



# Bathymetry Time Series Using High Spatial Resolution Satellite Images

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**Abstract:** The use of the new generation of remote sensors, such as echo sounders and Global Navigation Satellite System (GNSS) receivers with differential correction installed in a drone, allows the acquisition of high-precision data in areas of shallow water, as in the case of the channel of the Encañizadas in the Mar Menor lagoon. This high precision information is the first step to develop the methodology to monitor the bathymetry of the Mar Menor channels. The use of high spatial resolution satellite images is the solution for monitoring many hydrological changes and it is the basis of the three-dimensional (3D) numerical models used to study transport over time, environmental variability, and water ecosystem complexity.

Keywords: Mar Menor; spatio-temporal variability; Pleiades-1

## 1. Introduction

The issues of hydro-sedimentary processes in lagoons, estuaries, and enclosed bays are important in terms of coastal management, because they are generally characterized by a strong human presence that results in the release of contaminants and large quantities of nutrients into the ecosystem [1]. There are many methodologies to study the nutrients in ecosystems; one of them, possibly the most used, is the three-dimensional (3D) numerical model.

The use of numerical models as a tool to anticipate the effects of human impacts or climate change effects on marine ecosystems has increased in the last decade. They permit to simulate hydrodynamic process, tides, currents, wave action, sediment transport, pollutant dispersal or biological connectivity. A large list of models is available, like CSIRO [2], Delft3D-FLOW [3,4], MARS3D [5], MIKE 21/3 [6]; MIKE3 [7], MOHIDWater [8], 3D-MOHID [9], NHWAVE [10], REF-DIF1 [11], ROM [12,13], SELFE [14,15], SCHISM [16,17], SHYFEM [18,19], SHORECIRC [20], SWAN [21] or TELEMAC [22]. Some of these models are mainly applied to shallow coastal waters. A synthesis of concepts and recommendations for model selection, mainly applied to coastal lagoons modelling, is presented by [23].

In the Mar Menor lagoon, ROMS has been used to evaluate the water renewal conditions, and a time series spanning a period longer than one year (between 2010 and 2012) provides a record of the seasonal variability of the lagoon hydrodynamics, which was used to validate the hydrodynamic model implemented [24]. SHYFEM has been used to compare the water exchange and mixing in 10



Mediterranean lagoons (including the Mar Menor) in terms of water exchange and mixing behavior. The authors found that the exchange with the open sea highly influenced the transport time scales, although the wind can also enhance the exchange mechanisms in lagoons with more than one inlet [18]. The SHYFEM has been more specifically used for the Mar Menor lagoon under different dredging scenarios of the inlets, covering different dredging depths and extensions. From this study, it can be concluded that the impacts of dredging the channels on the hydrodynamics of a coastal lagoon increase with the magnitude of the dredging activities. In this sense, there is a threshold for the magnitude of these activities beyond which important environmental effects can be expected [25,26]. Additionally, the SHYFEM has been used to model environmental parameters such as salinity and temperature [26] as well as the connectivity between the Mar Menor lagoon and the Mediterranean using a Lagrangian model [27,28], so as for predicting the effects of climate change [29].

The 3D numerical models need precise and updated initial conditions (Xo, Yo, Zo) to run them [1]. In all the cases, but mainly for applications in coastal zones, the applicability and usefulness of the results are highly dependent on the quality and spatial resolution of the bathymetric information and model grid setup, which determines not only the water column depth, but also information on bed roughness and the incorporation of specific structures and features that affect water and sediment transport dynamics. Furthermore, although the numerical models are very reliable, bottom conditions are dynamic and can change due to natural dynamics, human activities, disasters induced by cut-off lows (CoLs), such as flooding [30], and changing of the coastline [31]. Therefore, it is crucial to regularly update the information and to improve both the accuracy of bathymetric measurements and the interpolation techniques to build high-resolution grids [32–36].

One of the approaches to get extensive bathymetric information in inland and coastal water masses with an increasing spatial resolution is developing different tools based in remote sensing techniques [37–41].

In several places around the world, the CoLs contribute to increases in precipitation and in extreme rainfall events [42]. In Spain, this atmospheric phenomenon has been known traditionally as the "gota fría", and also recently (due to the influence of meteorologists) as "dana"; both refer to CoLs. In recent years, CoLs have become more frequent in South East (SE) Spain and they are modifying the morphology of the coastline as well as the depth, seagrass, and bottom of aquatic ecosystems, particularly of the Mar Menor lagoon. An upgrade was designed to update the depth data and thus maintain the accuracy of the 3D numerical models, using the best techniques already available [43] or in development [44].

During the last semester of 2019, there were two CoLs over the Spanish Mediterranean coast that caused environmental disasters and human life loss.

Remote sensing has been used to monitor these environmental disasters [45]. Besides, remote sensing can also show the initial conditions in unaffected areas where the water is clear and the energy from the sun reaches the bottom. The energy in the water reaches the optical sensor placed above the water. The relationship between the water-leaving reflectance just above the surface and the water-leaving reflectance just below the surface is constant [46-48]. If the water is clear, the energy arrives from the bottom through the water column [49], in this case it is possible to determinate the water depth. The use of high-resolution satellite imagery over variable bottom types has allowed the determination of water depth with a new algorithm, which uses the relation between the neperian algorithms of two bands. The influence of the bottom is determined by a simple model, in which the bottom optical properties and the water optical properties are included [50,51]. It is noteworthy that the first studies on water depth by remote sensing in the 1970s were by simple regression methods with one band or multiple regression methods using two bands [52–54]. The second stage was to assess the influence of different types of bottom [55]. On the other hand, the water depth using passive optical sensors were compared with depth water obtained from LiDAR data [56]. According to the potential of the remote sensing, the images from different sensors have been used to determinate the bathymetric maps: Landsat 5 [57,58], Landsat 7 [59,60], Landsat 8 [61,62], ASTER [63], IKONOS [64], Worldview 2 [65], SPOT-4 [66], GeoEve-1 [67], and Pleiades [68]. In previous works carried out in the Mar Menor lagoon, this approach using remote sensing has been used to calculate the extinction coefficient of light (K). Then, after demonstrating its spatial heterogeneity, it was used to determine the concentration of chlorophyll *a* in the water column in shallow areas where the reflectance from the benthic meadows interferes with the signal of the water column [69]. Therefore, the calibration of the model requires field information provided by sounding equipment installed in ships, which cannot measure depths of less than 0.6 m [67]. In the last few years this limitation has been overcome due to the use of an Unmanned Surface Vehicle (USV) [64], which provides data from depths of just a few centimeters. The images from new sensors in satellites [70,71], planes, and new platforms (such as drones) have provided new and better tools for the management of coastal areas since the 1980s [72]. All remote sensing advances, tools, and techniques have been evaluated so that they can be adapted to new scenarios and to water management necessities [73–75]. However, the management of each water body has specific problems that have to be analyzed in terms of the local conditions in order to get the best results in the least possible time. Often, it is very difficult to provide solutions because there is not enough information, above all when studying areas such as the one that concerns us, with great human and environmental pressure. Good management needs a fast and efficient methodology as the basis for prediction models in near real-time, such as the monitoring Mar Menor geoportal in near real-time. In any case, the first step is to get good and precise initial data to feed the prediction models. Therefore, the aim of the present work is the use of high and very high-resolution spatial data, as well as time series, to assess the sand movement and volume of shallow water in natural channels that connect the Mar Menor lagoon and the Mediterranean Sea.

#### 2. Study Area and Materials

#### 2.1. Study Area

The Mar Menor is a hypersaline lagoon with a surface area of 135.76 km<sup>2</sup>, a volume of about  $0.65 \times 10^9$  m<sup>3</sup>, and a maximum depth of 7 m—located in the SE of the Iberian Peninsula, between the parallels 37°38′ and 37°50′ North latitude and the meridians 0°43′ and 0°57′ West longitude. There are five open channels or "golas" connecting the Mar Menor lagoon and the Mediterranean Sea. These can be seen in Figure 1c. The Encañizadas area is located at the north end of La Manga, on the southern edge of the Salinas de San Pedro Regional Park, and is made up of a group of small islets, surrounded by large flooded areas crossed by a network of shallow channels. Due to its morphology, it experiences periodic waterlogging related to tidal flows and changes in atmospheric pressure; historically, its topography and hydrology have been modified for the conditioning of fishing apparatus, known as "Encañizada". Currently, there are three open inlets (El Ventorrillo, La Torre, and El Charco) and one channel closed by a floodgate since 1980 (New); the Estacio channel was an ancient "Encañizada" drained for the construction of the port of Tomas Maestre in 1973 and is currently the deepest (5 m; Figure 1d); and the Marchamalo channel was an old artificial "Encañizada" that never worked due to its tendency to fill with sediments (Figure 1e) [76]. Currently, it is a navigable but has a shallow channel (1 m).

The Mar Menor lagoon was created by a sequence of consecutive oscillations of the sea level that occurred in the Quaternary. It is located at the bottom of a river basin bordered by mountains that enclose the Campo de Cartagena, a vast plain of 1350 km<sup>2</sup> with a low inclination to the SE. The contribution of continental water is made up of six watercourses, which are dry most of the year and contribute to the natural process of filling the Mar Menor lagoon [77]. In relation to this, the most outstanding geomorphological elements determining the dynamics of the lagoon are the sand barrier, continuous shore, islands and volcanic outcrops, and underwater area of the Mar Menor lagoon [77,78].

The erosion factors are mainly natural driving forces—winds, storms, waves, and a rise in sea level. The sand barrier was created by marine currents and the effect of the wind and waves. This effect is the reason why the Encañizada gola bottom is moving and changing [25,79]. The exchange between Mediterranean and lagoon waters can be affected because of changes in the inlet's geomorphology,

and these changes can affect the biological productivity, the species richness, and the complexity of the lagoon communities [80–82], and, therefore, affecting the environmental and economic balance of this area. Precise knowledge of the spatio-temporal evolution of the opening of this golas is very useful to improve the results of the hydrodynamic models that are being implemented for the Mar Menor lagoon and its connection with the Mediterranean Sea. According to the work of the Technical University of Cartagena (UPCT) [24], the "El Estacio" channel is the most important in the regulation of the water exchange of the Mar Menor lagoon, since it accounted for 60% of the total annual volume interchanged at that time (2010–2011), while the "Encañizadas" channel contributed 33% and the "Marchamalo" channel the remaining 7%. However, the Encañizadas can account for up to the 80% if the net balance is considered [25]. These percentages have significant seasonal fluctuations depending on the sand deposits caused by the storms in the area and the maintenance work that is carried out regularly on the golas. The water renewal rate of the volume of the Mar Menor lagoon has been calculated as 318 days [27], although this can be very variable depending on the weather conditions and freshwater inputs due to runoff or groundwater discharge [83].



**Figure 1.** Study area in the Mar Menor lagoon. (**a**) Sentinel 3 image of the ESA, (**b**) Sentinel 2 image of the Mar Menor on 1 January 2019, and Pleiades 1A images on 28 November 2019 of the (**c**) Encañizadas, (**d**) Estacio, and (**e**) Marchamalo channels.

## 2.2. In Situ Depth Data Points Obtained Using an Unmanned Surface Vehicle (USV)

The Murcia Institute of Agri-Food Research and Development (IMIDA) is working with two types of USV, developed with different aims: An Unmanned Surface Water Vehicle (USWV) to study bathymetry in shallow water, and a Remotely Operated Underwater Vehicle (ROV) to study water quality and the lagoon bottom.

These drones are based on ArduPilot or ArduPilotMega (APM). The latter is an autopilot of free code used in the control of several types of drones, such as unmanned helicopters, unmanned aircraft, and unmanned aquatic vehicles. In recent years, more and more unmanned automatic data acquisition equipment has been applied to ocean observation, and the USWV is one example. Due to its high

degree of automation and flexibility, it can be applied to the surveying and mapping of water depth. The Massachusetts Institute of Technology has developed a prototype ARTEMIS [84]; it is a small unmanned sounding vessel used in water-depth measurement. This idea has been developed into different types of small unmanned ships, boats, or catamarans [85–87], such as the IMIDA06 USV [88] used in this study (Figure 2).



**Figure 2.** The topo-bathymetric work (**a**) CN Los Alcázares from a Pleiades image 20 December 2016; (**b**) IMIDA06 USV; (**c**) USV data collection software; and (**d**) Bathymetric points recorded on 30 January 2018.

The drones were used to perform the bathymetric work in five coastal areas of the Mar Menor lagoon and in the three channels of communication with the Mediterranean Sea (Figure 3). Special attention was given to the Encañizadas channel, where seven bathymetries have been performed in the last 10 years by different Spanish organizations (Table 1): Instituto Hidrográfico de la Marina (IHM), Ministerio de Agricultura, Pesca y Alimentación (MAGRAMA), Instituto Español de Oceanografía (IEO), UPCT and IMIDA.

Organization	GPS	Photogrammetry	Sensor	MDS (m)	Date
IHM (MM)			Single probe		1967
IGN (out MM)		0.5 p/m <sup>2</sup>	Leica (ALS50)	2	4/10/2009
MAGRAMA (MM)		40 p/m <sup>2</sup>	Seabat 8124 (200 kHz)	1	4/10/2009
UPCT (En.)	Leica (TPS400)		ADCP (500 kHz)	1	5/08/2010
UPCT (Ma.)	Leica (TPS400)		River Surveyor M9	1	19/08/2011
UPCT (Es.)	Leica (TPS400)		River Surveyor M9	1	19/08/2011
IEO (En.)	Garmin (GPS72)	30 p/m <sup>2</sup>	Hondex (PS-7)	2	15/05/2014
IGN (out MM)		$0.5 \text{ p/m}^2$	Leica (ALS50)	2	4/10/2016
IEO (MM)	Trimble (SPS 751)		GeoSwath (500 kHz)	1	13/06/2017
IMIDA (En.)	Leica (1200)		Airmar (50 kHz)	1	12/03/2017
IMIDA (En.)	Leica (1200)		Airmar (50 kHz)	1	7/11/2017
IMIDA (En)	Leica (1200)		Airmar (50 kHz)	1	24/04/2018
IMIDA (En.)	Leica (1200)		ECT400S (450 kHz)	1	19/11/2019
IMIDA (Ma.)	Leica (1200)		ECT400S (450 kHz)	1	29/11/2019
IMIDA (Es.)	Leica (1200)		ECT400S (450 kHz)	1	30/11/2019

Table 1. Dates and characteristics of the field data used in the study of the channels.



**Figure 3.** The five data collection areas where the Unmanned Surface Vehicle (USV) was used to provide data for calibrations of the algorithms designed to obtain bathymetries with Pleiades satellite images: CN Los Alcazares; CN Los Urrutias; Gola Marchamalo; Gola Estacio; and Gola Encañizadas.

#### 2.3. Satellite Data

The images used in this work were provided by Airbus (Table 2) and correspond to the Pleiades satellites successfully launched on 16 December 2011. These are two optical satellites: Pleiades-1A and Pleiades-1B. Their orbits are differentiated by a 180° separation and they work together to provide daily images from anywhere on the planet. The Pleiades sensors have a nominal swath width of 20 km at nadir and two spectral configurations (Figure 4): a panchromatic configuration, with a spatial resolution of 0.5 m, and a multispectral configuration: visible (VI) and near infrared (NIR), both with a spatial resolution of 2 m.

The physical model used in this project (Appendix A) is based in the premise that the relationship between the water-leaving reflectance just above the surface  $R(0+,\lambda)$  and the water-leaving reflectance just below the surface  $R(0-,\lambda)$  is constant. The water-leaving reflectance just below the surface is a lineal combination reflectance from the water column and reflectance from the bottom. The water is clear; thus, the energy arrives from the bottom through the water column and the water column is homogenous [89]. The relationship between the water-leaving reflectance and the depth is

$$z = \mathbf{m} \cdot \frac{\ln R_{Blue}}{\ln R_{Green}} + \mathbf{n}$$

where m and n do not vary with depth (they are constant when the properties of the water column and the bottom are constant, too), but  $R_{Blue}$  and  $R_{Green}$  do vary [60,90], since they are the reflectances below the surface with bottom influence and the reflectance below the surface without bottom influence (that is, in deep water), respectively.



Figure 4. Spectral resolution of the Pleiades sensors (http://www.gstdubai.com/satelliteimagery/pleiades-1a.html).

In this study, 22 Pleiades images—taken between 15 October 2012 and 5 December 2019—were used (Table 2) (Figure A1).

Satellite	Sensor	PAN	VIS	IRC	Res (m)	Date
Pleiades 1A	HiRi	1	3	1	0.5/2	15/10/2012
Pleiades 1A	HiRi	1	3	1	0.5/2	1/07/2014
Pleiades 1B	HiRi	1	3	1	0.5/2	23/05/2014
Pleiades 1A	HiRi	1	3	1	0.5/2	12/10/2015
Pleiades 1A	HiRi	1	3	1	0.5/2	1/07/2016
Pleiades 1A	HiRi	1	3	1	0.5/2	15/10/2016
Pleiades 1A	HiRi	1	3	1	0.5/2	20/12/2016
Pleiades 1B	HiRi	1	3	1	0.5/2	21/03/2017
Pleiades 1B	HiRi	1	3	1	0.5/2	28/06/2017
Pleiades 1A	HiRi	1	3	1	0.5/2	9/11/2017
Pleiades 1A	HiRi	1	3	1	0.5/2	5/12/2017
Pleiades 1B	HiRi	1	3	1	0.5/2	15/04/2018
Pleiades 1B	HiRi	1	3	1	0.5/2	29/04/2018
Pleiades 1B	HiRi	1	3	1	0.5/2	18/05/2018
Pleiades 1A	HiRi	1	3	1	0.5/2	29/11/2018
Pleiades 1A	HiRi	1	3	1	0.5/2	4/09/2019
Pleiades 1A	HiRi	1	3	1	0.5/2	16/09/2019
Pleiades 1B	HiRi	1	3	1	0.5/2	24/09/2019
Pleiades 1B	HiRi	1	3	1	0.5/2	29/09/2019
Pleiades 1A	HiRi	1	3	1	0.5/2	28/11/2019
Pleiades 1A	HiRi	1	3	1	0.5/2	5/12/2019

Table 2. Dates and characteristics of the images used to track the shallow water (<3 m).

## 2.4. LiDAR Data

The airborne laser imaging detection and ranging (LiDAR) data correspond to the Spanish National Plan of Aerial Orthophotography (PNOA) of 2009 and 2016, of the Spanish National Geographic Institute (IGN). A model ALS50 (Leica Geosystems AG, Heerbrugg, Switzerland) was used, with a low point density (0.5 points/m<sup>2</sup>) but completely covering the Mar Menor lagoon. A digital surface model (DSM) was obtained by triangulation at a spatial resolution of 2 m. The LiDAR data (Figures 5

and 6) were processed with LAStools (Rapidlasso GmbH, Gilching, Germany) and ArcGiS 10.7 (ESRI, Redlands, CA, USA) software.



Figure 5. LiDAR data (2008) around the Mar Menor lagoon and MAGRAMA bathymetry in 2008.



**Figure 6.** LiDAR data (2016) around the Mar Menor lagoon and IEO bathymetry [14] in 2016 (no data for a depth <-3 m, blue line).

## 2.5. Echosounder Data (2008/2009 MAGRAMA and 2016/2017 IEO and IHM)

Until 2016, for the Mar Menor lagoon there was only a single bathymetry, performed by the IHM in 1967 and published in the nautical chart No. 471A; this bathymetry was obtained with a single-beam probe. Subsequently, in 2008/2009, the Ministry of Agriculture, Fisheries, and Food performed the ultrasound mapping of the coast of the Region of Murcia (Figure 5, Supplementary Materials S1), using a Seabat 8124 multi-beam sonar systems (Reson Ltd., Slangerup, Denmark).

In order to obtain a detailed bathymetry of the seabed of the Mar Menor lagoon, the IEO and IHM carried out a data collection campaign from 26 April to 13 June 2017. The objective of this was to conduct a bathymetric analysis of the interior of the Mar Menor lagoon (at depths > 3 m) and a characterization of the seabed (Figure 6) with the GeoSwath 500 Plus interferometric probe (Kongsberg Gruppen ASA, Kongsberg, Norway), with lateral scan sonar registers by IEO [43].

## 3. Methodology

#### 3.1. Ground-Based Data

During the field experiment, the IMIDA06 USV (Figure 2) used a sounding AIRMAR (Airmar Technology Corporation, Milford, CT, USA), a bi-frequency of 50/200 kHz, 600 W of power, and a GNSS receiver that synchronously collected water depth and position data with high accuracy. The GNSS RTK receiver was a GPS-GALILEO-GLONASS type and the horizontal positioning (for a Real Time Kinematic, RTK, model) was connected to the sounding data to get good depth and position data. Bathymetry is defined as the depth from the bottom to the zero value. In Spain (Royal Decree 1071/2007 Spanish law), the zero value is the sea surface in Alicante city (Spain). The measuring range is 0.5 m to 100 m and the accuracy is ±10 mm. The GNSS receiver was an Emlid Reach RS model (Emlid Ltd., Hong Kong, China), the horizontal positioning accuracy (in the RTK model) was about 10 mm, and the vertical positioning accuracy about 20 mm. The USV was developed by Inntelia (IPH Ltd., Huelva, Spain). Table 3 presents the specifications of USV components.

Part Name	Model/Number	Specifications	Function
Boat hull	Inntelia	Fiberglass	Navigation
Motors	BlueRobotics	Thruster-R1	Main actuator
Propellers	T-200	350 W	Propulsion
Microcontroller	Pixhawk 3.2.1	Open hardware	Navigation control
Battery (LiPo)	Tattu 22.2V	15C 4500 mAh	Power supply
Radio receiver	UHF	1 W (-30 dB)	Radio command
GNSS	EMLID	NMEA 0183	GPS antenna
Graupner	12 ch. PCM	2.4 GHz	Control station
Software	APM	3.2	Mission planner
Echo sounder	Airmar/DST 700	50/200 kHz	Echo sounder

Table 3. Specifications of the Unmanned Surface Vehicle (USV) components (IMIDA06).

In the bathymetries of the three golas of December 2019, a floating drone was used with an Echologger ECT400S (EofE Ultrasonics Ltd., Kyounggi-Do, Korea), acoustic frequency of 450 kHz, sampling frequency of 100 kHz, width of the conical beam 5° (±3 dB), precision of 1 mm, a receiver and GNSS RTK antenna (GPS, GALILEO, and GLONASS), as well as lithium polymer battery power. It was controlled manually with a UHF Radio Modem system with 1 W of power (+30 dBm) for connection and data transmission. The coordinates of the ground control points (GCP) were acquired through a traditional technique, by means of a Leica-Geosystems Station TPS1200 (Leica Geosystems AG, Hauptsitz, Heerbrugg, Switzerland).

## 3.2. Remote Sensing Data Processing

#### 3.2.1. Pre-Processing

Firstly, all images available were checked on OneAtlas (Airbus DS Geo, Toulouse, France); secondly, the decoding was done (Figure 7). Each type of image was downloaded applying a specific pre-processing (Table 2). The first step finished when all images had been pre-processed. In remote sensing imagery, atmospheric correction can remove the atmospheric influence (mainly caused by

atmospheric molecules and light scattering particles) and illumination factors. The target reflectance spectrum after correction will be similar to the ground-measured spectra [90–92].



Figure 7. Workflow diagram with Pleiades images.

## 3.2.2. Water Area

The water mask was generated using the normalized difference water index (NDWI), using green and infrared bands [93] in each reflectance image. From Figure 7, if NDWI is higher than 0, the pixel is water and if NDWI is less than 0, it is not water. Each water mask was applied to its reflectance image and all reflectance water images were obtained.

## 3.2.3. Water Depth Extraction Model

Reflectance water images near in time to the bathymetric data were chosen to get linear algorithms [90]. The operations between bands were performed using ENVI 4.8 software. The relationship between the raster data and the sounding data was processed with ArcMap 10.7. The m and n values of the model were obtained by EXCEL software, as well as the statistical study. The bathymetry derived from satellite data was compared with the bathymetric obtained from field data, and the normalized root mean square error (RMSE) was analyzed.

$$RMSE = \sqrt{\frac{\sum_{1}^{n} (Bathymetry \ Field - Bathymetry \ Satellite)^{2}}{n}}$$

## 3.2.4. Bathymetric Map and Water Amount Time Series Derived from Satellite Data

The bathymetry on each date was calculated using a water depth extraction model, and the water amount was obtained by multiplying the pixel size by the bathymetry.

## 3.2.5. Integration of Altimetry Data for the Mar Menor Lagoon

The bathymetric datasets were integrated in an ArcGiS 10.7 geodatabase containing the LiDAR data of two years (2008 and 2016). Using all the data, a DEM was generated with a spatial resolution of 2 m, to perform the volume calculation for each isobath (m) and thus to compute the volume (m<sup>3</sup>) of the Mar Menor lagoon. The volume between each isobath was calculated, starting at the deepest level

and finishing at the isobath corresponding to the maximum level of the Mar Menor lagoon. The final result corresponds to the height–volume relationship that is the current capacity curve. The applied methodology is represented in the flowchart in Figure 8.



Figure 8. Flowchart showing the integration of the LiDAR and bathymetric data of Mar Menor.

The Digital Terrain Model (DTM) was derived by triangulation from the spatial resolutions of 1 m (Figure 9, Appendix A). Considering the vertex of the National Geodetic Network closest to each reservoir, the supporting points were taken with bi-frequency GPS. The datasets were processed with ArcGiS 10.7 (ESRI, Redlands, CA, USA) software. Subsequently, the points cloud with known X, Y, Z coordinates was calculated in the official terrestrial space reference system (ETRS89) and in the official vertical spatial reference system in Spain (EVRS89), based on the Earth Gravitational Model (EGM2008) of the National Geospatial-Intelligence Agency (NGA) EGM Development Team. Subsequently, a Triangular Irregular Network (TIN) mesh was generated using ArcGiS 10.7 (ESRI, Redlands, CA, USA) and this was used to calculate the volume of the Mar Menor lagoon.

## 4. Results

The results show that multispectral optical imagery with very high spatial (i.e., 0.5 m) and temporal (in this study, 22 images) resolution is suitable for the tracking of water-level fluctuations of the Mar Menor golas and shallows at a fine scale. The results are a function of the different steps of the project; these are summarized as follows:

IMIDA06 USV has been developed by IPH and IMIDA and has provided a good bathymetry map (for 13 March 2017), with a spatial resolution of 50 cm and an error equal to  $\pm 3$  cm (Figure 9).





**Figure 9.** Water depth estimated on 13 March 2017, using an USV, Pleiades, and ground data (red and blue points correspond to GPS data, while yellow points correspond to USV data). The red line corresponds to the coast in 1999.

The NDWI values allow to evaluate the change of the water body and golas, as well as the annual movement of sediments and land area (Figure 10, Appendix A).



**Figure 10.** Map showing the differences among the dates 15 October 2012, 23 May 2014, and 21 March 2017. It was produced using Pleiades images of the Encañizada channel.

At the same time, the Water Depth Extraction Model [90] allows to obtain a bathymetric map from different satellite images and the parameters of this model (m and n). These parameters are not constant and depend on the water clarity (Figure 11).



**Figure 11.** Water Depth Extraction Model for two different water quality conditions: (**a**) 13 March 2017, less water transparency, and (**b**) 29 November 2019, more water transparency.

The relationship between both methods showed a good fit. On one hand, the slope of the regression is close to 1 (0.84 in the case of low transparency and 1.02 with clear waters), indicating the accuracy of the satellite image estimations and field data. The RMSE was analyzed at three different dates (Figure 12): one, five, and ten days of difference between the field data and the image. Firstly, for 2019, the RMSE<sub>2019</sub> is 0.179 m. Secondly, the RMSE<sub>2017</sub> is 0.186 m. Finally, in 2019, the image precedes the CoLs and the field data are subsequent to it; RMSE<sub>2019CoLs</sub> = 0.311 m. Besides, the slope of the line is reasonably close to 1: 0.92, 0.84, and 1.02. The RMSE values represent the accuracy of the measurements and dispersion of the data, while the slope is the true measure of accuracy.



**Figure 12.** Relationships between field and image data at three different dates: (**a**) 21 March 2017; (**b**) 19 November 2019; and (**c**) 29 November 2019.

The bathymetric map obtained from the Pleaides satellite images of 11 November 2017 (Figure 13) allowed to calculate the exchange water between Mar Menor and the Mediterranean Sea. The exchanged volume increased the water level at Mar Menor due to the low atmospheric pressures registered from 15 October 2012 to 28 November 2019 (Figure 14).



**Figure 13.** Water depth determined using the USV and the Pleiades images of 11 November 2017, as well as ground data (red and blue points correspond to GPS data, while yellow points to USV data). The red line corresponds to the coast in 1999.



**Figure 14.** Water exchange section  $(m^2)$  in the Gola de las Encañizadas on different dates (values in red and blue) and for different water levels  $(m^{-2})$  induced by low atmospheric pressures (values in black).

The bathymetry, the barometric pressure, and the wind of the Mar Menor are responsible for the currents and therefore for the distribution of sediments in the lagoon. Therefore, an updated and high-resolution bathymetric characterization is especially important in order to know its evolution and its dynamics at different spatio-temporal scales. In addition, an accurate bathymetry is decisive to know the total volume of water in the lagoon and for hydrodynamic calculations, relevant to the determination of the water exchanges between Mar Menor and the Mediterranean Sea, through the three active golas. Mar Menor is a coastal lagoon with a high degree of hydrodynamic confinement and, as such, is the final recipient of all the transport processes that take place in the large (1350 km<sup>2</sup>) watershed of Campo de Cartagena. The soil uses in the coastal area has undergone significant

transformations in the last decade, which have contributed to the acceleration of these transport processes and, consequently, the bottom of the lagoon have experienced an increasing accumulation of



Figure 15. Bathymetric evolution of Mar Menor between 2008 and 2009 and between 2016 and 2017.

Isobath (m)	Surface Area (2008) (km <sup>2</sup> )	Vol (2008) (10 <sup>9</sup> m <sup>3</sup> )	VT (2008) (10 <sup>9</sup> m <sup>3</sup> )	Surface Area (2016) (km <sup>2</sup> )	Vol (2016) (10 <sup>9</sup> m <sup>3</sup> )	VT (2016) (10 <sup>9</sup> m <sup>3</sup> )
-7	9.720	0.003	0.003	0.393	0.000	0.000
-6	39.312	0.021	0.024	45.498	0.018	0.018
-5	82.262	0.064	0.088	84.780	0.068	0.086
-4	99.449	0.092	0.180	102.564	0.095	0.181
-3	110.138	0.105	0.285	112.532	0.108	0.288
-2	117.504	0.114	0.399	119.432	0.116	0.404
-1	126.044	0.121	0.521	126.335	0.123	0.527
0	138.127	0.132	0.653	135.202	0.130	0.657

**Table 4.** Estimates of the surface areas and volumes per isobath, as well as the total cumulative volume of Mar Menor in 2008 and 2016.

VT = Total cumulative volume; Vol = Volume per isobath.

Organization/Bibliography	Area	Volume (10 <sup>9</sup> m <sup>3</sup> )	Section (m <sup>2</sup> )	Depth (m)	Date
UMU [49]	Mar Menor	0.580			1980–1988
IEO [58]	Mar Menor	0.580	-	-	1988
UPCT [4]	Mar Menor	0.725	-	-	2011
UMU/ISMAR [7]	Mar Menor	0.600			2009-2010
IMIDA	Mar Menor	0.653	-	-	2009
IMIDA	Mar Menor	0.657	-	-	2016
IEO [14]	Mar Menor	0.654	-	-	2017
UPCT [4]	Encañizada	-	207	-0.40	5/08/2010
UPCT [4]	Estacio	-	208	-5.20	13/06/2011
UPCT [4]	Marchamalo	-	21	-1.12	13/06/2011
IEO	Encañizada	-	145	-0.35	15/05/2014
IMIDA	Encañizada	-	11	-0.10	21/03/2017
IMIDA	Encañizada	-	54	-0.20	28/11/2018
IMIDA	Encañizada	-	120	-0.44	19/11/2019
IMIDA	Marchamalo	-	21	-0.78	29/11/2019
IMIDA	Estacio	-	204	-5.10	30/11/2019

Table 5. The volume of the Mar Menor and characteristics of the exchange channels.

Mar Menor underwent a generalized silting up (an average of +18 cm) between 2008 and 2016, mainly the result of sediment contributions due to the very frequent extreme precipitation events [47]. Very intense rainfall events with a high return period occurred in the Campo of Cartagena in the last decade. As examples, the event of 27–28 September 2009 corresponded to a 200-year return period, where the CA12 rain gauge (located in La Palma) recorded 268 mm in 30 h. Then, during the event registered between 17 and 19 December 2016 in Campo of Cartagena (with a 500-year return period), the rainfall exceeded 50 mm in one hour in the Torre Pacheco (TP42) and San Javier (TP22) rain gauges [47]. Finally, the CoLs event named *dana* of 12–15 September 2019 generated maximum hourly rainfall intensities of 70.4 mm and 60.6 mm at the Torre Pacheco and Pozo Estrecho rain gauges, respectively. The maximum daily rainfall of 217.8 mm recorded by the Torre Pacheco rain gauge (12 September 2019 12:00 h to 13 September 2019 12:00 h) corresponded to a 500-year return period and generated severe flash floods in the Campo of Cartagena area. The estimations of return periods were based on [48] by applying the SQRT-ETmax cumulative distribution function.

#### 5. Discussion

The deployment of an IMIDA06 USV is an example of how the free disposition of research findings can stimulate scientific advances, with the contribution of anyone, anywhere in the world. The images and their distribution are low-cost, which has allowed the development of new prototypes that, in turn, will be surpassed by novel work. Now it is necessary to take advantage of the developments from the recent years and make them useful to the community. In this sense, the IMIDA06 USV, as a source of field data, has allowed the development of a satellite image tracking system that is able to provide valuable data that helps solve environmental, social, and economic problems.

The Mar Menor lagoon and its area of influence is a complex and dynamic environmental, economic, and social system [94]; therefore, distinct types of information are necessary for its good management. The use of 3D numerical models represents a solution to obtain data and predict situations in the short term. However, model inputs condition the results and, therefore, special attention should be paid to their accuracy and precision; e.g., the initial conditionals (Xo, Yo, Zo) have to be precise enough. Remote sensing provides information on large coastal water bodies, such as the Mar Menor lagoon, besides providing good precision for Xo and Yo. However, knowing the depth, even in the shallowest waters, requires transparent waters that allow the sunlight to reach and be reflected by the bottom.

The initial idea was to use an unmanned aerial vehicle or unscrewed aerial vehicle (UAV), but in this area the use of a UAV is prohibited due to military activities. However, it is possible to use high spatial resolution satellite images to obtain a real-time bathymetry with an acceptable error that can be applied in different fields, such as scientific research, technology, and management.

At the same time, the idea was to select several places where very high spatial resolution bathymetric data could be obtained, using an aquatic drone, and then relate this data to the high spatial resolution images from satellites.

During this work, different water situations were induced by CoLs. This allowed producing distinct equations according to the transparency of the water, and a good adjustment with an acceptable RMSE was obtained. The methodology was improved by introducing a new step, since after calculating the relationship between LnBlue and LnGreen, the analysis of the ratio values is necessary for choosing the appropriate algorithm.

The changes in the depth of the lagoon depend mainly on the sediment loads mainly transported by runoff during flash-flood events, which deposit soil and other particles on the bottom. This contribution can reach 1863 t yr<sup>-1</sup> [95]. Contributions from the Mediterranean or exchanges through the inlets have not been studied in depth, but there is evidence that they can be significant throughout the recent history of the Mar Menor lagoon, especially during extreme storm events [96]. Sedimentation rates have changed during the evolution of the Mar Menor lagoon linked to human uses in the drainage basin, and have increased with time, from 30 mm/century before the Middle Age to more than 30 cm/century in the 20th century, particularly due to deforestation and agricultural practices [97]. To the best of our knowledge, no work had been done that allowed to know the recent changes and the spatial variability of the sedimentation rates in the lagoon.

In the present work, the lowest RMSE value obtained was 0.179 m and increased over time; in 2017 it was 0.186 m and after the CoLs of 2019 there was a random movement and deposition of sand according to the RMSE value.

In the Encañizada gola the deposition of sand meant that the channel, through which water enters and leaves the Mar Menor lagoon, has gradually closed over time (Figure 16). In the Marchamalo and Estacio golas the deposition of sand meant that the channels were shallower compared to the last bathymetry carried out by the UPCT, in 2011 (Figure 17).



Figure 16. The monitoring of the deposition of sediment in the Encañizada (1997–2019).



**Figure 17.** The monitoring of the sediment deposit: (**a**,**c**,**e**) El Estacio channel; (**b**,**d**,**f**) Marchamalo channel; and (**g**,**h**) Encañizada channel. The yellow areas correspond to depths less than 0.5 m. The rectangle red represents the zone of comparison between the bathymetries in the channels.

It was demonstrated that multispectral images are enough to obtain a precise bathymetry in shallow water. In this case, the shallow water is defined by the fact that the sunlight reaches the bottom of the lagoon. It was possible because, as shown in the bathymetric map, the maximum depth in the Mar Menor lagoon is 7 m; however, a depth of 4 m is presented in the photic zone, and this is the depth of the lagoon that it has been determined using optical sensors from different platforms: satellite, aircraft, or UAV.

The hyper-high spatial resolution of the sensor allows to obtain enough spatial detail, while the high resolution provides a temporal vision to observe details of the influence of the tides. Since the acquisition of the images is always at the same time and in the same place, the variation in the amount of water can only be due to movement of sand or movement of water.

The generated maps provide much information for the management of such a complicated area, but the most important thing is not only the information itself, but that it can be generated in quasi real time, allowing stakeholders to have accurate data in a short time for decision making. The depth model allows the use of images, obtained by aerial drones, of very high spatial resolution (18 cm) or the ones that the stakeholder prefers, since the resolution depends on the drone flight height; it can even reach 3 cm. As the gola is a small area, the stakeholder could have the bathymetric map a few hours after the flight. The low cost of obtaining spatial information from an area with great dynamism between water and land will allow the development of predictive models of soil movements and of increases or decreases in the amount of water in the gola. These models will provide the stakeholders enough information to be able to take decisions about the area with the aim of enhancing its conservation and the use of its resources. If the studies had been limited to the Encañizadas gola alone (Supplementary Materials S2), the sight of the importance of the other golas and the movement of water through them, between the Mediterranean Sea and the Mar Menor lagoon, would not have been known. Therefore, increasing knowledge of the movement of the sand in all the golas is fundamental for the area, in both economic and environmental terms (Figures A1 and A3).

#### 6. Conclusions

Once again, remote sensing was shown to be a tool, perhaps the most effective and cheapest one, for obtaining precise data. It may be used in 3D numerical models and in the management of the aquatic environment. It is necessary to remember that the Mar Menor lagoon is a dynamic ecosystem and the quantities of water and sand in each location within it are changing over time. Thus, although the bathymetric map generated throughout this work is the most up-to-date, it will be necessary to validate it from time to time due to the dynamic nature of the lagoon.

The results obtained show that the sedimentation processes occurring in the golas have produced a considerable reduction in their storage capacity, according the bathymetry. Although no significant changes have been detected in the current water renewal time in the lagoon compared to those estimated in the 1980s [25,27,83], the salinity has shown a tendency to approach the lower values of Mediterranean salinity rather than an increase [26]. It would be important to know how the observed changes can affect the capacity for water exchange between the Mar Menor lagoon and the Mediterranean Sea. In addition, climate change will negatively affect the frequency and intensity of natural hazards, such as extreme rainfall events in the area. Therefore, more frequent and severe flash flood events are expected for the contributing basins of the Mar Menor lagoon, which would affect their hydrodynamic [29] and sedimentation rates.

The bathymetries obtained from high-resolution satellite images, such as Pleiades, are demonstrated as very useful tools for complementing the bathymetries obtained with echo sounders in areas of shallow (<2 m) and clear water (Figure A2). Consequently, the precision of the calculation of the water volume of the Mar Menor lagoon has improved over time with the incorporation of new, much more precise technologies, such as interferometric echo sounders.

**Supplementary Materials:** These are available online at https://geoportal.imida.es/agua, (S1), together with a video on YouTube (S2): https://youtu.be/TdIUD8-K2Zw.

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## Appendix A

The physical model used in this project is presented as follows:

$$R(0-\lambda) = R_{water}(0-\lambda) + R_{bottom}(0-\lambda).$$

The influence of the bottom is determined by a simple model, in which the bottom optical properties and the water optical properties are included [49,50]:

$$R_{bottom}(0-,\lambda) = [R_{bottom}(0-,\lambda) - R_{water}(0-,\lambda)]\exp(-2Zk_d)$$

where Z is the depth and  $K_d$  the downwelling diffuse attenuation coefficient.

$$R^{-}(z, \lambda) = R^{-}_{w}(\lambda) + [R_{b}(\lambda) - R^{-}_{w}(\lambda)] \exp[-2K_{d}(\lambda)z]$$

$$[R^{-}(z,\lambda) - R^{-}_{w}(\lambda)] = [R_{b}(\lambda) - R^{-}_{w}(\lambda)] \exp[-2K_{d}(\lambda)z]$$

$$R^{-}(z, \lambda) \left[1 - \frac{R - w(\lambda)}{R - (z, \lambda)}\right] = R_{b}(\lambda) \left[1 - \frac{R - w(\lambda)}{R_{b}(\lambda)}\right] \exp\left[-2K_{d}(\lambda)z\right]$$

 $\left[1 - \frac{R-w(\lambda)}{R-(z,\lambda)}\right] \sim 1; \left[1 - \frac{R-w(\lambda)}{R_b(\lambda)}\right] \sim 1$  because the water is clear, thus  $R^-_w(\lambda)$  is very low [89]:

$$R^{-}(z, \lambda) = R_{b}(\lambda) \cdot \exp[-2K_{d}(\lambda)z]$$
$$\ln[R^{-}(z, \lambda)] = \ln [R_{b}(\lambda) \exp[-2K_{d}(\lambda)z]$$
$$\ln[R^{-}(z, \lambda)] = \ln [R_{b}(\lambda)] - 2K_{d}(\lambda)z$$

For the i band it is  $\ln [\mathbb{R}^{-}(z, \lambda i)] = \ln [R_{b}(\lambda i)] - 2K_{d}(\lambda i)z$ For the j band it is  $\ln [\mathbb{R}^{-}(z, \lambda j)] = \ln [R_{b}(\lambda j)] - 2K_{d}(\lambda j)z$ The ratio between  $\ln [\mathbb{R}^{-}(z, \lambda i)]$  and  $\ln [\mathbb{R}^{-}(z, \lambda j)]$  is

$$\left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] = \frac{\ln[R_b(\lambda i)] - 2zK_d(\lambda i)}{\ln[R_b(\lambda j)] - 2zK_d(\lambda j)}$$

 $K_d(\lambda i) \sim K_d(\lambda j) \sim K_d$  because the water is clear [89]:

$$\begin{bmatrix} \frac{\ln[\mathbf{R} - (z, \lambda \mathbf{i})]}{\ln[\mathbf{R} - (z, \lambda \mathbf{j})]} \end{bmatrix} = \frac{\ln[R_b(\lambda \mathbf{i})] - 2zK_d}{\ln[R_b(\lambda \mathbf{j})] - 2zK_d}$$
$$\begin{bmatrix} \frac{\ln[\mathbf{R} - (z, \lambda \mathbf{i})]}{\ln[\mathbf{R} - (z, \lambda \mathbf{j})]} \end{bmatrix} \begin{bmatrix} \ln[R_b(\lambda \mathbf{j})] - 2zK_d \end{bmatrix} = \ln[R_b(\lambda \mathbf{i})] - 2zK_d$$
$$\begin{bmatrix} \frac{\ln[\mathbf{R} - (z, \lambda \mathbf{i})]}{\ln[\mathbf{R} - (z, \lambda \mathbf{j})]} \end{bmatrix} \cdot \begin{bmatrix} \ln[R_b(\lambda \mathbf{j})] - \begin{bmatrix} \frac{\ln[\mathbf{R} - (z, \lambda \mathbf{i})]}{\ln[\mathbf{R} - (z, \lambda \mathbf{j})]} \end{bmatrix} 2zK_d = \ln[R_b(\lambda \mathbf{i})] - 2zK_d$$
$$\begin{bmatrix} \frac{\ln[\mathbf{R} - (z, \lambda \mathbf{i})]}{\ln[\mathbf{R} - (z, \lambda \mathbf{j})]} \end{bmatrix} \cdot \begin{bmatrix} \ln[R_b(\lambda \mathbf{j})] - \begin{bmatrix} \frac{\ln[\mathbf{R} - (z, \lambda \mathbf{i})]}{\ln[\mathbf{R} - (z, \lambda \mathbf{j})]} \end{bmatrix} 2zK_d$$

$$\begin{split} \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= -2zK_d + \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] 2zK_d \\ \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= -2zK_d + \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] 2zK_d \\ \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= 2zK_d [-1 + \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right]\right] \\ \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= 2zK_d [\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \left[\frac{L_a(X, \lambda i)}{L_a(X, \lambda i)}\right] \\ \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= 2z \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot K_d \\ \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= 2z \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot K_d \\ \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= 2z \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot K_d \\ \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= 2z \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot K_d \\ \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= 2z \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot K_d \\ \frac{\left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= 2z \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot K_d \\ \frac{\left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= 2z \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot K_d \\ \frac{\left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \ln[R_b(\lambda i)] &= 2z \left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot K_d \\ \frac{\left[\frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}\right] \cdot \ln[R_b(\lambda j)] - \frac{\ln[R - (z, \lambda i)]}{\ln[R - (z, \lambda j)]}} = 2zK_d \\ \frac{\ln[R - (z, \lambda i)]}{2K_d} = \frac{\ln[R - (z, \lambda i)]}{2K_d} = z \\ z = m \cdot \frac{\ln Rj}{\ln R} + n \end{aligned}$$

where m and n do not vary with depth (they are constant when the properties of the water column and the bottom are constant, too), but Ri (Green band) and *R*j (Blue band) do vary.

## Appendix B



**Figure A1.** Pleiades images time series of the study: (a) First image available (15 October 2012); (b) The sediment accumulation process is evident; (c) During the first cut-off lows (CoLs); (d) Sediment accumulation process after the first CoLs; (e) Start of depth restoration; (f) Final of depth restoration; (g) During the second CoLs; (h) Sediment accumulation process after the second CoLs; (i) During the third CoLs (5 December 2019).



Figure A2. Bathymetry time series using Pleiades images between 15 October 2012 and 28 November 2019.



**Figure A3.** Start of depth restoration in the Encañizada: (**a**) Pleiades image 28 June 2017; and (**b**) aerial image 15 August 2017, after the completion of maintenance work.

## References

- De Pascalis, F.; Ghezzo, M.; Umgiesser, G.; De Serio, F.; Mossa, M. Use of Shyfem open source hydrodynamic model for time scales analysis in a semi-enclosed basin. In Proceedings of the 2016 IEEE Workshop on Environmental, Energy, and Structural Monitoring Systems (EESMS), Bari, Italy, 13–14 June 2016. [CrossRef]
- 2. Wild-Allen, K.; Andrewartha, J. Connectivity between estuaries influences nutrient transport, cycling and water quality. *Mar. Chem.* 2016, *185*, 12–26. [CrossRef]
- 3. Yin, Y.; Karunarathna, H.; Reeve, D.E. Numerical modelling of hydrodynamic and morphodynamic response of a meso-tidal estuary inlet to the impacts of global climate variabilities. *Mar. Geol.* **2019**, 407, 229–247. [CrossRef]
- 4. Lesser, G.R.; Roelvink, J.A.; van Kester, J.; Stelling, G.S. Development and validation of a three-dimensional morphological model. *Coastal Eng.* **2004**, *51*, 883–915. [CrossRef]
- Lazure, P.; Garnier, V.; Dumas, F.; Herry, C.; Chiet, M. Development of a hydrodynamic model of the Bay of Biscay. Validation of hydrology. *Cont. Shelf Res.* 2009, 29, 985–997. [CrossRef]
- 6. Sedigh, M.; Tomlinson, R.; Cartwright, N.; Etemad-Shahidi, A. Numerical modelling of the Gold Coast Seaway area hydrodynamics and littoral drift. *Ocean Eng.* **2016**, *121*, 47–61. [CrossRef]
- Gunn, K.; Stock-Williams, C. On validating numerical hydrodynamic models of complex tidal flow. *Int. J. Mar. Energy* 2013, *3*, e82–e97. [CrossRef]
- 8. Garneau, C.; Duchesne, S.; St-Hilaire, A. Comparison of modelling approaches to estimate trapping efficiency of sedimentation basins on peatlands used for peat extraction. *Ecol. Eng.* **2019**, *133*, 60–68. [CrossRef]
- 9. De Pablo, H.; Sobrinho, J.; Garcia, M.; Campuzano, F.; Juliano, M.; Neves, R. Validation of the 3D-MOHID hydrodynamic model for the Tagus coastal area. *Water* **2019**, *11*, 1713. [CrossRef]
- 10. Ma, G.; Shi, F.; Kirby, J.T. Shock-capturing non-hydrostatic model for fully dispersive surface wave processes. *Ocean Modell.* **2012**, *43*–44, 22–35. [CrossRef]
- 11. Kirby, J.T.; Dalrymple, R.A. A parabolic equation for the combined refraction-diffraction of stokes waves by mildly varying topography. *J. Fluid Mech.* **1983**, *136*, 453–466. [CrossRef]
- 12. Penven, P.; Marchesiello, P.; Debreu, L.; Lefevre, L. Software tools for pre- and post-processing of oceanic regional simulations. *Environ. Modell. Softw.* **2008**, *23*, 660–662. [CrossRef]
- Grifoll, M.; Del Campo, A.; Espino, M.; Mader, J.; González, M.; Borja, Á. Water renewal and risk assessment of water pollution in semi-enclosed domains: Application to Bilbao Harbour (Bay of Biscay). *J. Mar. Syst.* 2013, 109–110, S241–S251. [CrossRef]

- 14. Guerreiro, M.; Fortunato, A.B.; Freire, P.; Rilo, A.; Taborda, R.; Freitas, M.C.; Andrade, C.; Silva, T.; Rodrigues, M.; Bertin, X.; et al. Evolution of the hydrodynamics of the Tagus estuary (Portugal) in the 21st century. *J. Integr. Coast. Z. Manag.* **2014**, *15*, 65–80. [CrossRef]
- 15. Zhang, Y.; Baptista, A.M. SELFE: A semi-implicit Eulerian–Lagrangian finite-element model for cross-scale ocean circulation. *Ocean Modell.* **2008**, *21*, 71–96. [CrossRef]
- Zhang, Y.; Ye, F.; Stanev, E.V.; Grashorn, S. Seamless cross-scale modelling with SCHISM. *Ocean Modell.* 2016, 102, 64–81. [CrossRef]
- Ye, F.; Zhang, Y.; Friedrichs, M.; Wang, H.V.; Irby, I.; Shen, J.; Wang, Z. A 3D, cross-scale, baroclinic model with implicit vertical transport for the Upper Chesapeake Bay and its tributaries. *Ocean Modell.* 2016, 107, 82–96. [CrossRef]
- Umgiesser, G.; Ferrarin, C.; Cucco, A.; De Pascalis, F.; Bellafiore, D.; Ghezzo, M.; Bajo, M. Comparative hydrodynamics of 10 Mediterranean lagoons by means of numerical modeling. *J. Geophys. Res. Oceans* 2014, 119, 2212–2226. [CrossRef]
- Cucco, A.; Quattrocchi, G.; Olita, A.; Fazioli, L.; Ribotti, A.; Sinerchia, M.; Tedesco, C.; Sorgente, R. Hydrodynamic modelling of coastal seas: The role of tidal dynamics in the Messina Strait, Western Mediterranean Sea. *Nat. Hazards Earth Syst. Sci.* 2016, *16*, 1553–1569. [CrossRef]
- 20. Van Dongeren, A.; Svendsen, I. Absorbing-generating boundary condition for shallow water models. *J. Waterw. Port Coast.* **1997**, 123, 303–313. [CrossRef]
- 21. Booij, N.; Ris, R.; Holthuijsen, L.H. A third-generation wave model for coastal regions: 1. Model description and validation. *J. Geophys. Res.* **1999**, *104*, 7649–7666. [CrossRef]
- Gaeta, M.G.; Samaras, A.G.; Federico, I.; Archetti, R.; Maicu, F.; Lorenzetti, G. A coupled wave-3-D hydrodynamics model of the Taranto Sea (Italy): A multiple-nesting approach. *Nat. Hazards Earth Syst. Sci.* 2016, *16*, 2071–2083. [CrossRef]
- 23. Chubarenko, B.; Koutitonsky, V.G.; Neves, R.; Umgiesser, G. Modeling concepts. In *Coastal Lagoons: Ecosystem Processes and Modeling for Sustainable Use and Development*; Gönenc, I.E., Wolflin, J.P., Eds.; CRC Press: Boca Ratón, FL, USA, 2005; pp. 231–306.
- 24. López-Castejón, F. Caracterización de la Hidrodinámica Del Mar Menor y Los Flujos de Intercambio Con El MEDITERRÁNEO Mediante Datos in Situ y Modelado Numérico. Ph.D. Thesis, Universidad Politécnica de Cartagena, Cartagena, Spain, 2017.
- 25. Garcia-Oliva, M.; Perez-Ruzafa, A.; Umgiesser, G.; McKiver, W.; Ghezzo, M.; De Pascalis, F.; Marcos, C. Assessing the hydrodynamic response of the mar menor lagoon to dredging inlets interventions through numerical modelling. *Water* **2018**, *10*, 959. [CrossRef]
- García-Oliva, M.; Marcosa, C.; Umgiesserb, G.; McKiverb, W.; Ghezzo, M.; De Pascalisb, F.; Pérez-Ruzafa, A. Modelling the impact of dredging inlets on the salinity and temperature regimes in coastal lagoons. *Ocean Coast. Manag.* 2019, 180, 104913. [CrossRef]
- 27. Ghezzo, M.; De Pascalis, F.; Umgiesser, G.; Zemlys, P.; Sigovini, M.; Marcos, C.; Pérez-Ruzafa, A. Connectivity in three European coastal lagoons. *Estuaries Coasts* **2015**, *38*, 1764–1781. [CrossRef]
- Pérez-Ruzafa, A.; Ghezzo, M.; De Pascalis, F.; Quispe, J.; Hernández-García, R.; Muñoz, I.; Vergara-Chen, C.; Umgiesser, G.; Marcos, C. Connectivity between coastal lagoons and sea: Reciprocal effects on assemblages' structure and consequences for management. *Estuar. Coast. Shelf Sci.* 2019, 216, 171–186. [CrossRef]
- De Pascalis, F.; Pérez-Ruzafa, A.; Gilabert, J.; Marcos, C.; Umgiesser, G. Climate change response of the Mar Menor coastal lagoon (Spain) using a hydrodynamic finite element model. *Estuar. Coast. Shelf Sci.* 2012, 114, 118–129. [CrossRef]
- 30. European Union. Directive 2007/60/EC of the European Parliament and of the Council of 23 October 2007 on the assessment and management of flood risks. *Off. J. Eur. Parliam.* **2007**, 2455, 27–34.
- Ibarra, A.D. Análisis y Evolución de las Playas de la Región de Murcia (1956–2013). Ph.D. Thesis, Universidad de Murcia, Murcia, Spain, 2015. Available online: http://hdl.handle.net/10201/51348d (accessed on 13 December 2019).
- 32. Lyard, F.; Genco, M.L. Optimization methods for bathymetry and open boundary conditions in a finite element model of ocean tides. *J. Comput. Phys.* **1994**, *114*, 234–256. [CrossRef]
- 33. Chen, X. A cartesian method for fitting the bathymetry and tracking the dynamic position of the shoreline in a three-dimensional, hydrodynamic model. *J. Comput. Phys.* **2004**, 200, 749–768. [CrossRef]

- 34. Du Bois, P.B. Automatic calculation of bathymetry for coastal hydrodynamic models. *Comput. Geosci.* 2011, 37, 1303–1310. [CrossRef]
- 35. Ardani, S.; Kaihatu, J.M. Optimization of bathymetry estimates for nearshore hydrodynamic models using bayesian methods. *J. Waterw. Port Coast.* **2018**, *144*, 04018024. [CrossRef]
- Ye, F.; Zhang, Y.J.; Wang, H.V.; Friedrichs, M.A.M.; Irby, I.D.; Alteljevich, E.; Valle-Levinson, A.; Wang, Z.; Huang, H.; Shen, J.; et al. A 3D unstructured-grid model for Chesapeake Bay: Importance of bathymetry. *Ocean Modell.* 2018, 127, 16–39. [CrossRef]
- 37. Jo, Y.-H.; Sha, J.; Kwon, J.-I.; Jun, K.; Park, J. Mapping bathymetry based on waterlines observed from low altitude Helikite remote sensing platform. *Acta Oceanol. Sin.* **2015**, *34*, 110–116. [CrossRef]
- 38. Chénier, R.; Faucher, M.-A.; Ahola, R. Satellite-derived bathymetry for improving Canadian hydrographic service charts. *ISPRS Int. J. Geo Inf.* **2018**, *7*, 306. [CrossRef]
- 39. Legleiter, C.J.; Harrison, L.R. Remote sensing of river bathymetry: Evaluating a range of sensors, platforms, and algorithms on the upper Sacramento River, California, USA. *Water Resour. Res.* **2018**, *55*, 3. [CrossRef]
- 40. Brêda, J.P.L.F.; Paiva, R.C.D.; Bravo, J.M.; Passaia, O.A.; Moreira, D.M. Assimilation of satellite altimetry data for effective river bathymetry. *Water Resour. Res.* **2019**, *55*, 9. [CrossRef]
- 41. Sagawa, T.; Yamashita, Y.; Okumura, T.; Yamanokuchi, T. Satellite derived bathymetry using machine learning and multi-temporal satellite images. *Remote Sens.* **2019**, *11*, 1155. [CrossRef]
- 42. Favre, A.; Hewitson, B.; Lennard, C.; Cerezo-Mota, R.; Tadross, M. Cut-off Lows in the South Africa region and their contribution to precipitation. *Clim. Dyn.* **2013**, *41*, 2331. [CrossRef]
- 43. IEO. Estudio del Fondo Marino de la Laguna Costera del Mar Menor; IEO: Madrid, Spain, 2018.
- 44. Erena, M.; Domínguez, J.A.; Aguado-Giménez, F.; Soria, J.; García-Galiano, S. Monitoring coastal lagoon water quality through remote sensing: The mar menor as a case study. *Water* **2019**, *11*, 1468. [CrossRef]
- 45. Caballero, I.; Ruiz, J.; Navarro, G. Sentinel-2 satellites provide near-real time evaluation of catastrophic floods in the west mediterranean. *Water* **2019**, *11*, 2499. [CrossRef]
- 46. Austin, R.W. The remote sensing of spectral radiance from below the ocean surface. In *Optical Aspects of Oceanography*; Jerlov, N.G., Steemann-Nielsen, E., Eds.; Academic Press: London, UK, 1974; pp. 317–344.
- 47. Cánovas-García, F.; García-Galiano, S.; Alonso-Sarría, F. Assessment of satellite and radar quantitative precipitation estimates for real time monitoring of meteorological extremes over the Southeast of the Iberian Peninsula. *Remote Sens.* **2018**, *10*, 1023. [CrossRef]
- 48. De Fomento, M. *Dirección General de Carreteras. Máximas Lluvias Diarias en la España Peninsular;* Ministerio de Fomento: Madrid, Spain, 1999.
- 49. Kirk, J.T.O. *Light and Photosynthesis in Aquatic Ecosystem*, 1st ed.; Cambridge University Press: New York, NY, USA, 1994.
- 50. Kutser, T. Quantitative detection of chlorophyll in cyanobacterial blooms by satellite remote sensing. *Limnol. Oceanogr.* **2004**, *49*, 2179–2189. [CrossRef]
- 51. Philpot, W.D. Radiative transfer in stratified waters: A single scattering approximation for irradiance. *Appl. Optics* **1987**, *26*, 4123–4132. [CrossRef] [PubMed]
- 52. Polcyn, F.C.; Brown, W.L.; Sattinger, I.J. *The Measurement of Water Depth by Remote Sensing Techniques*; Universidad de Michigan: Ann Arbor, MI, USA, 1970.
- 53. Hasell, P.G. *Michigan Experimental Multispectral Mapping System a Description of the M7 Airborne Sensor and ITS;* NASA: Washington, DC, USA, 1974.
- 54. NOAA. Hydrologic Optics. Vol. V. Properties. Colo Pacific Marine Environmental Lab. Honolulu; NOAA: Silver Spring, MD, USA, 1976.
- 55. Lyzenga, D.R. Passive remote sensing techniques for mapping water depth and bottom features. *Appl. Opt.* **1978**, *17*, 379. [CrossRef] [PubMed]
- 56. Lyzenga, D.R. Shallow-water bathymetry using combined lidar and passive multispectral scanner data. *Int. J. Remote Sens.* **1985**, *6*, 115–125. [CrossRef]
- 57. Martitorena, S.; Morel, A.; Gentili, B. Diffuse reflectance of oceanic shallow waters: Influence of water depth and bottom albedo. *Limnol. Oceanogr.* **1994**, *39*, 1689–1703. [CrossRef]
- 58. Bierwirth, P.N.; Lee, T.J.; Burne, R.V. Shallow seam floor reflectance and water depth derived by unmixing multispectral imagery. *Photogramm. Eng. Remote Sens.* **1993**, *59*, 331–338.

- Dierssen, H.M.; Zimmerman, R.C.; Leathers, R.A.; Downes, T.V.; Davis, C.O. Ocean color remote sensing of seagrass and bathymetry in the bahamas banks by high resolution airborne imagery. *Limnol. Oceanogr.* 2003, 48, 444–455. [CrossRef]
- 60. Stumpf, R.P.; Holderied, K.; Sinclair, M. Determination of water depth with high-resolution satellite imagery over variable bottom types. *Limnol. Oceanogr.* 2003, *48*, 547–556. [CrossRef]
- 61. Olayinka, D.N.; Okolie, C.J. Bathymetric mapping of nigerian coastal waters from optical imagery. *J. Eng. Res. Kuwait* 2017, *22*, 112–132.
- 62. Setiawan, K.T.; Adawiah, S.W.; Marini, Y.; Winarso, G. Bathymetry data extraction analysis using Landsat 8 Data. *Int. J. Remote Sens.* **2016**, *13*, 79–86. [CrossRef]
- 63. Vanderstraete, T.; Goossens, R.; Ghabour, T.K. Can ASTER-Data Be Used for Bathymetric Mapping of Coral Reefs in the Red Sea Using Digital Photogrammetry, 2nd ed.; Workshop EARSeL Special Interest Group on Remote Sensing for Developing Countries, 18-20 September; Gunter, M., Matthias, B., Eds.; University Bonn: Bonn, Germany, 2003; pp. 212–218.
- 64. Liang, J.; Zhang, J.; Ma, Y.; Zhang, C.Y. Derivation of bathymetry from high-resolution optical satellite imagery and usv sounding data. *Mar. Geod.* **2017**, *40*, 466–479. [CrossRef]
- 65. Liu, S.; Gao, Y.; Zheng, W.; Li, X. Performance of two neural network models in bathymetry. *Remote Sens. Lett.* **2015**, *6*, 321–330. [CrossRef]
- 66. Hernandez, W.J.; Armstrong, R.A. Deriving bathymetry from multispectral remote sensing data. *J. Mar. Sci. Eng.* **2016**, *4*, 8. [CrossRef]
- 67. Mohamed, H.; Negm, A.; Zahran, M.; Saavedra, O.C. Bathymetry determination from high resolution satellite imagery using ensemble learning algorithms in shallow lakes: Case study el-burullus lake. *Int. J. Environ. Sci. Dev.* **2016**, *7*, 296–301. [CrossRef]
- 68. Simon, R.N.; Tormos, T.; Danis, P.A. Very high spatial resolution optical and radar imagery in tracking water level fluctuations of a small inland reservoir. *Int. J. App. Earth Obs.* **2016**, *38*, 36–39. [CrossRef]
- 69. Pérez-Ruzafa, A.; Gilabert, J.; Bel-Lan, A.; Moreno, V.; Gutierrez, J.M. New approach to chlorophyll a determination in shallow coastal waters by remote sensing. *Sci. Mar.* **1996**, *60*, 19–27.
- 70. Yunus, A.P.; Dou, J.; Song, X.; Avtar, R. Improved bathymetric mapping of coastal and lake environments using Sentinel-2 and Landsat-8 images. *Sensors* **2019**, *19*, 2788. [CrossRef]
- Cahalanea, C.; Mageeb, A.; Monteysc, X.; Casalb, G.; Hanafind, J.; Harrise, A. Comparison of Landsat 8, RapidEye and Pleiades products for improving empirical predictions of satellite-derived bathymetry. *Remote Sens. Environ.* 2019, 233, 111414. [CrossRef]
- 72. Flener, C. Estimating deep water radiance in shallow water: Adapting optical bathymetry modelling to shallow river environments. *Boreal Environ. Res.* **2013**, *18*, 488–502.
- 73. Edwards, A.J. Applications of Satellite and Airborne Image Date to Coastal Management. Coastal region and Small Island, papers 4; UNESCO: Paris, France, 1999; p. 185.
- 74. Committee on Earth Observing Satellites (CEOS). Feasibility Study for an Aquatic Ecosystem Earth Observing System. 2018. Available online: http://ceos.org/document\_management/Publications/Feasibility-Study-for-an-Aquatic-Ecosystem-EOS-v.2-hi-res\_05April2018.pdf (accessed on 12 December 2019).
- 75. Jawak, S.D.; Vadlamani, S.S.; Luis, A.J. A synoptic review on deriving bathymetry information using remote sensing technologies: Models, methods and comparisons. *Adv. Space Res.* **2015**, *4*, 147–162. [CrossRef]
- Ferrandiz, C. La encañizada de Calnegre en la manga del mar menor y su formación en el siglo XVIII. *Murgetana* 1976, 45, 87–101.
- Del Río, V.D. Estudio Ecológico del Mar Menor. Geología, (Proyecto nº 1005 Medio Marino); IEO: Madrid, Spain, 1990.
- 78. Lillo, M.J. Geomorfología del mar menor. Pap. Dep. Geogr. 1981, 8, 9-48.
- 79. Serra, J. *Mar Menor (Spain). EUROSION Case Study, Institut de Ciència i Tecnologia Ambientals;* Universitat Autónoma de Barcelona: Bellaterra, Spain, 2013.
- 80. Pérez-Ruzafa, A.; Mompeán, M.C.; Marcos, C. Hydrographic, geomorphologic and fish assemblage relationships in coastal lagoons. *Hydrobiologia* **2007**, *577*, 107–125. [CrossRef]
- 81. Pérez-Ruzafa, A.; Marcos, C. Fisheries in coastal lagoons: An assumed but poorly researched aspect of the ecology and functioning of coastal lagoons. *Estuar. Coast. Shelf Sci.* **2012**, *110*, 15–31. [CrossRef]

- Pérez-Ruzafa, A.; Pérez-Ruzafa, I.M.; Newton, A.; Marcos, C. Chapter 15—Coastal lagoons: Environmental variability, ecosystem complexity, and goods and services uniformity. In *Coasts and Estuaries: The Future;* Wolanski, E., Day, J., Elliott, M., Ramachandran, R., Eds.; Elsevier: Amsterdam, The Netherlands, 2019; pp. 253–276.
- 83. Pérez-Ruzafa, A. Estudio Ecológico y Bionómico de Los Poblamientos Bentónicos Del Mar Menor (Murcia, SE de España). Ph.D. Thesis, Universidad de Murcia, Murcia, Spain, 1989.
- 84. Mėžinė, J.; Ferrarin, C.; Vaičiūtė, D.; Idzelytė, R.; Zemlys, P.; Umgiesser, G. Sediment transport mechanisms in a lagoon with high river discharge and sediment loading. *Water* **2019**, *11*, 1970. [CrossRef]
- 85. Manley, J.E. *High Fidelity Hydrographic Surveys Using and Autonomous Surface Craft;* Massachusetts Institute of Technology: Cambridge, MA, USA, 1998.
- Manley, J.E.; Marsh, A.; Cornforth, W.; Wiseman, C. Evolution of the autonomous surface craft Autocrat. In Proceedings of the IEEE Conference and Exhibition on Oceans 2000 MTS, IEEE 2000, Providence, RI, USA, 11–14 September 2000; Volume 1, pp. 403–408.
- 87. Brown, H.C.; Kim, A.; Eustice, R.M. An overview of autonomous underwater vehicle research and testbed at PeRL. *Mar. Technol. Soc. J.* **2009**, *43*, 33–47. [CrossRef]
- 88. Erena, M.; Atenza, J.F.; García-Galiano, S.; Domínguez, J.A.; Bernabé, J.M. Use of drones for the topo-bathymetric monitoring of the reservoirs of the Segura river basin. *Water* **2019**, *11*, 445. [CrossRef]
- 89. Domínguez Gómez, J.A.; Chuvieco Salinero, E.; Sastre Merlín, A. Monitoring transparency in inland water bodies using multispectral images. *Int. J. Remote Sens.* **2009**, *30*, 1567–1586. [CrossRef]
- Lia, J.; Knappa, D.; Schillb, S.; Roelfsemac, C.; Phinnc, S.; Silmand, M.; Mascaroe, J.; Asnera, G. Adaptive bathymetry estimation for shallow coastal waters using Planet Dove satellites. *Remote Sens. Environ.* 2019, 232, 111302. [CrossRef]
- 91. Bernstein, L.; Adler-Golden, S.; Sundberg, R.; Levine, R.; Perkins, T.; Berk, A.; Ratkowski, A.; Felde, G.; Hoke, M. A new method for atmospheric correction and aerosol optical property retrieval for vis-swir multi- and hyperspectral imaging sensors: Quac (quick atmospheric correction). In Proceedings of the IEEE International Geoscience and Remote Sensing, Symposium, Seoul, Korea, 25–29 July 2005; Volume 5, pp. 3549–3552.
- Vanhellemont, Q.; Ruddick, K. ACOLITE for SENTINEL-2: Aquatic applications of MSI imagery. In Proceedings of the ESA Living Planet Symposium, Prague, Czech Republic, 9–13 May 2016. ESA Special Publication SP-740.
- 93. McFeeters, S.K. The use of the normalized difference water index (NDWI) in the delineation of open water features. *Int. J. Remote Sens.* **1996**, *17*, 1425–1432. [CrossRef]
- 94. Más, J. El Mar Menor. Relaciones, Diferencias y Afinidades Entre la Laguna Costera y el Mar Mediterráneo Adyacente; IEO: Madrid, Spain, 2006.
- García-Pintado, J.; Martínez-Mena, M.; Barberá, G.; Albaladejo, J.; Castillo, V. Anthropogenic nutrient sources and loads from a Mediterranean catchment into a coastal lagoon: Mar Menor, Spain. *Sci. Total Environ.* 2007, 37, 220–239. [CrossRef] [PubMed]
- Dezileau, L.; Perez-Ruzafa, A.; Blanchemanche, P.; Degeai, J.P.; Raji, O.; Martinez, P.; Marcos, C.; Von Grafenstein, U. Extreme storms during the last 6500 years from lagoonal sedimentary archives in the Mar Menor (SE Spain). *Clim. Past* 2016, *12*, 1389–1400. [CrossRef]
- 97. Pérez-Ruzafa, A.; Marcos, C.; Pérez-Ruzafa, I.; Ros, J.D. Evolución de las características ambientales y de los poblamientos del Mar Menor (Murcia, SE España). *An. Biol.* **1987**, *12*, 53–65.



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