



Article Characteristics of Unorganized Hydrogen Sulfide Dispersion for Industrial Building Layout Optimization

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Abstract: Hydrogen sulfide (H₂S) is the main toxic pollutant emitted to the atmosphere from autocoating wastewater. Its unorganized dispersion poses a health challenge for workers. Defining safe working distance, which transfers the H₂S occupational exposure limit into industrial construction design regulation, would be a useful approach for reducing H₂S exposure risk. Therefore, in this study, an H₂S dispersion prediction, within 25 m, was performed by a computational fluid dynamics (CFD) method to explore the influence of temperature and wind speed on H₂S dispersion. With the temperature changes from 288 K to 303 K, the H₂S concentration at different observing points decreased. With wind speed changes from 2 m/s to 20 m/s, the plume layer structure was studied in the whole process. According to the H₂S distribution characteristics, when the sedimentation tank treatment capacity is less than or equal to 10 m³/h, the safe working distance of H₂S unorganized dispersion is 10 m. Hence, when there are workplaces within 10 m of the tank, closed measures should be taken for the sedimentation tank, or the manufacturer layout should be optimized to protect the environment and human health.

Keywords: occupational exposure; human health risk; hydrogen sulfide; air pollution; unorganized dispersion; safe working distance

1. Introduction

Hydrogen sulfide (H_2S), a toxic gas with a rotten egg smell, widely exists in industrial production processes [1]. In fact, H_2S is commonly found in industrial sites such as geothermal power plants, livestock farms, paper mills, pharmaceutical plants, sewage treatment plants, landfills, and automobile factories [2–4], being a potential health hazard for workers. Therefore, analyzing different types of H_2S dispersion in different industrial plants is significant.

The H₂S potential health risks are divided into acute toxicity (high H₂S concentration) and chronic damage (low H₂S concentration) [5]. Mature operation specifications and safety measures have been conducted for acute poisoning prevention, including carbon-based material adsorption, metal catalytic oxidation, electrochemical treatment, and biological treatment [6,7]. Although the majority of H₂S would be systematically treated [8], there is still a considerable amount of H₂S that would be freely emitted to the atmosphere, due to concentration differences, pressure differences, or other reasons. As a result, many industrial sites are chronically filled with low-concentration H₂S. It was illustrated that chronic exposure to low levels of H₂S is associated with increased mortality and morbidity for respiratory diseases, disorders of the peripheral nervous system, heart failure, and diseases of the veins [9,10]. Therefore, workers who are exposed to low levels of H₂S for a long time are at great risk of H₂S chronic damage to their health.

Occupational exposure standards for low-concentration H_2S have been set. According to WHO, having about 7.5 mg/m³ H_2S in the atmosphere, people start to have discomfort symptoms, such as eye irritation and bad breathing [11]. The Chinese occupational exposure limits for hazardous agents in the workplaces stipulates that the maximum allowable



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Copyright: © 2022 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (https:// creativecommons.org/licenses/by/ 4.0/). concentration of H_2S is 10 mg/m³ [12]. Although the limits are specific, there are still big difficulties in implementing the policy, due to the lack of effective control measures of lowconcentration H_2S in workplaces [13]. In this case, defining a safe working distance, which transfers the H_2S occupational exposure limit into industrial construction design regulation, would be a useful approach for reducing the potential H_2S exposure risk. Modeling the dispersion of low-concentration H_2S in different industrial application scenarios becomes an economical and efficient way to visualize H_2S dispersion. The simulation models would be conducive to avoiding potential health risk areas with high H_2S concentration, optimizing plant layout and ventilation design, and reducing the health damage of H_2S to humans.

Current studies on outdoor low-concentration H_2S dispersion simulation mainly treated a factory as a point source of H_2S , using the air pollution model aermod or calpuff, combined with a GIS (Geographic Information System) system [14–16], to investigate the effect of H_2S on the air quality of the surrounding area and the health effects on nearby residents [17], rather than workers, who stay much closer to the source. The dispersion distribution of H_2S was simulated at the kilometer scale [18]. However, the accuracy of the two models is not suitable for simulating low-concentration H_2S on the meter scale. For distances from 50 m to 250 m, the CFD method [19] was used to analyze high-concentration H_2S leakage events, which indicated that the CFD method could be suitable for shortdistance and low-concentration simulation.

At present, no simulation study about short-distance unorganized H_2S dispersion has been reported to define the safe working distance. Low-concentration H_2S is still a hidden danger for workers. Therefore, it is necessary to study the unorganized H_2S dispersion for occupational exposure in a short distance, smaller than 25 m, from the source (inside manufacture) by CFD method. Based on the gas chromatography–mass spectrometry (GC–MS) method, using a multi-component gas analyzer, the main odorous component of automobile coating wastewater was analyzed to be H_2S . Therefore, an outdoor automobile coating sewage sedimentation tank is taken as the simulation scenario in this study to establish a dispersion model of low-concentration H_2S . The model explores the influencing factors and dispersion characteristics of low-concentration H_2S in this scene.

With the research gaps mentioned, the aim of this study is to conduct dispersion model of unorganized H_2S and define a proper safe working distance. The implementation steps are: first, select the proper models for surface source no-barrier H_2S dispersion that closely match the experimental measurements; then, analyze the wind speed and temperature influences on short-distance outdoor H_2S dispersion; finally, acquire the safe distances of different wind speeds and temperature. This study will clearly illustrate the risks of unorganized H_2S no-barrier dispersion from sewage sedimentation tank to human and help to create a safe working distance for H_2S exposure management.

2. Materials and Methods

2.1. Description of the Study Area

We studied an auto parts paint washing sewage sedimentation tank in $20 \text{ m} \times 10 \text{ m}$ size, in Jinzhou city, Hubei Province (Figure 1a) and built a 3D sewage tank model (Figure 1b). In production processes, H₂S from auto parts paint dissolved in washing wastewater, conveyed to the sewage pool for centralized treatment. The sewage pool could be regarded as a surface emission source of odorous pollutants in a wide-open outdoor space. The odor of H₂S would affect the working environment. Note that the wastewater sedimentation tank, processing 10 m^3 /h, discharged regularly in work time. There was no barrier tall building near the sewage pool, except a small office and a sidewalk 4 m and 11 m away the sewage pool, respectively.



Figure 1. (a) The onsite view of the washing sewage sedimentation tank; (b) The 3D model of the onsite view; (c) The CFD model of the onsite view; (d) The schematic diagram of monitoring points. The numbers in parentheses represent x, y, z three-dimensional coordinates of monitoring points.

2.2. Experimental Measurement

Sampling dates were collected in winter 2020 and summer 2021, respectively. Three observing points (Figure 1d) were set at 0 m, 4 m, and 11 m, called P1, P2, and P3, to monitor H_2S concentration by the H_2S detector (SGA-608, SINGOAN). Active sampling by vacuum pump can quickly carry out real-time online detection of multi-component factors in the environment. We only needed one key to start the machine, and could take the initiative to sample and display the current measured value. The H_2S detected time of each observing point lasted 5 min, and the concentration average values are displayed in Figure 2. Meanwhile, the environmental parameters, including temperature and wind speed, were measured in continuous 1 min in both measurements. Additionally, the Cartesian coordinates of the three monitoring points were (10 m, 0 m, 1.5 m), (14 m, 0 m, 1.5 m), and (21 m, 0 m, 1.5 m), respectively. The measurement details and results were listed in Table 1.



Figure 2. Onsite data of H_2S at 1.5 m height. The dotted line is the occupational exposure standard for H_2S in China, which is 10 mg/m³ = 6.9 ppm.

Measurement No.	Collection Date	Sampling Locations	Temperature	Wind Speed
1	21 September 2020	P1, P2, P3	302.5–303.3 K	3.2–4.6 m/s
2	8 March 2021	P1, P2, P3	287.8–288.4 K	15.8–16.2 m/s

Table 1. Collection parameters of each measurement in experimental scene.

2.3. Parameters Calculation

On the basis of double membrane theory, the process of H_2S release can be divided into three parts, which are liquid-phase, gas–liquid interphase, and gas-phase mass transfer. The sewage pool was assumed to have an interphase between liquid and gas phases. The three phases and the two films between them formed a five-layer structure to complete the three procedures mentioned above. All the phases and films were considered to be in a dynamic equilibrium. There were three assumptions in this theory: firstly, the interphase was assumed to have no mass transfer resistance; secondly, the sewage was regarded as H_2S saturated solution; thirdly, H_2S concentration in gas phase was zero. Besides, it is known that interphase mass transfer should obey Henry's law. According to the setting above, the following equations were listed:

$$N = N_L = N_G \tag{1}$$

$$N = D \left(c_L - c_G \right) \tag{2}$$

$$c_i = Hp_i \tag{3}$$

$$N = D \left(c_L - H p_G \right) \tag{4}$$

where *N* is the total mass transfer rate, N_L is the mass transfer rate in the liquid phase, N_G is the mass transfer rate in the gas phase, *D* is the mass transfer coefficient, c_L is the species concentration in liquid phase, c_G is the species concentration in gas phase, c_i is the concentration of species *i* in liquid phase when it balances with the material partial pressure p_i in the gas phase, *H* is the Henry's law constant.

The formular of mass transfer coefficient is:

$$D = \frac{435.7T^{3/2}\sqrt{\frac{1}{M_A} + \frac{1}{M_B}} \times 10^{-4}}{P\left[\left(\sum A\right)^{\frac{1}{3}} + \left(\sum B\right)^{\frac{1}{3}}\right]^2}$$
(5)

where *T* is the environment temperature, *M* is the relative molecular mass, $\sum A$ and $\sum B$ are the molar volume per gram, *P* is the environment pressure.

The main purpose of this study is to investigate the H_2S flow characters in the air. A computational volume of 40 m × 30 m × 3 m in x, y, z directions was established, with the volume center at origin point (x = y = z = 0 m) (Figure 1d). The computational domain included the clean air velocity inlet (blue arrow), the H_2S mass inlet (yellow arrow), and four mixture flow outlets (left, right, opposite, and up). The H_2S mass inlet was on the bottom surface and H_2S initially diffused upward. The model structure was shown in Figure 1c.

For a flowing H₂S computational cell, the governing equations are mass and momentum conservation. The mass conservation is reflected by the species transport equations, based on the assumptions above:

$$\frac{\partial}{\partial t}(\rho Y_i) + \nabla \cdot (\rho u Y_i) = -\nabla \cdot J_i \tag{6}$$

where in a turbulent flow

$$J_i = -(\rho D_{i,m} + \frac{\mu_t}{Sc_t})\nabla Y_i \tag{7}$$

where *t* is time, ρ is fluid density, Y_i is the mass fraction of species *i*, *u* is the mean instantaneous velocity, J_i is the diffusion flux of the species *i*, $D_{i,m}$ is the diffusion coefficient of species *i* in the mixture, Sc_t is the turbulence Schmidt number generally being equal to 0.7, and μ_t is the turbulence viscosity. The odor dispersion is dependent on the species gradient, the diffusion coefficients, and the turbulence viscosity [20]. Using Reynolds-averaged Navier–Stokes (RANS) method would be a good solution for the equations system with the two governing equations. CFD software offers several turbulent models. It has three different types, including realizable, RNG (Re-normalization group), and standard. These models perform differently in different cases. So, it is recommended that different studies choose their suitable models according to their own circumstances [21,22]. Using simulation results (linear fitting curves) of wind speed at 4 m/s and 16 m/s to match the onsite measured data is a reasonable method for estimating the suitable turbulence model.

2.4. Simulation Scheme and Boundary Conditions

For outdoor dispersion, the environment conditions would be unstable. Wind speed might be the dominant factor affecting outdoor H_2S no-barrier dispersion in a short distance [23]. Ana et al. [24] noticed that, at the same monitoring point, the summer H_2S concentration was much higher than winter. It was assumed that wind speed affected H_2S unorganized dispersion by changing turbulent state and mass transfer. Additionally, temperature mainly changed H_2S diffusion by affecting mass transfer through temperature difference. Herein, this study would verify these two factors on H_2S dispersion, respectively.

Control variable method was adopted to design the simulation plan. To analyze the effect of temperature, we controlled the wind speed at 4 m/s and changed the environment temperature from 288 K to 303 K. Observing the H₂S concentration change could help us understand the influence of temperature on H₂S dispersion. Similarly, for wind speed study, we kept environment temperature at 288 K and changed the wind speed from 2 m/s to 20 m/s to illustrate the influence on H₂S distribution. So, here is a list of all the environmental conditions we need to simulate (Table 2). The H₂S emission rates for model simulation at different temperatures and wind speeds are also shown below.

Wind Speed (m/s)	Temperature (K)	H ₂ S Mass Flow Rate (kg/s)
1.0	288	1.50
1.5	288	1.51
2.0	288	1.62
	288	2.18
4.0	293	2.23
4.0	298	2.29
	303	2.35
6.0	288	2.62
8.0	288	3.07
10.0	288	3.49
12.0	288	3.88
14.0	288	4.25
16.0	288	4.61
18.0	288	4.95
20.0	288	5.29

Table 2. H₂S emission rate at different temperatures and wind speeds.

Except taking wind speed and temperature as variables, turbulent intensity and turbulent viscosity ratios also need to be calculated as inputs in k- ε model. The first step is to calculate the Reynolds number and the turbulent intensity. The second step is to acquire turbulent viscosity ratio from turbulent kinetic energy k and its dissipation rate ε . The turbulent viscosity ratio increases with the Reynolds number. The boundary conditions change with the wind speed. To build the H₂S dispersion model, the assumptions taken are as follows. First, the flow is considered to be three-dimensional turbulence. Second,

the wind speed value is considered to not vary with height, and the wind is parallel to the ground. Third, gauge pressures of all pressure outlets are set as 0. Additionally, the CFD simulation solver in this study was PISO (Pressure-Implicit with Splitting of Operators), which is a pressure-based solver.

3. Results and Discussion

3.1. Subsection H₂S Concentration Results and Model Validation

The results of H_2S concentration measured on site are shown in Figure 2. Because of the sedimentation tank without any waste gas treatment equipment, the high concentration value was not out of expectation. At 1.5 m height, the H_2S distribution of two measurements were similar, and the H_2S concentration decreased with the distance increase away from the H_2S source. However, the environment conditions of the two measurements were different, and the effect of temperature and wind speed on the H_2S dispersion needed deeper analysis.

Before simulating, a suitable model needed to be ensured. About k- ε model selection, Muhammad et al. [25] found that, in the gas dispersion process of hydrogen, the turbulence model influence was very small for high Reynold number regions, such as the air injection source, but in the region far away from the injection source, where turbulence was weaker, the influence was larger. It meant that the choice of different k- ε models would have a great impact on the simulation results for turbulence with a small Reynolds number. They reported that the RNG and the standard models provided a better estimate of the region where turbulence is fully developed. To validate this opinion and find out the appropriate model for H₂S dispersion, realizable, RNG, and standard turbulence models were selected to simulate H_2S dispersion at 4 m/s and 16 m/s, because the simulation data could be fitted in six lines (Figure 3) to compare with the onsite data in Figure 2. There was a low accuracy for the standard model because the gap between the curves and the measurement points were large, since, for 4 m/s, the analogue results were generally lower than the field measurements, but for 16 m/s, they were higher than the field measurements. On the contrary, the performances of realizable were better in a high wind speed condition, but its results error of 4 m/s was larger than standard. The realizable model was still not accurate enough. The concentration curves of the RNG model under 4 m/s and 16 m/s had better down-trend synchronicity and were close to the measurement points. The result errors of RNG were less than 15%.



Figure 3. Validation of Realizable, RNG, and Standard model.

From the comparison above, it can be seen that there were significant differences between the three models, with RNG performing the best. Meanwhile, the applicability of the RNG model for the outdoor dispersion simulation of gaseous pollutants has been fully verified in the studies about dispersion around buildings [26–28]. Herein, the RNG model was chosen to simulate under various temperature and wind speeds.

3.2. H₂S Dispersion Simulation with Different Temperature

From the mass transfer coefficient formula (5), it was known that H_2S emission increases with temperature rise. Although, for the sewage sedimentation tank, the increase in H_2S production due to temperature rise is small [29]. Besides, Archana and Melanie [30] considered that the increase of H_2S production had little effect on the H_2S dispersion movement. The studies above illustrated that the temperature change would affect the H_2S concentration in two aspects: the production and the dispersion. As temperature rises, the production and the dispersion of H_2S would be enhanced simultaneously. Therefore, in order to further eliminate the interference, the H_2S mass flow rate was fixed at 2.18 kg/s (environment condition: 4 m/s and 288 K) to analyze the effect on dispersion caused by the enhancement of the molecular thermal.

The H_2S concentration decreased with the temperature rise in Figure 4, which meant the enhancement effect of high temperature on dispersion is greater than that on production. There were more H_2S coming out than coming in at the observing points. This was the reason for the H_2S concentration decrease. However, the differences were small, in a range from 0.16 to 0.21 ppm. The ratio of mean deviation to mean value of H_2S concentration reflected the temperature influence on H_2S dispersion (Table 3). Among the three monitoring points, the ratio was no more than 3.5%, which explained that the annual temperature change had little effect on H_2S dispersion. So, the big differences during the two measurements at the three points were mainly due to wind speed.



Figure 4. H_2S concentration at P1, P2, and P3 (0 m, 4 m, 11 m), with the temperature change (from 288 K to 303 K). The wind speed was constant at 4 m/s.

Table 3. Mean value, mean deviation, and ratio of mean deviation to mean value of H_2S concentration with different temperatures at P1, P2, and P3.

H ₂ S Concentration (ppm)	P1	P2	P3
Mean value	7.88	3.48	1.45
Mean deviation	0.055	0.063	0.051
Ratio	0.7%	1.8%	3.5%

3.3. H₂S Dispersion Simulation with Different Wind Speeds

Having wind speeds from 1 m/s to 20 m/s as the variable, the simulated H_2S plumes were shown in Figures 5 and 6. The legend shows the H_2S concentration fraction of the initial concentration Q, which was the initial H_2S concentration above the sewage pool closing to the water surface.



Figure 5. H₂S plume of wind speed 1 m/s, 1.5 m/s, and 2 m/s, controlling temperature at 298 K.

The change of H_2S concentration fraction revealed the dispersion tendency of H_2S . Generally, with wind speed increasing, H_2S concentration declined with height increase in y direction, influenced range enlarged with the increase of distance in x direction, and the H_2S plume eventually stabilized at a limit height about 3 m during the wind speed changes. The whole plume shape was consistent with the report of Abdullah and Weiming [31] about leaked gas having the same, or similar, density as ambient air. For more details, the research range was divided into two parts. Region 1 was the range above the sewage pool (x coordinate from -10 m to 10 m), and region 2 was the rest (x coordinate from 10 m to 25 m).

Region 1 mainly observed the wind speed influence on H_2S emission. It was evaluated by the H_2S high concentration range (H_2S concentration above 0.5Q) change. Before 2 m/s, the turbulence caused by wind was not strong enough to produce a big mass transfer coefficient for strong mass transfer. So, the H_2S high concentration range was small. At the outlet of the tank, the H_2S concentration reached 0.9Q. That is because the velocity ratio between odorous compounds velocity at the source outlet and wind speed was small [32]. The H_2S compounds tended to disperse vertically upward. This characteristic was also noticed in Figure 7, where, before 2 m/s, the H_2S concentrations at P1, P2, and P3 were close. However, from 4 m/s, the stronger turbulence improved the mass transfer, and the vertical dispersion was improved. Comparing the H_2S plumes of different wind speeds in Figures 5 and 6, the H_2S emission was strengthened with wind speed, increasing from 2 m/s to 10 m/s. High wind speeds let the ambient air have more power to carry the H_2S climbing up, presenting a larger influence height.



Figure 6. H₂S plume of wind speed changing from 4 m/s to 20 m/s, controlling temperature at 298 K.



Figure 7. H_2S concentration at P1, P2, and P3 with wind speed change only (from 2 m/s to 20 m/s). The temperature was a constant at 298 K.

Surprisingly, the growth trend of the high-concentration H_2S range in region 1 changed in high wind speed condition. When the wind speeds were faster than 10 m/s, the highconcentration H_2S influence range shrank. This phenomenon was studied in depth, and the reason was found to be the change of flow field (Figure 8).



Figure 8. Streamline graphics of wind speed 4 m/s, 10 m/s, and 18 m/s case. The diagonal line in the dotted box represents the slope at the right end of the streamline in the box.

In the three cases (4 m/s, 10 m/s, and 18 m/s) in Figure 8, the slope of streamline above region 1 was positive, which meant air flowed upwards. The slope was minimum at 4 m/s and maximum at 10 m/s. The slope of 18 m/s was in the middle, which was slightly smaller than 10 m/s (Figure 9). The characteristics of air flow were exactly the reason why the influence range of high-concentration H_2S first increased and then decreased.



Figure 9. Slope comparation of streamlines in region 1 and region 2 in the three cases (4 m/s, 10 m/s, and 18 m/s).

Region 2 mainly observed the influence of wind speed on the H_2S dispersion. When the wind speed was under 2 m/s, due to the small amount of H_2S emission, the whole region 2 was H₂S low-concentration range (H₂S concentration under 0.5Q). H₂S tended to disperse perpendicularly to the jet direction because the horizontal velocity was also small. Additionally, the edge of the plume was unsmooth. With wind speed augmenting to 4 m/s, the H₂S concentration changed quickly. This concentration rapidly changed, due to wind speed rise, which was also illustrated in Shen's research [33]. It was displayed by the concentration gradient in Figure 7. Anyway, the plume edge began to be smooth. It meant that more H₂S dispersed down-wind, instead of flying up, which increased the H₂S fraction in region 2. With the further increase of wind speed, the height of the influence range gradually rose again and stabilized at about 3 m under 8 m/s. After 10 m/s, the H_2S influence range was basically unchanged. This phenomenon was also found in the outdoor H₂S dispersion study of a landfill [34]. The whole H₂S influence range fluctuation in Figure 6 and the streamlines state in Figure 8 explain that, as the distance between the observation point and the leak source increased, the influence of air flow on the distribution of H_2S became more obvious, which was consistent with the report of Majid et al [35].

Meanwhile, the H₂S plume structure was also changed. With the increase of wind speed from 4 m/s to 12 m/s, the thickness of each concentration layer increased. However, this trend reversed from 14 m/s. As the wind speed continued to increase, the layers whose fractions were higher than 0.5Q became thinner, and other layers whose fractions were smaller than 0.5Q became thicker. The H₂S distribution changed because of the flow field change, which was the same as region 1 (Figure 9). In the three cases, focusing on the slopes of streamlines in region 2, it was noticeable that the changes in characteristics were similar to the slopes in region 1. Using *k* to represent the slope, we could obtain that $k_4 < k_{18} < k_{10}$. The change of streamline slope conducted the H₂S layer thickness change.

3.4. Safe Distance for Outdoor Sedimentation Tank

In certain industries, such as landfills, sewage treatment plants, and animal husbandry, workers face a higher risk of H₂S exposure. Therefore, H₂S occupational exposure is highly regarded, and relevant laws and regulations have been formulated in various countries [36,37].

As mentioned above, China specifies a H_2S maximum allowable concentration (MAC), which is 10 mg/m³. Uniformly converting the units of H_2S MAC to ppm for data analysis, 10 mg/m³ corresponded to 6.9 ppm. The initial measured max concentration Q was 13.34 ppm. So, the MAC of H_2S equaled 0.54Q. Hence, an area having H_2S concentration fraction higher than 0.54Q could be regarded as the unsafe region.

The area bounded by a 0.54Q curve and the x-axis was an unsafe region in Figure 10. The unsafe region of 2 m/s was about 11 m long and 0.5 m high, which was the distance away from the office monitoring point P2. As for lower wind speed, such as 1 m/s and 1.5 m/s, whose unsafe region were supposed to be smaller than 2 m/s, being similar with the H₂S plume change in region 1, the unsafe region first developed and then shrank with wind speed increase. According to the onsite data, mostly the wind speed in Jingzhou city was under 8 m/s, whose corresponding safe region was 17.2 m long and 1.5 m height. This is the most basic distance for workers to be safe, but the general safe region was supposed to be larger than 17.2 m. Under 12 m/s, the unsafe region reached its maximum range, which was 19.5 m long and 1.7 m height. With wind speed higher than 12 m/s, the unsafe region began to shrink.



Figure 10. Unsafe regions where H_2S concentration was higher than 10 mg/m³.

Anyway, analyzing the partial enlarged drawing in Figure 10, it was found that the slope of H_2S concentration under different wind speeds had the same characteristics as the streamlines of different wind speeds. The unsafe region border slope before 10 m/s was always greater than the slope after 10 m/s, which meant H_2S dispersed further down-wind after 10 m/s. It was concluded that, under the worst environment conditions of 12 m/s, the safe distance was over about 20 m long and 1.8 m height. Therefore, the office at point 2 was suitable for workers only when the wind speed was less than 2 m/s. Additionally, the sidewalk at point 3 could be used as a daily road. In summary, for this model scenario, a work office was better be set at least 10 m away from the wastewater pool (down-wind direction). The adjusted sewage pool model is shown below in Figure 11.



Figure 11. Modified model of building layout near automobile sewage pool.

4. Conclusions

For surface source unorganized H_2S dispersion modeling, RNG is the most suitable k- ε turbulence model that most closely matches the experimental measurements. Using the RNG model to explore the influences of temperature and wind speed on H_2S dispersion, the results show that temperature change indeed leads to different H_2S production and

different H_2S concentration at the monitoring points, but the effect is much smaller than wind speed for H_2S dispersion. The different wind speed changes the flow field. Taking the wind speed of 12 m/s as the dividing line, H_2S dispersion shows two states: enhanced (less than 12 m/s) and unchanged (more than 12 m/s).

In this case, the CFD simulation provides a visualization for H_2S dispersion at short distance outdoors. Between 288 K to 303 K, the H_2S concentration outside the sewage sedimentation tank does not reach the occupational exposure limit when the wind speed is less than 2 m/s. However, when the wind speed increases from 4 m/s to 12 m/s, the safety distance increases from 7 m to the maximum of 10 m. Therefore, for the sewage sedimentation tank with sewage treatment capacity less than or equal to 10 m³/h, it is recommended that workshops or offices be built 10 m away from the pollution source. This guidance would have value for practical industrial environment layouts.

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References

- 1. Li, W.; Yang, W.; Li, J. Characterization and prediction of odours from municipal sewage treatment plant. *Water Sci. Technol.* 2018, 2017, 762–769. [CrossRef] [PubMed]
- 2. Guarrasi, J.; Trask, C.; Kirychuk, S. A Systematic Review of Occupational Exposure to Hydrogen Sulfide in Livestock Operations. *J. Agromed.* **2015**, *20*, 225–236. [CrossRef] [PubMed]
- Colomer, F.L.; Morató, H.E.; Iglesias, E.M. Estimation of hydrogen sulfide emission rates at several wastewater treatment plants through experimental concentration measurements and dispersion modeling. *J. Air Waste Manag. Assoc.* 2012, 62, 758–766. [CrossRef] [PubMed]
- 4. Liu, Y.; Lu, W.; Wang, H.; Gao, X. Odor impact assessment of trace sulfur compounds from working faces of landfills in Beijing, China. *J. Environ. Manag.* 2018, 220, 136–141. [CrossRef]
- 5. Yalamanchili, C.; Smith, M.D. Acute hydrogen sulfide toxicity due to sewer gas exposure. *Am. J. Emerg. Med.* 2008, 26, 518.e5–518.e7. [CrossRef]
- Jiang, Q.; Li, T.; He, Y.; Wu, Y.; Zhang, J.; Jiang, M. Simultaneous removal of hydrogen sulfide and ammonia in the gas phase: A review. *Environ. Chem. Lett.* 2022, 20, 1403–1419. [CrossRef]
- Bu, H.; Carvalho, G.; Yuan, Z.; Bond, P.; Jiang, G. Biotrickling filter for the removal of volatile sulfur compounds from sewers: A review. *Chemosphere* 2021, 277, 130333. [CrossRef]
- Jiang, Z.A.; Wang, Y.P.; Men, L.G. Ventilation control of tunnel drilling dust based on numerical simulation. J. Cent. South Univ. 2021, 25, 1342–1356. [CrossRef]
- Nuvolone, D.; Petri, D.; Pepe, P.; Voller, F. Health effects associated with chronic exposure to low-level hydrogen sulfide from geothermoelectric power plants. A residential cohort study in the geothermal area of Mt. Amiata in Tuscany. *Sci. Total Environ.* 2019, 659, 973–982. [CrossRef] [PubMed]
- Haouzi, P.; Tubbs, N.; Cheung, J.; Judenherc-Haouzi, A. Methylene Blue Administration During and After Life-Threatening Intoxication by Hydrogen Sulfide: Efficacy Studies in Adult Sheep and Mechanisms of Action. *Toxicol. Sci.* 2019, 168, 443–459. [CrossRef] [PubMed]
- Selene, C.H.; Chou, J. Concise International Chemical Assessment Document 53: Hydrogen sulfide: Human health aspects. In IPCS Concise International Chemical Assessment Documents; World Health Organization: Geneva, Switzerland, 2003; p. 53.
- GBZ 2.1-2019; Occupational Exposure Limits for Hazardous Agents in the Workplace—Part 1: Chemical Hazardous Agents. Available online: http://www.nhc.gov.cn/wjw/pyl/202003/67e0bad1fb4a46ff98455b5772523d49.shtml (accessed on 13 September 2022).
- Walewska, A.; Szewczyk, A.; Koprowski, P. Gas Signaling Molecules and Mitochondrial Potassium Channels. *Int. J. Mol. Sci.* 2018, 19, 3227. [CrossRef] [PubMed]
- 14. Somma, R.; Granieri, D.; Troise, C.; Terranova, C.; De Natale, G.; Pedone, M. Modelling of hydrogen sulfide dispersion from the geothermal power plants of Tuscany (Italy). *Sci. Total Environ.* **2017**, *583*, 408–420. [CrossRef] [PubMed]

- Moreno-Silva, C.; Calvo, D.C.; Torres, N.; Ayala, L.; Gaitán, M.; González, L.; Rincón, P.; Susa, M.R. Hydrogen sulphide emissions and dispersion modelling from a wastewater reservoir using flux chamber measurements and AERMOD[®] simulations. *Atmos. Environ.* 2020, 224, 117263. [CrossRef]
- 16. Gulia, S.; Kumar, A.; Khare, M. Performance evaluation of CALPUFF and AERMOD dispersion models for air quality assessment of an industrial complex. *J. Sci. Ind. Res.* 2015, 74, 302–307.
- Wu, C.; Yang, F.; Brancher, M.; Liu, J.; Qu, C.; Piringer, M.; Schauberger, G. Determination of ammonia and hydrogen sulfide emissions from a commercial dairy farm with an exercise yard and the health-related impact for residents. *Environ. Sci. Pollut. Res.* 2020, *27*, 37684–37698. [CrossRef]
- 18. Asadollahfardi, G.; Mazinani, S.; Asadi, M.; Mirmohammadi, M. Mathematical and experimental study of hydrogen sulfide concentrations in the Kahrizak landfill, Tehran, Iran. *Environ. Eng. Res.* **2019**, *24*, 572–581. [CrossRef]
- Zhang, B.; Chen, G. Hydrogen Sulfide Dispersion Consequences Analysis in Different Wind Speeds: A CFD Based Approach. In Proceedings of the International Conference on Energy and Environment Technology, Washington, DC, USA, 16–18 October 2009; pp. 365–368.
- 20. Lin, X.; Barrington, S.; Gong, G.; Choinière, D. Simulation of odour dispersion downwind from natural windbreaks using the computational fluid dynamics standard k-ε model. *Can. J. Civ. Eng.* **2009**, *36*, 895–910. [CrossRef]
- 21. Tominaga, Y.; Stathopoulos, T. Numerical simulation of dispersion around an isolated cubic building: Comparison of various types of k-ε models. *Atmos. Environ.* **2009**, *43*, 3200–3210. [CrossRef]
- Moen, A.; Mauri, L.; Narasimhamurthy, V.D. Comparison of k-ε models in gaseous release and dispersion simulations using the CFD code FLACS. *Process. Saf. Environ. Prot.* 2019, 130, 306–316. [CrossRef]
- 23. Abdul-Wahab, S.A.; Chan, K.; Elkamel, A.; Ahmadi, L. Effects of meteorological conditions on the concentration and dispersion of an accidental release of H₂S in Canada. *Atmos Environ.* **2014**, *82*, 316–326. [CrossRef]
- Godoi, A.; Grasel, A.M.; Polezer, G.; Brown, A.; Potgieter-Vermaak, S.; Scremim, D.C.; Yamamoto, C.I.; Godoi, R.H.M. Human exposure to hydrogen sulphide concentrations near wastewater treatment plants. *Sci. Total Environ.* 2018, 610–611, 583–590. [CrossRef] [PubMed]
- Saeed, M.; Yu, J.; Abdalla, A.A.A.; Zhong, X.-P.; Ghazanfar, M.A. An assessment of k-ε turbulence models for gas distribution analysis. *Nucl. Sci. Tech.* 2017, 28, 146. [CrossRef]
- 26. Mirzaei, F.; Mirzaei, F.; Kashi, E. Turbulence Model Selection for Heavy Gases Dispersion Modeling in Topographically Complex Area. J. Appl. Fluid. Mech. 2019, 12, 1745–1755. [CrossRef]
- 27. Ben Ramoul, L.; Korichi, A.; Popa, C.; Zaidi, H.; Polidori, G. Numerical study of flow characteristics and pollutant dispersion using three RANS turbulence closure models. *Environ. Fluid. Mech.* **2019**, *19*, 379–400. [CrossRef]
- Lateb, M.; Masson, C.; Stathopoulos, T.; Bédard, C. Comparison of various types of k-ε models for pollutant emissions around a two-building configuration. J. Wind. Eng. Ind. Aerodyn. 2013, 115, 9–21. [CrossRef]
- 29. Zwain, H.M.; Nile, B.K.; Faris, A.M.; Vakili, M.; Dahlan, I. Modelling of hydrogen sulfide fate and emissions in extended aeration sewage treatment plant using TOXCHEM simulations. *Sci. Rep.* **2020**, *10*, 22209. [CrossRef] [PubMed]
- Nagaraj, A.; Sattler, M.L. Correlating emissions with time and temperature to predict worst-case emissions from open liquid area sources. *Air Repair* 2005, 55, 1077–1084. [CrossRef] [PubMed]
- Alakalabi, A.; Liu, W. Numerical investigation into the effects of obstacles on heavy gas dispersion in the atmosphere. In Proceedings of the XII International Conference on Computational Heat, Mass and Momentum Transfer (ICCHMT 2019), Preston, UK, 8 November 2019.
- 32. Maïzi, A.; Dhaouadi, H.; Bournot, P.; Mhiri, H. CFD prediction of odorous compound dispersion: Case study examining a full scale waste water treatment plant. *Biosyst. Eng.* **2010**, *106*, 68–78. [CrossRef]
- Shen, S.; Wu, B.; Xu, H.; Zhang, Z.-Y. Assessment of Landfill Odorous Gas Effect on Surrounding Environment. Adv. Civ. Eng. 2020, 2020, 8875393. [CrossRef]
- Feng, Y.; Eun, J.; Moon, S.; Nam, Y. Assessment of gas dispersion near an operating landfill treated by different intermediate covers with soil alone, low-density polyethylene (LLDPE), or ethylene vinyl alcohol (EVOH) geomembrane. *Environ. Sci. Pollut. Res.* 2022, 1–16. [CrossRef]
- 35. Bayatian, M.; Azari, M.R.; Ashrafi, K.; Jafari, M.J.; Mehrabi, Y. CFD simulation for dispersion of benzene at a petroleum refinery in diverse atmospheric conditions. *Environ. Sci. Pollut. Res.* **2021**, *28*, 32973–32984. [CrossRef] [PubMed]
- List of Highly Hazardous Chemicals, Toxics and Reactives (Mandatory). Available online: https://www.osha.gov/laws-regs/ regulations/standardnumber/1910/1910.119AppA (accessed on 13 September 2022).
- 37. Costigan, M.G. Hydrogen sulfide: UK occupational exposure limits. Occup. Environ. Med. 2003, 60, 308–312. [CrossRef] [PubMed]