

## Article

# Total Organic Carbon Concentration and Export in a Human-Dominated Urban River: A Case Study in the Shenzhen River and Bay Basin

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**Abstract:** In order to understand the organic carbon dynamics in urban rivers, the present study monitored the total organic carbon (TOC) concentration and export in the Shenzhen River and Bay basin. The results show that the average TOC concentrations ranged from 7.04 to 17.50 mg/L in the study area, which exhibited pronounced spatial and temporal variations due to urbanization level, rainfall–runoff, and effluent of wastewater treatment plants (WWTPs). The TOC concentrations of rainwater were averaged at 4.03 mg/L during 2011–2012, which was higher than that of some urban river basins in developed countries. As an average rainfall year, the total TOC export in 2012 was  $11.2 \times 10^6$  kg/yr in the study basin, of which 37.5% was contributed by the effluent of WWTPs, 14.1% by wet deposition, and 48.4% by the surface non-point sources and endogenous pollution. The areal yield of TOC in the Shenzhen River and Bay basin was  $23.73 \times 10^3$  kg/(km<sup>2</sup>.yr) in 2012, which was 2.86 times the Pearl River's average value and 6.43 times the global average value. According to the predicted values of linear regression, the TOC concentration showed a gradual downward trend ( $R = 0.87$ ,  $p < 0.001$ ,  $n = 14$ ) during the period 2006–2019, which also induced a decreasing TOC export ( $R = 0.23$ ,  $p > 0.05$ ,  $n = 14$ ).

**Keywords:** TOC; urban river; carbon load; Shenzhen River and Bay



**Citation:** Huang, Y.; Zhang, L.; Ran, L. Total Organic Carbon Concentration and Export in a Human-Dominated Urban River: A Case Study in the Shenzhen River and Bay Basin. *Water* **2022**, *14*, 2102. <https://doi.org/10.3390/w14132102>

Academic Editor: Thomas Hein

Received: 28 May 2022

Accepted: 27 June 2022

Published: 30 June 2022

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## 1. Introduction

Rivers connect terrestrial and marine ecosystems and have essential ecosystem service functions in the carbon and hydrological cycle [1,2]. The annual carbon transported to the ocean via global streams and rivers is 0.8–1.2 Pg on average, accounting for about 25% of the net productivity of terrestrial ecosystems [3,4], which is on the same order as the carbon released by fossil fuel combustion of 5.2 Pg/yr and the net absorption of marine CO<sub>2</sub> of 1.7–2.8 Pg/yr [5–8]. River systems are an essential component of the global carbon cycle [2,5–8].

In recent decades, great attention has been paid to riverine carbon export due to its importance and complexity [9,10]. TOC is helpful to advance understanding of the role of freshwaters in contributing to regional carbon budgets, given that recent literature suggests that freshwater ecosystems may play a more important role than previously recognized in carbon transformations and retention [10–12]. Natural biological and geochemical activities in soils and bedrock, as well as anthropogenic activities, are responsible for the TOC export and its composition in stream and rivers [9]. The TOC concentrations in many Chinese rivers have increased or maintained at a high-level due to the effect of natural conditions and human pressures [2]. Zhang et al. (2009) revealed that the range of POC concentration was 0.14–6.33 mg/L, and DOC was 1.01–3.78 in a human-disturbed mountainous river in the Pearl River basin [13], which was two to three times the TOC concentration in

Xijiang River (the main branch of Pearl River) as reported by Wei et al. (2011) [14]. For some other large rivers in China, TOC concentration was 4.26–9.51 mg/L in the Haihe River, 8.09 mg/L in the Yangtze River, 3.0–16.27 mg/L in the Yellow River, and 5.05–26.62 mg/L in the Majia River [15–17]. It is well documented that land use profoundly affects lateral fluxes of carbon in surface waters [18–20]. Shih et al. (2010) [18] and Kaushal et al. (2014) [21] found that land use influences the TOC load, and a shift from agriculture or forest to urban will increase TOC load by three or four times. Kalev and Toor (2020) [22] reported that the TOC concentration of stormwater runoff could reach 12.5 mg/L, and DOC is the dominant component (86.2% of TOC) in the urban catchment of Manatee River in Southwest Florida. Wu et al. (2016) [23] determined that rainfall is the dominant factor for TOC load, and land use is also a principal factor for TOC load in a peri-urban watershed of Taihu Lake (China). Increases in riverine organic carbon concentrations have been observed across the northern hemisphere over the past few decades [14–27] due largely to climate change and the associated warming and increasing variation of precipitation, acidification, and land-use change [28–30]. For example, TOC concentrations with an average increase from 12.0 to 15.1 mg/L during the period 1993–2017 have been observed in boreal river systems of Finland, and discharge-associated land-use change is the most common driver, whereas pH and temperature were less important drivers [27]. Raymond et al. (2008) also reported that anthropogenic perturbations have greatly enhanced the fluxes of water and carbon from the Mississippi River over the past 100 years [31].

Urbanization or land use affects TOC concentration and profoundly affects TOC output by affecting rainfall–runoff and drainage in the Pearl River basin [13,14]. According to the nature–society dualistic cycle, the water cycle can be divided into natural and social cycles [32,33]. The natural water cycle is formed by the four paths of precipitation–slope–channel–underground, which is a typical collective structure [32–34]. With the development of human society, the social cycle forms complex paths such as water intake, water supply, water use, drainage, sewage collection and treatment, and recycling, which is also a typical dissipative structure [32–34].

Although there are some studies on TOC export in developed countries, carbon load in developing countries, especially for TOC export in urban rivers with a human-dominated hydrological cycle, has not been fully reported. Their contribution to the net ecosystem carbon load is still uncertain. This study aims to better understand (1) the primary characteristics of TOC concentration in rivers and rainwater and (2) TOC export and its primary sources in a rapidly urbanizing area of the Shenzhen River and Bay basin.

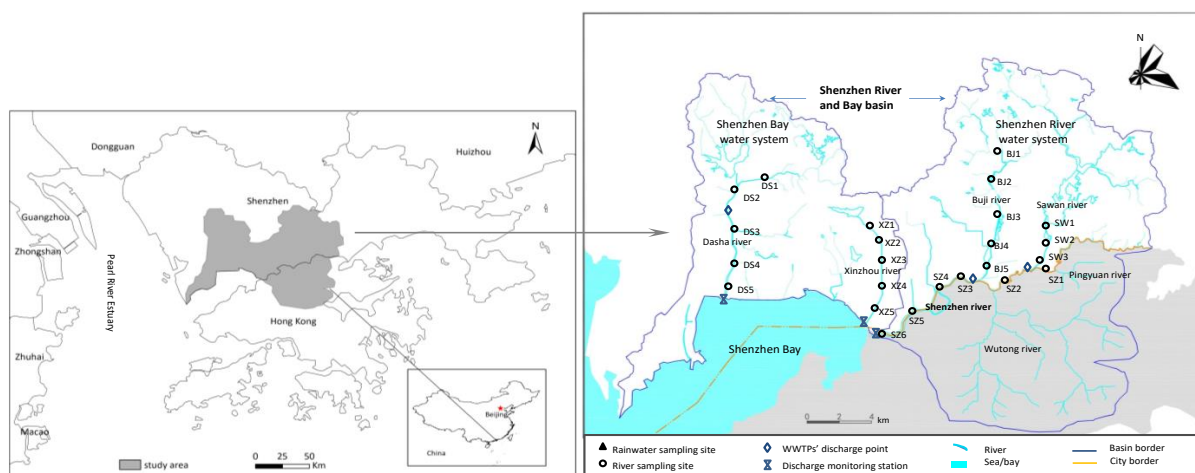
## 2. Materials and Method

### 2.1. Study Area

Shenzhen City is located in the central and southern districts of Guangdong Province, bordering Daya Bay to the East, the Pearl River Estuary to the west, Dongguan City and Huizhou City to the north, and territories of Hong Kong to the south. The geographical coordinates are 22°26′59″ to 22°51′49″ North latitude and 113°45′42″ to 114°37′21″ East longitude. The plane shape of the land area is long and narrow, which is wide from east to west and narrow from north to south. The territory has complex terrain and various geomorphologic types. It belongs to a wide geomorphologic landscape dominated by hills combined with low mountains, terraces, and plains. The city covers an area of 1952.48 km<sup>2</sup>. Shenzhen has a subtropical marine monsoon climate on the South China Sea coast. The weather is mild, and the annual average temperature is 22.3 °C. The rainy season starts from April until September, with an average yearly rainfall of 1925 mm.

Shenzhen has numerous streams and rivers. Most rivers originate from the city's Yangtai Mountain and coastal mountains and flow into the Pearl River Estuary, Shenzhen Bay, Dapeng Bay, and Daya Bay, respectively. Among them, the Shenzhen River and Bay basin was the fastest-growing region and had completed industrialization and urbanization processes during the period 1980–2010. Shenzhen River and Bay basin includes the Shenzhen River and Shenzhen Bay water system (Figure 1). The basin's total area is 472 km<sup>2</sup>: 348.5 km<sup>2</sup> on the

Shenzhen side and the other 123.5 km<sup>2</sup> on the Hong Kong side. The rivers of Shenzhen River and Bay ultimately flow into Shenzhen Bay. Shenzhen Bay is a semi-closed bay with a water surface area of 90 km<sup>2</sup>. The Shenzhen River and Bay basin is a high-intensity residential and commercial land with a GDP of USD 215 billion (about 50% of GDP of Shenzhen City, and the GDP of the Shenzhen River basin on Hong Kong side is not considered), and the population density was  $1.32 \times 10^4$  people/km<sup>2</sup> in 2019 (Shenzhen side).



**Figure 1.** Study area and sampling location in the Shenzhen River and Bay basin.

Although the Shenzhen River and Bay basin is located in subtropical China with high precipitation, there is an area with a high level of economy and society and a large population. Local water resources cannot meet the increased demand for production and living needs. The local water-gathering catchment can only provide around 15% of the total water demand. More than 85% of the water sources consumed in the Shenzhen River and Bay basin are transferred from the Dongjiang River. Two forms of the water cycle are existed. One is the natural water cycle, in which natural rainfall and runoff are deducted from reservoir water storage [32,33]. Another is the social water cycle, in which external water is transferred, and local water storage in reservoirs enters the water supply system [32,33]. Moreover, the water amount of the social water cycle is twice or more than that of the natural water cycle. The local water cycle and its associated carbon cycle and other ecological processes can be changed profoundly by enhancing the social water cycle process [32,33].

## 2.2. Water Samples

In order to understand the TOC concentration and load of rivers, a total of 24 water sampling sites among the major streams and rivers were collected in this study (Figure 1), including 5 monitoring sites in the Dasha River (DS1, DS2, DS3, DS4, and DS5), Bujia River (BJ1, BJ2, BJ3, BJ4, and BJ5), and Xinzhou River (XZ1, XZ2, XZ3, XZ4, and XZ5), respectively, and 3 monitoring sites in the Shawan River (SW1, SW2, and SW3) and 6 monitoring sites in the Shenzhen River (SZ1, SZ2, SZ3, SZ4, SZ5, and SZ6). No sampling was performed in the Pingyuan River and Wutong River, the branches of Shenzhen River on Hong Kong side. Water samplings were collected from May 2011 to Dec 2012. The sampling frequency was generally monthly in 2011 but increased to biweekly in the rainy season from July to August and decreased to three months in 2012. In total, 312 samples were collected in this study.

There are three wastewater treatment plants in the study area. Water samples of the WWTPs' effluent were also collected in 2011 at the same sampling frequency with the river systems. Because the variation in TOC concentration among the effluent of the three WWTPs was very small, no further sampling was carried out in 2012.

### 2.3. Chemical Analysis

Water samples from river sites and treated sewage were analyzed for pH, TOC, and chemical oxygen demand (COD). Because a significant correlation between COD and TOC was reported [17], this study also analyzed COD for further use. pH was determined using a micro-glass electrode (pH-Electrode, Sentix21-3, WTW, Munich, Germany). Water samplings were kept unfiltered for TOC and COD analysis. The TOC concentrations were determined with a TOC-VSCH analyzer (Shimadzu Corp., Tokyo, Japan) in this study. Before analysis, inorganic carbon was stripped off by adjusting the pH to 2 with HNO<sub>3</sub> and sparging with CO<sub>2</sub>-free synthetic air. Duplication and blank samples were used to control the quality, and the analytical error (1 $\sigma$ ) was below 1%. COD<sub>Cr</sub> was analyzed using a standard method of Chinese EPA (rapid digestion spectrophotometry HJ 399-2007). In this method, potassium dichromate solution was added to the sample and silver sulfate was used as catalyst in a medium with strong sulfuric acid. After high-temperature digestion, we determined the chemical oxygen demand value by spectrophotometry. The 24 samples from October 2011 were affected by engineering construction, and 13 samples in other months were discarded in quality control.

### 2.4. Data Collection

Historical data of rainfall, runoff and water quality, and effluent of WWTPs data in the Shenzhen River and Bay basin during the period 2006~2019 were collected from Shenzhen Water Bureau (<http://swj.sz.gov.cn>, accessed on 12 May 2021) and Shenzhen Eco-environment Bureau (<http://meeb.sz.gov.cn>, accessed on 6 May 2021). The urbanization levels (the ratio of the built-up area to the total urban area) among different river basins were collected from Shenzhen Planning and Natural Resource Bureau (<http://pnr.sz.gov.cn/>, accessed on 15 April 2022).

The rainfall amount, TOC concentration, and associated chemical data of rainwater were provided by Wang et al. (2013) and Qin et al. (2015) [35,36], which include 26 precipitation events in the Shenzhen University Town (in the middle of the Dasha River) from June 2011 to May 2012.

### 2.5. Statistical Analysis

Analysis of variance (ANOVA) was used to test the difference in TOC among the different rivers. Comparisons were made using the least significant difference (LSD) method with  $p < 0.05$ .

Correlation analysis was applied to test the relationship between rainfall volume and rainwater TOC. Linear regression was used to analyze the relationship between TOC and COD<sub>Cr</sub> concentration. Pearson's correlation coefficients were calculated to analyze the correlation between rainfall volume and rainwater TOC, COD<sub>Cr</sub> concentration, and TOC export over time, respectively, to detect the statistical significance.

## 3. Results and Discussion

### 3.1. TOC Concentrations on the Water's Surface

The averaged TOC concentrations ranged from 7.04 to 17.50 mg/L, with volume-weighted averages at 12.14 mg/L in 2011 and 12.25 mg/L in 2012 in the study basin (Figure 2). We also found significant spatial variations of TOC concentration. Generally, the TOC in the upstream area was lower than that in the midstream area and the highest in the downstream area within the same river (ANOVA,  $p < 0.05$  in Shenzhen River, and  $p > 0.05$  in the other four rivers). The reason was primarily related to urbanization, such as the low urbanization level in the upstream basin (53%, 22%, 31%, and 29% of urbanization level in the Buji River, Dasha River, Shawan River, and Xinzhou River, respectively) and the high urbanization level in the downstream basin (64%, 43%, 45%, and 62% of urbanization level in the Buji River, Dasha River, Shawan River, and Xinzhou River, respectively). Furthermore, high urbanization degree basins, such as the Buji River (urbanization level 64%) and the Xinzhou River (urbanization level 62%), have higher (ANOVA,  $p < 0.05$ ) TOC concentrations than the Dasha

River (urbanization level 43%) and Shawan River (urbanization level 45%). Our results support that the urbanization process increased the TOC load, which is consistent with the report that the development of urbanization induces the increasing water consumption and discharge, generally resulting in higher TOC concentrations and increased carbon export [37]. Figure 3 shows the river-averaged TOC concentration during different sampling periods. Averaged TOC concentrations in the wet season were 11.44, 9.86, and 13.70 mg/L for Shenzhen River, Dasha River, and Xinzhou River, respectively, which were higher than (ANOVA,  $p < 0.05$ ) that of 8.76, 7.60, and 11.06 mg/L in the dry season for the three rivers, respectively. In addition, TOC concentration was 12.25 mg/L in the wet season and 16.18 mg/L in the dry season for Buji River due to intense human activities. Seasonal variations were obvious (ANOVA,  $p < 0.05$ ) in the TOC concentration of the above four rivers during the study period, possibly due to the combined influence of the natural–social water cycle on TOC concentration [28,29]. Our results were consistent with the findings of [2,38,39] and partly supported the findings of Davies (2013), which reported that the TOC concentration in the wet season significantly exceeded that of the dry season in the Bonny Estuary, Nigeria [40]. The result of TOC concentration in this study was higher than that of the Pearl River (6.65 mg/L) [2], urban River Kelvin in the UK (8.28 mg/L) [41], and close to that of the Manatee River watershed in Florida, USA (12.52 mg/L) [22] but lower than that of Chongqing City (>43.6 mg/L) [42], Sacramento River, California (26.5 mg/L) [4], and the urban river Ouseburn in the UK (14.11 mg/L) [7]. The above analyses also showed that the TOC concentrations in surface waters vary by natural conditions and anthropogenic activities [10,21,43–46]. In recent decades, urbanization and its induced land uses have been the most important factors that increased the concentration of TOC in our study area.

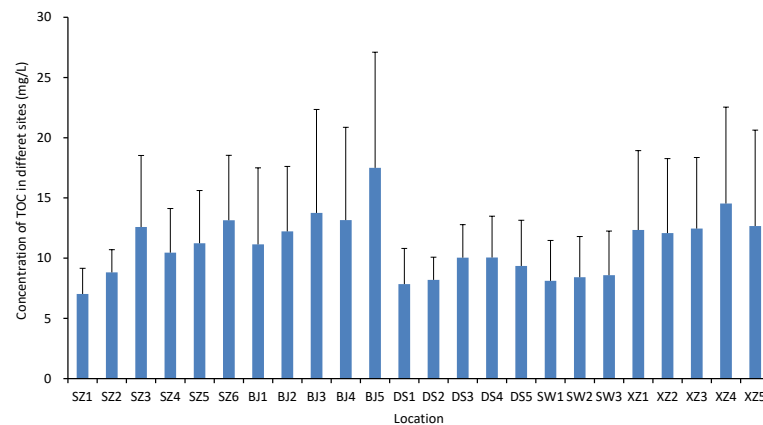


Figure 2. TOC concentration at different sites in the Shenzhen River and Bay.

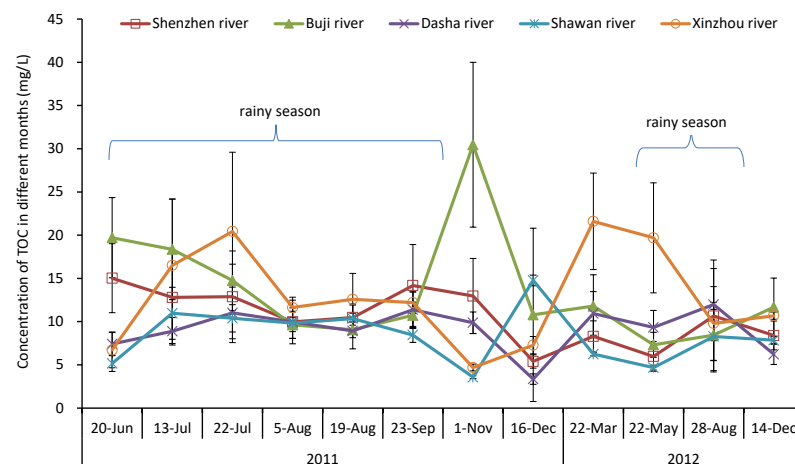
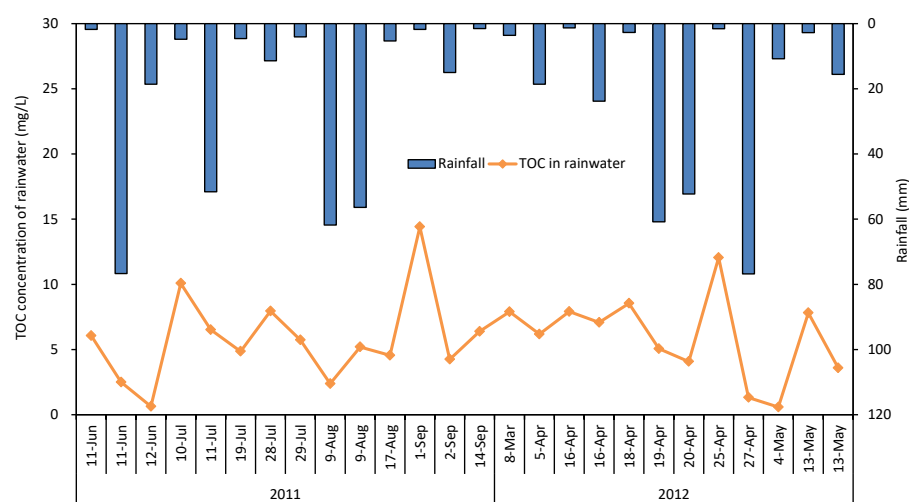


Figure 3. TOC concentrations in the Shenzhen River and Bay in different months.



### 3.2. TOC Concentrations in Rainwater

The TOC concentration in rainwater ranged from 0.58 to 14.41 mg/L, with an average of 4.12 mg/L in 2011 and 3.96 mg/L in 2012 (Figure 4). The highest value of 14.41 mg/L appeared on September 1 of the dry year 2011. There was no antecedent rainfall for a long time of period. Compared with some other studies, our findings were higher than that of Kalev and Toor (2020) in the urban Manatee River of Florida (2.6 mg/L) [22]. Antecedent rainfall can purify the atmosphere, so as to reduce the TOC concentration in rainwater [35]. The present study also found that the early rainfalls on June 11 have higher TOC concentrations than later rain on June 12 in 2011. It is supposed to be without rain in the early stage for many days. In that case, the subsequent rain generally has a relatively high TOC concentration due to air pollutant accumulation, shortage of self-purification (e.g., no rain), and rainfall characteristics (e.g., light rainfall) [36]. For example, the rain on July 10 and September 1 in 2011 and April 25 in 2012 had a relatively high TOC compared with the other rainfall events. Correlation analysis showed that the TOC concentration of rainwater decreased with rainfall (volume), showing a significant negative correlation ( $R = -0.51$ ,  $p < 0.05$ ,  $n = 26$ ).



**Figure 4.** Seasonal TOC characteristics of rainwater in Shenzhen City.

### 3.3. TOC Export and Areal Yield

From the perspective of the social water cycle [32,33], the water discharge of the wastewater treatment plants was relatively stable with slight seasonal variations; the monthly discharge was 5.6–9.5% of the annual discharge in the study basin (Figure 5). On the contrary, the runoff discharge of the natural water cycle had noticeable seasonal differences, in which the rainfall and runoff from April to September accounted for about 80.5–82.6% of the total annual, resulting in significant seasonal differences in the total discharge of the Shenzhen River and Bay basin (Figure 5).

Taking the average rainfall year 2012 (annual rainfall: 1923 mm) as an example, the yearly runoff resources of the Shenzhen River and Bay basin were  $490 \times 10^6 \text{ m}^3/\text{yr}$ , of which  $91 \times 10^6 \text{ m}^3/\text{yr}$  was stored in reservoirs and  $399 \times 10^6 \text{ m}^3/\text{yr}$  was directly discharged into Shenzhen Bay. In this year, the urban water supply was  $569 \times 10^6 \text{ m}^3/\text{yr}$ , of which  $91 \times 10^6 \text{ m}^3/\text{yr}$  was supplied by local reservoirs and  $478 \times 10^6 \text{ m}^3/\text{yr}$  was supplied by the Dongsheng Water Supply Project (Figure 6). The total effluent of WWTPs to Shenzhen Bay was  $511 \times 10^6 \text{ m}^3/\text{yr}$ . The total discharge of the WWTPs and river runoff reached  $911 \times 10^6 \text{ m}^3/\text{yr}$ . According to the flow-averaged TOC data of rivers, the total export of TOC can be calculated, which reached  $11.2 \times 10^6 \text{ kg}/\text{yr}$ , of which 37.5% was contributed by the effluent of WWTPs, 14.1% by wet deposition, and 48.4% by the surface non-point sources and endogenous pollution. Therefore, wet deposition, surface non-point sources, and endogenous pollution were dominant, which were critical ways to decrease TOC load.

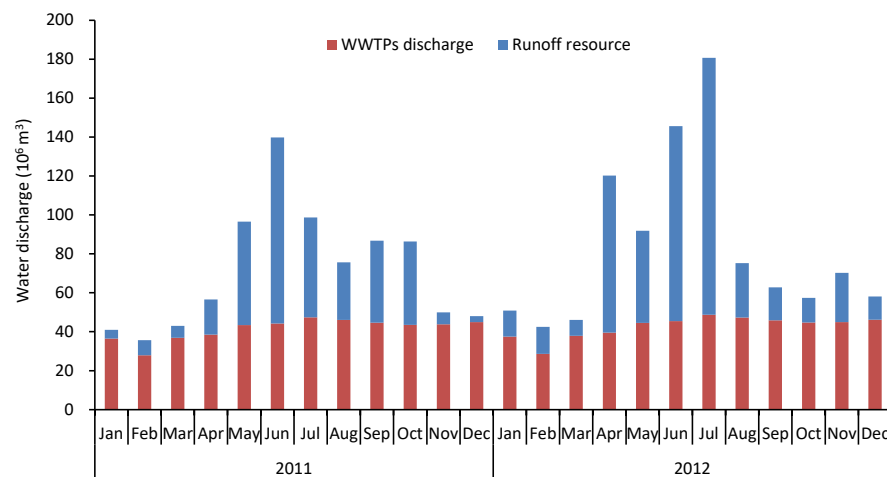


Figure 5. Runoff and effluent of treated sewage in the Shenzhen River and Bay basin in 2011–2012.

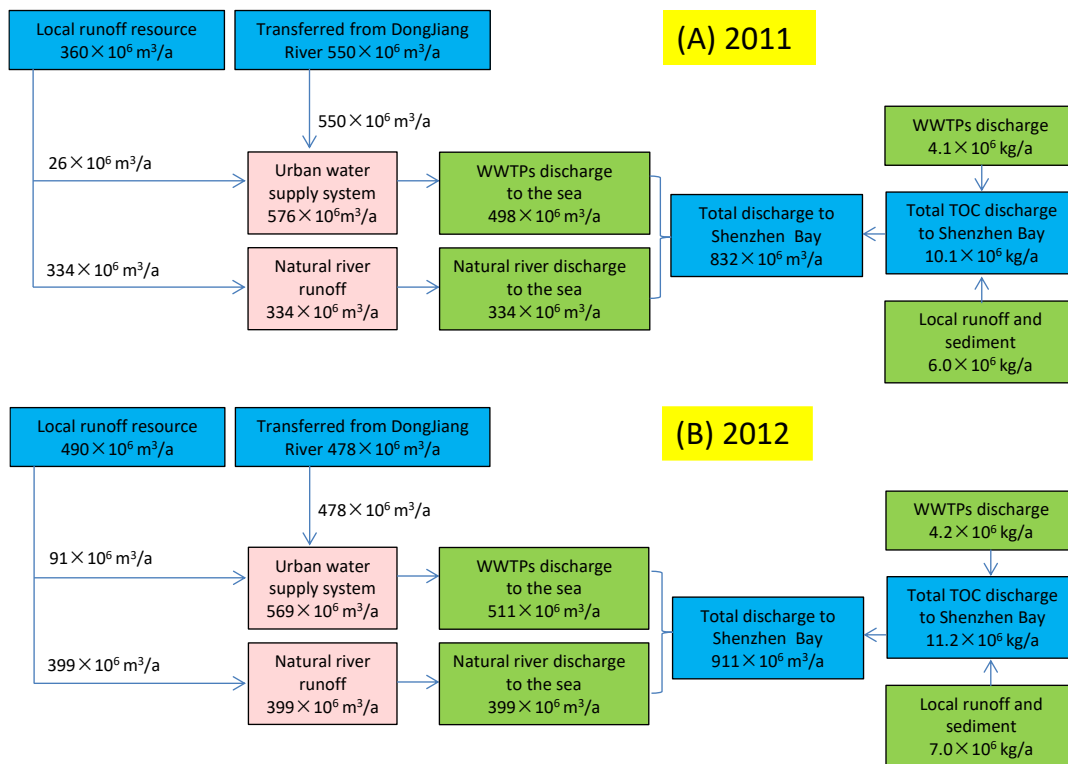


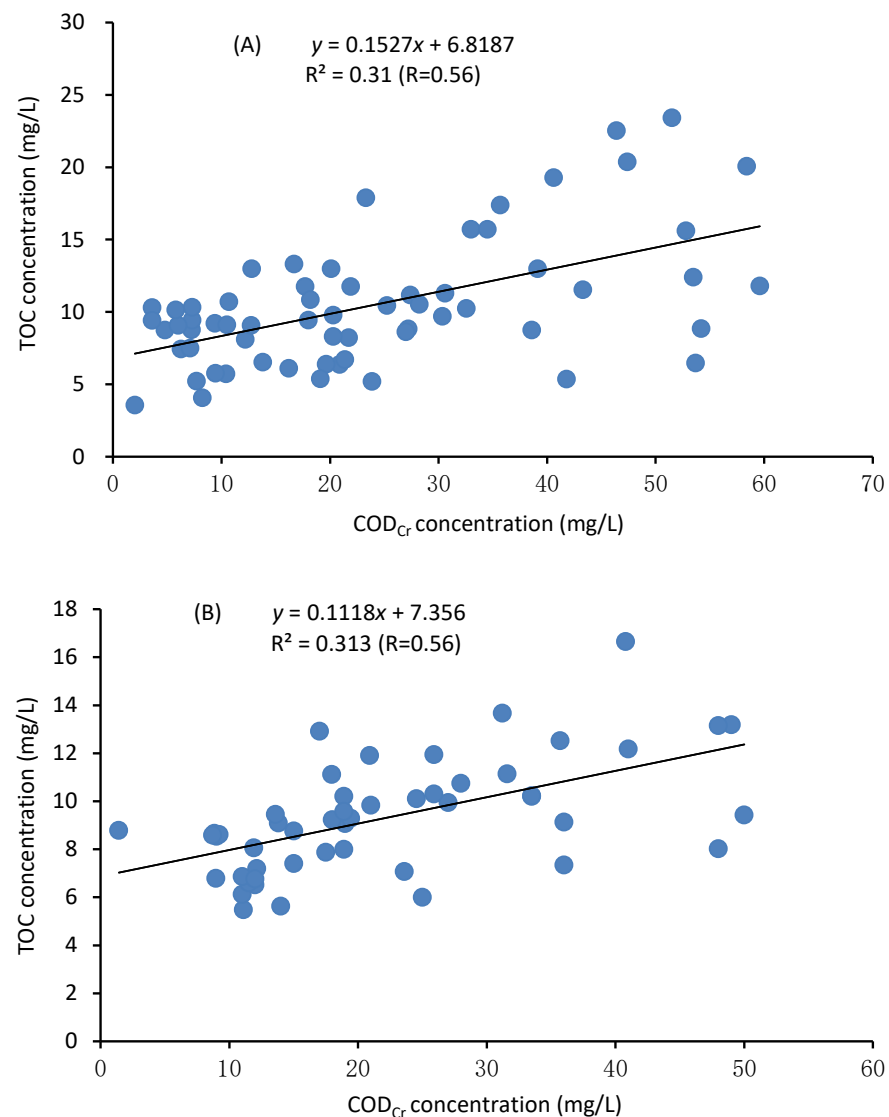
Figure 6. Water cycle and TOC export of the Shenzhen River and Bay basin in 2011 and 2012.

As a typical dry year in 2011, its annual rainfall was only 1412 mm. The water supply from the local reservoir was only  $26 \times 10^6 \text{ m}^3/\text{yr}$ , and about 95% of the urban water supply was dependent on the Dongsheng Water Supply Project (Figure 6). The total TOC export was  $10.1 \times 10^6 \text{ kg}/\text{yr}$ , the contribution of the effluent of WWTPs was  $4.1 \times 10^6 \text{ kg}/\text{yr}$ , and the local river runoff and sediment was  $6.0 \times 10^6 \text{ kg}/\text{yr}$ . The contribution of wet deposition, surface non-point sources, and endogenous pollution was 59.4% of the total TOC export, which was slightly lower than that of the year 2012.

The areal yield of TOC of the Shenzhen River and Bay basin was  $23.73 \times 10^3 \text{ kg}/(\text{km}^2.\text{yr})$  in 2012, which was 2.86 times that of the average value of the Pearl River and 6.43 times that of the global average value [2], due to the critical effect of urbanization and the nature–society dualistic cycle. The areal yield of TOC in this study was higher than that of some big rivers, such as  $15.2 \times 10^3 \text{ kg}/(\text{km}^2.\text{yr})$  in the Yellow River,  $11.83 \times 10^3 \text{ kg}/(\text{km}^2.\text{yr})$  in the Yangtze

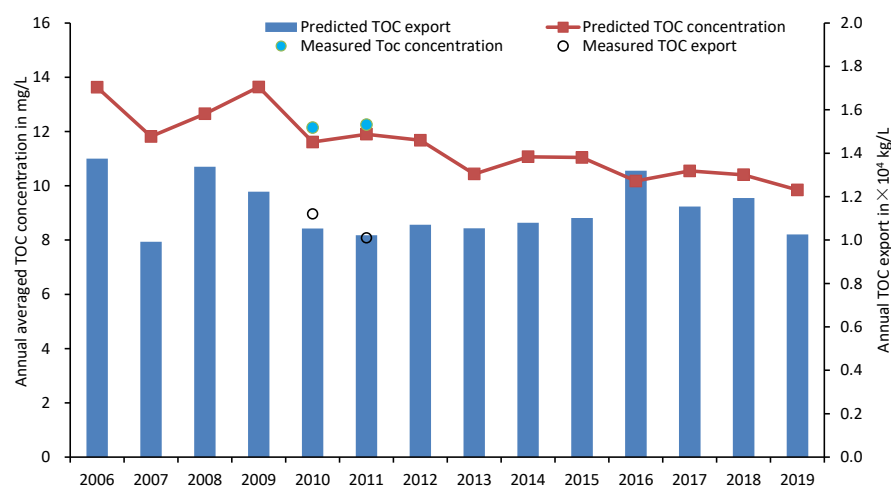
River, and  $4.74 \times 10^3$  kg/(km<sup>2</sup>.yr) in the India River [2], which was also higher than that of some urban rivers, such as the Manatee River in Florida ( $6.60 \times 10^3$  kg/(km<sup>2</sup>.yr) [22], Sacramento River in California ( $15 \times 10^3$  kg/km<sup>2</sup>.yr) [4], and Atlantic provinces in Canada ( $1.61$ – $12.35 \times 10^3$  kg/(km<sup>2</sup>.yr) [9]. Given the high areal yield of TOC in this study, differences in hydrologic conditions and land use may explain some of the differences between yields in the literature [47].

Many studies have shown a significant correlation between COD and TOC, and the regression equation can be used to predict the concentration of TOC [48,49]. The regression equation between concentrations of TOC and COD in the Shenzhen River basin (mainstream) and Shenzhen Bay basin (mainstream) was established in this study ( $C_{TOC} = 0.1527C_{COD} + 6.8187$ ,  $R = 0.56$  and  $p < 0.05$ ,  $n = 64$ ;  $C_{TOC} = 0.1188 C_{COD} + 7.356$ ,  $R = 0.56$  and  $p < 0.05$ ,  $n = 55$ , respectively, Figure 7) and used the monthly monitored COD data during the period 2006–2019 to predict the TOC concentrations, respectively. The error between the predicted concentration of TOC and the actual concentration was within 1% in 2011 and 3% in 2012 (residuals analysis). The present study calculated the annual TOC export in the Shenzhen River and Bay basin according to the TOC values and river flow. The error between the predicted total TOC load and the actual value in 2011–2012 was 8.9% in 2011 and 5.9% in 2012 (Figure 8).



**Figure 7.** The relationship between the concentrations of TOC and COD<sub>Cr</sub> in Shenzhen River basin (A) and Shenzhen Bay basin (B).





**Figure 8.** The annual TOC concentrations and export in the Shenzhen River and Bay during the period 2006–2019.

According to linear regression, the TOC concentration showed a gradual downward trend over time during 2006–2019 ( $R = 0.87$ ,  $p < 0.001$ ,  $n = 14$ ), especially after 2012 since the ecological restoration project had been finished in the primarily streams and rivers (Figure 8). Therefore, the TOC load in the Shenzhen River and Bay basin also decreased gradually over time from 2006 to 2019 ( $R = 0.23$ ,  $p > 0.05$ ,  $n = 14$ ). However, the TOC load did not reach a significant level due to the multiple effects of runoff, the effluent of WWTPs, and TOC concentration.

#### 4. Conclusions

After several decades of rapid development, the average TOC concentrations in the Shenzhen River and Bay basin remained high because of the substantial effects of urbanization and human activities. The TOC concentrations exhibited pronounced spatial variations due to the influence of urbanization level and showed temporal variations due to the combined impacts of rainfall–runoff and effluent of WWTPs. It is also worth noting that the TOC concentrations in rainwater remained high and, together with the binding effect of surface non-point sources and endogenous pollution, it has become the priority control source for TOC reduction.

The areal yield of TOC in the Shenzhen River and Bay basin was several times the Pearl River’s average TOC yield and the global average TOC yield. According to linear regression analyses, the TOC concentration and export showed a gradual downward trend over time. Further studies should focus on the composition characteristics of TOC (i.e., POC and DOC) and the concentration and export of TOC linked to nutrients (i.e., organic N, TN, organic P, and TP). In the near future, decoupling of the carbon economy to achieve sustainable development is a critical issue in this region.

**Author Contributions:** Conceptualization, Y.H. and L.R.; methodology, Y.H.; software, L.Z.; validation, Y.H., L.R. and L.Z.; formal analysis, L.Z.; investigation, L.Z.; resources, L.Z.; data curation, L.Z.; writing—original draft preparation, Y.H.; writing—review and editing, Y.H. and L.R.; visualization, L.Z.; supervision, L.R.; project administration, Y.H. and L.R.; funding acquisition, Y.H. and L.R. All authors have read and agreed to the published version of the manuscript.

**Funding:** This research was funded by [the Shenzhen Science and Technology Project] grant number [JSGG20170824102728529], and [the Research Grants Council of Hong Kong] grant number [17300621].

**Institutional Review Board Statement:** Not applicable.

**Informed Consent Statement:** Not applicable.

**Data Availability Statement:** The authors would exclude this statement because the study did not report any data.

**Acknowledgments:** This study was supported by the Shenzhen Science and Technology Project (grant: JSGG20170824102728529). L.R. acknowledged support from the Research Grants Council of Hong Kong (grant: 17300621). Thanks are also given to Peking University and Tsinghua University for their data support of rainwater and river TOC. Three anonymous reviewers are also acknowledged for their critical comments and suggestions.

**Conflicts of Interest:** The authors declared no conflict of interest.

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