2). To obtain an unbiased accuracy estimation, we compared our results with two types of benchmark IHS datasets. We used only the IHS pixels corresponding to the 2012–2016 period for the comparison with the Liu_2018 data but used all IHS pixels for the comparison with the sample data to guarantee that the comparisons were carried out on consistent time scales. Any duplicate IHS pixels that were identified multiple times were removed before the comparisons.

We found that 150 out of 193 patches (OA = 77.7%) from the Liu_2018 data (Figure 6A) and 797 out of 864 patches (OA = 92.2%) from the sample data (Figure 6D) overlapped with our IHS pixels. We also compared the two reference datasets as a further reference, although they represented different periods. We identified 307 patches (OA = 35.5%) from the sample data that overlapped with the Liu_2018 data (Figure 6B), while as a comparison, the IHS pixels of 2012–2016 (OA = 82.4%) overlapped with 712 patches (Figure 6C). We applied an ROA \geq 0.75 as a criterion to evaluate the state of IHS overlap. The results showed that 84.8% of overlapped patches (with 2012–2020 IHS pixels, Figure 6D) from the sample data and 58% of overlapped patches (with 2012–2016 IHS pixels, Figure 6A) from Liu_2018 met the standard. Approximately 83.1% overlapped patches (with 2012–2016 IHS pixels, Figure 6B) from the sample data met the standard.



Figure 6. The density histogram shows the patch-level rate of the overlapping area (ROA) that was cross-validated between the Liu_2018 data and the IHS pixels of 2012–2016 (**A**), the sample data and the Liu_2018 data (**B**), the sample data and the IHS pixels of 2012–2016 (**C**), the sample data and the IHS pixels of 2012–2020 (**D**). N represents the number of overlapping reference IHS patches; OA represents the overall accuracy in terms of the fraction of validated patches from the reference data.

3.3. Comparison with Corporate Inventory Data

According to the spatial extents of the remotely sensed IHS pixels (2012–2020), we obtained a total of 5443 corporations with explicit attributes corresponding to the industrial sectors and energy consumption amounts (Table 3). Integrating IHS pixels from multiple years can increase the coverage area and thereby increase the number of cross-validated corporations. We found that the IHS pixels contained similar manufacturing industry (i.e., Sector_C) and energy-related industry (i.e., Sector_D) percentages, suggesting very similar capacities of the thermal anomalies for detecting IHS objects from these two domains. Because the manufacturing industry (i.e., Sector_C) occupied approximately 95.0% of the total number of analyzed corporations, it is reasonable that most of the identified corporations belonged to Sector_C.

IHS)16–2020)	-2020) IHS (2012–2		Liu_2018 (2012–2016)		Total Inventory	
Sector	Count	Percent	Count	Percent	Count	Percent	Corporations	
В	5	100%	0	0	1	20.0%	5	
С	1579	65.5%	2068	85.7%	904	37.5%	2412	
D	97	62.2%	123	78.8%	44	28.2%	156	
Total	1681	65.3%	2191	85.2%	949	36.9%	2573	

Table 3. Comparison of cross-validated corporations and their compositions (R).

As shown in Figure 7, the manufacturing of nonmetallic mineral products (e.g., cement, glass, graphite, and carbon products) and the chemical industry were the top two subcategories among the identified Sector_C corporations, followed by the smelting and pressing of ferrous metals (e.g., steel produce), fuel-processing industries (e.g., petroleum and coal), and the manufacturing of metal products. Electric heat production (i.e., thermal power plants) was the primary subcategory representing Sector_D.

3.4. BRT Modeling Performance Evaluation

We used 661 IHS pixels that overlapped with the gridded carbon emission dataset characterizing the 2016–2020 period to establish the regression model. The results showed that the BRT model performed reasonably well, as approximately 61.9% of the variance was explained, and the IOA was 0.83, indicating a high goodness of fit between the model predictions and corporate inventory data (Figure 8A). Across the five years, the model accuracy in terms of the MAPE was 0.16, while the RMSE was 0.92. Figure 8B shows the relative importance rankings of all explanatory variables. As the figure shows, the population density (22.2%) and two thermal anomaly related features, namely, the duration (19.5%) and the FRP (19.6%), as well as the nighttime light (13.2%), had larger contributions to the predictive power of the BRT model than the other three analyzed variables.



13 14 15 17 18 19 20 21 22 23 25 26 27 29 30 31 32 33 34 35 36 38 42 44 Sub-category of industrial sector

Figure 7. The composition of the subcategorical industry sectors fell in the industrial heat source (IHS) pixels of 2016–2020 (**A**). As a comparison, we also listed the composition for IHS pixels of 2012–2016 (**B**) and the Liu_2018 data (**C**). Details of each sector can be found in Appendix A Table A1.



Figure 8. Scatter plot showing the accuracy assessment between the model-fitted carbon emissions (unit: ton carbon) and the amount of corporate inventory data (**A**). A log transformation (log10) was applied to remove the skewness of the original data. (**B**) The bar plot shows the relative importance values of the explanatory variables (the abbreviations for these explanatory variables can be found in Table 1).

4. Discussion

In this study, we evaluated the capacity of remotely sensed thermal anomaly products to identify IHSs. Based on the DBSCAN approach and a land cover dataset, we generated the wall-to-wall IHS distribution in Shandong Province and generally found that the modeled IHS results were well aligned with the IHSs in the two independent reference datasets and were confirmed by high-resolution satellite imagery. These IHS pixels were mostly located in the outskirts of cities (Figure 9), but some were spatially continuous within cities dominated by manufacturing (e.g., Linyi city, Figure 9E). Isolated IHS objects were generally easy to detect due to the relatively cool background conditions and clustered thermal anomaly patterns. In most cities, the IHS objects located in our analysis. We considered that this may have been because our analysis was conducted at a larger grain size than the reference data, which will increase the opportunity to retain more fire spots in the identification.



Figure 9. Spatial distributions of industrial heat source (IHS) pixels in the cities with the top 5 highest GDPs in Shandong Province (i.e., Jinan (**A**), Qingdao (**B**), Yantai (**C**), Weifang (**D**) and Linyi (**E**)).

We also considered some factors that may have influenced the consistency between our results and the reference datasets. First, because we filtered out thermal points located in farmlands and forests according to the ESA_CCI land cover product, some IHS patches in the Liu_2018 product located in farmlands and forests were not identified in our study. Second, the accuracy of our IHS identification results was partially limited by the quality of the land cover mapping, but such a filtering process was useful for reducing the confusion and uncertainty that arise from biomass burning during wildfires or anthropogenic fires. Since the DBSCAN approach was physically based and the raw data (i.e., fire spots and landcover maps) could be obtained easily for other regions, our methods can be expanded directly to other regions. In addition, in contrast to some other studies in which combined datasets covering multiple years of observation were used [23,25,28], our approach to mapping IHS annually was straightforward. This method can be beneficial in aiding policymakers who track the temporal dynamics of IHS abundance, but we admit that the consequently shorter period decreases the accumulation of thermal points around any given IHS and increases the potential for omission errors.

Through a comparison with corporate-level inventory data, our results clearly showed that the identified IHS pixels mostly belonged to the EIIS, especially to the heavy industries mainly belonging to the chemical industry and the manufacturing of nonmetallic mineral products such as the cement, glass, graphite, and carbon production industries. Corporations of Sector_C accounted for most of the total carbon emissions due to the large number of individual corporations in this sector, but the individual corporations of Sector_D had remarkably higher mean annual emission amounts than other industrial sectors (Figure 10A). From the perspective of fuel sources, coal and coke consumption contributed the majority of the carbon emissions associated with the identified IHSs (Figure 10B). The EIIS corporations that use coal and coke as fuel resources usually emit considerable waste heat into the environment during their processing operations [57], while corporations that rely primarily on electricity to support their operations may generate little heat waste, and facilities such as constructions or shelters can also shield industrial heat emissions. Although our results can cover these corporation components, we believe that the spatial points identified on behalf of the spatial locations of these corporations fell in the IHS pixels by chance or were not specifically identified as members of the EIIS. Thus, we concluded that the identified IHS pixels could capture the EIIS corporations that contributed to the majority of the carbon emissions.

The BRT model explained a considerable amount of variance in the total carbon emissions estimated using the bottom-up inventory approach, suggesting that it is feasible to estimate carbon emissions by combining machine learning approaches with environmental variables that are linked with carbon emissions. We selected explanatory variables that were sensitive to social economics and human activity footprints for model training. The training data predictors were obtained from global gridded products that were carefully investigated and produced with broad public credibility and are well accepted by the academic community. Benefiting from the strong ability of the BRT model to handle complex and nonlinear relationships between predictors and response variables (Figure 11), we found that the population density, FRP, and NTL showed negative relationships and notable nonmonotonicity with carbon emissions, while the tendency curve between duration and carbon emissions showed a positive relationship. A

25

20

Carbon Emission (×10⁸ ton)

Total Emission

Fuel Source

Coke

Electrity Gas

Industrial Sector



Industrial Sector

Figure 10. Comparisons of fuel and energy consumption, total carbon emissions (**A**), and mean annual carbon emissions (**B**) among different industrial sectors. The total carbon emissions of Sector_B are relatively low in (**A**) since the number of identified IHS in Sector_B was very small compared with that in Sector_C and Sector_D. The high variation of mean annual carbon emissions in (**B**) was caused by some high-emission corporations.

0.5

For the population density, we generally observed a decreasing carbon emissions trend as the population number increased (Figure 11A). This result was consistent with the general reality that high-emission industries are usually located in urban periphery regions, where the population density is not as high as that in downtown areas where low-emissions corporations are usually distributed. Regarding the duration, we found that IHS objects with relatively long thermal anomaly durations led to increased carbon emissions; however, we also noted that this relationship curve fluctuated considerably (Figure 11B). We believe the underlying reason for this fluctuation was that multiple corporations are often located within a given IHS pixel, and the operating temperature of corporations belonging to different industrial sectors may differ substantially in their processing periods [58]. These discrepancies may introduce uncertainties, as we could not attribute the identified thermal anomalies to specific corporations. The mean FRP, a measure of thermal emissivity, and the observed thermal anomaly duration were found to be important variables that could improve the fit of the carbon emissions (Figure 11C). These two thermal-related variables represented rather complex response curves. Ignoring the rugged curve regions with large uncertainties, we generally found that carbon emissions increased sharply as the FRP increased but easily reached a saturation situation. For NTL, we found that areas of high brightness corresponded to low carbon emissions (Figure 11D). The underlying reason for this is similar to that of population density; high-emission industries are usually located in urban periphery regions, where the NTL is not as high as that in downtown areas. Moreover, regions without any nighttime lights do not mean the regions have no CO₂ emissions [59]. For example, some EIIS corporations only operate during the daytime but stop operating and turn off all lights at night.

Although the modeling framework was straightforward and the outputs were validated, we noted that the estimation approach could be further improved in future studies. First, based on corporate inventory, we learned that the mean carbon emission amount was considerably differentiated among industrial sectors. We believe that the model per-

formance could be improved by involving the attribute information of the identified IHS objects. Such information can perhaps be obtained from the POI services provided by web platforms or databases such as Google Maps, Baidu Maps, and OpenStreetMap. Ideally, these databases could be promoted to form a more comprehensive characterization of IHS objects; however, it should also be noted that mismatch and omission issues may reduce their usage. Second, we estimated only the carbon emissions associated with the identified IHS objects, which may account for only a part of the overall carbon emissions. Future studies should focus on improving this omission problem during IHS identification, especially when identifying IHSs containing EIIS corporations. Simultaneously, to ensure spatial consistency with the explanatory variables, we identified IHSs at a relatively coarse spatial resolution. This resolution could be further improved by applying high-resolution satellite imageries, especially nighttime thermal infrared imageries. In addition, though the BRT model is simple and novel, we noted that the model fitting can be further improved by using deep learning approaches regarding their high performance in handling classification and regression problems. However, caution is required due to the fact of their disadvantages such as greater computational burden, proneness to overfitting, and the empirical nature of model development.



Figure 11. Partial dependence curves representing the relationships between the carbon emission amounts and Population (**A**), Duration (**B**), Mean FRP (**C**) and Nighttime Light (**D**). The purple lines represent smoothed conditional means and the green areas represent 95% confidence intervals.

5. Conclusions

In summary, in this study, we described a thermal anomaly based approach to identifying IHS objects and modeling their carbon emissions using machine learning. Based on the DBSCAN algorithm, we identified approximately 414 IHS pixels $(1 \times 1 \text{ km})$ per year in Shandong Province from 2012 to 2020. The cross-validation showed that our results were consistent with the reference datasets. According to the corporation-level inventory dataset, we highlighted that the remotely sensed thermal anomalies were sensitive to the IHSs originating from EIIS-related corporations, especially thermal power plants, cement plants, the chemical industry, the steel production industry, and other nonmetallic mineral production industries (e.g., glass, graphite, and carbon production). We followed the bottom-up method to calculate the corporation-level carbon emissions and upscaled them to match the IHS pixels. Based on the BRT model, we found that the population density, NTL, and remotely sensed thermal features were useful for improving the model fitting of the carbon emissions in the identified IHS pixels. In this study, we proposed a novel and straightforward framework that can provide valuable support for understanding the spatiotemporal dynamics of IHSs, especially those related to the EIIS. We also provided an alternative approach for estimating industrial carbon emissions, and although this approach can account for only a limited fraction of the total emission amount, it can help monitor industrial activities in the shift toward a cleaner and lower-carbon environment. Future works need to improve the identification of IHS objects with low thermal emissivity to reduce the omission problems and involve multiple-source datasets at finer scales to characterize IHS object more comprehensively and accurately.

Author Contributions: Conceptualization, X.K., X.W. and M.J.; methodology, X.K.; software, X.K.; validation, X.W.; formal analysis, X.K.; data curation, X.W. and M.J.; writing—original draft preparation, X.K.; writing—review and editing, X.K., X.W., M.J. and Q.L.; visualization, M.J.; supervision, Q.L. All authors have read and agreed to the published version of the manuscript.

Funding: This research received no external funding.

Data Availability Statement: Not applicable.

Conflicts of Interest: The authors declare no conflict of interest.

Appendix A

Code	Industry Sector		
В	Mining		
06	Mining and Washing of Coal Industry		
07	Extraction of Petroleum and Natural Gas		
08	Ferrous Metal Mining and Selection Industry		
09	Nonferrous Metal Mining and Selection Industry		
10	Nonmetallic Mining and Selection Industry		
11	Mining Professional and Auxiliary Activities		
12	Other Mining Industry		
С	Manufacturing		
13	Processing of Food from Agricultural Products		
14	Manufacture of Foods		
15	Manufacture of Beverages		
16	Manufacture of Tobacco		
17	Manufacture of Textile		
18	Manufacture of Textile Wearing Apparel, Footwear, and Caps		

Table A1. A brief description of the subcategorical industrial sectors of mining, manufacturing, and energy industries that have the potential to be detected as industrial heat sources via thermal remote sensing.

Code	Industry Sector	
19	Manufacture of Leather, Fur, Feather, and Related Products	
20	Processing of Timber, Manufacture of Wood, Bamboo, Rattan, Palm, and Straw Products	
21	Manufacture of Furniture	
22	Manufacture of Paper and Paper Products	
23	Printing, Reproduction of Recording Media	
24	Manufacture of Articles for Culture, Education, and Sport Activities	
25	Processing of Petroleum, Coking, Processing of Nuclear Fuel	
26	Manufacture of Raw Chemical Materials and Chemical Products	
27	Manufacture of Medicines	
28	Manufacture of Chemical Fibers	
29	Manufacture of Rubber and Plastics	
30	Manufacture of Nonmetallic Mineral Products	
31	Smelting and Pressing of Ferrous Metals	
32	Smelting and Pressing of Nonferrous Metals	
33	Manufacture of Metal Products	
34	Manufacture of General Purpose Machinery	
35	Manufacture of Special Purpose Machinery	
36	Manufacture of Automobile	
37	Manufacture of Railways, Shipbuilding, Aerospace, and Other Transportation Equipment	
38	Manufacture of Electrical Machinery and Equipment	
39	Manufacture of Communication Equipment, Computers, and Other	
40	Manufacture of Measuring Instruments and Machinery for Cultural Activity and Office Work	
41	Manufacture of Artwork and Other Manufacturing	
42	Recycling and Disposal of Waste	
43	Repair and Installation of Machinery and Equipment	
D	Production and Supply of Electricity, Gas, and Water	
44	Production and Supply of Electric Power and Heat Power	
45	Production and Supply of Gas	
46	Production and Supply of Tap Water	

Table A2. Conversion factor of power generation coal to standard coal equivalent (SCE) and CO_2 emission factor for different fuel types.

Energy Type	Conversion Factor to SCE (Unit: tSCE/t)	Carbon Emission Factor (×10 ⁴ tC/10 ⁴ tSCE)		
	CF _{SCE}	CEF		
Raw Coal	0.7143	0.7559		
Coke	0.9714	0.855		
Natural Gas *	1.33	0.4483		
Electricity *	-	0.272		

* The unit of natural gas conversion to CO_2 emissions is $t/10^3$ m³, and the unit of electricity conversion to CO_2 emissions is t C/10³ kWh.



Figure A1. The annual variation in the total masked area (i.e., settlement and bare area) between 2012 and 2020.



Figure A2. Examples of comparison between the industrial heat source (IHS) pixels of this study and two reference data (i.e., the sample data and Liu_2018 data). The dots represent the corporations falling in IHS pixels.

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