

Article

# Seasonal Variations of Atmospheric Pollution and Air Quality in Beijing

Wei Chen 1,2,\*, Lei Yan 2,\* and Haimeng Zhao 2

- <sup>1</sup> College of Geoscience and Surveying Engineering, China University of Mining and Technology, Beijing 100083, China
- <sup>2</sup> Beijing Key Lab of Spatial Information Integration and Its Applications, Peking University, Beijing 100081, China; E-Mail: zhaohaimeng@163.com
- \* Authors to whom correspondence should be addressed; E-Mails: chenw@cumtb.edu.cn (W.C.); lyan@pku.edu.cn (L.Y.); Tel.: +86-138-119-50643 (W.C.); +86-010-62759765 (L.Y.).

Academic Editor: Robert W. Talbot

Received: 30 September 2015 / Accepted: 4 November 2015 / Published: 18 November 2015

Abstract: New ambient air quality standards were released in 2012 and implemented in 2013 with real time monitoring data publication of six atmospheric pollutants: particulate matter (PM)<sub>2.5</sub>, PM<sub>10</sub>, O<sub>3</sub>, SO<sub>2</sub>, NO<sub>2</sub> and CO. According to the new standards, Beijing began to publicize real-time monitoring data of 35 monitoring stations in 2013. In this study, real time concentrations of all six atmospheric pollutants of all 35 monitoring stations were collected from September 2014 to August 2015 to investigate the spatial and temporal pattern of the air quality and atmospheric pollutants. By comparing the annual and seasonal variations of all six pollutants' concentrations, it was found that particulate matter, especially PM<sub>2.5</sub>, is still the major contributor to the deterioration of air quality in Beijing. Although the NO<sub>2</sub> and O<sub>3</sub> concentrations of some stations were still high under certain circumstances, their contributions to air quality index (AQI) were not comparable to those of PM2.5 and PM<sub>10</sub>. SO<sub>2</sub> and CO concentrations have dropped to well below the qualification standards. Winter and autumn were the most polluted seasons for all pollutants except O<sub>3</sub>, whose concentrations are higher in summer. South and southeast stations were the most polluted compared with the rest of the stations, especially for particulate matter. Wind profile analysis with heavy pollution situations indicates that low speed southwest or east wind situations have the higher possibility of heavy pollution, suggesting that it is highly possible that long-range transportation of air pollutants from south or east neighboring provinces played an important role in the worsening air conditions in Beijing.

**Keywords:** air quality; air quality standards; PM<sub>2.5</sub>; PM<sub>10</sub>; atmospheric pollution

### 1. Introduction

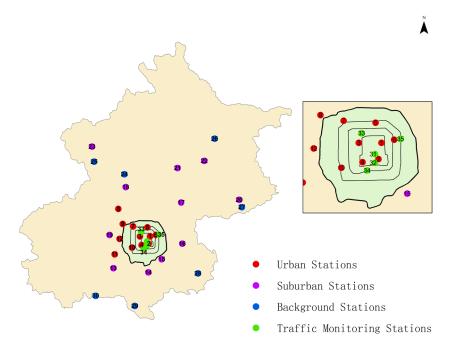
Air quality is a major concern for people living in Beijing [1,2], especially after several serious haze-fog events since 2011 [3,4]. However, the air quality evaluation standards released in 1996 (NAAQS-1996), did not take into account PM<sub>2.5</sub> (Particulate Matter with aerodynamic diameter less than 2.5  $\mu$ m) [5]. As a result, the air quality attainment rate of Beijing under NAAQS-1996 (Air Pollution Index, API < 100) has always been greater than 70% since 2008 [6]. This is inconsistent with public awareness since people are suffering by the worsening air pollution [7,8].

In China, air quality monitoring started from the mid-1980s while the release of daily air pollution levels began from 2000 under NAAQS-1996 by taking into account the daily average concentrations of PM<sub>10</sub>, SO<sub>2</sub> and NO<sub>2</sub>. Twelve years of daily monitoring data provided a effective data source for the analysis of air pollution level of major Chinese cities [9]. However, a series of studies suggested that PM<sub>2.5</sub> in major northern Chinese cities had been an important issue, as although the air pollution index (API) under NAAQS-1996 suggested that air quality was fine, the atmospheric conditions were poor [10–15]. In February 2012, China then released new air quality standards (NAAQS-2012) by taking into account six air pollutants: PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, O<sub>3</sub>, NO<sub>2</sub> and CO [16]. From January 2013, Beijing adopted NAAQS-2012 to publicize real time air monitoring data at one hour intervals along with 73 other cities by deploying 35 permanent air quality monitoring stations. These monitoring data provided unique tools to analyze the present atmospheric pollution levels for Beijing, the capital of China.

In this paper, we collected hourly air quality data of all 35 air quality monitoring stations in Beijing from September 2014 to August 2015 to give a comprehensive analysis of the air quality evaluations under the new national ambient air quality standards.

## 2. Data and Methods

From January 2013, Beijing began to publicize real time air quality monitoring data under National Ambient Air Quality Standards released in 2012 (NAAQS-2012) at one hour intervals on a municipal web platform [17]. In this paper, hourly monitoring concentrations of all six pollutants (PM<sub>2.5</sub>, PM<sub>10</sub>, O<sub>3</sub>, SO<sub>2</sub>, NO<sub>2</sub> and CO) from September 2014 to August 2015 were collected through deploying a web download program. In total, more than 300,000 hourly monitoring data were collected and analyzed in this paper. The 35 monitoring stations were divided into six groups: 12 urban stations, 11 suburban stations (four in north and seven in the south of Beijing), seven background stations (four in the north and three in the south) and five traffic monitoring stations, shown in Figure 1 and Table 1. Although the real-time web platform publicizes current monitoring data, the history data two weeks ahead are not provided publicly. Therefore, our web deployed program accessed to the web platform at an interval of half an hour to download the real-time data. Due to equipment failure or internet transfer error, some data were missing during the collection. According to NAAQS-2012, daily observations for at least 20 h were supposed to be valid, while the rest were abandoned to ensure the representativeness of the data.



**Figure 1.** Distribution and station IDs of the 35 air quality monitoring stations. The four rings in the map are the 2nd, 3rd, 4th, and 5th Ring Road in Beijing from inner to outside.

The 12 urban stations were deployed in the urban areas of Beijing, while the 11 suburban stations were located at the downtown areas of suburban counties in Beijing. The seven background stations were placed at those areas far away from human activities: Station 24 is the background for urban areas; Stations 25 and 26 are for monitoring the transportation from north and northwest directions; Stations 27 and 28 in the northeast and southeast mainly monitored the transportation from nearby Tianjin; Stations 29 and 30 in the south and southwest were for monitoring the data from the heavily polluted Hebei Province. Different to the previous 30 stations, the five traffic monitoring stations were placed just next to the main road as suggested in Table 1, while the other 30 stations were 150 m away from the main road according to NAAQS-2012.

**Table 1.** Category and locations of monitoring stations.

Station ID & Name	Category		
1 Dongsi	Urban Stations		
2 Temple of Heaven	Urban Stations		
3 West Park Officials	Urban Stations		
4 West Wanshou Nishinomiya	Urban Stations		
5 Olympic Sports Center	Urban Stations		
6 Agricultural Exhibition Hall	Urban Stations		
7 Wanliu	Urban Stations		
8 Northern New Area	Urban Stations		
9 Botanical Garden	Urban Stations		
10 Fengtai garden	Urban Stations		
11 Yungang	Urban Stations		
12 Shijingshan city	Urban Stations		
13 Liangxiang	Suburban Stations (South)		
14 Daxing	Suburban Stations (South)		
15 Yizhuang	Suburban Stations (South)		

Table 1. Cont.

Station ID & Name	Category		
16 Tongzhou	Suburban Stations (South)		
17 Shunyi	Suburban Stations (North)		
18 Changping	Suburban Stations (North)		
19 Mentougou	Suburban Stations (South)		
20 Pinggu	Suburban Stations (North)		
21 Huairou	Suburban Stations (North)		
22 Miyun	Suburban Stations (North)		
23 Yanqing	Suburban Stations (North)		
24 Dingling	Background Stations (North)		
25 Badaling	Background Stations (North)		
26 Miyun Reservoir	Background Stations (North)		
27 Donggaocun	Background Stations (North)		
28 Yongledian	Background Stations (South)		
29 Yufa	Background Stations (South)		
30 Liulihe	Background Stations (South)		
31 Qianmen East Street	Traffic Monitoring Stations		
32 Yongdingmen Inner Street	Traffic Monitoring Stations		
33 Xizhimen North Street	Traffic Monitoring Stations		
34 South 3rd Ring Road	Traffic Monitoring Stations		
35 East 4th Ring Road	Traffic Monitoring Stations		

**Table 2.** Concentration limits for air quality index (AQI) calculation.

IAQI	PM <sub>10</sub>	SO <sub>2</sub>	NO <sub>2</sub>	PM <sub>2.5</sub>	1-Hour Peask	CO
	$(\mu g/m^3)$	$(\mu g/m^3)$	$(\mu g/m^3)$	$(\mu g/m^3)$	$O_3 (\mu g/m^3)$	$(mg/m^3)$
50	50	50	40	35	160	2
100	150	150	80	75	200	4
150	250	475	180	115	300	14
200	350	800	280	150	400	24
300	420	1600	565	250	800	36
400	500	2100	750	350	1000	48
500	600	2620	940	500	1200	60

The AQI is an index that indicates the pollution level of the atmosphere, ranging from 0 to 500. The higher the AQI value is, the heavier the atmospheric pollution is. According to NAAQS-2012, PM<sub>2.5</sub>, PM<sub>10</sub>, ozone (O<sub>3</sub>), sulfur dioxide (SO<sub>2</sub>), nitrogen dioxide (NO<sub>2</sub>) and carbon monoxide (CO) were included into the calculation of the AQI. The first step in calculating the AQI is to calculate the IAQI (individual air quality index) for each pollutant. The IAQI of each pollutant mentioned above is calculated as follows:

$$IAQI_{P} = \frac{IAQI_{Hi} - IAQI_{Lo}}{BP_{Hi} - BP_{Lo}}(C_{P} - BP_{Lo}) + IAQI_{Lo}$$
(1)

where IAQI<sub>P</sub> is the individual air quality index for pollutant P (PM<sub>2.5</sub>, PM<sub>10</sub>, O<sub>3</sub>, SO<sub>2</sub>, NO<sub>2</sub> and CO), and  $C_P$  is the daily mean concentration of pollutant P.  $BP_{Hi}$  and  $BP_{Lo}$  are the nearby high and low values of  $C_P$  as shown in Table 2.  $IAQI_{Hi}$ , and  $IAQI_{Lo}$  are the individual air quality indexes in terms of  $BP_{Hi}$  and  $BP_{Lo}$  as shown in Table 2. The largest IAQI value is 500, and once the air pollutant's concentration

exceeds the highest limit in Table 2, the IAQI will be set to 500. After the calculation of each IAQI<sub>P</sub>, the AQI is then calculated by choosing the max IAQI<sub>P</sub> as follows:

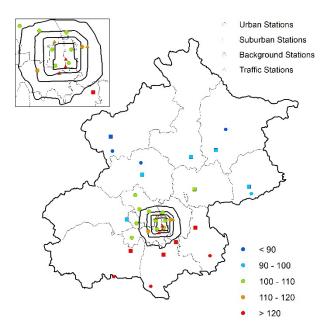
$$AQI = \max(IAQI_1, ..., IAQI_n)$$
 (2)

Equation (2) suggests that the AQI is not the sum contribution of all of the air pollutants but rather the maximum value of the IAQI. The air pollutant with a maximum IAQI when the AQI is larger than 50 is designated as the Primary Pollutant. Daily AQI less than 100 is supposed to be qualified according to NAAQS-2012.

### 3. Results

# 3.1. Air Quality Attainment Rate under the New National Ambient Air Quality Standards

The annual averaged AQI for all 35 stations is shown in Figure 2. A distinguished spatial pattern of AQI could be observed from Figure 2: south and southeast stations (both suburban and background) as well as three traffic monitoring stations in the south were those having the worst air qualities (annual AQI > 120). Urban stations and two north traffic monitoring stations followed with annual AQI in the range of 100 to 120, which is still exceeding the attainment level of 100. Suburban stations and north background stations in the north of Beijing had the best air qualities of AQI less than 100. The four remote north stations even had annual AQIs less than 90.



**Figure 2.** Annual average AQIs of the 35 monitoring stations.

The attainment rates (rates of AQI < 100 or API < 100) for the 35 monitoring stations were calculated and grouped into the six categories, as shown in Figure 3. It could be determined that the air quality attainment rates have all fallen by 15%–22% under NAAQS-2012 with south background stations having the largest attainment rate decrease. Under both standards, north background stations were those having the highest air quality attainment rates (88% for NAAQS-1996 and 73% for NAAQS-2012) while south background stations had the lowest air quality attainment rates (69% for NAAQS-1996 and 47%

for NAAQS-2012). These results suggested that the shift from NAAQS-1996 to NAAQS-2012 have had greatly decreased the air quality attainment rates for all stations by stricter standards but the spatial pattern remained similar: north background had the best air qualities while south background had the worst ones.

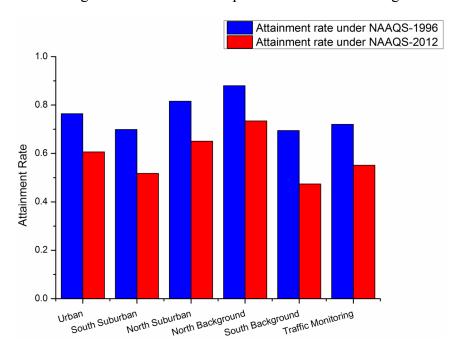
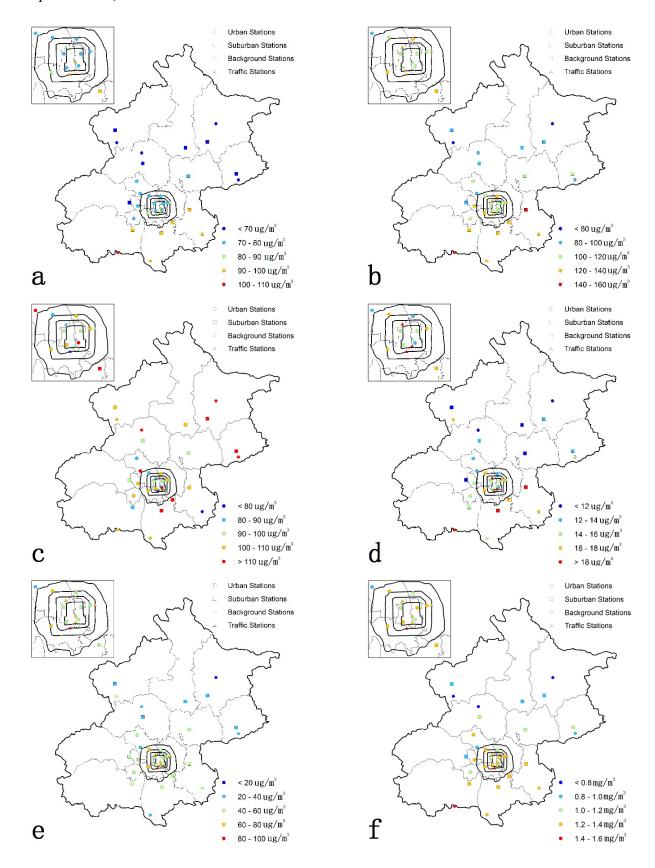


Figure 3. Annual AQI attainment rate of the six monitoring stations groups.

# 3.2. Annual Variations of Air Quality and Atmospheric Pollutants

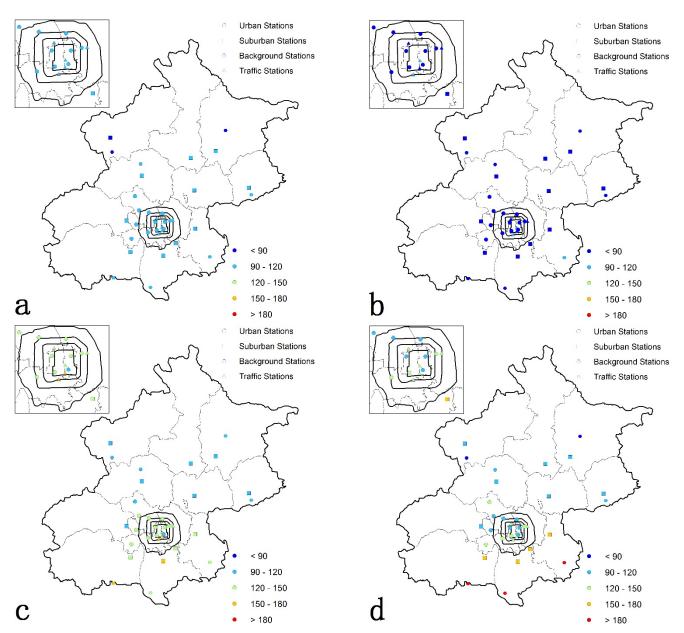
The annual mean concentrations of the five pollutants and mean daily maximum O<sub>3</sub> concentrations for the 35 monitoring stations were shown in Figure 4. The spatial patterns of PM<sub>2.5</sub> and PM<sub>10</sub> are similar to that of AQI: South and southeast stations were the most polluted while the north stations had the lower PM<sub>2.5</sub> and PM<sub>10</sub> concentrations. For O<sub>3</sub>, the situation is different and no obvious spatial pattern could be obtained except that the O<sub>3</sub> concentrations for all five traffic monitoring stations were all significantly lower than their nearby stations. For SO<sub>2</sub>, the spatial pattern is similar to that of PM<sub>10</sub> as shown in Figure 4d. Although the southeast stations had the higher concentrations of SO<sub>2</sub>, they were all far less than 50 μg/m<sup>3</sup>, with IAQI less than 50, suggesting that SO<sub>2</sub> has not been the major pollutants. NO<sub>2</sub>, a kind of vehicle exhaust, has a different spatial pattern that traffic monitoring stations and those stations near the main ring roads have had higher concentrations. For the rest stations, north stations have obvious lower concentrations of NO<sub>2</sub> than those of south stations. The pattern of CO is similar to that of PM<sub>10</sub>: stations at the urban, south and southeast had higher concentrations of CO than those north stations. However, the annual average CO of all 35 stations were less than 2 mg/m<sup>3</sup>, with IAQI of CO less than 50 according to Table 2. The spatial distribution of the six pollutants suggested that annual O<sub>3</sub>, SO<sub>2</sub> and CO concentrations were all qualified according to NAAQS-2012. PM<sub>2.5</sub>, PM<sub>10</sub> and NO<sub>2</sub>, were the major pollution pollutants as some monitoring stations' annual concentrations of these three pollutants exceeded the level of attainment (IAQI > 100).



**Figure 4.** Annual average concentrations of six pollutants ((a) PM<sub>2.5</sub>; (b) PM<sub>10</sub>; (c) O<sub>3</sub>; (d) SO<sub>2</sub>; (e) NO<sub>2</sub>; (f) CO) of the 35 monitoring stations.

# 3.3. Seasonal Variations of AQI

The air quality of Beijing shows an obvious seasonal pattern as in Figure 5. Summer had the best air quality with nearly all monitoring stations' average AQI less than 100. Spring followed with most stations average AQI in the range of 90–120. Air quality in winter deteriorated as most urban and south stations' air quality rose to larger than 120 with four even greater than 150. Winter's air quality had become even worse with southeast stations' air quality rising to more than 180.

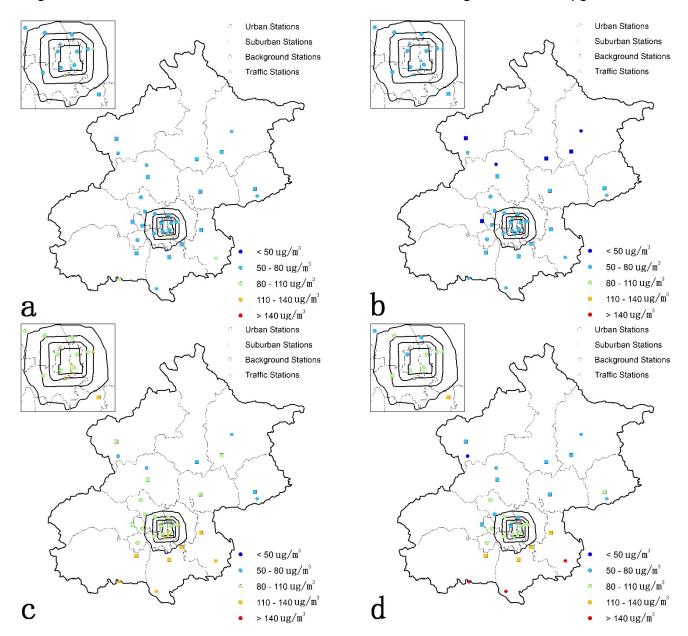


**Figure 5.** Seasonal variations of AQI for the 35 air quality monitoring stations ((a) spring; (b) summer; (c) autumn; (d) winter).

# 3.4. Seasonal Variations of PM<sub>2.5</sub>

The seasonal average PM<sub>2.5</sub> concentrations of all monitoring stations were less than 80  $\mu$ g/m<sup>3</sup>, three suburban stations in the north, one suburban station in the west and two background stations in the north

even had seasonal average PM<sub>2.5</sub> concentrations less than 50  $\mu$ g/m<sup>3</sup> as shown in Figure 6a,b. PM<sub>2.5</sub> concentrations increased sharply in autumn: most stations' PM<sub>2.5</sub> concentrations increased to more than 80  $\mu$ g/m<sup>3</sup> except six stations in the north as shown in Figure 6c. Stations in the south and southeast even had seasonal average PM<sub>2.5</sub> concentrations in the range of 110–140  $\mu$ g/m<sup>3</sup>. In the winter, PM<sub>2.5</sub> concentrations in the north background stations and suburban stations decreased while those in southeast and south background stations increased with three stations PM<sub>2.5</sub> concentrations greater than 140  $\mu$ g/m<sup>3</sup>.

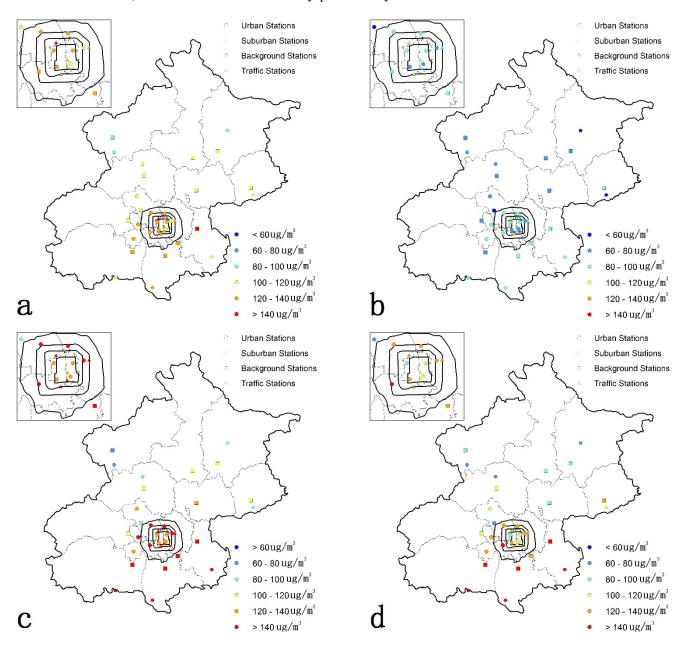


**Figure 6.** Seasonal variations of  $PM_{2.5}$  concentrations for the 35 air quality monitoring stations ((a) spring; (b) summer; (c) autumn; (d) winter).

## 3.5. Seasonal Variations of PM<sub>10</sub>

Most stations'  $PM_{10}$  concentrations in summer were less than  $100 \mu g/m^3$  except two traffic monitoring stations, as shown in Figure 7b. South stations, urban stations and traffic monitoring stations tended to have higher  $PM_{10}$  concentrations than north stations. Spring had higher average  $PM_{10}$  concentrations

than summer with most stations' average  $PM_{10}$  concentrations greater than  $100 \mu g/m^3$ . Autumn and winter, shown in Figure 7c,d had higher  $PM_{10}$  concentrations than those of summer and spring, with most urban and south stations' average  $PM_{10}$  concentrations higher than  $140 \mu g/m^3$ . By comparing winter and autumn, autumn was more seriously polluted by  $PM_{10}$ .

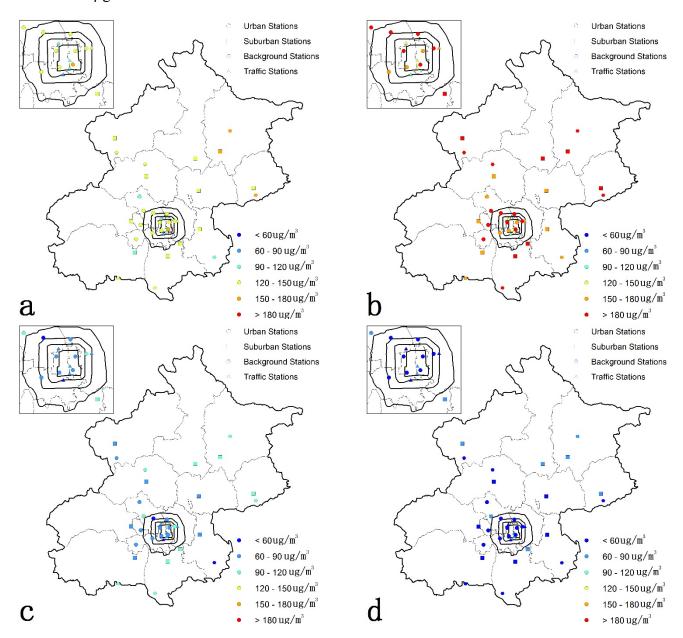


**Figure 7.** Seasonal variations of  $PM_{10}$  concentrations for the 35 air quality monitoring stations ((a) spring; (b) summer; (c) autumn; (d) winter).

# 3.6. Seasonal Variations of O<sub>3</sub>

Seasonal variations of  $O_3$  were different to those of  $PM_{2.5}$  and  $PM_{10}$ . Summer was the most polluted by  $O_3$  with most stations' averaged maximum daily  $O_3$  concentrations higher than  $150 \, \mu g/m^3$ . Only four traffic monitoring stations' averaged maximum daily  $O_3$  concentrations lower than  $120 \, \mu g/m^3$  shown in Figure 8b. Spring's averaged maximum daily  $O_3$  concentrations had fallen to less than  $120 \, \mu g/m^3$  for

most stations. Traffic monitoring stations and two stations in the west had the lowest  $O_3$  concentrations as shown in Figure 8a. The  $O_3$  concentrations dropped significantly to less than 90  $\mu g/m^3$  for autumn and even 60  $\mu g/m^3$  for winter.

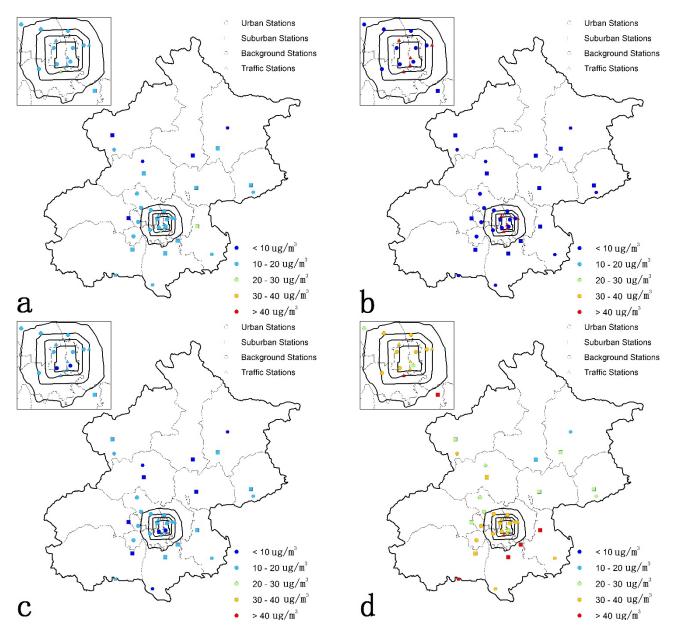


**Figure 8.** Seasonal variations of O<sub>3</sub> concentrations for the 35 air quality monitoring stations ((a) spring; (b) summer; (c) autumn; (d) winter).

# 3.7. Seasonal Variations of SO<sub>2</sub>

Figure 9 shows the seasonal variations of SO<sub>2</sub> concentrations for the 35 monitoring stations in Beijing. Concentrations of SO<sub>2</sub> in spring, summer and autumn were all very low to less than 20  $\mu$ g/m<sup>3</sup>. Spring and autumn had comparable SO<sub>2</sub> concentrations while summer had the lowest SO<sub>2</sub> concentrations with most stations' SO<sub>2</sub> concentrations lower than 10  $\mu$ g/m<sup>3</sup> except the five traffic monitoring stations whose average SO<sub>2</sub> concentrations higher than 40  $\mu$ g/m<sup>3</sup>. Winter time was the most SO<sub>2</sub> polluted season with most stations' SO<sub>2</sub> concentrations higher than 20  $\mu$ g/m<sup>3</sup> which is consistent with the heating period in

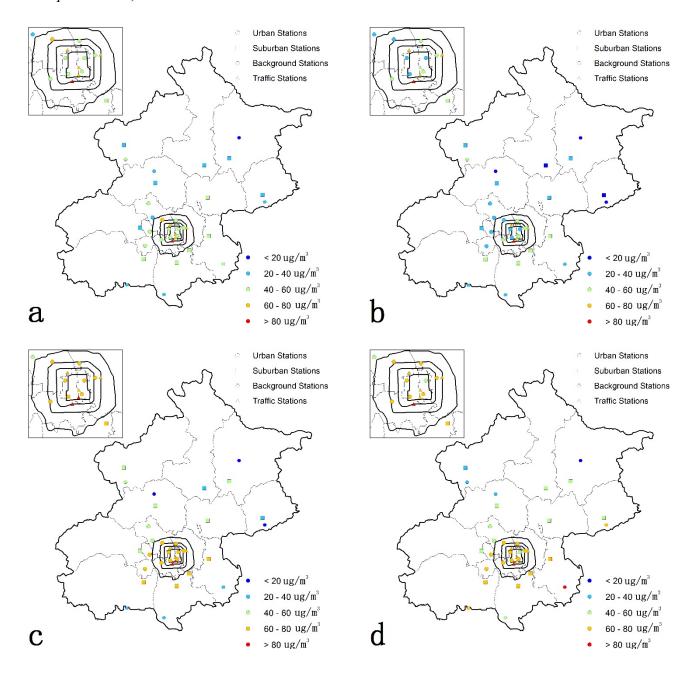
Beijing during the winter time with large amount of coal combustion. The south and southeast stations were affected most by  $SO_2$  pollution with average  $SO_2$  concentrations higher than 30 or even 40  $\mu$ g/m<sup>3</sup>.



**Figure 9.** Seasonal variations of SO<sub>2</sub> concentrations for the 35 air quality monitoring stations ((a) spring; (b) summer; (c) autumn; (d) winter).

# 3.8. Seasonal Variations of NO2

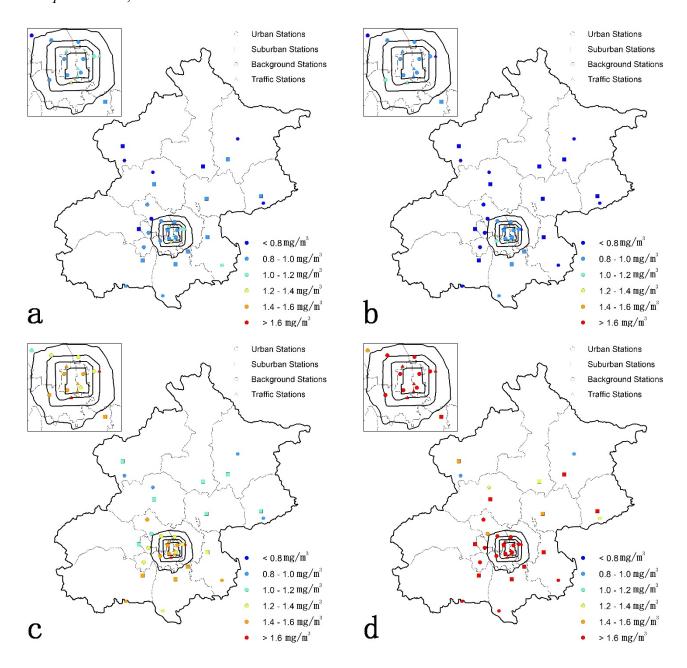
Autumn and winter had higher  $NO_2$  concentrations than those of spring and summer as shown in Figure 10. In autumn and winter,  $NO_2$  concentrations were higher in south stations and traffic monitoring stations than the rest stations in the north. In spring and summer, most stations'  $NO_2$  concentrations were less than  $40 \,\mu\text{g/m}^3$  except the traffic monitoring stations and stations. It is obvious that stations in the south were more polluted by  $NO_2$  than those in the north as the southern parts of Beijing were more affected by the traffic.



**Figure 10.** Seasonal variations of NO<sub>2</sub> concentrations for the 35 air quality monitoring stations ((a) spring; (b) summer; (c) autumn; (d) winter).

# 3.9. Seasonal Variations of CO

CO concentrations in spring and summer were very low to less than 1.0 mg/m³ or even 0.8 mg/m³ as shown in Figure 11. CO concentrations began to rise from winter to more than 1.2 mg/m³ for most monitoring stations in south area shown in Figure 11c. Winter was the most CO polluted season with most monitoring stations having average CO concentrations higher than 1.6 mg/m³ except seven stations in the north. The Maximum average CO concentration even reached 4.1 mg/m³, suggesting that the heating period in winter brought a lot of CO pollution for Beijing.



**Figure 11.** Seasonal variations of CO concentrations for the 35 air quality monitoring stations ((a) spring; (b) summer; (c) autumn; (d) winter).

# 4. Discussion

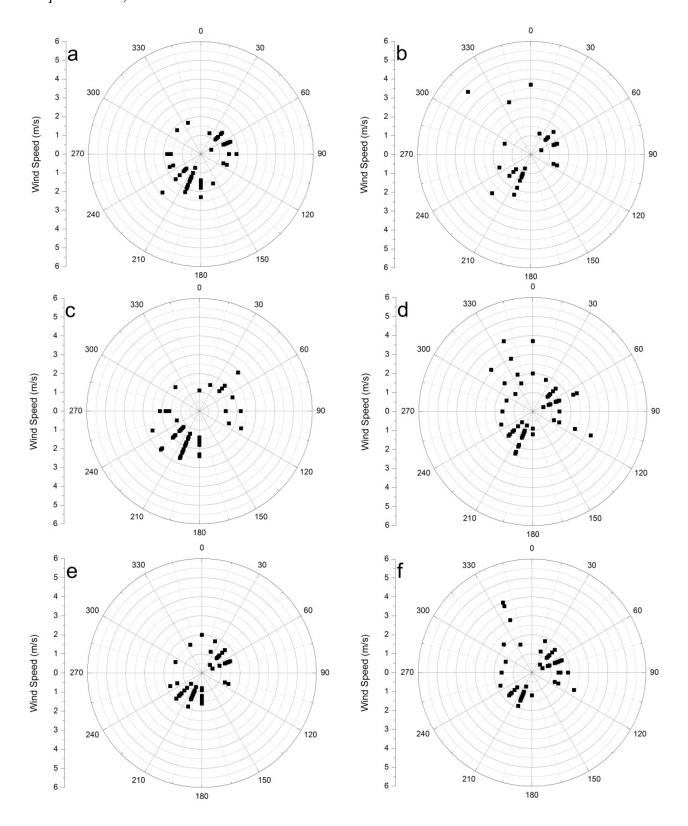
The real-time release of atmospheric pollution concentrations and air qualities for 35 permanent monitoring stations on the web-platform provided a powerful data source for governments and researchers to investigate the air pollution situations and make policies to deal with air pollutions. New ambient air quality standard has brought PM<sub>2.5</sub>, O<sub>3</sub> and CO into the air quality monitoring system. First of all, the introduction of PM<sub>2.5</sub> in the standard had brought down the air quality attainment rate by 15% to 22% for all monitoring stations, which is really a great problem for the evaluation of the work of local government as the air quality attainment rates have dropped to less than 70% for most monitoring stations.

Different to the daily report of API without concentration of each pollutant under NAAQS-1996, Beijing municipal government provided real-time concentrations of all six pollutants to the public. These

public data, although sometimes with some missing data due to internet transfer or data error, could help researchers to insightfully investigate the atmospheric pollution situations. As the most frequent primary pollutants, PM<sub>2.5</sub> and PM<sub>10</sub> shew similar spatial and seasonal patterns that south and southeast stations were polluted most during autumn and winter while the pollutions in summer and spring were reduced to a fairly good level. Except O<sub>3</sub>, the rest of the five pollutants showed similar temporal patterns that summer had the lowest pollutants concentrations while autumn and winter were the most polluted. O<sub>3</sub> concentrations tend to increase in humid summer time with strong solar radiations under strong photo-chemical reactions in the atmosphere. In dry winter and autumn, O<sub>3</sub> concentrations dropped greatly. Additionally, O<sub>3</sub> concentrations are higher in north stations without traffic activities, which is the major contribution to the NO<sub>2</sub> in the atmosphere. The high concentrations of nitrogen oxides in the regions with heavy traffic monitoring have greatly reduced the concentrations of O<sub>3</sub> by the chemical reaction process, while for those stations far away from traffic (mainly in north stations), the O<sub>3</sub> concentrations are higher with less concentrations of nitrogen monoxide as a catalyst. SO<sub>2</sub>, NO<sub>2</sub> and CO concentrations' spatial and temporal distributions were similar as south and south stations in winter were most serious polluted while spring and summer were less polluted. As a kind of vehicle exhaust, NO<sub>2</sub> concentrations of traffic monitoring stations just next to main streets were higher than other stations, demonstrating the impact of traffic. SO<sub>2</sub> and CO, usually considered to be products of coal combustion, were higher in the winter heating season and probably transported from the south neighboring province.

Table 2 shows the current operating air quality stand with its limits. By comparing the limit with qualified level (IAQI < 100), we would find that under most situations, the concentrations of CO and SO<sub>2</sub> fall in the range of IAQI < 100 or even 50, suggesting that CO and SO<sub>2</sub>, although included in the air quality monitoring system, were not the major reason for the air pollution. NO<sub>2</sub> and O<sub>3</sub>, however, sometimes would exceed the daily limits (concentrations when IAQI = 100). Particulate matter, especially PM<sub>2.5</sub>, was the frequent pollutant over the limit for south and southeast stations during winter and autumn. These results demonstrated that under the new ambient air quality standard, PM<sub>2.5</sub> is the major reason for the deterioration of air quality in Beijing. The high concentrations of PM<sub>2.5</sub> in south and southeast stations also suggested that except for a local source, the transportation of south neighboring province could not be neglected.

Wind fields play an important role in the dispersion of atmospheric pollutants, especially under the conditions of heavy pollution. In this paper, we selected the top 15% concentrations of all pollutants for one selected station (The situations of the rest stations are similar). The top 15% concentrations of the six pollutants with their associations with wind speed and wind directions of Station 1 (Dongsi, an urban station) are shown in Figure 12. The results suggest that the top 15% concentrations of all pollutants (high heavy pollutant concentrations) mainly occurred under the situation of low wind speed (less than 1.5 m/s) except for O<sub>3</sub>. For PM<sub>2.5</sub>, south and east low wind speed situations have the highest possibility of high PM<sub>2.5</sub> concentrations. For PM<sub>10</sub>, the situation is similar except that some large northwest wind speed may bring high PM<sub>10</sub> concentrations indicating that the dust source of PM<sub>10</sub>. For O<sub>3</sub>, SO<sub>2</sub>, NO<sub>2</sub> and CO, high pollutants' concentrations mainly occurred under southwest and east low speed wind conditions. These results suggested that low wind speed, with southwest or east wind situations have the most frequent heavy polluted occurrences in Beijing.



**Figure 12.** Wind profiles of the top 15% measurements of each pollutant for Station 1: Dongsi in urban Beijing ((a) PM<sub>2.5</sub>, (b) PM<sub>10</sub>, (c) O<sub>3</sub>, (d) SO<sub>2</sub>, (e) NO<sub>2</sub>, and (f) CO).

# 5. Conclusions

Beijing, the capital of China, has endured serious air pollution problems in the past decade. However, the previous air quality standard (NAAQS-1996) suggested that the air quality was fairly good which is

inconsistent with residents' feelings and the frequent fog-haze events after 2011. Therefore, China has released new air quality standards (NAAQS-2012) by adding PM<sub>2.5</sub>, O<sub>3</sub> and CO besides PM<sub>10</sub>, SO<sub>2</sub> and NO<sub>2</sub> into the air quality standards with real time requests. In this study, we collected the annual monitoring data of all six pollutants from September 2014 to August 2015 to have a comprehensive evaluation of the seasonal and spatial situations of the atmospheric pollution in Beijing.

Temporal distributions of the air pollutants demonstrated that autumn and winter were the season most polluted by PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub> and CO while summer was the one having the best air quality. O<sub>3</sub> showed an obviously different pattern that summer was polluted most by O<sub>3</sub> as the abundant solar radiation, and a humid atmosphere helped the generation of O<sub>3</sub> in the afternoon. Spatial distribution of these air pollutants suggested the stations in south and southeast areas were more seriously polluted by PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub> and CO in winter and autumn, while the atmospheric pollution concentrations in north stations were far lower. This kind of spatial pattern of air pollution demonstrated the transportation from south and southeast directions. Extra wind profile analysis with heavy pollution situations indicates that low speed southwest or east wind situations have the higher possibility of heavy pollution, suggesting that it is highly possible that long-range transportation of air pollutants from south or east sources could not be neglected. NO<sub>2</sub> concentrations of traffic monitoring stations were higher. The monitoring results revealed that PM<sub>2.5</sub> and PM<sub>10</sub> were the major contributors to the air pollution and that NO<sub>2</sub> and O<sub>3</sub> remained relatively important pollutants but not comparable to particulate matter. SO<sub>2</sub>, a kind of pollutant that used to be high in concentration at the beginning of 21st century, has become a pollutant with low concentration well under the air quality standard qualification. CO concentrations, with winter higher than the rest of the seasons, were qualified under most circumstances. Therefore, the major challenge for government of Beijing to deal with the serious problems of air pollution is how to reduce the concentrations of PM<sub>2.5</sub> both locally and transported a long way from southerly directions.

# Acknowledgments

This work was supported by the Fundamental Research Funds for the Central Universities (Grant No. 2014QD02).

## **Author Contributions**

All authors contributed a lot. Wei Chen defined the major working scheme, collected the data and wrote the paper. Lei Yan provided some key data of air pollution. Haimeng Zhao processed and analyzed the aerosol data.

## **Conflicts of Interest**

The authors declare no conflicts of interest

## References

1. Li, R.; Li, Z.; Gao, W.; Ding, W.; Xu, Q.; Song, X. Diurnal, seasonal, and spatial variation of PM<sub>2.5</sub> in Beijing. *Sci Bull.* **2015**, *60*, 387–395.

2. Guo, Y.; Feng, N.; Christopher, S.A.; Kang, P.; Zhan, F.B.; Hong, S. Satellite remote sensing of fine particulate matter (PM<sub>2.5</sub>) air quality over Beijing using MODIS. *Int. J. Remote Sens.* **2014**, *35*, 6522–6544.

- 3. Zhang, R.; Jing, J.; Tao, J.; Hsu, S.C.; Wang, G.; Cao, J.; Shen, Z. Chemical characterization and source apportionment of PM<sub>2.5</sub> in Beijing: Seasonal perspective. *Atmos. Chem. Phys.* **2013**, *13*, 7053–7074.
- 4. Zhang, A.; Qi, Q.; Jiang, L.; Zhou, F.; Wang, J. Population exposure to PM<sub>2.5</sub> in the urban area of Beijing. *Plos One* **2013**, *8*, E63486.
- 5. Wang, J. F.; Hu, M.G.; Xu, C.D.; Christakos, G.; Zhao, Y. Estimation of citywide air pollution in Beijing. *Plos One* **2013**, *8*, E53400.
- 6. Andrews, S.Q. Inconsistencies in air quality metrics: 'Blue Sky' days and PM<sub>10</sub> concentrations in Beijing. *Environ. Res. Lett.* **2008**, *3*, 034009.
- 7. Liu, Z.R.; Hu, B.; Liu, Q.; Sun, Y.; Wang, Y.S. Source apportionment of urban fine particle number concentration during summertime in Beijing. *Atmos. Environ.* **2014**, *96*, 359–369.
- 8. Baccarelli, A.A.; Zheng, Y.; Zhang, X.; Chang, D.; Liu, L.; Wolf, K.R.; Hou, L. Air pollution exposure and lung function in highly exposed subjects in Beijing, China: A repeated-measure study. *Part. Fibre Toxicol.* **2014**, *11*, 51.
- 9. Wang, L.; Zhang, P.; Tan, S.; Zhao, X.; Cheng, D.; Wei, W.; Pan, X. Assessment of urban air quality in China using air pollution indices (APIs). *J. Air Waste Manag. Assoc.* **2013**, *63*, 170–178.
- 10. Xu, P.; Chen, Y.F.; Ye, X.J. Haze, air pollution, and health in China. Lancet 2013, 382, 2067.
- 11. Gao, H.; Chen, J.; Wang, B.; Tan, S.C.; Lee, C.M.; Yao, X.; Shi, J. A study of air pollution of city clusters. *Atmos. Environ.* **2011**, *45*, 3069–3077.
- 12. Zhao, H.; Che, H.; Ma, Y.; Xia, X.; Wang, Y.; Wang, P.; Wu, X. Temporal variability of the visibility, particulate matter mass concentration and aerosol optical properties over an urban site in Northeast. China. *Atmos. Res.* **2015**, *166*, 204–212.
- 13. San Martini, F.M.; Hasenkopf, C.A.; Roberts, D.C. Statistical analysis of PM<sub>2.5</sub> observations from diplomatic facilities in China. *Atmos. Environ.* **2015**, *110*, 174–185.
- 14. You, M.Q. Addition of PM<sub>2.5</sub> into the national ambient air quality standards of China and the contribution to air pollution control: The case study of Wuhan, China. *Sci. World J.* **2014**, 768405.
- 15. Hu, J.; Ying, Q.; Wang, Y.; Zhang, H. Characterizing multi-pollutant air pollution in China: comparison of three air quality indices. *Environ. Int.* **2015**, *84*, 17–25.
- 16. Zhang, H.; Wang, Y.; Hu, J.; Ying, Q.; Hu, X.M. Relationships between meteorological parameters and criteria air pollutants in three megacities in China. *Environ.Res.* **2015**, *140*, 242–254.
- 17. Distribution of real time air quality of Beijing. Available on line: http://zx.bjmemc.com.cn/ (accessed on 12 November 2015).
- © 2015 by the authors; licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution license (http://creativecommons.org/licenses/by/4.0/).