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Impacts of Urban Layouts and Open Space on Urban Ventilation Evaluated by Concentration Decay Method

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Abstract: Previous researchers calculated air change rate per hour (*ACH*) in the urban canopy layers (UCL) by integrating the normal component of air mean velocity (convection) and fluctuation velocity (turbulent diffusions) across UCL boundaries. However they are usually greater than the actual *ACH* induced by flow rates flushing UCL and never returning again. As a novelty, this paper aims to verify the exponential concentration decay history occurring in UCL models and applies the concentration decay method to assess the actual UCL *ACH* and predict the urban age of air at various points. Computational fluid dynamic (CFD) simulations with the standard *k-ε* models are successfully validated by wind tunnel data. The typical street-scale UCL models are studied under neutral atmospheric conditions. Larger urban size attains smaller *ACH*. For square overall urban form ($L_x = L_y = 390$ m), the parallel wind ($\theta = 0^\circ$) attains greater *ACH* than non-parallel wind ($\theta = 15^\circ, 30^\circ, 45^\circ$), but it experiences smaller *ACH* than the rectangular urban form ($L_x = 570$ m, $L_y = 270$ m) under most wind directions ($\theta = 30^\circ$ to 90°). Open space increases *ACH* more effectively under oblique wind ($\theta = 15^\circ, 30^\circ, 45^\circ$) than parallel wind. Although further investigations are still required, this paper provides an effective approach to quantify the actual *ACH* in urban-like geometries.

Keywords: small open space; air change rate per hour (*ACH*); concentration decay method; urban age of air; computational fluid dynamic (CFD) simulation

1. Introduction

The increase in number of vehicles in cities and the ongoing urbanization worldwide are causing more concerns about urban air pollution. The urban canopy layer (UCL) is defined as the outdoor air volume below the rooftops of buildings [1]. Raising UCL ventilation capacity by the surrounding atmosphere with relatively clean air has been regarded as one effective approach to diluting environmental pollutants [2–8] and improving the urban thermal environment in the hot summer [9,10] as well as reducing human exposure to outdoor pollutants [11–13].

UCL ventilation is strongly correlated to urban morphologies. For two-dimensional street canyons [14–19], street aspect ratio (building height/street width, H/W) is the first key parameter to affect the flow regimes and UCL ventilation. Four flow regimes have been classified depending on different aspect ratios (H/W) [14–16], i.e., the isolated roughness flow regime (IRF, in which the aspect ratio is less than 0.1 to 0.125), the wake interference flow regime (WIF, with an aspect ratio of 0.1 to 0.67), the skimming flow regime (SF, with an aspect ratio of 0.67 to 1.67), and the multi-vortex regime in

two-dimensional deep street canyons with two or more vortices as $H/W > 1.67$. For three-dimensional UCL geometries, the plan area index λ_p and frontal area index λ_f [4–6,9,10] are regarded as the key urban parameters to influence the flow and pollutant dispersion in urban areas. In addition, the other factors including overall urban form and ambient wind directions [2,3,6–9], building height variations [5,20–25], thermal buoyancy force for weak-wind atmospheric conditions [10,18,26–30], etc. also play significant roles in the flow and pollutant dispersion in UCL models.

In recent years, various ventilation indices have been applied for UCL ventilation assessment, such as pollutant retention time and purging flow rate [2,7,23], age of air and ventilation efficiency [3–7,23], in-canopy velocity and exchange velocity [22,25], net escape velocity [31], etc. Similar to indoor ventilation [32–36], the key point of applying these ventilation indices is based on the UCL ventilation processes as below: The surrounding external air is relatively clean and can be transported into urban areas to aid pollutant dilution with physical processes of horizontal dilution, vertical transport, recirculation of contaminants, and turbulent diffusion. In this sense, the tracer gas technique originated from indoor ventilation sciences has been used to predict outdoor ventilation by analyzing the final steady-state pollutant concentrations or the transient concentration decay history, which illustrates the processes of pollutant dilution and ventilation. For example, the concept of purging flow rate (*PFR*) was adopted to assess the net flow rate of flushing the urban domain [2] or the entire urban canopy layer [23] induced by the convection and turbulent diffusion. Moreover, the urban age of air [3–7] represents how long the external air can reach a place after it enters the UCL space. Hang et al. [3] first adopted the homogeneous emission method into CFD simulations [32] to calculate the urban age of air for quantifying the characteristics of wind supplying external air into UCL space for pollutant dilution.

In particular, the air change rate per hour is one of the most widely used indoor ventilation concepts [32–36], calculated by $ACH = 3600Q_T / Vol$, where Q_T is the total volumetric flow rate and Vol is the room volume. Later it is adopted to quantify the volumetric air exchange rate of the entire UCL volume per hour, including two-dimensional street canyon ventilation assessment [37,38] and three-dimensional UCL ventilation modeling [39–44]. Previous researchers usually calculated outdoor *ACH* indexes based on the volumetric flow rate obtained by integrating the mean value of the component of the mean velocity normal to the UCL boundaries and the effective flow rate due to turbulence based on integrating half the standard deviation (rms-value) of the velocity fluctuations on street roofs [37–44]. Specially, vertical turbulent exchange across a street roof has been verified to significantly influence UCL ventilation and pollutant removal in urban-like geometries [37–40] because street roofs are open at the top with a large area. However, the mean velocity often exhibits recirculation and the velocity fluctuations across the UCL open roof are bidirectional and therefore air or pollutant can return to the given UCL space several times. Thus, these two *ACH* indexes do not represent the entire UCL air volume that is really exchanged *ACH* times in an hour by external air, i.e., they are not the actual air change rate per hour in UCL models calculated by the actual flow rate flushing or leaving this space and never returning again. There are efficiency problems in UCL ventilation assessment [3,5].

The purpose of this paper is to predict the actual or net air change rate per hour in UCL models and how it is related to urban morphologies and atmospheric conditions. The net or actual air change rate is defined as the exchange of air flushing the space and never returning to the UCL space again. It depends on several time scales, i.e., the overall turn over time (air volume (Vol) divided by the total volumetric flow rate (Q_T)), purging time for air moving from entrances to exits, time constant for local vortices and turbulent mixing. The concentration decay method has been widely used to predict the actual *ACH* and age of air in indoor environments in which the concentration temporal profile accords with the exponential decay law and the concentration decay history shows the pollutant dilution and ventilation processes in room ventilation [32–36,45,46]. However, to date, this method has not been introduced to quantify the actual or net *ACH* index and age of air in UCL models.

Therefore, as a novelty, this paper aims to verify whether the temporal decay profile of concentration in UCL models accords with the exponential decay law and explore the effectiveness of applying concentration decay method for outdoor ventilation assessment. As a start, the typical

medium-dense urban canopy layers ($H/W = 1$, $\lambda_f = \lambda_p = 0.25$) are first studied. The effects of urban size, ambient wind directions, overall urban forms, and open space arrangements are evaluated under neutral atmospheric conditions.

2. Methodology

2.1. Turbulence Models for Urban Airflow Modeling

As outlined by the review papers, CFD simulations have been widely applied in urban flow/dispersion modeling outdoor in the last two decades [47–49]. Large eddy simulations (LES) are known to perform better in predicting turbulence than the Reynolds-Averaged Navier-Stokes (RANS) approaches. However, there are still some challenges involved with widely applying LES because of the strongly increased computational requirements, the development of advanced sub-grid scale models, and the difficulty in specifying appropriate time-dependent inlet and wall boundary conditions [16,17,20,26,38,39]. Therefore, in quantitative work one has to adopt time-averaged turbulence models (i.e., RANS approaches). Actually, in spite of their deficiencies in predicting turbulence, steady RANS models have become the most popular CFD approaches for urban flow modeling, and the standard k - ε model is one of the most widely adopted in the literature [2,3,5–7,9–11,21–23], with successful validation by wind tunnel measurements and good performance in predicting mean flows. This paper selects the standard k - ε model and the CFD validation case is conducted to evaluate the reliability of CFD methodologies by wind tunnel data.

2.2. Model Description in the CFD Validation Case

As shown in Figure 1a, Brown et al. [50] measured turbulent airflows in a seven-row and 11-column cubic building array with a parallel approaching wind. Building width (B), building height (H), and street width (W) are the same ($B = H = W = 15$ cm, $H/W = 1$, building packing densities $\lambda_p = \lambda_f = 0.25$, $L_x = 13H$, $L_y = 21H$). The scale ratio to full-scale models is 1:200. x , y , and z are the stream-wise, span-wise (lateral), and vertical directions. $x/H = 0$ represents the location of windward street opening. $y/H = 0$ is the vertical symmetric plane of the middle column. Point V_i represents the center point of the secondary street No. i . The measured vertical profiles of time-averaged (or mean) velocity components (stream-wise velocity $\bar{u}(z)$, vertical velocity $\bar{w}(z)$), and turbulence kinetic energy $\kappa(z)$ at Points V_i are used to evaluate CFD simulations.

In the CFD validation case, the full-scale seven-row building array is numerically investigated ($B = H = W = 30$ m, $\lambda_p = \lambda_f = 0.25$, $H/W = 1$; see Figure 1b). According to the literature [23,51–53], the lateral urban boundaries hardly affect airflows in the middle column as the wind tunnel model is sufficiently wide in the lateral (y) direction ($L_y = 21H$). Thus, it is effective to only consider half of this middle column (shaded area in Figure 1a) to reduce the computational requirement. Figure 1b shows model geometry, grid arrangements, CFD domain, and boundary conditions. In particular, the distances of UCL boundaries to the domain inlet, domain outlet, and domain top are $6.7H$, $40.3H$, and $9.0H$, respectively. Zero normal gradient condition is used at the domain top, domain outlet, and two lateral domain symmetry boundaries. At the domain inlet, the measured power-law profile of time-averaged (or mean) velocity $U_0(z)$ in the upstream free flow is adopted (see Equation (1a)) (Brown et al. [51]). Moreover, the vertical profiles of turbulent kinetic energy $k(z)$ and its dissipation rate (ε) at the domain inlet are calculated by Equations (1b) and (1c) [23,51–53]:

$$U_0(z) = U_{ref} \times (z/H)^{0.16} \quad (1a)$$

$$k(z) = u_*^2 / \sqrt{C_\mu} \quad (1b)$$

$$\varepsilon(z) = C_\mu^{3/4} k^{3/2} / (\kappa_v z), \quad (1c)$$

where U_{ref} ($=3.0 \text{ ms}^{-1}$) is the reference velocity at the building height ($H = 30 \text{ m}$), C_μ is a constant ($=0.09$), u_* is the friction velocity ($=0.24 \text{ ms}^{-1}$), and κ_v is von Karman's constant ($=0.41$).

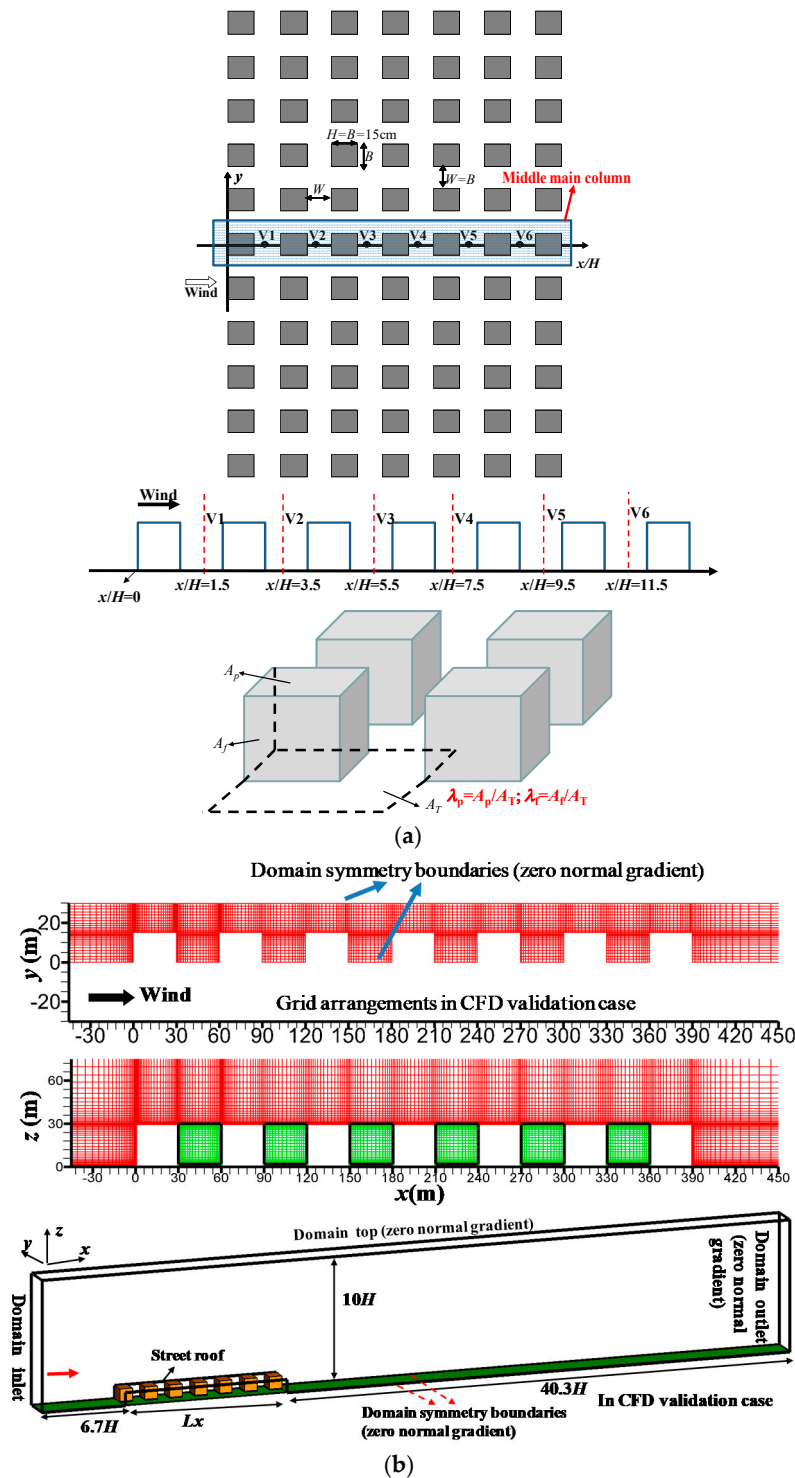


Figure 1. (a) Model description of wind tunnel test; (b) setups in CFD validation case.

For this CFD validation case, hexahedral cells of 531,657 and the minimum cell sizes of 0.5 m ($H/60$) are used (the medium grid). Following the CFD guidelines [54,55], four hexahedral cells exist below the pedestrian level ($z = 0 \text{ m}$ to 2 m). The grid expansion ratio from wall surfaces to the

surrounding is 1.15. With such grid arrangements, the normalized distance from wall surfaces y^+ ($y^+ = yu_\tau/\nu$) ranges from 60 to 1000 at most regions of wall surfaces within UCL space.

No slip boundary condition with standard wall function [56] is used at wall surfaces. The practice guidelines for setting the upstream and downstream ground are followed to reproduce a horizontally homogeneous atmospheric boundary layer surrounding urban areas [46,57]: The roughness height k_S and the roughness constant C_S are correlated by the aerodynamic roughness length z_0 :

$$k_S = \frac{9.793z_0}{C_S}. \tag{2}$$

Note that the distance between the center point P of the ground adjacent cell and the ground surface is $Y_P = 0.25$ m. Fluent 6.3 does not allow k_S to be larger than Y_P [57]. Thus, according to van Hooff and Blocken [46], a user-defined function is used to set the roughness constant $C_S = 4$ because Fluent 6.3 does not allow it to be greater than 1. Then as $C_S = 4$ and $z_0 = 0.1$ m, $k_S = 0.245$ m is smaller than $Y_P = 0.25$ m.

ANSYS FLUENT with the finite volume method is used to predict the steady isothermal urban airflows [56]. The SIMPLE scheme is adopted to couple the pressure to the velocity field. The first-order upwind scheme is not appropriate for all transported quantities since the spatial gradients of the quantities tend to become diffusive due to a large viscosity. Thus, as recommended by the literature [54,55], all transport equations are discretized by the second-order upwind scheme for better numerical accuracy. The under-relaxation factors for pressure term, momentum term, k , and ϵ terms are 0.3, 0.7, 0.5, and 0.5, respectively. The solutions do not stop until all residuals become constant. The fine grid with a minimum size of 0.2 m ($H/150$) is also adopted to perform a grid independence study. The above CFD settings on domain sizes, grid arrangements and boundary conditions fulfill the major requirements given by CFD guidelines [54,55]. In particular, steady CFD simulations are first carried out about 4000 iterations with the first-order upwind scheme, then continued with the second-order upwind scheme until all residuals became constant. The residuals reached the following minimum values or less: 10^{-4} for the continuity equation, 0.5×10^{-5} for the velocity components and k , 0.5×10^{-5} and 0.5×10^{-4} for pollutant concentration and ϵ . After solving the steady-state airflow field, the uniform initial concentration $C(0)$ is defined in the entire urban canopy layer, then the unsteady concentration decay history is calculated and recorded for ventilation assessment as the solved flow field remains constant.

2.3. Model Description in All CFD Test Cases

As displayed in Table 1 and Figure 2a–d, two groups of medium-dense UCL models ($H = B = W = 30$ m, $\lambda_p = \lambda_f = 0.25$) are investigated in CFD simulations. The row and column numbers are referred to the numbers of buildings along the stream-wise direction (x , main streets) and span-wise direction (y , secondary streets). Wind directions are represented by the angles ($\theta = 0^\circ$ to 90°) between the approaching wind and the main streets.

Table 1. Test cases investigated ($\lambda_p = \lambda_f = 0.25$, $H/W = 1$).

Case Name	Number of Rows/Columns, Urban Sizes L_x and L_y	Wind Direction(θ°)
Group I		
[5-5, θ°]	5 rows, 5 columns, $L_x = L_y = 270$ m	$0^\circ, 15^\circ, 30^\circ, 45^\circ$
[7-7, θ°]	7 rows, 7 columns, $L_x = L_y = 390$ m	
[10-5, θ°]	10 rows, 5 columns, $L_x = 570$ m, $L_y = 270$ m	$0^\circ, 15^\circ, 30^\circ, 45^\circ, 60^\circ, 75^\circ, 90^\circ$
Group II		
[5-5, θ° , Oij]	5 rows, 5 columns, $L_x = L_y = 270$ m, Oij = O21, O22, O23, O24, O33, O34	$0^\circ, 15^\circ, 30^\circ, 45^\circ$

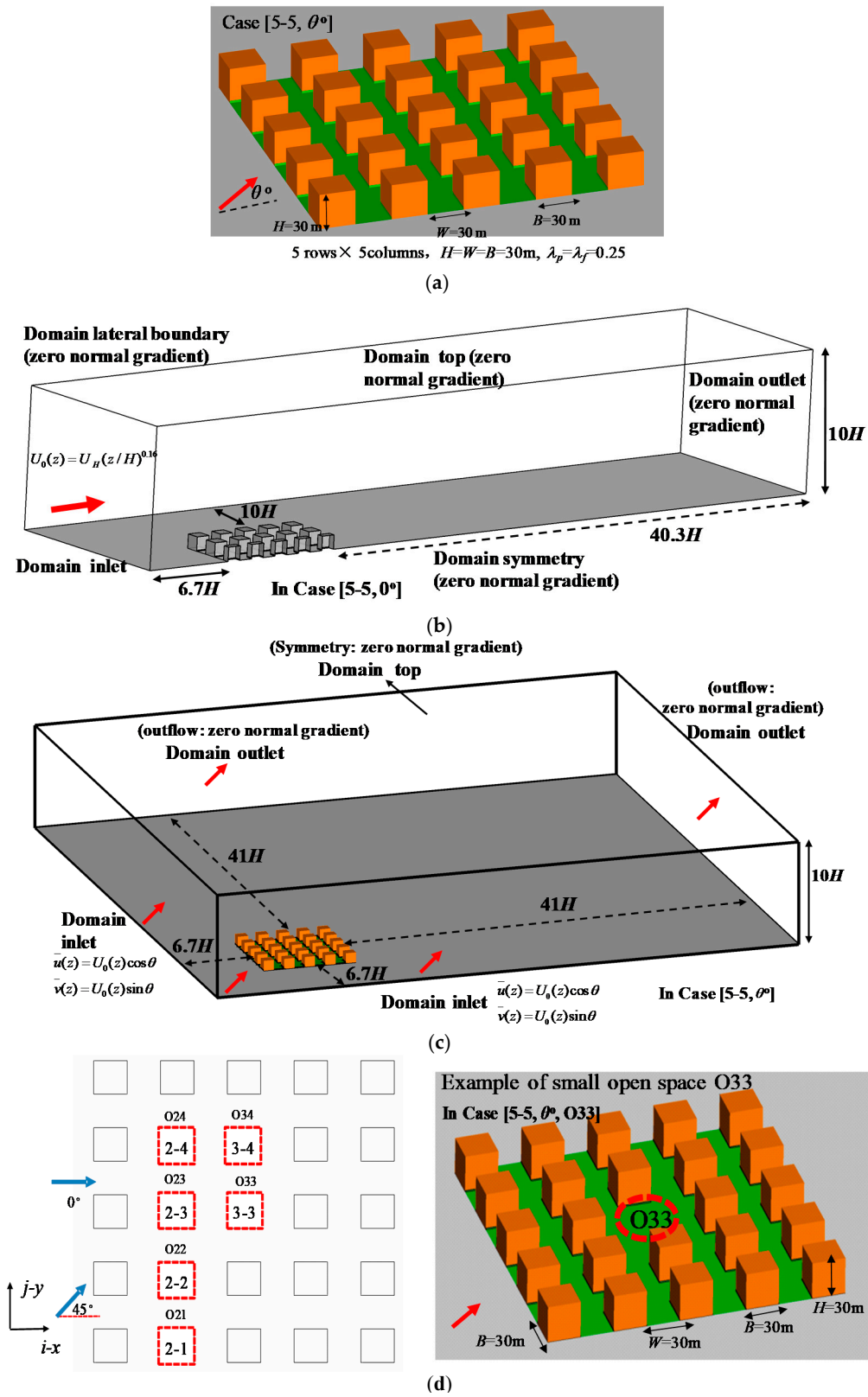


Figure 2. Cont.

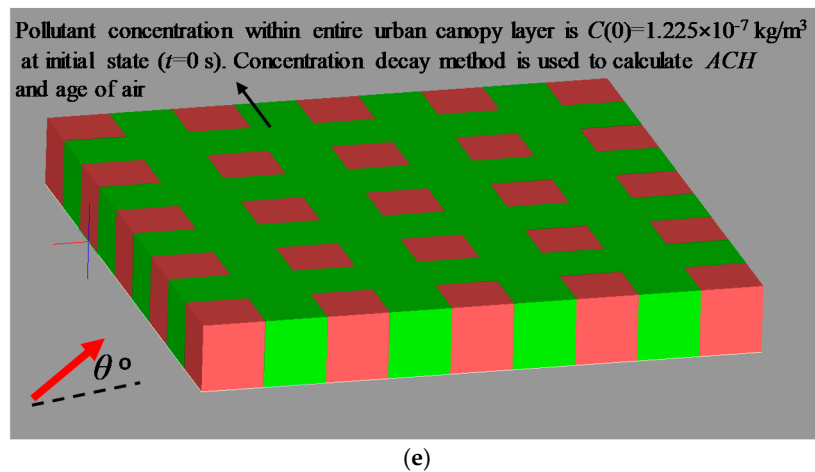


Figure 2. (a) Model description and (b) computational domain of Case [5-5, 0°] ($\theta = 0^\circ$); (c) Computational domain in Case [5-5, θ] ($\theta = 15^\circ, 30^\circ, 45^\circ$); (d) Model description of Case [5-5, θ , Oij]; (e) Definition of initial-state concentration field in the entire UCL space.

In Group I (Table 1 and Figure 2a), test cases without open space are named as Case [row number-column number, wind direction θ°]. For cases with a parallel approaching wind ($\theta = 0^\circ$, Figure 2b), three cases are included (i.e., Case [5-5, 0°], Case [7-7, 0°], Case [10-5, 0°]) with only half of computational domain simulated. The distances from UCL boundaries to the domain top, domain outlet, domain inlet, and the domain lateral boundary are $9H$, $40.3H$, $6.7H$, and $10H$, respectively. Zero normal gradient boundary condition is used at the domain outlet, domain top, and the lateral domain boundary. At the domain inlet, Equation (1) is used to provide the boundary condition.

For test cases with oblique wind ($\theta = 15^\circ, 30^\circ, 45^\circ, 60^\circ, 75^\circ, 90^\circ$, Figure 2c), the full CFD domains are used. There are two domain inlets and two domain outlets. The distances from UCL boundaries to the domain top, domain outlets, domain inlets are $9H$, $41H$ and $6.7H$. At two domain inlets, the vertical profiles of time-averaged (or mean) velocity components in Equations (3a)–(3c) and turbulent quantities defined in Equations (1b) and (1c) are used to define boundary conditions:

$$\bar{u} = U_0(z) \cos \theta \tag{3a}$$

$$\bar{v} = U_0(z) \sin \theta \tag{3b}$$

$$\bar{w}(z) = 0. \tag{3c}$$

At the domain outlets and domain top, the zero normal gradient boundaries are used.

In Group II (Table 1 and Figure 2d), for UCL models with five rows and five columns, open space arrangements are included. In total 24 test cases of Case [5-5, θ° , Oij] are investigated ($\theta = 0^\circ, 15^\circ, 30^\circ, 45^\circ$). Here Oij represents only one building of position i-j (2-1, 2-2, 2-3, 2-4, 3-3, or 3-4) (see Figure 2d) is removed to attain an open space effect. Computational domain and boundary conditions are similar to Group I.

For all the test cases in Table 1, the grid arrangements are similar to the medium grid of the CFD validation case. The total number of hexahedral cells is from 531,657 to 3,360,096. The above CFD settings on domain sizes, grid arrangements, and boundary conditions fulfill the major requirements recommended by the best CFD guidelines.

2.4. ACH Indexes and Age of Air by Concentration Decay Method

2.4.1. Volumetric Flow Rates and the Corresponding ACH [23,37–42]

To analyze the flow field, the velocity components and wind speed are all normalized by the reference velocity in the upstream free flow at the same height, i.e., Equation (1a) at the domain inlet $U_0(z) = U_{ref} \times (z/H)^{0.16}$. For example, for velocity at $z = 15 \text{ m} = 0.5H$, the velocity is normalized by dividing it by $U_0(z = 15 \text{ m}) = 2.685 \text{ ms}^{-1}$.

To quantify the ventilation capacity, the mean volumetric flow rates are calculated by integrating the normal air velocity across all UCL boundaries (Equation (4a)); moreover, the effective flow rate through street roofs due to turbulent exchange is defined by integrating the fluctuation velocity across open street roofs (Equation (4b)) [23,37–42]:

$$Q = \int_A \vec{V} \cdot \vec{n} dA \quad (4a)$$

$$Q_{roof}(turb) = \pm \int_{A_{roof}} 0.5\sigma_w dA, \quad (4b)$$

where, in Equation (4a), \vec{V} is the velocity vector, consisting of three time-averaged (or mean) velocity components ($\vec{V} = (\bar{u}, \bar{v}, \bar{w})$ in x, y, z directions), and \vec{n} and A are the normal direction and area, respectively, of UCL boundaries. In Equation (4b), A_{roof} is the area of street roofs, $\sigma_w = \sqrt{\overline{w'w'}} = \sqrt{2k/3}$ is the fluctuation velocity based on the approximation of isotropic turbulence in all k - ϵ turbulent models where u', v', w' are the stream-wise, span-wise, vertical velocity fluctuations ($u' = v' = w'$) and turbulent kinetic energy is $k = \frac{1}{2}(\overline{u'u'} + \overline{v'v'} + \overline{w'w'})$.

It is worth noting that the values of $+0.5$ and -0.5 in Equation (4b) represent the fact that turbulent fluctuations induce the same upward and downward air fluxes across street roofs [23,37–42]. As pollution sources exist within UCL space, the literature confirmed that turbulent fluctuations across street roofs significantly contribute to pollutant removal since the upward pollutant flux through street roofs usually exceeds the downward pollutant flux because the pollutant concentration below street roofs is higher than that above them [21,31]. Due to the much greater total area of street roofs, the effective flow rates by turbulent exchange across street roofs $Q_{roof}(turb)$ are also important for the urban canopy layer ventilation compared to mean flows across UCL boundaries (Equation (4a)). Thus, this paper mainly adopts $Q_{roof}(turb)$ to assess the effect of turbulent fluctuations on UCL ventilation [23,37–44].

Due to the flow balance, the total outflow rate induced by mean flows across UCL boundaries (Q_{out}) equals that entering UCL boundaries (Q_{in}). They are named the total mean volumetric flow rates Q_T (see Equation (5)) [23,37–44], which is the sum of all volumetric rates through all street openings and street roofs:

$$Q_T = |Q_{in}| = |Q_{out}|. \quad (5)$$

Then ACH by Q_T and $Q_{roof}(turb)$ are used to quantify the volumetric air exchange rate per hour, as below [23,37–44]:

$$ACH_T = 3600Q_T / Vol \quad (6a)$$

$$ACH_{turb} = 3600Q_{roof}(turb) / Vol, \quad (6b)$$

where Vol is the entire UCL volume, and $Q_{roof}(turb)$ is defined in Equation (4b).

2.4.2. Actual or Net ACH and Urban Age of Air by the Concentration Decay Method

In urban ventilation sciences, the flow rates across a control surface are usually defined and recorded by integrating the normal velocity through it. Generally the complicated circulating flow is

generated in an urban area, hence there are fluid particles coming back into UCL space across UCL boundaries. Moreover turbulent fluctuations induce upward and downward fluxes across the open street roofs, representing a large amount of fluid particles leave and re-enter UCL space. Thus, similar to the indoor environment [35,36], fluid particles flowing across UCL boundaries can be divided into two groups, one is for air leaving UCL space for never returning again and the other for air returning or revisiting UCL space (it has been in the UCL space before). The first group is much more significant to UCL ventilation and pollutant dilution. Therefore, only a fraction of the flow rates defined in Equations (4a) and (4b) contributes to flushing UCL space by external air or diluting pollutants within it. Furthermore, the *ACH* indexes defined in Equation (6) do not represent the entire UCL air volume that is really exchanged *ACH* times in an hour by external air.

To attain and assess the actual and net *ACH* induced by mean flows (ACH_T) and turbulent exchange (ACH_{turb}) across UCL boundaries, the concentration decay method is introduced into CFD simulations of UCL ventilation modeling. Similarly with indoor *ACH* by concentration decay method [32,45,46], if the predicted temporal decay profile of spatial mean concentration in the entire UCL volume accords with the exponential decay law, this decay rate is correlated to air change rate per hour in UCL models.

After the steady-state flow field is solved, a uniform initial time-averaged concentration of tracer gas (CO, carbon monoxide) is defined in the entire UCL air volume at time of $t = 0$ s ($C(0) = 1.225 \times 10^{-7}$ kg/m³, see Figure 2e). Then the transient concentration decay process $C(t)$ is numerically simulated and recorded while the flow field keeps steady.

The unsteady governing equation of time-averaged concentration $C(t)$ is:

$$\frac{\partial C}{\partial t} + \bar{u}_j \frac{\partial C}{\partial x_j} - \frac{\partial}{\partial x_j} ((D_m + D_t) \frac{\partial C}{\partial x_j}) = 0, \tag{7}$$

where \bar{u}_j is the time-averaged (or mean) velocity components ($\bar{u}, \bar{v}, \bar{w}$), and D_m and D_t are the molecular and turbulent diffusivity of pollutants or tracer gas. Here $D_t = \nu_t / Sc_t$, ν_t is the kinematic eddy viscosity and Sc_t is the turbulent Schmidt number ($Sc_t = 0.7$) [3–7,21–23].

The decay duration of all test cases is sufficiently long (400 s). The default time step is $dt = 1$ s. In the CFD validation case, time steps of $dt = 0.5$ s and 2 s are also tested. The second-order upwind scheme is used for Equation (7) with the under-relaxation factors of 1.0. For each time step, the residual of Equation (7) reaches a value below 10^{-10} . For Equation (7), the inflow pollutant concentration at the domain inlet is set as zero and zero normal flux condition is used at all wall surfaces; moreover, zero normal gradient condition is applied at the domain top, domain outlet, and domain lateral boundaries.

The net and actual *ACH* index is calculated by the decay rate of spatial mean concentration in the entire UCL space ($\langle C(t) \rangle / C(0)$) (Equation (8)) [32–34]:

$$\langle C(t) \rangle = \frac{\int C(t) dx dy dz}{Vol} \tag{8a}$$

$$ACH = \frac{3600}{t} \left| \ln \frac{\langle C(t) \rangle}{C(0)} \right|, \tag{8b}$$

where $C(0) = 1.225 \times 10^{-7}$ kg/m³ is a uniform initial concentration, Vol is the entire volume of UCL space, $\langle C(t) \rangle$ is the spatial mean concentration in the entire UCL space, and 3600 denotes one hour or 3600 s.

Moreover, the concentration decay method can be used to predict urban age of air (τ_p s) at a given point (see Equation (9)) [32,36]:

$$\tau_p = \frac{\int_0^\infty C(t) dt}{C(0)}. \tag{9}$$

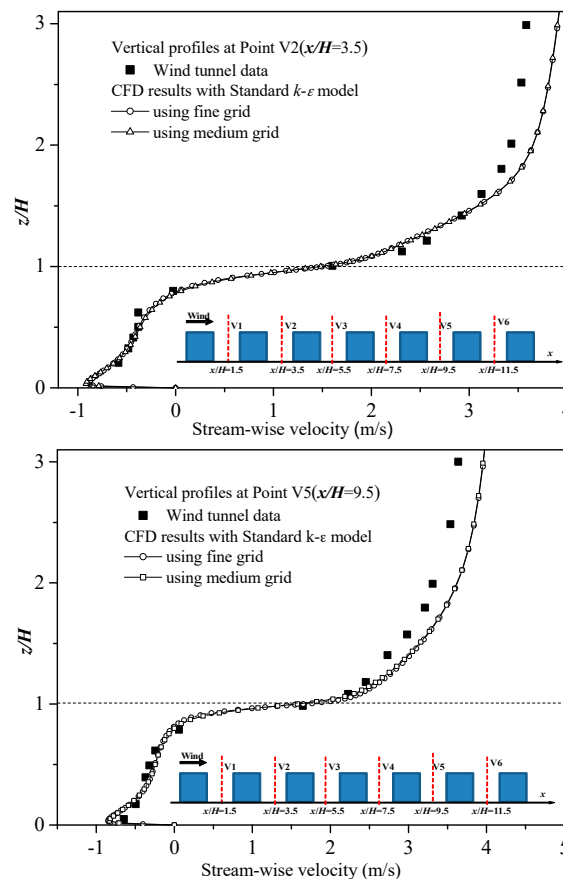
Obviously, the physical meanings of Equations (8) and (9) are that the larger decay rate (K) of $\langle C(t) \rangle / C(0)$ represents greater *ACH* and the pollutant dilution rate of the entire UCL volume; moreover,

a larger K at a point denotes that it requires a shorter time for external air to reach this point (i.e., the age of air is smaller).

3. Results and Discussion

3.1. Evaluation of CFD Simulations Using Wind Tunnel Data

To validate CFD simulations, Figure 3 compares wind tunnel data and CFD results in the CFD validation case by using the fine and medium grid, including time-averaged stream-wise velocity $\bar{u}(z)$ and turbulent kinetic energy $k(z)$ at Point V2 and Point V5. Compared to wind tunnel data, the standard $k-\epsilon$ model with present grid arrangements can provide results of mean flows $\bar{u}(z)$ in good agreement with wind tunnel data, but does a little worse at predicting $k(z)$ and can only predict the shape of vertical profile well. Such findings are similar with the same CFD validation studies the literature [23,51–53]. In addition, CFD results with the fine grid are only slightly different from those with the present medium grid. Given the results from the validation tests, we conclude that the application of the standard $k-\epsilon$ model with present grid arrangement is acceptable for the purposes of our research. There are relatively large discrepancies of stream-wise velocity between the simulated and measured results at the higher area. The possible reason is the ratio of vertical grid size to the stream-wise grid size is relatively large at much higher levels above the building roofs (i.e., at the higher area). However, the simulation results below and near the building roofs are performed in good quality, which is more important for UCL ventilation assessment.



(a) Validation case: vertical profiles of $\bar{u}(z)$ at Point V2 and Point V5

Figure 3. Cont.

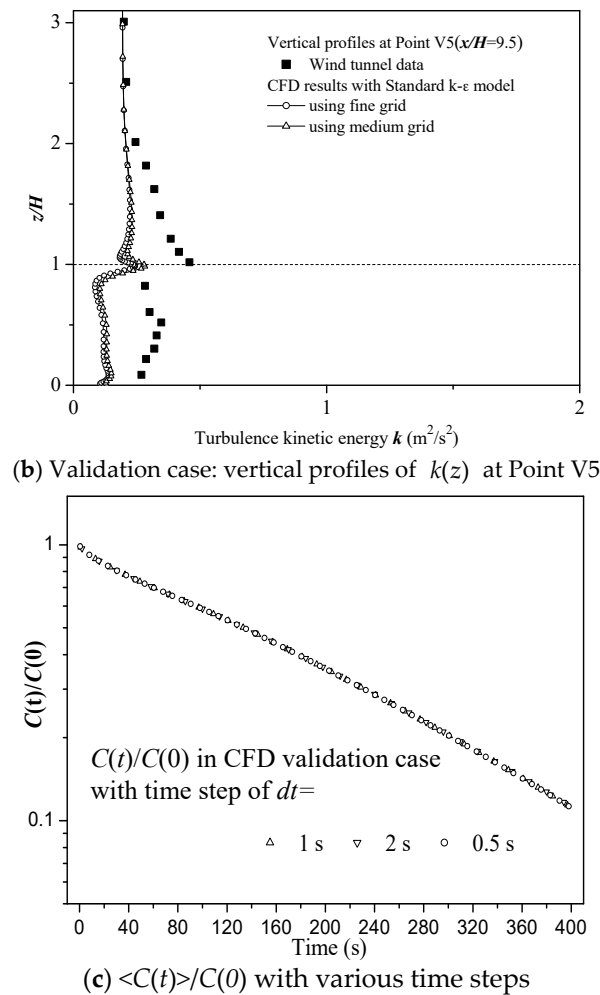


Figure 3. CFD results and experimental data in the CFD validation case: (a,b) Vertical profiles of $\bar{u}(z)$ and $k(z)$ at Point V2 and/or Point V5. (c) $\langle C(t) \rangle / C(0)$ with various time steps.

To further quantitatively evaluate the CFD models with the present grid arrangement and standard $k-\epsilon$ model, several statistical performance metrics are calculated, including the mean value, the standard deviation (named *St dev.* here), the fraction of predictions (here it is the CFD result) within a factor of two of observations (wind tunnel experiment data) named *FAC2*, the normalized mean error (*NMSE*), the fraction bias (*FB*), and the correlation coefficient (*R*). Results of $\bar{u}(z)$ at point V2 and V5 and $k(z)$ at point V5 are shown in Table 2. According to the recommended reference criteria [52], for $\bar{u}(z)$ prediction, a good performing simulation model is supposed to meet the statistical metrics standards as below: $FAC2 \geq 0.5$; $NMSE \leq 1.5$; $-0.3 \leq FB \leq 0.3$. All the results satisfy the recommended criteria except the *FB* value of $k(z)$ at point V5, which exceeds the upper limit, while its *FAC2* just reaches the threshold. Furthermore, $k(z)$ has a quite low *R* between wind tunnel data and CFD results. Results suggest a poorer quality of numerical prediction of $k(z)$ than $\bar{u}(z)$. As for $\bar{u}(z)$, although it is overestimated as the *FR* is positive, the value of *R* is particularly high (0.93) at both V2 and V5, which implies a credible prediction of $\bar{u}(z)$. Assessing models' acceptance requires considering all relevant performance metrics rather than one specific index, so the CFD models are considered to have quite a satisfactory performance on the whole.

Table 2. Statistical performance metrics in CFD validation case.

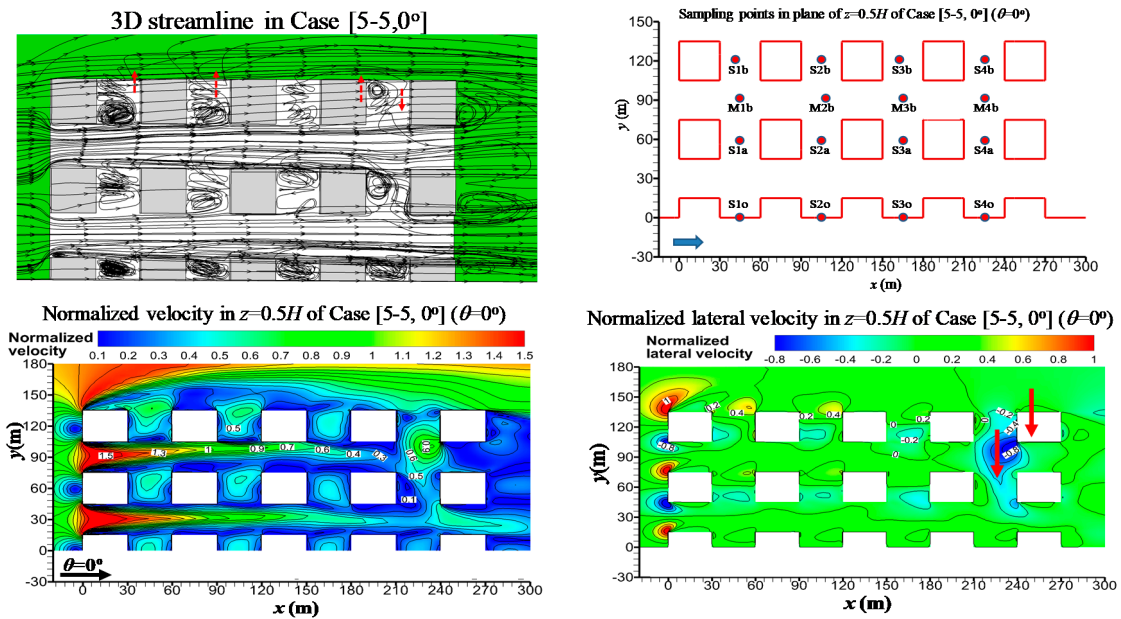
Variable (Position)	Cases	$k(z)$ (V5)	$\bar{u}(z)$ (V2)	$\bar{u}(z)$ (V5)
Average				
	Wind tunnel	0.31	1.61	1.58
	CFD	0.17	1.66	1.69
Standard deviation				
	Wind tunnel	0.07	1.76	1.70
	CFD	0.05	1.83	1.85
FAC2		0.50	0.93	0.93
NMSE		0.52	0.01	0.02
FB		0.61	-0.03	-0.07
R		0.13	0.93	0.93

Finally, with this CFD setup, Figure 3c shows the effects of time steps on the decay history of spatial mean concentration ($\langle C(t) \rangle / C(0)$). The decay rates for three time steps ($dt = 0.5, 1, 2$ s) are almost the same, confirming that the present time step ($dt = 1$ s) is good enough for ACH prediction.

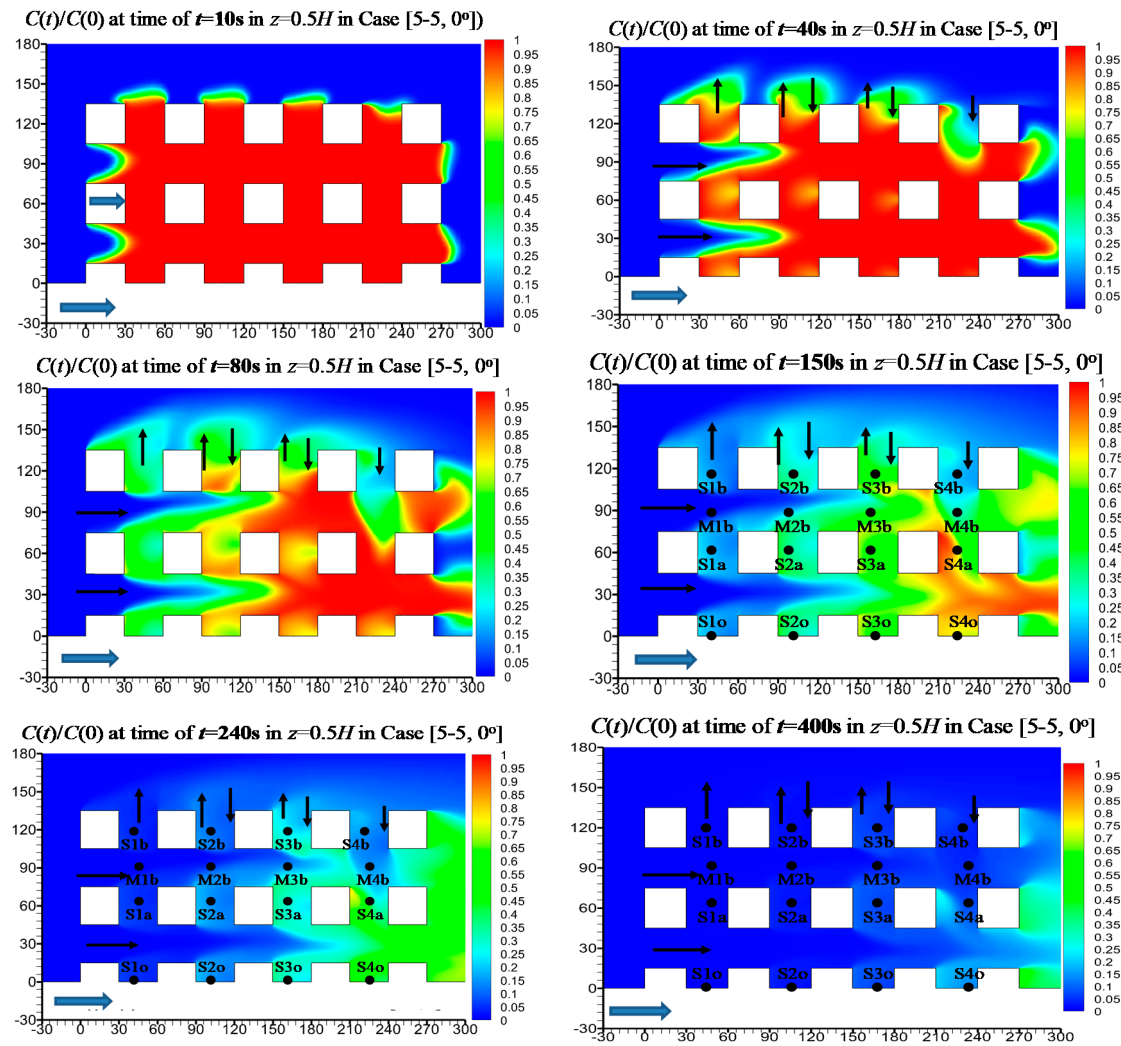
3.2. Flow and Concentration Decay in an Example Case [5-5, 0°] ($\theta = 0^\circ$)

As an example to analyze the concentration decay history related to UCL ventilation, Figure 4 shows 3D streamline, normalized velocity ($V = \sqrt{\bar{u}^2 + \bar{v}^2 + \bar{w}^2}$), normalized lateral velocity (\bar{v}) in the plane of $z = 0.05H$, and normalized concentration ($C(t)/C(0)$) at a time of 10 s to 400 s in $z = 0.5H$ in Case [5-5, 0°]. Here the velocity and lateral velocity are normalized by the freestream velocity at the same height of $z = 0.05H$ (see Equation (1)), similar to the literature [17,18]. Then Figure 5 displays the concentration history ($C(t)/C(0)$) and the age of air (τ_p) at various points in $z = 0.5H$ in Case [5-5, 0°]. In the main streets (Figure 4a), the flow is channeled toward downstream regions. In the secondary streets 3D vortices and helical flows exist in building wake regions. Across the lateral UCL boundary (at $y = 135$ m), there are lateral helical airflows leaving or re-entering UCL volume. Points with bigger concentration decay rates (K) experience better ventilation, a greater dilution rate, and smaller age of air (τ_p). Figures 4b and 5a,b show that the ventilation in upstream regions is better than in downstream regions (Figure 5a), and that in the main streets it is better than in the secondary streets (Figure 5a); those (Point S1b–S4b) near the lateral UCL boundary are better than those (Point S1o–S4o) in urban center regions (Figure 5b). Figure 5b displays that K at Point S4b is near to Point S3b, and that at Point S4b near the UCL lateral boundaries it is much bigger than at Point S4o because there is a strong helical inflow across lateral boundaries, bringing in external air to help with pollutant dilution at Point S4b (Figure 4a). Similarly, in Figure 5c, the τ_p at Points M1b to M4b in the main streets is much smaller than Points S1a to S4a and Points S1o to S4o in the secondary streets. Moreover, air at Points S1b to S4b near lateral UCL boundaries is younger than Points S1o to S4o far from lateral UCL boundaries. More importantly, the τ_p at Points S4b, S4a, and S4o (89.5 to 249.4 s) verifies that the lateral inflow across UCL lateral boundaries significantly improves the ventilation at Point S4b.

There are qualitative differences between the decay curves in Figure 5a,b. Analyzing the behavior of the decay curves in Figure 5a, for example, after some time the decay rates at Points S1a and M1b are similar to each other. This is a manifestation that there is a coupling interaction between the two locations (more explanation can be found by comparing Figure 11.2 in [32]). However, in Figure 5b all curves exhibit different decay rates, which shows that is not feedback (no backflow against the wind direction) that connects the different locations (see also Chapter 11 in [32]).

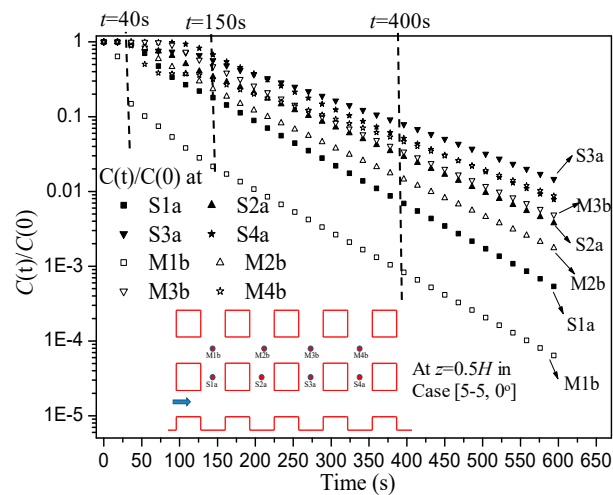


(a) 3D streamline and normalized velocity in $z = 0.05H$ in Case [5-5, 0°] ($\theta = 0^\circ$)

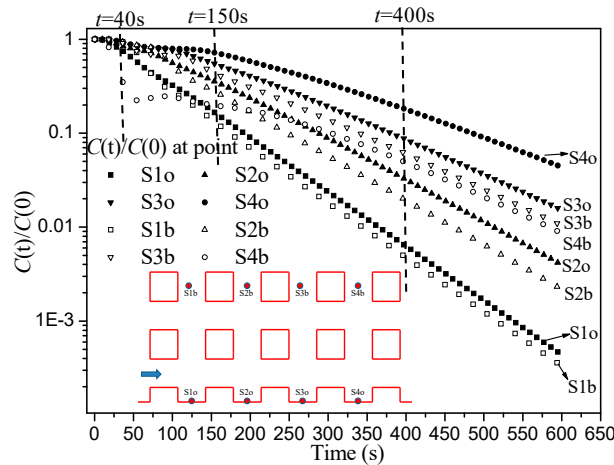


(b) $C(t)/C(0)$ at time of 10 s to 400 s at $z = 0.5H$ in Case [5-5, 0°] ($\theta = 0^\circ$)

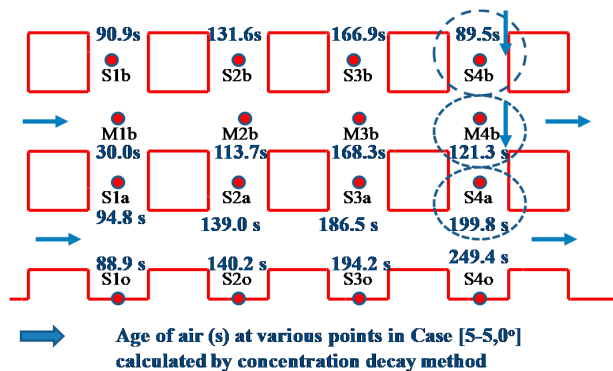
Figure 4. In Case [5-5, 0°] ($\theta = 0^\circ$): (a) 3D streamline and normalized velocity in $z = 0.05H$; (b) $C(t)/C(0)$ at time of 10 s to 400 s in plane of $z = 0.5H$.



(a) $C(t)/C(0)$ at Points S1a–S4a and M1b–M4b at $z = 0.5H$ in Case [5-5, 0°]



(b) $C(t)/C(0)$ at Points S1o–S4o and S1b–S4b at $z = 0.5H$ in Case [5-5, 0°]



(c) Age of air at these points at $z = 0.5H$ in Case [5-5, 0°]

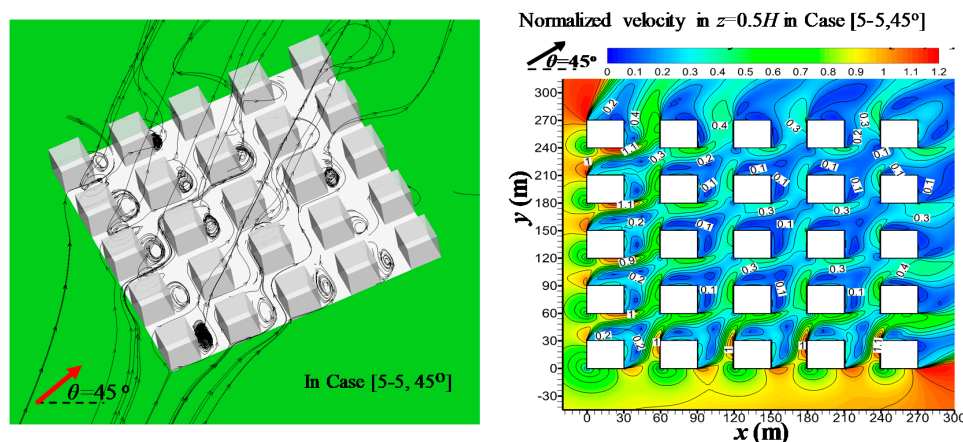
Figure 5. $C(t)/C(0)$ in Case [5-5, 0°]: (a) at Points S1a–S4a and M1b–M4b in $z = 0.5H$, (b) at Points S1o–S4o and S1b–S4b in $z = 0.5H$. (c) Age of air at these points.

3.3. Effect of Overall Urban Form and Ambient Wind Direction on UCL Ventilation

As an example, Figure 6 shows 3D streamline, normalized velocity, $C(t)/C(0)$, and τ_p at various points in $z = 0.5H$ in Case [5-5, 45°]. Here the velocity is normalized by the freestream velocity at the same height of $z = 0.5H$. Obviously wind brings clean air into an urban area downstream for pollutant dilution. There are recirculation regions with small wind speed (Figure 6a). Thus, the concentration decay rate (the age of air) is relatively small (large) in downstream regions and recirculation regions (Figure 6b,c).

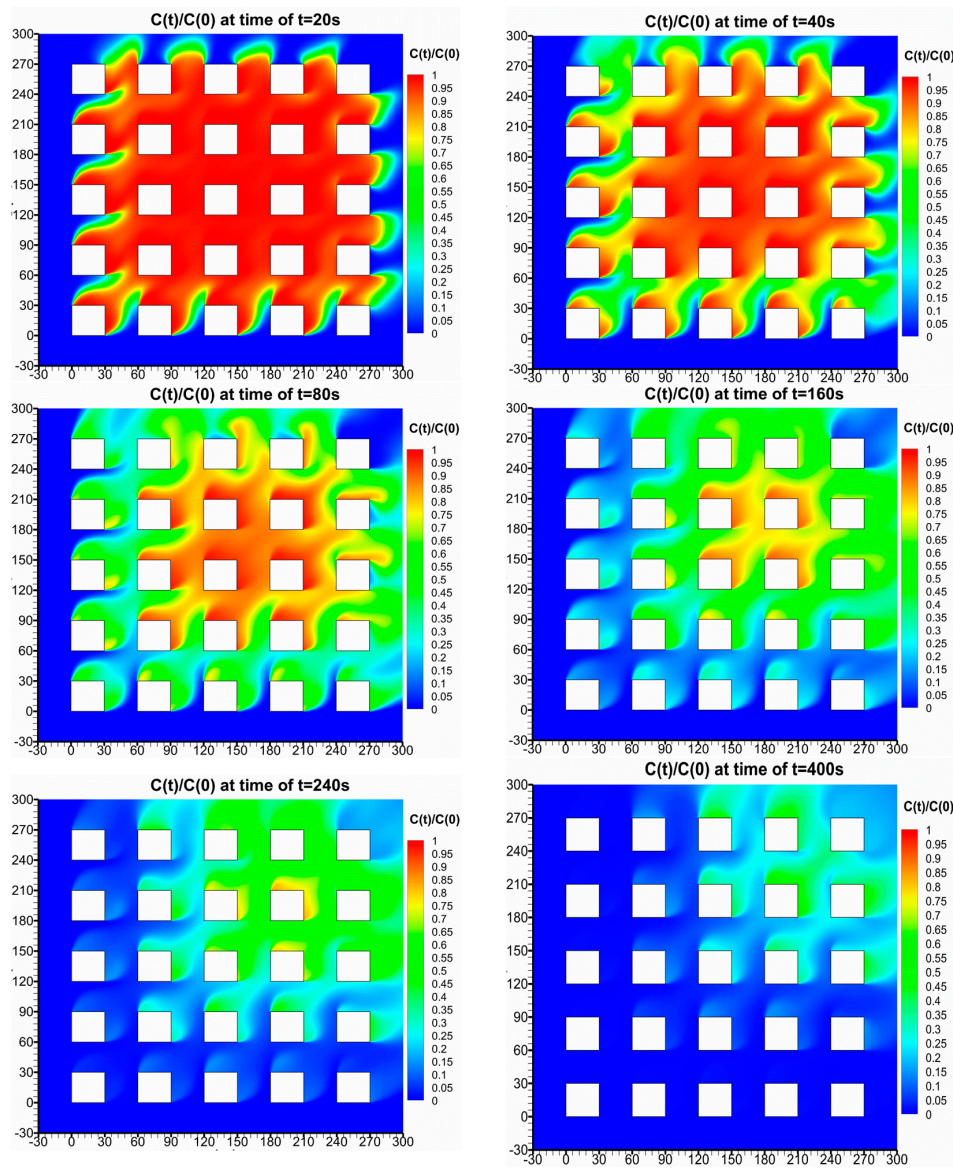
Then Figure 7a,b further displays the normalized velocity in Case [7-7, θ°] (square overall urban form) and Case [10-5, θ°] (rectangular overall urban form) with similar total UCL air volume. Obviously ambient wind directions significantly influence the flow pattern for both overall urban forms (Figure 7a,b). As $\theta = 0^\circ$, there are lateral flows across lateral UCL boundaries. With oblique winds, the flows enter UCL across two sides in upstream regions and leave across the other two toward downstream. The velocity is relatively small in recirculation regions and the presence of street crossings produces considerable momentum and scalar exchange between neighbor streets. In particular, Figure 7a confirms that, for square overall urban form, UCL models with $\theta = 45^\circ$ experience more recirculation regions and quicker wind reduction than those with $\theta = 0^\circ$. It is consistent with the literature that building arrays with $\theta = 45^\circ$ produce greater flow resistances and worse UCL ventilation performance than $\theta = 0^\circ$ [2,7,23,58,59]. Such characteristics with $\theta = 45^\circ$ produce adverse effects on its ventilation performance in contrast to $\theta = 0^\circ$.

Then Figure 7c,d displays the net ACH by concentration decay method, ACH_T and ACH_{turb} in cases of Group I. Since Case [5-5, θ°] and Case [7-7, θ°] are of square overall urban form, the flows with $\theta = 0^\circ, 15^\circ, 30^\circ, 45^\circ$ are the same with those with $\theta = 90^\circ, 75^\circ, 60^\circ, 45^\circ$. Obviously for Case [5-5, θ°] ($L_x = L_y = 270$ m) and Case [7-7, θ°] ($L_x = L_y = 390$ m), $\theta = 0^\circ$ attains smaller ACH_T and ACH_{turb} but bigger net ACH than $\theta = 15^\circ, 30^\circ, 45^\circ$. Therefore, the parallel wind ($\theta = 0^\circ$) experiences the best overall UCL ventilation with the highest net ventilation efficiency. It can be explained that UCL models with non-parallel approaching wind (for example $\theta = 45^\circ$ in Figure 7a) experience more recirculation regions and weaker wind due to the greater blockage induced by buildings, which can reduce the flow rates flushing through UCL space that never return or decrease the actual or net air change rate of the entire UCL space. Then for Case [10-5, θ°] with a rectangular overall urban form ($L_x = 570$ m, $L_y = 270$ m), as θ varies from 0° to 90° , ACH_T rises from 16.1 h^{-1} to 29.0 h^{-1} , ACH_{turb} first increases from $\theta = 0^\circ$ to $\theta = 45^\circ$ then decreases to $\theta = 90^\circ$, and the net ACH rises from 14.3 h^{-1} to 28.7 h^{-1} . Thus $\theta = 90^\circ$ obtains the best overall UCL ventilation in Case [10-5, θ°], in which the approaching wind is parallel to the shorter urban size ($L_y = 270$ m). Finally, the net ACH by the concentration decay method are always much smaller than the sum of ACH_{turb} and ACH_T . This verifies that UCL ventilation efficiency is limited. As discussed in Section 2.4.2, the recirculation flows in UCL space and turbulent fluctuations across street roofs induce a significant fraction of fluid particles to return or revisit UCL space across UCL boundaries after they leave it. However, UCL ventilation mainly depends on the flow rates flushing UCL space and leaving it to never return. Thus only a fraction of ACH_{turb} and ACH_T contributes to UCL ventilation and the actual air change rate is limited due to the ventilation efficiency problems.

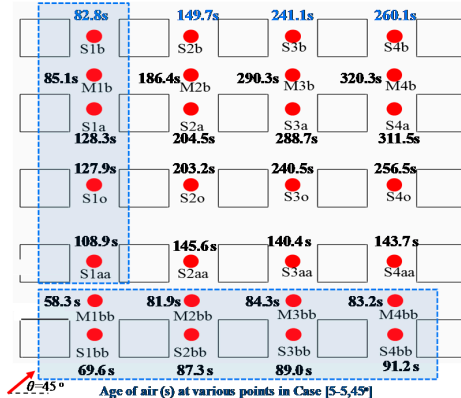


(a) 3D streamlines and the normal velocity at $z = 0.5H$ in Case [5-5, 45°]

Figure 6. Cont.



(b) $C(t)/C(0)$ at $z = 0.5H$ at time of 20 s to 400 s in Case [5-5, 45°]



(c) Age of air at some example points at $z = 0.5H$ in Case [5-5, 45°]

Figure 6. In Case [5-5, 45°]: (a) 3D streamline and the normal velocity in $z = 0.5H$; (b) $C(t)/C(0)$ in $z = 0.5H$ at time of 20 s to 400 s; (c) age of air at some example points at $z = 0.5H$.

By comparing the net ACH in these three cases (Figure 7d), ACH in Case [5-5, θ°] ($\sim 21.1\text{--}30.2\text{ h}^{-1}$) are larger than those in Case [7-7, θ°] ($\sim 16.8\text{--}18.6\text{ h}^{-1}$) for all wind directions because Case [5-5, θ°] has smaller urban size and requires shorter time for the approaching wind to flow through and will be exchanged more times by the external air within one hour. Rectangular urban form (Case [10-5, θ°]) attains greater ACH ($\sim 21.1\text{--}28.7\text{ h}^{-1}$) than the square urban form (Case [7-7, θ°]) ($\sim 16.8\text{--}18.6\text{ h}^{-1}$) for most ambient wind directions except $\theta = 0^\circ$ and 15° (~ 14.3 and 16.9 h^{-1}). It can be explained by two example cases: for Case [10-5, 0°] there are 10 rows of buildings for the approaching wind to flush, much longer than Case [10-5, 90°], in which there are only five rows of buildings for wind to flow through.

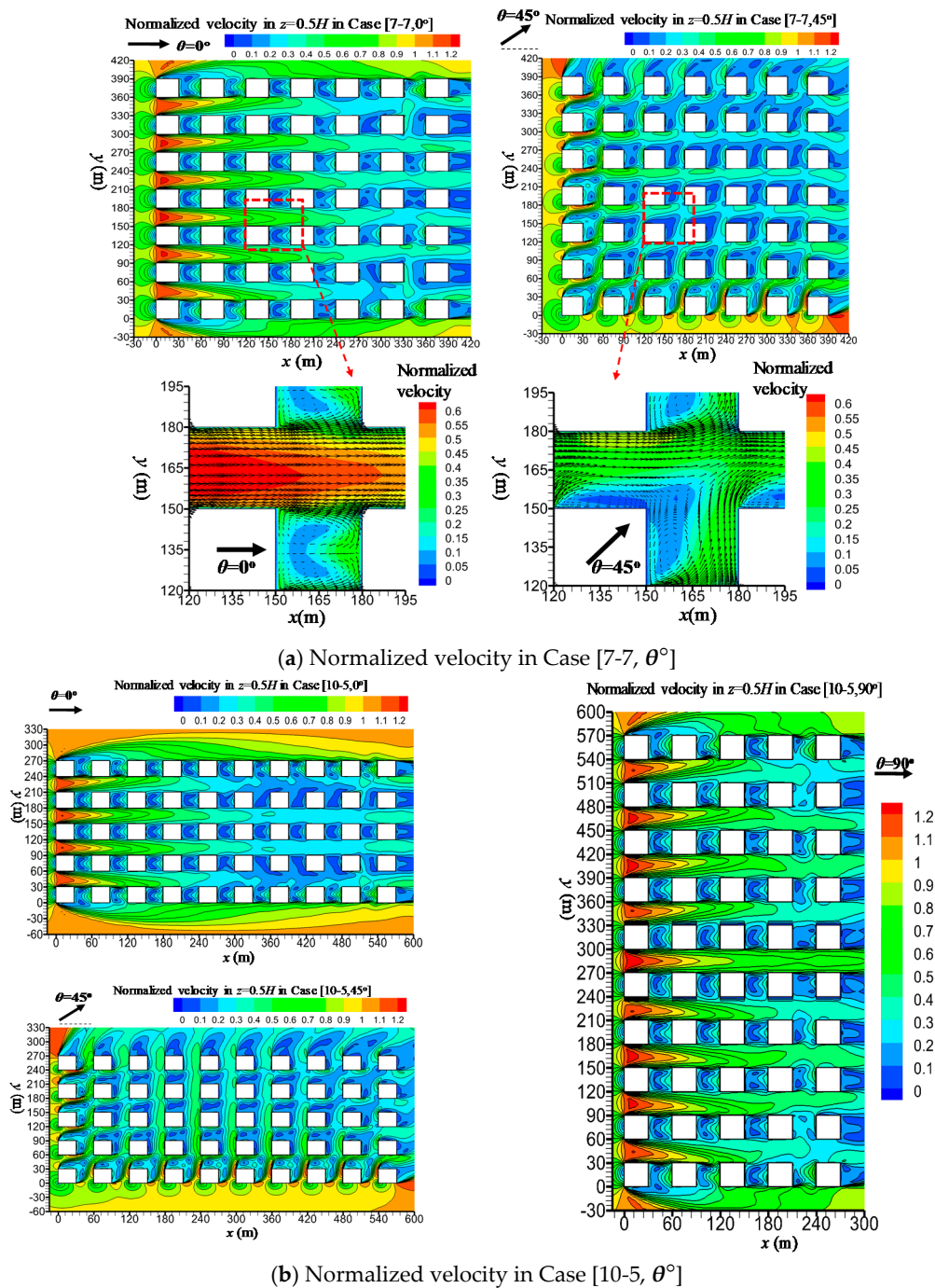
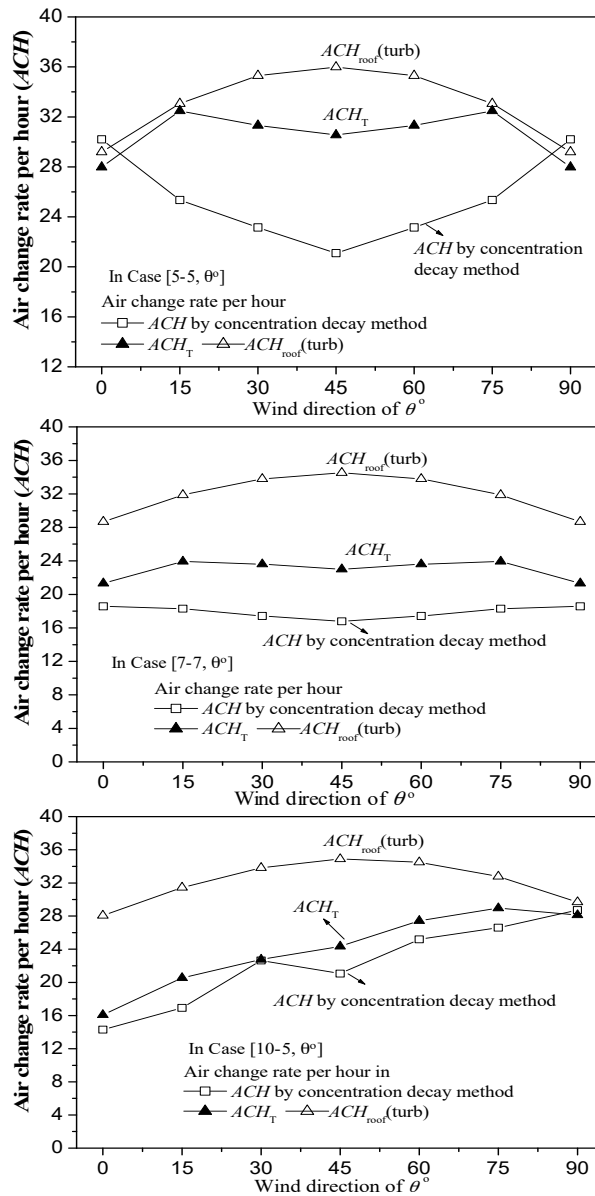
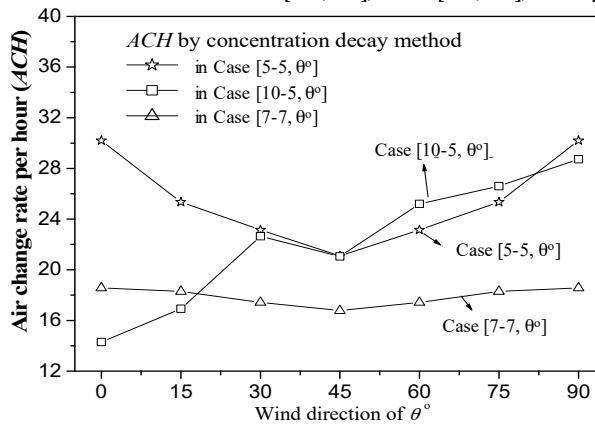


Figure 7. Cont.



(c) Various ACH indexes in Case [5-5, θ°], Case [7-7, θ°], Case [10-5, θ°]

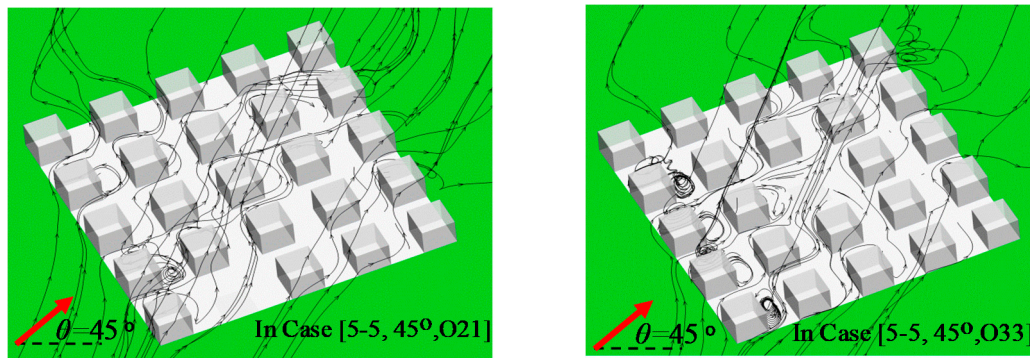


(d) Net ACH by concentration decay method in these three cases

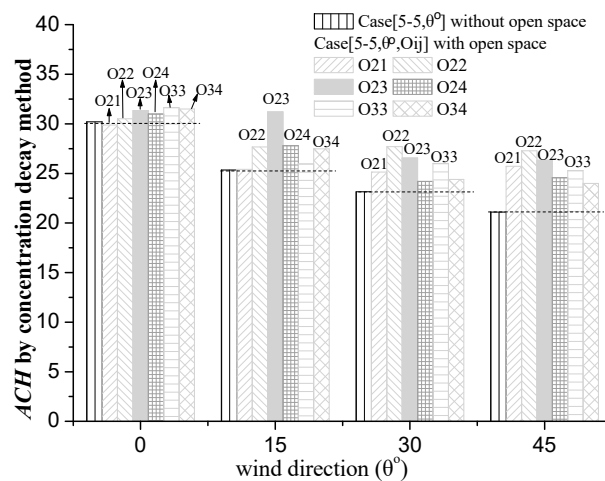
Figure 7. Normalized velocity in (a) Case [7-7, θ°]; (b) case [10-5, θ°]; (c) Various ACH indexes in Case [5-5, θ°], Case [7-7, θ°], Case [10-5, θ°]; (d) Net ACH by concentration decay method in these three cases.

3.4. Effect of Open Space Arrangements and Ambient Wind Direction

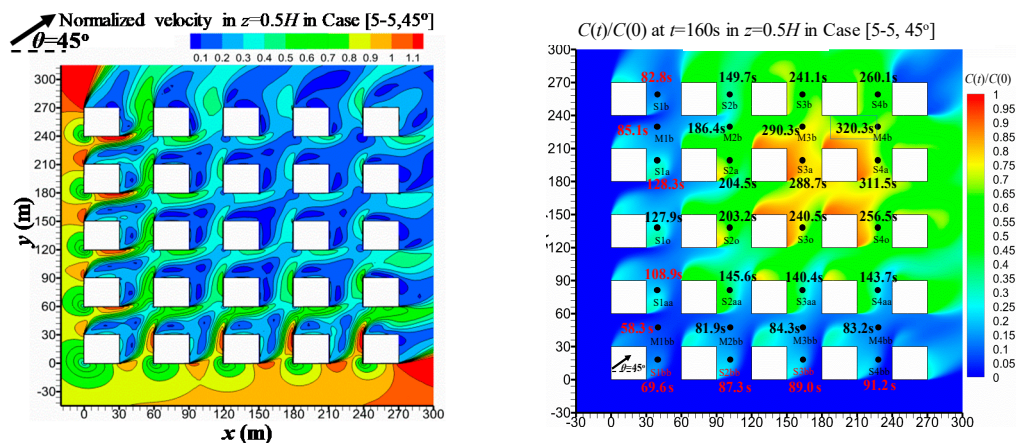
Open space arrangements have been regarded as one possible way to improve UCL ventilation. As shown in Table 1 (Group II), this subsection investigates 24 test cases with six kinds of open space arrangements under four wind directions ($\theta = 0^\circ, 15^\circ, 30^\circ, 45^\circ$). Note that Oij represents the building of position i-j is removed for better ventilation. Figure 8a displays the 3D streamline in two example cases; Figure 8b further summarizes the net ACH by the concentration decay method in all 24 test cases.



(a) 3D streamline in Case [5-5, 45°, O21] and Case [5-5, 45°, O33]



(b) The net ACH in all 24 test cases



(c) $C(t)/C(0)$ at $t = 160s$, age of air at various points and normalized velocity in $z = 0.5H$ in Case [5-5, 45°]

Figure 8. Cont.

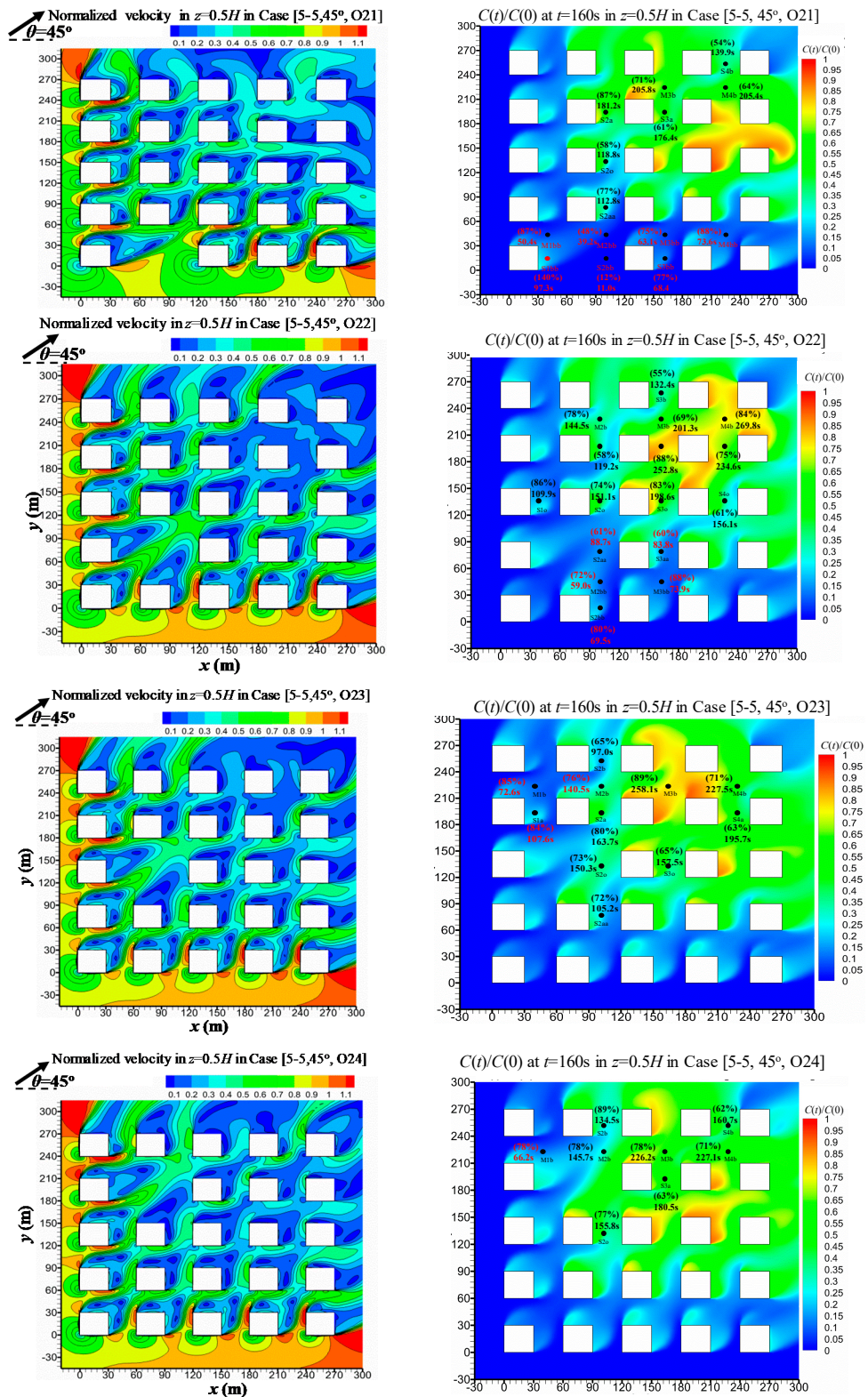
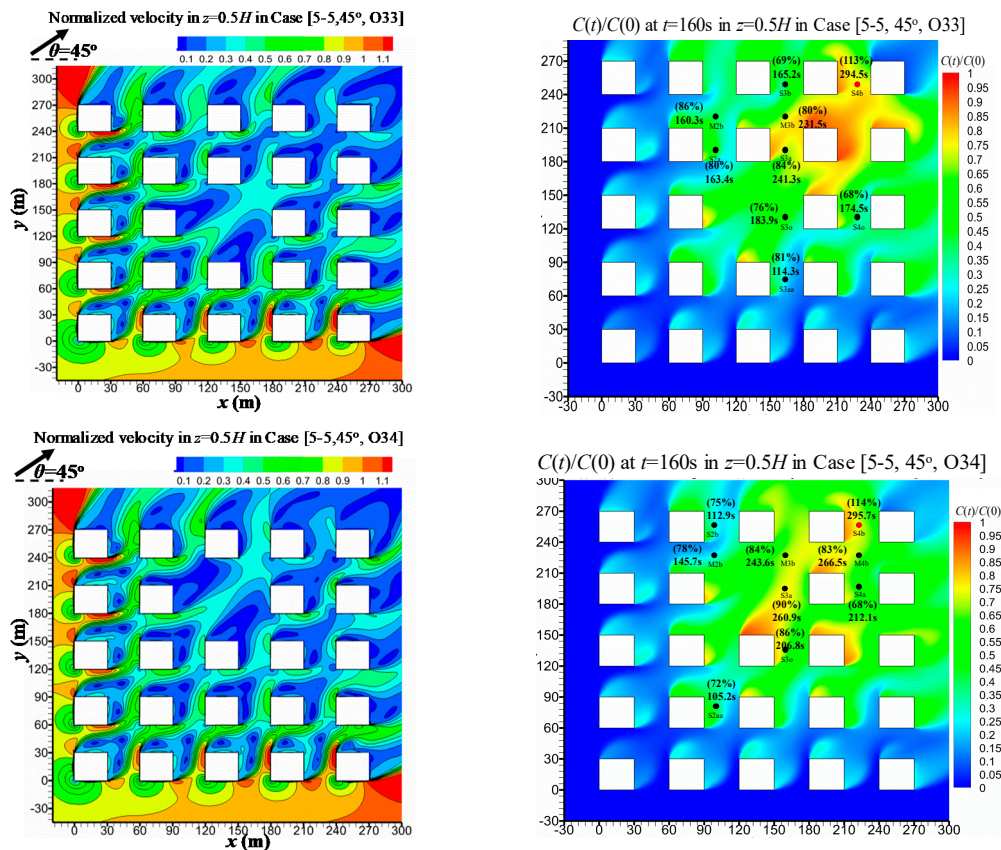


Figure 8. Cont.



(d) $C(t)/C(0)$ at $t = 160$ s, age of air and normalized velocity at $z = 0.5H$ in Case [5-5, 45°, Oij]

Figure 8. (a) 3D streamline in Case [5-5, 45°, O21] and Case [5-5, 45°, O33]; (b) the net ACH in all 24 test cases. (c,d) $C(t)/C(0)$ at $t = 160$ s, age of air at various points, and normalized velocity at $z = 0.5H$ in Case [5-5, 45°] and Case [5-5, 45°, Oij].

Obviously, with oblique wind ($\theta = 15^\circ, 30^\circ, 45^\circ$) most open space arrangements improve UCL ventilation more than $\theta = 0^\circ$ (see Figure 8b). The improvement of UCL ventilation is the best as $\theta = 45^\circ$. Thus $\theta = 45^\circ$ is emphasized in the below analysis. Figure 8c,d shows $C(t)/C(0)$ at $t = 160$ s, age of air (τ_p) at various points and normalized velocity in $z = 0.5H$ in Case [5-5, 45°] and Case [5-5, 45°, Oij]. Because the open space arrangements act as a kind of ventilation corridor and reduce the total flow resistances induced by buildings, wind speed increases in regions near/surrounding open space and in its downstream regions; subsequently, the concentration decay processes speed up and the ventilation becomes better. The percentage data of τ_p in Figure 8d refer to the ratio between τ_p with open space (Case [5-5, 45°, Oij] in Figure 8d) and that without open space (Case [5-5, 45°] in Figure 8c). This percentage at some points can be from 12% to 65%, showing that τ_p significantly decreases in the regions near these points. Obviously, the effects of open space on flow pattern satisfy the variations of τ_p (Figure 8c,d) and overall ACH (Figure 8b).

4. Conclusions

As a novelty, this paper confirms the temporal decay profile of concentration in UCL models accords with the exponential decay law, and it is effective to introduce the concentration decay method into CFD simulations to predict the net air change rate (ACH) flushing UCL space and never returning. Street-scale (~ 100 m), medium-dense ($\lambda_f = \lambda_p = 0.25, H/W = 1$), urban-like geometries are studied under neutral atmospheric conditions. The standard $k-\epsilon$ model is first successfully evaluated by wind tunnel data. Then the flow pattern, the concentration decay rate and age of air (τ_p) at some points in UCL space, the net ACH are analyzed.

With a parallel approaching wind ($\theta = 0^\circ$), UCL models with larger urban sizes ($Lx = Ly = 390$ m) experience smaller ACH ($\sim 16.8\text{--}18.6$ h $^{-1}$) than smaller UCL models ($Lx = Ly = 270$ m, $ACH \sim 21.1\text{--}30.2$ h $^{-1}$). The urban age of air τ_p at points near upstream/lateral UCL boundaries is smaller (or the air is younger) than far from them. For the square overall urban form, the parallel wind attains greater net ACH than non-parallel wind ($\theta = 15^\circ, 30^\circ, 45^\circ$). For the rectangular overall urban form ($Lx = 570$ m, $Ly = 270$ m), ACH is greater than the square overall urban form ($Lx = Ly = 390$ m) under most wind directions ($\sim 21.1\text{--}28.7$ h $^{-1}$ as $\theta = 30^\circ$ to 90°), and the ventilation is the best as the approaching wind is parallel to the shorter urban size (~ 28.7 h $^{-1}$ as $\theta = 90^\circ$). With open space arrangements, the dilution capacity near/surrounding open space and in its downstream regions is enhanced; moreover, most open space arrangements are more effective at improving UCL ventilation under oblique wind directions ($\theta = 15^\circ, 30^\circ, 45^\circ$) than $\theta = 0^\circ$, and the best ventilation improvement by open space appears as $\theta = 45^\circ$.

Similar to the purging flow rate [2,23], ACH calculated by the concentration decay approach has been proven effective to evaluate the effects of urban morphologies on the overall UCL ventilation capacity induced by mean flows and turbulent diffusions, which seems to be a better ventilation index than ACH calculated by the volumetric flow rates integrating the normal mean velocity or fluctuation velocity across UCL boundaries. Turbulent diffusions across open street roofs have been proven to significantly contribute to UCL ventilation, but further investigations are still required to analyze the net ventilation efficiency by mean flows and turbulence.

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Author Contributions: Qun Wang did the major CFD simulations. Yuanyuan Lin performed the CFD validation study. Mats Sandberg and Shi Yin contributed to the research discussion and writing the introduction. Jian Hang finished the manuscript and acts as the supervisor of the first author and corresponding author.

Conflicts of Interest: The authors declare no conflict of interest.

Nomenclature

A	area of a surface (m 2)
ACH	air change rate per hour by concentration decay method
ACH_T	ACH calculated by Q_T for entire UCL volume
ACH_{turb}	ACH calculated by $Q_{roof}(turb)$ for entire UCL volume
B, H, L, W	building width, building height, total length, street width
$C, \langle C \rangle$	time-averaged pollutant concentration and its spatial mean value
K_c, ν_t	turbulent eddy diffusivity of pollutant and momentum $K_c = \nu_t / S_{ct}$
λ_p	building area density (or plan area index)
λ_f	frontal area density (or frontal area index)
k, ϵ	turbulent kinetic energy and its dissipation rate
\vec{n}	normal direction of street openings or canopy roofs
Q	flow rate through street openings or street roofs
Q_{in}, Q_{out}	total inflow and outflow rate by mean flows across UCL boundaries
Q_T	total ventilation flow rate by mean flows
$Q_{roof}(turb)$	effective flow rate across street roofs by turbulence
S_{ct}	turbulent Schmidt number
σ_w	fluctuation velocity on street roofs
τ_p	age of air (s)
$U_0(z)$	velocity profiles used at CFD domain inlet for ventilation cases
U_{ref}	reference velocity at $z = H$ at the domain inlet
\vec{u}_j, x_j	velocity and coordinate components

\bar{V}_j	velocity vector
Vol	control volume
x, y, z	stream-wise, span-wise, vertical directions
$\bar{u}, \bar{v}, \bar{w}$	stream-wise, lateral, vertical velocity components
u', v', w'	stream-wise, lateral, vertical velocity fluctuations

References

- Oke, T.R. *Boundary Layer Climates*, 2nd ed.; Methuen Publishing: London, UK, 1987; p. 289.
- Bady, M.; Kato, S.; Huang, H. Towards the application of indoor ventilation efficiency indices to evaluate the air quality of urban areas. *Build. Environ.* **2008**, *43*, 1991–2004. [[CrossRef](#)]
- Hang, J.; Sandberg, M.; Li, Y.G. Age of air and air exchange efficiency in idealized city models. *Build. Environ.* **2009**, *44*, 1714–1723. [[CrossRef](#)]
- Buccolieri, R.; Sandberg, M.; Di Sabatino, S. City breathability and its link to pollutant concentration distribution within urban-like geometries. *Atmos. Environ.* **2010**, *44*, 1894–1903. [[CrossRef](#)]
- Hang, J.; Li, Y.G. Age of air and air exchange efficiency in high-rise urban areas. *Atmos. Environ.* **2011**, *45*, 5572–5585. [[CrossRef](#)]
- Ramponi, R.; Blocken, B.; Laura, B.; Janssen, W.D. CFD simulation of outdoor ventilation of generic urban configurations with different urban densities and equal and unequal street widths. *Build. Environ.* **2015**, *92*, 152–166. [[CrossRef](#)]
- Hang, J.; Luo, Z.W.; Sandberg, M.; Gong, J. Natural ventilation assessment in typical open and semi-open urban environments under various wind directions. *Build. Environ.* **2013**, *70*, 318–333. [[CrossRef](#)]
- Kim, J.J. Assessment of observation environment for surface wind in urban areas using a CFD model. *Atmosphere* **2015**, *25*, 449–459. (In Korean)
- Ashie, Y.; Kono, T. Urban-scale CFD analysis in support of a climate-sensitive design for the Tokyo Bay area. *Int. J. Climatol.* **2011**, *31*, 174–188. [[CrossRef](#)]
- Yang, X.; Li, Y.G. The impact of building density and building height heterogeneity on average urban albedo and street surface temperature. *Build. Environ.* **2015**, *90*, 146–156. [[CrossRef](#)]
- Hang, J.; Luo, Z.; Wang, X.; He, L.; Wang, B.; Zhu, W. The influence of street layouts and viaduct settings on daily CO exposure and intake fraction in idealized urban canyons. *Environ. Pollut.* **2017**, *220*, 72–86. [[CrossRef](#)] [[PubMed](#)]
- Luo, Z.; Li, Y.G.; Nazaroff, W.W. Intake fraction of nonreactive motor vehicle exhaust in Hong Kong. *Atmos. Environ.* **2010**, *44*, 1913–1918. [[CrossRef](#)]
- Ng, W.; Chau, C. A modeling investigation of the impact of street and building configurations on personal air pollutant exposure in isolated deep urban canyons. *Sci. Total Environ.* **2014**, *468*, 429–448. [[CrossRef](#)] [[PubMed](#)]
- Meroney, R.N.; Pavegeau, M.; Rafailidis, S.; Schatzmann, M. Study of line source characteristics for 2-D physical modelling of pollutant dispersion in street canyons. *J. Wind Eng. Ind. Aerodyn.* **1996**, *62*, 37–56. [[CrossRef](#)]
- Li, X.X.; Liu, C.H.; Leung, D.Y.C.; Lam, K.M. Recent progress in CFD modelling of wind field and pollutant transport in street canyons. *Atmos. Environ.* **2006**, *40*, 5640–5658. [[CrossRef](#)]
- Li, X.X.; Liu, C.H.; Leung, D.Y.C. Numerical investigation of pollutant transport characteristics inside deep urban street canyons. *Atmos. Environ.* **2009**, *43*, 2410–2418. [[CrossRef](#)]
- Zhang, Y.W.; Gu, Z.; Lee, S.C.; Fu, T.M.; Ho, K.F. Numerical simulation and in situ investigation of fine particle dispersion in an actual deep street canyon in Hong Kong. *Indoor Built Environ.* **2011**, *20*, 206–216. [[CrossRef](#)]
- Lin, L.; Hang, J.; Wang, X.X.; Wang, X.M.; Fan, S.J.; Fan, Q.; Liu, Y.H. Integrated effects of street layouts and wall heating on vehicular pollutant dispersion and their reentry into downstream canyons. *Aerosol Air Qual. Res.* **2016**, *16*, 3142–3163. [[CrossRef](#)]
- Dong, J.L.; Tan, Z.J.; Xiao, X.M.; Tu, J.Y. Seasonal changing effect on airflow and pollutant dispersion characteristics in urban street canyons. *Atmosphere* **2017**, *8*, 43. [[CrossRef](#)]
- Gu, Z.; Zhang, Y.; Cheng, Y.; Lee, S.C. Effect of uneven building layout on air flow and pollutant dispersion in non-uniform street canyons. *Build. Environ.* **2011**, *46*, 2657–2665. [[CrossRef](#)]

21. Hang, J.; Li, Y.; Sandberg, M.; Buccolieri, R.; Di Sabatino, S. The influence of building height variability on pollutant dispersion and pedestrian ventilation in idealized high-rise urban areas. *Build. Environ.* **2012**, *56*, 346–360. [[CrossRef](#)]
22. Panagiotou, I.; Neophytou, M.K.A.; Hamlyn, D.; Britter, R.E. City breathability as quantified by the exchange velocity and its spatial variation in real inhomogeneous urban geometries: An example from central London urban area. *Sci. Total Environ.* **2013**, *442*, 466–477. [[CrossRef](#)] [[PubMed](#)]
23. Lin, M.; Hang, J.; Li, Y.G.; Luo, Z.W.; Sandberg, M. Quantitative ventilation assessments of idealized urban canopy layers with various urban layouts and the same building packing density. *Build. Environ.* **2014**, *79*, 152–167. [[CrossRef](#)]
24. Lee, D.; Kim, J.J. A study on the characteristics of flow and reactive pollutants' dispersion in step-up street canyons using a CFD model. *Atmosphere* **2015**, *25*, 473–482. (In Korean) [[CrossRef](#)]
25. Chen, L.; Hang, J.; Sandberg, M.; Claesson, L.; Di Sabatino, S.; Wigo, H. The impacts of building height variations and building packing densities on flow adjustment and city breathability in idealized urban models. *Build. Environ.* **2017**, *118*, 344–361. [[CrossRef](#)]
26. Li, X.X.; Britter, R.E.; Norford, L.K. Transport processes in and above two-dimensional urban street canyons under different stratification conditions: results from numerical simulation. *Environ. Fluid Mech.* **2015**, *15*, 399–417. [[CrossRef](#)]
27. Nazarian, N.; Kleissl, J. Realistic solar heating in urban areas: Air exchange and street-canyon ventilation. *Build. Environ.* **2016**, *95*, 75–93. [[CrossRef](#)]
28. Cui, P.Y.; Li, Z.; Tao, W.Q. Wind-tunnel measurements for thermal effects on the air flow and pollutant dispersion through different scale urban areas. *Build. Environ.* **2016**, *97*, 137–151. [[CrossRef](#)]
29. Wang, X.X.; Li, Y.G. Predicting urban heat island circulation using CFD. *Build. Environ.* **2016**, *99*, 82–97. [[CrossRef](#)]
30. Fan, Y.F.; Hunt, J.C.R.; Li, Y.G. Buoyancy and turbulence-driven atmospheric circulation over urban areas. *J. Environ. Sci.* **2017**. [[CrossRef](#)]
31. Hang, J.; Wang, Q.; Chen, X.Y.; Sandberg, M.; Zhu, W.; Buccolieri, R.; Di Sabatino, S. City breathability in medium density urban-like geometries evaluated through the pollutant transport rate and the net escape velocity. *Build. Environ.* **2015**, *94*, 166–182. [[CrossRef](#)]
32. Etheridge, D.; Sandberg, M. *Building Ventilation: Theory and Measurement*; John Wiley & Sons: New York, NY, USA, 1996; pp. 573–633.
33. Lim, E.S.; Ito, K.; Sandberg, M. New ventilation index for evaluating imperfect mixing conditions—Analysis of Net Escape Velocity based on RANS approach. *Build. Environ.* **2013**, *61*, 45–56. [[CrossRef](#)]
34. Li, J.Q.; Ward, I.C. Developing computational fluid dynamics conditions for urban natural ventilation study. In Proceedings of the Building Simulation, Beijing, China, 3–6 September 2007.
35. Zaki, S.A.; Hagishima, A.; Tanimoto, J. Experimental study of wind-induced ventilation in urban building of cube arrays with various layouts. *J. Wind Eng. Ind. Aerodyn.* **2012**, *103*, 31–40. [[CrossRef](#)]
36. Padilla-Marcos, M.Á.; Meiss, A.; Feijó-Muñoz, J. Proposal for a simplified CFD procedure for obtaining patterns of the age of air in outdoor spaces for the natural ventilation of buildings. *Energies* **2017**, *10*, 1252. [[CrossRef](#)]
37. Li, X.X.; Liu, C.H.; Leung, D.Y.C. Development of a $k-\epsilon$ model for the determination of air exchange rates for street canyons. *Atmos. Environ.* **2005**, *39*, 7285–7296. [[CrossRef](#)]
38. Liu, C.H.; Leung, D.Y.C.; Barth, M.C. On the prediction of air and pollutant exchange rates in street canyons of different aspect ratio using large-eddy simulation. *Atmos. Environ.* **2005**, *39*, 1567–1574. [[CrossRef](#)]
39. Moonen, P.; Dorer, D.; Carmeliet, J. Evaluation of the ventilation potential of courtyards and urban street canyons using RANS and LES. *J. Wind Eng. Ind. Aerodyn.* **2011**, *99*, 414–423. [[CrossRef](#)]
40. Hang, J.; Sandberg, M.; Li, Y. Effect of urban morphology on wind condition in idealized city models. *Atmos. Environ.* **2009**, *43*, 869–878. [[CrossRef](#)]
41. Hang, J.; Li, Y.G. Ventilation strategy and air change rates in idealized high-rise compact urban areas. *Build. Environ.* **2010**, *45*, 2754–2767. [[CrossRef](#)]
42. Liu, J.; Luo, Z.; Zhao, J.; Shui, T. Ventilation in a street canyon under diurnal heating conditions. *Int. J. Vent.* **2012**, *11*, 141–154. [[CrossRef](#)]
43. Yang, L.; Li, Y.G. City ventilation of Hong Kong at no-wind conditions. *Atmos. Environ.* **2009**, *43*, 3111–3121. [[CrossRef](#)]

44. Yang, L.; Li, Y.G. Thermal conditions and ventilation in an ideal city model of Hong Kong. *Energy Build.* **2011**, *43*, 1139–1148. [[CrossRef](#)]
45. Gao, N.P.; Niu, J.L.; Perino, M.; Heiselberg, P. The airborne transmission of infection between flats in high-rise residential buildings: Tracer gas simulation. *Build. Environ.* **2008**, *43*, 1805–1817. [[CrossRef](#)]
46. Hooff, V.T.; Blocken, B. CFD evaluation of natural ventilation of indoor environments by the concentration decay method: CO₂ gas dispersion from a semi-enclosed stadium. *Build. Environ.* **2013**, *61*, 1–17. [[CrossRef](#)]
47. Fernando, H.J.S.; Zajic, D.; Di Sabatino, S.; Dimitrova, R.; Hedquist, B.; Dallman, A. Flow, turbulence, and pollutant dispersion in urban atmospheres. *Phys. Fluids* **2010**, *22*, 051301. [[CrossRef](#)]
48. Di Sabatino, S.; Buccolieri, R.; Salizzoni, P. Recent advancements in numerical modelling of flow and dispersion in urban areas: a short review. *Int. J. Environ. Pollut.* **2013**, *52*, 172–191. [[CrossRef](#)]
49. Blocken, B. Computational fluid dynamics for urban physics: Importance, scales, possibilities, limitations and ten tips and tricks towards accurate and reliable simulations. *Build. Environ.* **2015**, *91*, 219–245. [[CrossRef](#)]
50. Brown, M.J.; Lawson, R.E.; DeCroix, D.S.; Lee, R.L. *Comparison of Centerline Velocity Measurements Obtained Around 2D and 3D Buildings Arrays in a Wind Tunnel*; Report LA-UR-01-4138; Los Alamos National Laboratory: Los Alamos, NM, USA, 2001; p. 7.
51. Lien, F.S.; Yee, E. Numerical modelling of the turbulent flow developing within and over a 3-D building array, part I: A high-resolution Reynolds-averaged Navier-Stokes approach. *Bound. Layer Meteorol.* **2004**, *112*, 427–466. [[CrossRef](#)]
52. Santiago, J.L.; Martilli, A.; Martin, F. CFD simulation of airflow over a regular array of cubes Part I: Three-dimensional simulation of the flow and validation with wind-tunnel measurements. *Bound. Layer Meteorol.* **2007**, *122*, 609–634. [[CrossRef](#)]
53. Hang, J.; Li, Y.G. Wind conditions in idealized building clusters: Macroscopic simulations by a porous turbulence model. *Bound. Layer Meteorol.* **2010**, *136*, 129–159. [[CrossRef](#)]
54. Tominaga, Y.; Mochida, A.; Yoshie, R.; Kataoka, H.; Nozu, T.; Yoshikawa, M.; Shirasawa, T. AIJ guidelines for practical applications of CFD to pedestrian wind environment around buildings. *J. Wind Eng. Ind. Aerodyn.* **2008**, *96*, 1749–1761. [[CrossRef](#)]
55. Franke, J.; Hellsten, A.; Schlunzen, H.; Carissimo, B. The COST732 Best Practice Guideline for CFD simulation of flows in the urban environment: A summary. *Inter. J. Environ. Pollut.* **2011**, *44*, 419–427. [[CrossRef](#)]
56. FLUENT V6.3. User's Manual. Available online: <http://www.fluent.com> (accessed on 20 August 2017).
57. Blocken, B.; Stathopoulos, T.; Carmeliet, J. CFD simulation of the atmospheric boundary layer: Wall function problems. *Atmos. Environ.* **2007**, *41*, 238–252. [[CrossRef](#)]
58. Kim, J.J.; Baik, J.J. A numerical study of the effects of ambient wind direction on flow and dispersion in urban street canyons using the RNG $k-\epsilon$ turbulence model. *Atmos. Environ.* **2004**, *38*, 3039–3048. [[CrossRef](#)]
59. Kanda, M. Large-eddy simulations on the effects of surface geometry of building arrays on turbulent organized structures. *Bound. Layer Meteorol.* **2006**, *18*, 151–168. [[CrossRef](#)]

