

## Article

# Variability of Carbon Export in the Lower Mississippi River during an Extreme Cold and Warm Year

Lee Potter<sup>1</sup> and Y. Jun Xu<sup>1,2,\*</sup> <sup>1</sup> School of Renewable Natural Resources, Louisiana State University, Baton Rouge, LA 70803, USA<sup>2</sup> Coastal Studies Institute, Louisiana State University, Baton Rouge, LA 70803, USA

\* Correspondence: yjxu@lsu.edu; Tel.: +1-225-578-4168

**Abstract:** The Mississippi River (MR) discharges on average 474 km<sup>3</sup> of water annually into the Northern Gulf of Mexico (NGOM) with a large quantity of carbon, playing a vital role in the ecosystem's food chain and water quality. In this study, we analyzed exports of dissolved inorganic carbon (DIC) and organic carbon (DOC) from January 2021 to December 2021, during which the contiguous United States experienced one of the coldest winters as well as the hottest summer on record. Bi-weekly in situ river measurements and water sampling were conducted in the lower MR at Baton Rouge in Louisiana, USA, approximately 368 km from the river's mouth. We found that the MR transported 12.61 Tg C of DIC and 4.54 Tg C of DOC into the NGOM during the study period. Much of the DOC mass export occurred during the winter (~38%), while much of the DIC mass export took place in the spring months (~35%). The seasonality of DOC and DIC exports was affected by their concentrations, water temperature, and discharge. DIC concentrations were significantly higher in the fall (32.0 mg L<sup>-1</sup>) than those during the winter (20.4 mg L<sup>-1</sup>), while DOC concentrations were highest during the winter months (11.3 mg L<sup>-1</sup>) and varied seasonally, however, not significantly. Partial pressure of dissolved carbon dioxide (*p*CO<sub>2</sub>) in the MR averaged 1703 ± 646 μatm peaking in the summer at 2594 μatm and reaching a low in the winter at 836 μatm. Outgassing of CO<sub>2</sub> (FCO<sub>2</sub>) peaked in the spring averaging 3.43 g C m<sup>2</sup> d<sup>-1</sup> and was lowest in the winter at 1.62 g C m<sup>2</sup> y<sup>-1</sup>. Our findings validate our initial hypotheses that seasonal variability and weather extremes strongly affect terrestrial-aquatic carbon transfer, and that climate change will likely intensify carbon export from the Mississippi River Basin.

**Keywords:** carbon outgassing; seasonality; dissolved inorganic carbon; dissolved organic carbon; Mississippi River

**Citation:** Potter, L.; Xu, Y.J.Variability of Carbon Export in the Lower Mississippi River during an Extreme Cold and Warm Year. *Water* **2022**, *14*, 3044. <https://doi.org/10.3390/w14193044>

Academic Editor: Cesar Andrade

Received: 5 August 2022

Accepted: 24 September 2022

Published: 27 September 2022

**Publisher's Note:** MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.



**Copyright:** © 2022 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (<https://creativecommons.org/licenses/by/4.0/>).

## 1. Introduction

Riverine carbon transport is a significant component of the global carbon cycle currently not well constrained. Despite covering a relatively low portion of the global area, rivers are a critical conduit connecting two of the world's largest carbon reservoirs [1–3]. Globally rivers transport inland waters carrying terrestrially derived carbon and essential nutrients to coastal environments [4–6]. Riverine carbon is transported through outgassing to the atmosphere as carbon dioxide gas (FCO<sub>2</sub>) or laterally exported to the ocean. Lateral export of river carbon includes particulate organic carbon (POC, 0.17–0.24 Pg C/y), dissolved organic carbon (DOC, 0.20–0.36 Pg C/y), dissolved inorganic carbon (DIC, 0.38–0.41 Pg C/y), and particulate inorganic carbon (PIC, 0.05–0.17 Pg C/y) [2,7–10]. Global fluxes of riverine dissolved carbon are typically greater than that of particulate carbon, indicating the importance of an accurate estimation of their total flux [4,6,8]. As a significant component of the global carbon budget, recent studies have aimed to estimate global dissolved carbon export from rivers [2,4]. However, our knowledge of the seasonality of the riverine carbon transport is still limited for predicting future riverine carbon dynamics.

Dissolved inorganic carbon (DIC) is of critical importance to fluvial and coastal ecosystem health originating from soil respiration, groundwater, and carbonate weathering [7,11,12]. DIC is utilized by phytoplankton and submerged aquatic vegetation through photosynthesis and regulates aquatic CO<sub>2</sub> concentration. DIC remains the largest terrestrial carbon pool delivered to coastal ecosystems by fluvial transport [4,8,11]. Consumption and oxidation of DOC is a significant energy source within aquatic food webs, with implications for the health of coastal and estuary environments [13–15]. DOC sources primarily from the decomposition of terrestrial vegetation, runoff of detrital, and autochthonous DOC from in-stream microorganisms [16–18]. Past estimations of dissolved carbon have typically utilized long-term data sets to model carbon transport [2,19]. Many of the data sets used in these studies only account for a particular carbon form, with very few studies collecting direct measurements of dissolved carbon and fewer collecting multiple carbon forms [2,4,19]. Past models tend to have a wide discrepancy in total export as they often fail to account for seasonal variations which have significant impacts on river hydrology and environmental factors [2,4,6]. The restraints from these global models have emphasized the need for high resolutions studies on dissolved carbon transport with a direct collection of river carbon.

In addition to lateral carbon transport, rivers are known to act as a carbon source to the atmosphere through the outgassing of CO<sub>2</sub>. Large rivers are often supersaturated with CO<sub>2</sub> with respect to the atmosphere as respiration from aquatic organisms and oxidation of DOC typically overcome in-stream primary production [5,20,21]. As a greenhouse gas, CO<sub>2</sub> outgassing from rivers presents a critical component of atmospheric CO<sub>2</sub> with considerable uncertainty [2,5]. Globally inland waters are estimated to outgas between 0.23–1.80 Pg of CO<sub>2</sub> to the atmosphere annually, potentially equivalent to the total carbon exported laterally [4,6,20,22,23]. The variability in estimation has largely been due to low-resolution data and estimations of riverine partial pressure of CO<sub>2</sub> (*p*CO<sub>2</sub>) based on DIC concentration, total alkalinity, and river pH [24]. Evasion of CO<sub>2</sub> is known to vary both spatially and temporally due to changes in environment, hydrology, and land-use conditions [20,25,26]. These variations are particularly prevalent in today's large rivers in subtropical climate regions [27]. Subtropic rivers have a relatively high *p*CO<sub>2</sub> due to warmer climate, high precipitation, and increasing biogeochemical activity [21,27,28]. Subtropical regions have been subject to environment and hydrologic shifts as climate change has increased global precipitation and atmospheric temperatures [29,30]. Currently, climate change is often observed as changing mean conditions, less often considering the drastic change in weather highs and lows. As these changes continue in severity, seasonal variability is predicted to decline, with summers becoming hotter and longer and winters shorter and milder with more extreme heat waves and cold snaps [31]. Therefore, studies are needed to quantify seasonal variation in riverine carbon transport during the extreme years.

In recent years, intensifying effects of climate change have resulted in increased precipitation, temperatures, and discharge across the Mississippi River Basin [32]. Rising temperatures throughout the year have resulted in longer summers and shorter winter seasons with extreme weather events increasing in intensity [31,33,34]. Current models are projecting these trends to continue, with water temperature in the Mississippi River expected to increase between 1–8 °C over the next 80 years [35]. Additionally, river discharge has continued to follow a similar trend increasing steadily over the past century due mainly to land-use change [18,33,36]. River discharge is predicted to increase 11–60% in the coming century resulting in more regular flooding events [37,38]. These changes are predicted to increase riverine dissolved carbon exported to the Gulf of Mexico [18,39]. Recent modeling work by Tian et al. [33] has suggested DOC export has increased by nearly 40% in the last 100 years as much of the basin has undergone severe land use change and urbanization. Similarly, a study by Ren et al. [18] found DIC is predicted to increase by up to 65% in the coming century due largely to climate changes, including increased atmospheric temperature and precipitation. Previous studies simplified seasonal export from the Mississippi River into a wet and dry season where the bulk of carbon

is exported during the wet season and the least in the dry season [39,40]. With seasonal climate continuing to evolve, a more comprehensive evaluation of the temporal variability of carbon export from the Mississippi River is critical to accurately model future carbon fluxes and evaluate changes in the export and of carbon from this extensive river network.

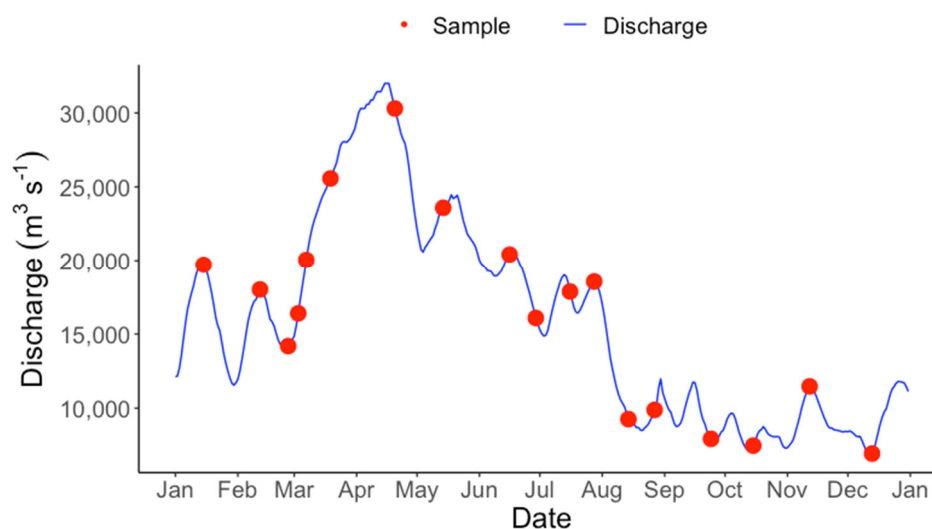
With the above introduction in mind, we conducted this study measuring riverine DIC, DOC, and  $p\text{CO}_2$  from January to December of 2021, a year with extreme weather variability across the Mississippi River Basin. Our primary goal was to determine if there was a significant relationship between seasonality and carbon export in the lower Mississippi River near its mouth to the Gulf of Mexico. Additionally, we aimed to provide an updated estimation of the total export from the Mississippi River during a year with unprecedented weather extremes. Our objectives were to: (1) determine if a substantial seasonal variation in carbon export was present during an extreme cold winter and hot summer; (2) confirm the increasing trend in the export of dissolved carbon from the Mississippi River; and (3) provide direct measurements to assist in future modeling of carbon flux estimations. With these objectives in mind, we aimed to test the hypothesis that riverine carbon export will be highest in the warm and wet spring and lowest in the dry and cold winter seasons.

## 2. Materials and Methods

This study was conducted in the Lower Mississippi River at Baton Rouge in southern United States ( $30^{\circ}26'23.5''$  N  $91^{\circ}11'30.4''$  W). The Mississippi River drains 3.2 million  $\text{km}^2$  of land, roughly 41% of the contiguous United States, and discharges each year, on average over the past four decades, a total of  $673 \text{ km}^3$  of freshwater into the Gulf of Mexico through its mainstem channel ( $474 \text{ km}^3$ ) and its distributary the Atchafalaya River ( $199 \text{ km}^3$ ) [41,42]. The study site was on the mainstem and was chosen as it rests roughly 370 km from the river's outlet into the Gulf of Mexico, eliminating the potential effect of tidal change and the resulting salinity. Over the last century, the Mississippi River Basin has undergone rapid urbanization and land-use change, severely impacting terrestrial loading and transport of carbon [18,36]. Large wetland and forest areas have been converted into urban and agricultural regions across the Mississippi Basin. These landscape changes have increased runoff and nutrient addition from wastewater effluent and agricultural fertilizers to the Mississippi River [32,43]. The lower portion of the Mississippi River has been heavily engineered to run through a series of levees, effectively cutting the river off from its floodplain. This modification has increased the speed of carbon transport, resulting in less organic carbon being stored along the rivers banks [44,45]. The faster rate of transportation has also decreased the time for organic carbon to be oxidized to  $\text{CO}_2$  on this route, increasing overall organic carbon export [45].

Climate conditions in this region are humid and subtropical, with hot and humid summers and mild winters rarely reaching sub-freezing temperatures. Our study location is in a 16 km straight shot of the river, where the water is well mixed. According to a river channel study of the Mississippi River by Wang and Xu [46], the average width of the Mississippi River at Baton Rouge was about 1200 m and the average depth at the location was 10 m. Our sampling location was on a local barge approximately 80 m into the river, where conditions were turbulent and under constant flow.

We conducted monthly and bi-weekly in situ measurements at the sample site between January and December of 2021 (Figure 1). Our sample schedule ensured we collected all samples and measurements between 9:00 and 9:30 AM Central Standard Time (CST). During each sampling trip, we took in situ measurements of partial pressure of carbon dioxide ( $p\text{CO}_2$ ) (C-Sense<sup>TM</sup> sensor Turner Designs, San Jose, CA, USA) and measured dissolved oxygen (DO), water temperature, conductivity using a YSI 556 multi-probe meter (YSI Inc., Yellow Springs, OH, USA).  $\text{NO}_2 + \text{NO}_3$  concentrations were taken from USGS gauge station # 07374000 ([https://waterdata.usgs.gov/nwis/uv?site\\_no=07374000](https://waterdata.usgs.gov/nwis/uv?site_no=07374000), accessed on 31 July 2022), choosing an hourly reading closest to our sample time. We collected water samples approximately 80 m from the riverbank in well-flowing river water. Samples were taken approximately 30–50 cm below the surface using a grab sampler.



**Figure 1.** Discharge of the Mississippi River at Study site (USGS # 07374000) over study period January–December 2021. Red dots represent sample dates.

Each trip, the grab sampler was rinsed three times with river water, in addition to rinsing the acid rinsed HDPE sample bottle three times with river water. Once thoroughly rinsed, we collected a composite sample consisting of three individual water samples into the HDPE sample bottle. We then filtered river samples into a pre-cleaned 40 mL glass vial using a 0.2  $\mu\text{m}$  filter with a capped pierceable lid and immediately placed the sample on ice. Following the sample, trip samples were placed in a fridge until delivered to LSU Wetland Biogeochemistry Analytical Services (WBAS) laboratory in Baton Rouge LA, United States for DIC and DOC analysis. Water samples were analyzed using a Total Organic Carbon Analyzer (TOC-L CHS/CSN Shimadzu, Kyoto, Japan) using the 680  $^{\circ}\text{C}$  combustion catalytic oxidation method with nondispersive infrared sensor (NDIR) detection. The instrument was calibrated using TC and IC standards, TOC was then calculated from the TC (Total Carbon) and IC (Inorganic Carbon) concentrations.

We calculated the mass flux of DIC and DOC using supplemental discharge data from USGS# 07374000. We multiplied discharge data by average monthly DIC and DOC concentrations to determine annual total DIC and DOC mass loadings ( $L_{\text{annual}}$ ) in teragrams per day (Tg/d). Average monthly DIC/DOC concentration values were assumed to represent the month due to the frequency of samples:

$$L_{\text{annual}} = \sum_{\text{month}} Q \times \left( \frac{\sum L_{\text{daily}}}{\sum Q_{\text{daily}}} \right) \quad (1)$$

whereby  $L_{\text{daily}}$  is the daily load of sampled DIC and DIC from each sample date (Tg C  $\text{d}^{-1}$ ).  $Q_{\text{daily}}$  is the total daily discharge ( $\text{km}^3 \text{d}^{-1}$ ) of each sample date in each month.  $Q$  is the average total discharge of a given sample month ( $\text{km}^3 \text{mo}^{-1}$ )  $\text{CO}_2$  outgassing ( $\text{FCO}_2$  in  $\text{mmol CO}_2 \text{m}^2 \text{h}^{-1}$ ) estimates between the river and air were calculated using the following equation from Cai and Wang [47]:

$$\text{FCO}_2 = K_T K_H (p\text{CO}_2 \text{ water} - p\text{CO}_2 \text{ air}) \quad (2)$$

$K_T$  is the gas transfer velocity (m/d), we elected to use 3.9 as was consistent with a recent study conducted at the same location by Reiman and Xu [39]. The value is slightly higher than 3.0 found by Raymond et al. [5] for 10th order streams. However, as gas transfer velocity is highly variable and not well constrained, we elected to remain consistent with

previous finding confirmed by Dubois et al. [48]. We calculated  $K_H$  using the following equation developed by Weiss [49]:

$$\ln K_H = A_1 + A_2 \left( \frac{100}{T} \right) + A_3 \left( \frac{T}{100} \right) + S \left[ B_1 + B_2 \left( \frac{T}{100} \right) + B_3 \left( \frac{100}{T} \right)^2 \right] \quad (3)$$

$T$  is the water temperature in kelvin and constants  $A_1$ ,  $A_2$ ,  $A_3$ ,  $B_1$ ,  $B_2$ , and  $B_3$  are given to be  $A_1 = -58.0931$ ,  $A_2 = 90.5069$ ,  $A_3 = 22.2940$ ,  $B_1 = 0.027766$ ,  $B_2 = -0.025888$ ,  $B_3 = 0.0050578$ .  $S$  is the salinity, which was assumed to be 0 as the Mississippi River is considered freshwater.  $p\text{CO}_2$  water is the value measured using the C-Sense sensor (Turner Designs, San Jose, CA, USA). Due to damaged equipment during the summer month, we used a regression model ( $r^2 = 0.792$ ) shown below to account for three  $p\text{CO}_2$  readings during June and July. Where  $Q$  is river discharge ( $\text{m}^3 \text{s}^{-1}$ ) at the time of sampling (09:00 CST) and  $\text{DO}$  is recorded dissolved oxygen at time of sampling ( $\text{mg L}^{-1}$ ).

$$p\text{CO}_2 = 0.02352 \times Q - 302.23305 \times \text{DO} + 3786.45046 \quad (4)$$

$p\text{CO}_2$  air was assumed to be  $410 \mu\text{atm}$  for all calculations; however, due to the local oil refinery and infrastructure, atmospheric  $\text{CO}_2$  at Baton Rouge can often be higher than average conditions. While temporal changes in atmospheric  $\text{CO}_2$  are known, a constant value was assumed for all calculations of  $\text{FCO}_2$ .

Discharge data were averaged into monthly means for flow-weighted mass export calculations. A Pearson's correlation matrix was used to look for significant ( $p < 0.05$ ) correlations across the overall data, specifically between carbon measurements and ambient water parameters. Linear regressions were used for modeling the relationships between carbon measurements and river discharge and temperature across the entire study. With seasons as factors, a one-way analysis of variance (ANOVA) was chosen to determine if seasonal variation occurred across any of our carbon data or highly correlated variables. All ANOVAs were followed by a post hoc Tukey test to determine the significant differences across factors of the ANOVA. While some data management and organization were done using Microsoft Excel, all statistical analysis was done in R version 4.2 and ran in R studio (Boston, MA, USA).

### 3. Results

#### 3.1. River Discharge and Ambient Conditions

During the study period from January 2021 through December 2021, discharge of the Mississippi River at Baton Rouge (USGS# 07374000) averaged  $15,852 \text{ m}^3 \text{ s}^{-1}$  (Std. Dev.  $\pm 6897 \text{ m}^3 \text{ s}^{-1}$ ). Discharge ranged from a maximum of  $31,998 \text{ m}^3 \text{ s}^{-1}$  in April to a low of  $6909 \text{ m}^3 \text{ s}^{-1}$  in December (Figure 1). Spring flow was highest during the study year accounting for 37.3% of total annual discharge. Comparatively, summer, fall, and winter, 19.9%, 14.2%, and 28.6% of annual discharge. Discharge during 2021 was on average 6% lower than the previous 16 years at the same site ( $16,859 \pm 8247 \text{ m}^3 \text{ s}^{-1}$ ).

Annual mean water temperature in the Mississippi River at Baton Rouge in 2021 ( $18.76, \pm 8.48 \text{ }^\circ\text{C}$ ) was nearly identical to the previous 12-year average from 2008–2020 ( $18.31, \pm 8.68 \text{ }^\circ\text{C}$ ). While mean temperature was consistent with previous years, a seasonal breakdown revealed increased temperature variability (Table 1). The 12-year mean water temperatures during spring ( $21.0 \text{ }^\circ\text{C}$ ) and summer ( $28.7 \text{ }^\circ\text{C}$ ) were both  $1 \text{ }^\circ\text{C}$  higher in this study. Long term averages for fall ( $15.4 \text{ }^\circ\text{C}$ ) and winter ( $8.1 \text{ }^\circ\text{C}$ ) deviated more drastically with winter temperatures 16% lower ( $6.82, \pm 3.18 \text{ }^\circ\text{C}$ ) and fall 28% higher ( $19.7, \pm 7.09 \text{ }^\circ\text{C}$ ) than previous years. Fall water temperature remained above  $10 \text{ }^\circ\text{C}$  for the entire season for the first time in the past decade, stressing the rising temperatures and unique weather of this study. River temperature ranged from a three-decade low of  $2.3 \text{ }^\circ\text{C}$  in mid-February to a high of  $30.5 \text{ }^\circ\text{C}$  in late July. Water temperatures were considerably lower in February following a disturbance in arctic oscillation resulting in a drastic temperature drop across the Mississippi River Basin which brought water temperatures to a near record low. With

seasonal water temperatures on average greater during our study period it is of note that the previous 12 years have shown an already increasing trend in water temperature.

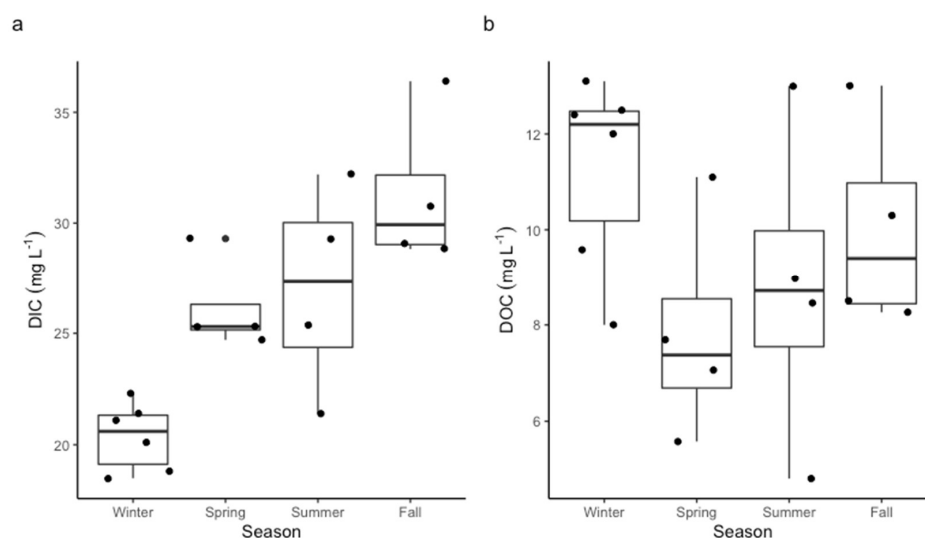
**Table 1.** Average concentration of water measurements across seasons with ( $\pm$ standard deviation). Turbidity measured in NTU (Nephelometric Turbidity Units) cDOM,  $\text{NH}_4$ , Chl-a, and Phycocyanin measured in AFU (Arbitrary Fluorescence Units).

Season	Temp °C	DO mg L <sup>-1</sup>	pH	Turbidity NTU	cDOM AFU	$\text{NH}_4$ AFU	Chl-a AFU	Phyco AFU
Winter	6.82 $\pm$ 3.18	11.43 $\pm$ 1.31	8.06 $\pm$ 0.7	92.85 $\pm$ 36.55	59.31 $\pm$ 5.13	83.75 $\pm$ 7.24	230.62 $\pm$ 45.02	127.32 $\pm$ 29.08
Spring	22.6 $\pm$ 4.92	6.81 $\pm$ 0.73	7.89 $\pm$ 0.22	41.6 $\pm$ 17.54	59.56 $\pm$ 2.62	83.55 $\pm$ 3.59	148.85 $\pm$ 17.05	77.35 $\pm$ 27.45
Summer	29.0 $\pm$ 0.95	6.07 $\pm$ 0.39	7.79 $\pm$ 0.09	85.32 $\pm$ 77.91	61.47 $\pm$ 12.04	88.64 $\pm$ 19.32	175.95 $\pm$ 55.08	97.05 $\pm$ 61.37
Fall	19.7 $\pm$ 7.09	8.21 $\pm$ 1.56	7.84 $\pm$ 0.06	30.37 $\pm$ 20.04	54.97 $\pm$ 4.67	76.41 $\pm$ 6.49	130.73 $\pm$ 24.05	54.38 $\pm$ 19.24

Other ambient parameters also showed seasonal differences. Dissolved oxygen concentration was highest in the winter season averaging ( $11.4 \pm 1.3 \text{ mg L}^{-1}$ ) and lowest in the summer at ( $6.2 \pm 0.4 \text{ mg L}^{-1}$ ) (Table 1). cDOM ( $61.47, \pm 12.04 \text{ AFU}$ ) and  $\text{NH}_4$  ( $88.64, \pm 19.32 \text{ AFU}$ ) showed a slight increase in the summer months following a spike in discharge in mid-July. Chlorophyll-a ( $230.62, \pm 45.02 \text{ AFU}$ ) and phycocyanin ( $127.32, \pm 29.08 \text{ AFU}$ ) were highest in the winter season correlating positively with low temperatures and high turbidity ( $92.85, \pm 36.55 \text{ NTU}$ ).  $\text{NO}_3+\text{NO}_2$  concentrations were significantly correlated to river discharge across our study period ( $r^2 = 0.29, p < 0.05$ ). Concentrations of  $\text{NO}_3+\text{NO}_2$  were highest during the spring season ( $1.15 \pm 0.18 \text{ mg L}^{-1}$ ) and lowest during the late summer ( $0.82 \pm 0.41 \text{ mg L}^{-1}$ ).

### 3.2. Seasonal Trend of DIC and DOC Concentrations

Mean concentrations of dissolved organic and inorganic carbon had a strong seasonal distribution within the Mississippi River. DIC concentrations increased throughout the year from a minimum average in the winter at ( $20.4 \pm 1.5 \text{ mg L}^{-1}$ ) (standard deviation) rising steadily through the spring and summer to a mean fall concentration of ( $32.0 \pm 3.9 \text{ mg L}^{-1}$ ) (Table 2). DOC concentrations inversely were highest in the winter months, correlating significantly with the low winter temperatures ( $p < 0.05$ ). Discharge was significantly and inversely correlated with DIC concentrations ( $p < 0.05$ ). Based on a one-way ANOVA and post hoc Tukey test, DOC concentration had no significant seasonal concentration variation. While not significant, DOC did decrease in the spring and summer with rising discharge and temperatures. Meanwhile, DIC displayed a significant difference in mean concentration between winter and the remaining three seasons (Figure 2).



**Figure 2.** Boxplot of seasonal variability of (a) DIC; (b)DOC. Dots are sample points from each field trip.

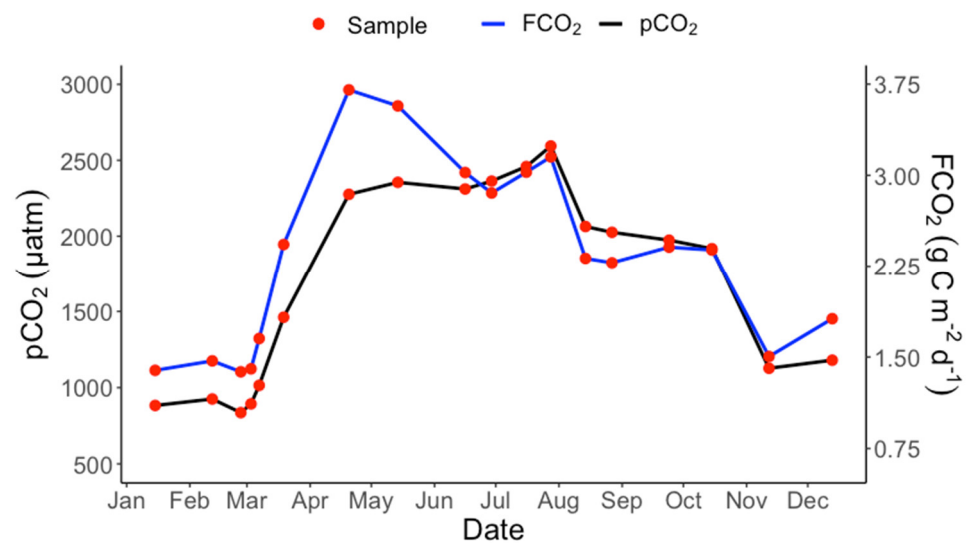
The DIC: DOC ratio within the Mississippi River increased steadily throughout the year from a low in March (1.4) directly following the cold winter event to a high (6.1) in late August following a month-long heatwave with temperatures averaging  $>29$  °C. The ratio of DIC: DOC remained below 2.0 for one of the winter samplings dates (19 March), which was on the latter end of the season when temperatures and discharge of the river had begun to rise. The ratio remained above 2.0 for the remainder of the year aside from one sample point in July following a two-week period where river discharge rose 30%.

**Table 2.** DIC, DOC, NO<sub>3</sub> and NO<sub>2</sub> and river discharge across all sample dates.

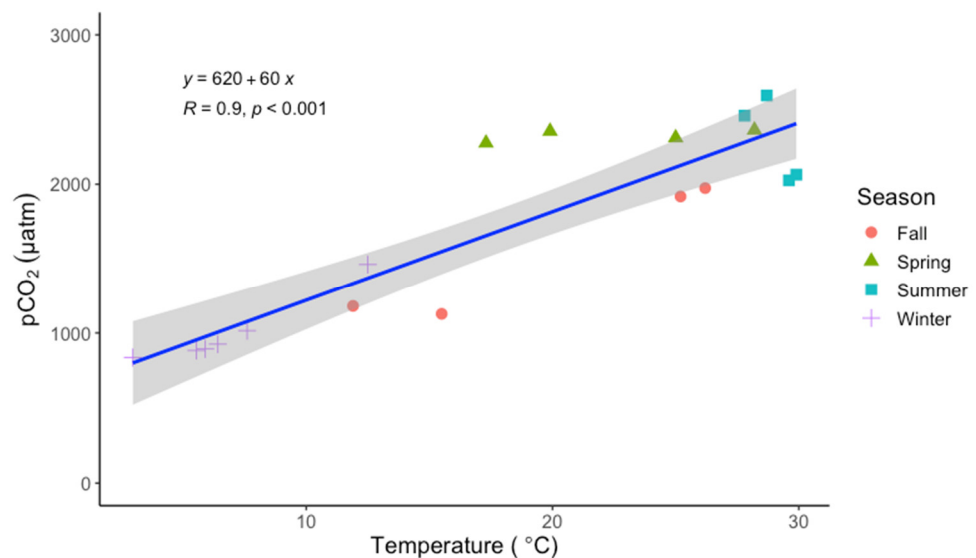
Sample Date	DIC mg L <sup>-1</sup>	DOC mg L <sup>-1</sup>	NO <sub>3</sub> + NO <sub>2</sub> mg L <sup>-1</sup>	Q m <sup>3</sup> s <sup>-1</sup>
15 January 2021	18.8	9.58	0.98	19,708.53
12 February 2021	20.1	12.4	1.06	18,009.51
26 February 2021	22.3	12.5	1.09	14,186.74
03 March 2021	21.1	12	1.04	16,395.45
07 March 2021	18.5	13.1	0.92	20,048.33
19 March 2021	21.4	8	1.11	25,570.11
20 April 2021	25.3	5.58	1.15	30,299.03
14 May 2021	24.7	7.06	1.16	23,587.93
16 June 2021	25.3	7.69	1.37	20,416.45
29 June 2021	29.3	11.1	0.93	16,083.97
16 July 2021	21.4	13	1.37	17,867.93
28 July 2021	25.36	8.97	1.09	18,547.53
14 August 2021	32.2	8.46	0.7	9259.61
27 August 2021	29.3	4.81	0.37	9882.58
24 September 2021	29.1	8.5	0.56	7928.72
15 October 2021	30.76	10.3	0.6	7475.65
12 November 2021	28.83	13.01	1.12	11,468.32
13 December 2021	36.4	8.26	1.24	6937.63

### 3.3. Seasonal Variation in Partial Pressure and Outgassing of CO<sub>2</sub>

The partial pressure of CO<sub>2</sub> in the Mississippi River was supersaturated with respect to the atmosphere (greater than 410 µatm) across all 18 sample dates, averaging (1703 ± 646 µatm, Figure 3). *p*CO<sub>2</sub> varied widely across seasons rising from low winter (1002 ± 233 µatm) concentrations to a high in the spring (2326 ± 40 µatm) summer (2223 ± 283 µatm) then steadily declining into the late fall (1408 ± 442 µatm). *p*CO<sub>2</sub> reached a low of 836 µatm directly following a drop in water temperature to near record low of 2.3 °C. *p*CO<sub>2</sub> was near twice the winter concentration in all other seasons. *p*CO<sub>2</sub> had a significant and positive correlation with water temperature and inverse relationship with DO ( $p < 0.05$ , Figures 4 and 5). Dissolved oxygen correlated negatively and significantly with *p*CO<sub>2</sub> ( $p < 0.05$ , Figure 6) and positively with Chlorophyll-a ( $p < 0.01$ , Table 2). River turbidity, Chlorophyll-a and cyanobacteria presence were all highest in the winter season at 92.85 NTU, 230.62 AFU and 127.32 AFU. These parameters saw a drastic decline in the spring season at nearly half their winter values.



**Figure 3.** Seasonal trend of  $p\text{CO}_2$  and  $\text{CO}_2$  outgassing from the Mississippi River at Baton Rouge, Louisiana.



**Figure 4.** Positive correlation between  $p\text{CO}_2$  and water temperature in the Mississippi River across the study period.

Outgassing rates followed a similar trend to  $p\text{CO}_2$ , averaging  $2.32 \pm 0.77 \text{ g m}^{-2} \text{ d}^{-1}$  and ranging from a winter low of  $1.62 \text{ g m}^{-2} \text{ d}^{-1}$  to a spring high of  $3.43 \text{ g m}^{-2} \text{ d}^{-1}$  (Figure 3). Rates remained much higher in the warmer spring and summer months, nearly double the winter, and fall. The highest rate of  $\text{CO}_2$  outgassing occurred in April  $3.70 \text{ g m}^{-2} \text{ d}^{-1}$  in concurrence with the peak discharge of  $31,998 \text{ m}^3 \text{ s}^{-1}$  witnessed in our study.

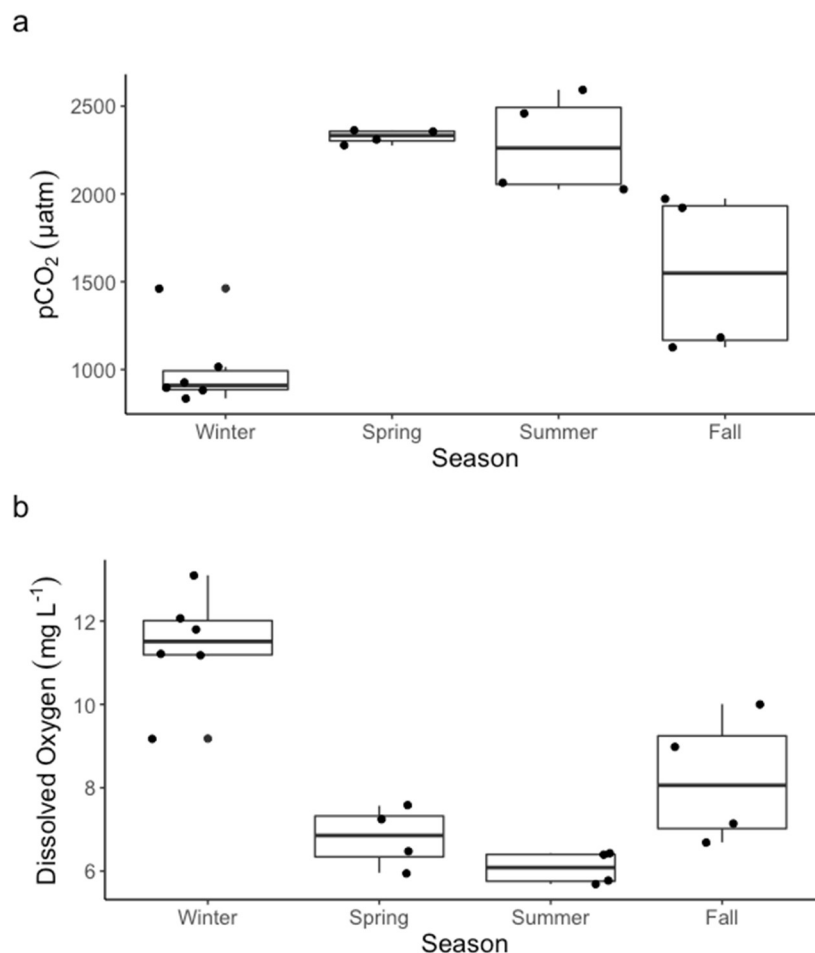
In total, the Lower Mississippi outgassed  $865 \text{ g C m}^{-2}$  (Table 3) totaling  $267.18 \text{ Gg C}$  for the study year. The bulk of this occurred in the spring and summer, i.e.,  $169.92 \text{ Gg C}$  (or 63.6%). Outgassing was calculated based on direct  $p\text{CO}_2$  measurements and water temperature which both followed a similar seasonal trend. The least amount of outgassed  $\text{CO}_2$  occurred in the winter season ( $43.28 \text{ Gg C}$ ) and this corresponded with low river temperature and  $p\text{CO}_2$ .

### 3.4. Mass Export of DIC and DOC

During our study period, a total of  $12.61 \text{ Tg}$  of DIC was exported to the Northern Gulf of Mexico from the lower Mississippi River (Table 4). During this period the greatest export occurred in the spring at 37.9%. The lowest export was in the fall where only 18.1%



(Table 3) of the annual load was exported. This ratio was nearly identical to that of the river discharge that accounted for 37.3% of annual flow in the spring compared with 14.2% in the fall. However, this was inverse when looking at concentration of DIC which increased throughout the year and was highest in the fall ( $32.0, \pm 3.9 \text{ mg L}^{-1}$ ) (Figure 5).



**Figure 5.** Boxplot of seasonal variation in riverine (a)  $p\text{CO}_2$ ; (b) Dissolved oxygen concentration. Dots showing sample points.

**Table 3.** Total mass export of dissolved inorganic carbon (DIC) dissolved organic carbon (DOC), CO<sub>2</sub> outgassing (FCO<sub>2</sub>), and river discharge (Q). with seasonal percentage of the annual total.

Season	DIC Tg	DOC Tg	FCO <sub>2</sub> g C m <sup>-2</sup> d <sup>-1</sup>	Q km <sup>3</sup>
Winter	2.88 (22.8%)	1.56 (34.5%)	1.62 (16.2%)	143.1 (28.6%)
Spring	4.78 (37.9%)	1.31 (28.9%)	3.43 (35.9%)	186.4 (37.3%)
Summer	2.68 (21.2%)	0.91 (20.1%)	2.72 (27.7%)	99.5 (19.9%)
Fall	2.28 (18.1%)	0.75 (16.5%)	2.03 (20.3%)	71.0 (14.2%)
Total	12.61	4.54	864.6	500

During the same period the Mississippi River exported 4.54 Tg of DOC, about one 3rd the DIC load. DOC load (1.56 Tg) was greatest in the winter, accounting for 34.5% of the annual total (Table 3). DOC correlated significantly with both cyanobacteria and

chlorophyll-a which were both significantly higher in the winter ( $p < 0.05$ ) and nearly half in the drier and warmer fall months.

**Table 4.** Comparison of previous studies of dissolved carbon and outgassing from the Mississippi River.

Study Period	Q	DOC		DIC		$p\text{CO}_2$	$\text{FCO}_2$	Reference
	$\text{Km}^3 \text{ yr}^{-1}$	$\text{Tg C yr}^{-1}$	$\mu\text{mol L}^{-1}$	$\text{Tg C yr}^{-1}$	$\mu\text{mol L}^{-1}$	$\mu\text{atm}$	$\text{g C m}^{-2} \text{ yr}^{-1}$	
2021	500	4.54	$806 \pm 218$	12.61	$2129 \pm 419$	$1703 \pm 646$	864.6	This Study
2015–2018	548	3.95	$607 \pm 158$	12.25	$1782 \pm 585$	$1500 \pm 743$	654	[39]
2009–2010	550	1.6	$296 \pm 54$	-	-	-	-	[50]
2006–2008	457	1.88	$307 \pm 28$	13.6	$2421 \pm 480$	-	-	[40]
2000–2001	374	1.51	$375 \pm 42$	-	-	$1362 \pm 267$	$1077 \pm 407$	[48]
1971–2000	-	$2.6 \pm 0.4$	-	$18.8 \pm 3.4$	-	-	-	[33] *

Note: \* Findings based on simulated models.

#### 4. Discussion

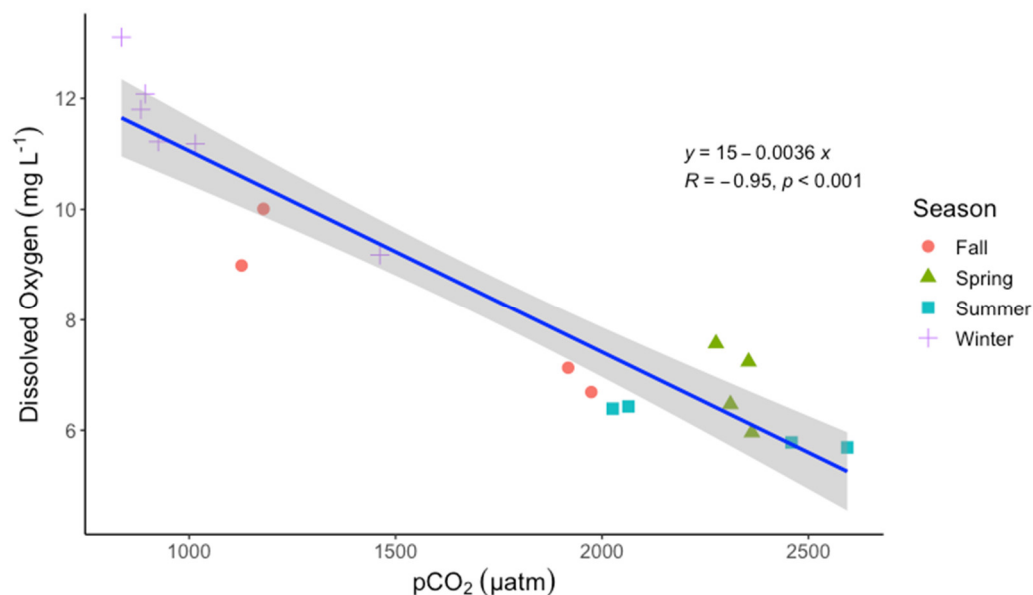
Our study took place during a year with significant seasonal variability. 2021 was one of the hottest years on record across the Mississippi River Basin and was above average in annual precipitation. Severe weather extremes throughout the year contributed to conditions being above the long-term average. One such example was intense variability in winter conditions as a severe weather event brought on by a destabilization of the polar vortex event provided much of the basin one of the coldest winters on record. Across the basin the bulk of the annual precipitation occurred in the spring and summer with conditions becoming much drier and hotter leading into the fall season. The abnormal precipitation and temperature contributed and were emphasized by the seasonal variation in riverine carbon transport.

**Table 5.** Pearson correlation coefficients for measured parameters. Phycocyanin (Phyco), chlorophyll-a (Chl-a), colored dissolved organic matter (cDOM), ammonium ( $\text{NH}_4$ ), turbidity (N), water temperature (T), specific conductivity (SC), dissolved oxygen (DO), pH, partial pressure of carbon dioxide ( $p\text{CO}_2$ ), dissolved inorganic carbon (DIC), dissolved organic carbon (DOC) and discharge Q. Only significant relationships displayed ( $p < 0.05$ ).

	Phyco	Chl a	cDOM	$\text{NH}_4$	N	T	SC	DO	pH	$p\text{CO}_2$	DIC	DOC	Q
Phyco	1.00												
Chl a	0.94	1.00											
cDOM	0.69	0.63	1.00										
$\text{NH}_4$	0.67	0.62	0.99	1.00									
N	0.79	0.70	0.58	0.63	1.00								
T	-	-0.59	-	-	-	1.00							
SC	-0.57	-0.51	-	-	-0.54	-	1.00						
DO	-	0.57	-	-	-	-0.97	-	1.00					
pH	-	-	-	-	-	-	-	-	1.00				
$p\text{CO}_2$	-	-	-	-	-	0.90	-	-0.95	-0.15	1.00			
DIC	-0.78	-0.77	-0.48	-0.48	-0.71	0.54	0.86	-0.48	-0.09	-	1.00		
DOC	0.49	0.55	-	-	-	-	-	-	0.05	-0.49	-	1.00	
Q	-	-	-	-	-	-	-0.91	-	0.20	-	-0.66	-	1.00

Findings from this study indicate significant seasonal variability in dissolved carbon export from the Mississippi River to the Northern Gulf of Mexico. As shown in previous studies by Reiman and Xu [39] & Joshi [41], DIC export varied significantly with temperature and discharge [51]. Extreme and abnormal temperature, precipitation, and discharge variability across the study period magnified this effect. One such event during the winter was followed by river temperatures far below the 14-year seasonal average at near-record lows with river temperature below  $5^\circ\text{C}$  for over a week. Following this event, the highest

DOC concentration (13.1 mg/L) and lowest  $p\text{CO}_2$  (836  $\mu\text{atm}$ ) from our study were recorded. Recent studies on carbon transport by Raymond and Bauer [15] and Bianchi et al. [17] suggested instream processing of organic carbon increases from headwaters to the coast [39]. The abnormally high winter DOC was considered to be attributed to a decrease in stream processing of DOC to DIC as low temperature hindered biogeochemical processes. As a net heterotrophic system [52] decreased winter processing of DOC is consistent with previous findings by Reiman and Xu [53] of diel variability of  $p\text{CO}_2$  during the summer months compared to spring. While winter discharge was around one quarter of the annual, over 1/3rd of total DOC was exported this season (Table 3).



**Figure 6.** Inverse relationship of river dissolved oxygen content and  $p\text{CO}_2$ .

DIC was the primary form of carbon in the Mississippi River, accounting for approximately 75% of total dissolved carbon exported in our study. While DIC was negatively correlated to river discharge ( $r^2 = -0.66$   $p < 0.05$ ), it remained the dominant form of carbon through all seasons. DIC concentration increased throughout the year with increasing water temperature and decreasing discharge. DIC concentration was correlated most significantly with discharge and water temperature (Table 5), as found in previous works by Tian et al. [33] & Reiman and Xu [39]. DIC and  $p\text{CO}_2$  were not correlated significantly, as in the fall season,  $p\text{CO}_2$  decreased while DIC increased. A study by Abril et al. [24] found the use of DIC or total alkalinity to calculate  $p\text{CO}_2$  empirically resulted in an overestimation of overall  $p\text{CO}_2$  from rivers. These findings suggest that in the Mississippi, DO and discharge may be a better fit for determining  $p\text{CO}_2$  as nearly all study parameters significantly impact DIC.

Nearly 40% of total carbon exports occurred during the spring season. This was consistent with the annual river flush, where river discharge was highest and high concentrations of soil DOC and nutrients from agriculture fertilizers are released through runoff from high precipitation and headwater snow melt [54]. While no correlation was seen between nitrates and either carbon species, the positive correlation with discharge and  $\text{NO}_2 + \text{NO}_3$  may be evidence of increased fertilizer leaching during the spring flush and increased biological activity within the river. Rivers are known to be more heterotrophically active in the spring and summer months as water temperatures rise [55,56]. During these warmer months, organic carbon can be mineralized into DIC at a more rapid rate, leading to the drop in DOC as seen in the spring and summer. While DOC and DIC did not have a clear significant inverse relationship, our study did find high  $p\text{CO}_2$  in the spring and summer to be consistent with rising DIC and lower DOC. As seen in previous studies by both Joshi

et al. [41] and Shen et al. [50] of the Mississippi River this was likely due to the rising temperatures throughout the year in the region and likely increased DOC mineralization.

This study is the first that investigated carbon export during one of the coldest winters and hottest falls in the Mississippi River. Severe winter conditions during our study resulted in an intense drop in temperature lasting for three weeks between 13 February and 7 March 2021. The winter weather event coincided with some of the lowest  $p\text{CO}_2$  measurements recorded in the Lower Mississippi River and our study's highest recorded concentration of DOC. The increase in riverine DOC may be attributed to several factors. (1). As a net heterotrophic system, the low temperatures can result in less instream processing of DOC to DIC, decreasing the DIC: DOC ratio and river  $p\text{CO}_2$  concentration [40,57] (2). The increased flow from the spring flush is known to flush a large amount of soil organic carbon into the river [18,40]. The increased flow also acts as a diluting effect on DIC, which could partly explain the sudden shift in the DIC: DOC ratio [33,58]. While previous studies such as Reiman and Xu [39] have reported similar seasonal trends, the sudden and pronounced change in weather resulted in ambient winter conditions significantly below the long-term average. While we did not observe a net sink of  $\text{CO}_2$ , the low  $p\text{CO}_2$  resulted in less overall  $\text{CO}_2$  outgassed during the winter. A recent study by Reiman and Xu [53] showed a diel variation in  $p\text{CO}_2$  at the exact location, suggesting values may have been even lower in the middle of the day. Additionally, the rising discharge from the early spring flush corresponding with peak DOC concentrations resulted in the most significant seasonal and annual reported export of DOC to the Northern Gulf of Mexico to date (Tables 3 and 4). When ignoring the extreme winter weather and using only monthly measurements DOC concentrations and  $p\text{CO}_2$  were 20% lower and 10% higher respectively. These findings help to confirm the significant role of weather events on long terms carbon export sums.

In addition to one of the coldest records, our study also witnessed one of the warmest falls. Water temperatures during the fall season of 2021 remained above  $10\text{ }^\circ\text{C}$  for the entirety of the fall season, a shift from the past decade where temperatures typically drop around late November. The concentration of DIC was higher during our study ( $25.57 \pm 5.03\text{ mg L}^{-1}$ ) than in the most recent update of carbon export in the Mississippi River from 2019 [39] ( $21.40 \pm 7.02\text{ mg L}^{-1}$ , Table 4). DIC in our study followed the season trend found in previous findings by Cai et al. and Raymond et al. of increasing throughout the year with rising temperature and decreasing discharge [36,40]. Due to a warmer and drier fall than last year, DIC concentration in the fall averaged above  $30\text{ mg L}^{-1}$  throughout the season. As previously found, DIC concentration in the Mississippi is primarily attributed to carbonate dissolution of soil  $\text{CO}_2$  in the Upper Mississippi [48]. High river discharge can reduce contact time with minerals and dilute the overall concentration of DIC [58,59]. Under the changing seasonal conditions in our study, DIC concentrations were subject to more extreme variability, resulting in a much higher DIC: DOC ratio in the fall season. While conditions during our study were considered irregular, climate change is expected to shift seasonal conditions to extremes. Based on their analysis of extreme temperature indicators in the Mississippi River Basin during 1948–2017, Tavakol et al. [34] postulated that the region has a growing risk of extreme temperatures with more frequent and longer hot events. Summer and fall are expected to become hotter and drier, with spring precipitation and flooding events increasingly regular [60]. While our study has shown a clear seasonal trend and increased overall export of carbon from the Mississippi River, it is only representative of a single year of data collection in the river. To further support these findings, more work is needed in rivers throughout different climate regions and over longer study periods. While winters are expected to become milder, the findings from our study on dissolved carbon transport have implications for predicting future seasonal and annual variation. As these conditions continue to change, our findings point to the importance of considering these events in carbon export models to assess the role of large rivers more accurately in the carbon budget.

## 5. Conclusions

This study demonstrates a significant seasonal effect on riverine carbon export in a large river basin. This effect is likely to become more variable as climate change continues to impact basin weather resulting in longer and hotter summers followed by shorter and more mild winters. Our study found that the bulk of carbon export occurs in the spring season in concurrence with the highest discharge of the river and rising temperatures. The unique weather patterns from our study confirmed a strong effect of extreme weather on carbon transport in the short term, with rippling effects throughout the year. The lower Mississippi River exported 4.54 Tg of DOC and 12.61 Tg of DIC. CO<sub>2</sub> outgassing (864.6 g C m<sup>-2</sup> yr<sup>-1</sup>) was greater during our study than all previous findings, supporting the effect of rising water temperature and discharge on outgassing. High riverine pCO<sub>2</sub> during the spring (3.43 g C m<sup>2</sup>) led to outgassing rates more than double that of winter (1.62 g C m<sup>2</sup>). DIC showed no significant increase compared to previous findings. However, seasonality strongly affected DIC as concentration increased with warming temperatures and drier conditions throughout the year ultimately 50% higher in the fall (32.0 mg L<sup>-1</sup>) than the winter 20.37 mg L<sup>-1</sup>. The impact of shifting seasonal conditions is seen in changing DIC: DOC export ratios throughout the year. While our finding saw an increase in DOC export due to low temperatures, moving forward, DIC will increase in concentration with rising temperatures owing to increased in-situ processing of DOC and carbonate weathering rates throughout the basin. Outgassing rates of CO<sub>2</sub> are most significant in the high flow warmer conditions and will likely increase in the future as these conditions intensify. While our study reaffirmed the river as a carbon source to the atmosphere, the low pCO<sub>2</sub> observed during the winter may suggest a negative feedback loop if extreme winter weathers events persist, resulting in lower winter outgassing or a temporal carbon sink. Our findings imply that seasonality is critical in assessing total carbon load from large rivers such as the Mississippi River. As seasonal variation in the Mississippi River Basin continues to intensify under the changing climate, dissolved carbon transport to the Gulf of Mexico will become more seasonally dependent. As the weather becomes more unpredictable under the changing climate, more work on short-term events will be needed to confirm their effect on total seasonal transport. These findings will help future models better account for river carbon budgets under changing seasonal conditions and further validate these findings.

**Author Contributions:** L.P.: Field measurements, Data collection & analysis, Writing—review and editing. Y.J.X.: Conceptualization, Methodology, Supervision, Funding acquisition, Writing—review and editing. All authors have read and agreed to the published version of the manuscript.

**Funding:** This study was supported by a U.S. National Science Foundation grant (Award#: 1937826) and a U.S. Geological Survey 104b Award (Award#: G16AP00056). The study also benefited from a U.S. Department of Agriculture Hatch Fund project (Project#: LAB94459).

**Institutional Review Board Statement:** Not applicable.

**Informed Consent Statement:** Not applicable.

**Data Availability Statement:** Not applicable.

**Acknowledgments:** The authors are grateful to the associate editor and two anonymous reviewers for their valuable feedback and suggestions, which were important and helpful to improve the quality of this article.

**Conflicts of Interest:** The authors declare no conflict of interest.

## References

1. Meybeck, M. Carbon, nitrogen, and phosphorus transport by world rivers. *Am. J. Sci.* **1982**, *282*, 401–450. [[CrossRef](#)]
2. Li, M.; Peng, C.; Wang, M.; Xue, W.; Zhang, K.; Wang, K.; Shi, G.; Zhu, Q. The carbon flux of global rivers: A re-evaluation of amount and spatial patterns. *Ecol. Indic.* **2017**, *80*, 40–51. [[CrossRef](#)]
3. Battin, T.J.; Luyssaert, S.; Kaplan, L.A.; Aufdenkampe, A.K.; Richter, A.; Tranvik, L.J. The boundless carbon cycle. *Nat. Geosci.* **2009**, *2*, 598–600. [[CrossRef](#)]

4. Cole, J.J.; Prairie, Y.T.; Caraco, N.F.; McDowell, W.H.; Tranvik, L.J.; Striegl, R.G.; Duarte, C.M.; Kortelainen, P.; Downing, J.A.; Middelburg, J.J.; et al. Plumbing the global carbon cycle: Integrating inland waters into the terrestrial carbon budget. *Ecosystems* **2007**, *10*, 172–185. [[CrossRef](#)]
5. Raymond, P.A.; Hartmann, J.; Lauerwald, R.; Sobek, S.; McDonald, C.; Hoover, M.; Butman, D.; Striegl, R.; Mayorga, E.; Humborg, C.; et al. Global carbon dioxide emissions from inland waters. *Nature* **2013**, *503*, 355–359. [[CrossRef](#)]
6. Meybeck, M.; Vörösmarty, C. Global transfer of carbon by rivers. *Glob. Chang. Newsl.* **1999**, *37*, 18–19.
7. Meybeck, M. Concentrations of river waters in major elements and contributions in solution awx oceans. *Rev. Geol. Dyn. Geogr. Phys.* **1979**, *21*, 215–246.
8. Ludwig, W.; Amiotte Suchet, P.; Probst, J.L. River discharges of carbon to the world's oceans: Determining local inputs of alkalinity and of dissolved and particulate organic carbon. *Sci. Terre Des. Planètes Comptes Rendus L'académie Des. Sci.* **1996**, *323*, 1007–1014.
9. Huang, T.H.; Fu, Y.H.; Pan, P.Y.; Chen, C.T.A. Fluvial carbon fluxes in tropical rivers. *Curr. Opin. Environ. Sustain.* **2012**, *4*, 162–169. [[CrossRef](#)]
10. Hedges, J.; Keil, R.G.; Benner, R. What happens to terrestrial organic matter in the ocean? *Org. Geochem.* **1997**, *27*, 195–212. [[CrossRef](#)]
11. Jiang, L.-Q.; Carter, B.R.; Feely, R.A.; Lauvset, S.K.; Olsen, A. Surface ocean pH and buffer capacity: Past, present and future. *Sci. Rep.* **2019**, *9*, 18624. [[CrossRef](#)] [[PubMed](#)]
12. Drake, T.W.; Raymond, P.A.; Spencer, R.G. Terrestrial carbon inputs to inland waters: A current synthesis of estimates and uncertainty. *Limnol. Oceanogr. Lett.* **2018**, *3*, 132–142. [[CrossRef](#)]
13. Hope, D.; Billett, M.F.; Cresser, M.S. A review of the export of carbon in river water: Fluxes and processes. *Environ. Pollut.* **1994**, *84*, 301–324. [[CrossRef](#)]
14. Bianchi, T.S. The role of terrestrially derived organic carbon in the coastal ocean: A changing paradigm and the priming effect. *Proc. Natl. Acad. Sci. USA* **2011**, *108*, 19473–19481. [[CrossRef](#)] [[PubMed](#)]
15. Raymond, P.A.; Bauer, J.E. Use of <sup>14</sup>C and <sup>13</sup>C natural abundances for evaluating riverine, estuarine, and coastal DOC and POC sources and cycling: A review and synthesis. *Org. Geochem.* **2001**, *32*, 469–485. [[CrossRef](#)]
16. Chaplot, V.; Mutema, M. Sources and main controls of dissolved organic and inorganic carbon in river basins: A worldwide meta-analysis. *J. Hydrol.* **2021**, *603*, 126941. [[CrossRef](#)]
17. Bianchi, T.S.; Filley, T.; Dria, K.; Hatcher, P.G. Temporal variability in sources of dissolved organic carbon in the lower Mississippi river. *Geochim. Cosmochim. Acta* **2004**, *68*, 959–967. [[CrossRef](#)]
18. Lohrenz, S.E.; Hopkinson, C.S.; Yang, J.; Tao, B.; Pan, S.; He, R.; Ren, W.; Tian, H.; Cai, W.-J.; Huang, W.-J. Century-long increasing trend and variability of dissolved organic carbon export from the Mississippi River basin driven by natural and anthropogenic forcing. *Glob. Biogeochem. Cycles* **2016**, *30*, 1288–1299. [[CrossRef](#)]
19. Meybeck, M. Riverine transport of atmospheric carbon: Sources, global typology and budget. *Water Air Soil Pollut.* **1993**, *70*, 443–463. [[CrossRef](#)]
20. Lauerwald, R.; Laruelle, G.G.; Hartmann, J.; Ciais, P.; Regnier, P.A. Spatial patterns in CO<sub>2</sub> evasion from the global river network. *Glob. Biogeochem. Cycles* **2015**, *29*, 534–554. [[CrossRef](#)]
21. Richey, J.E.; Melack, J.M.; Aufdenkampe, A.K.; Ballester, V.M.; Hess, L.L. Outgassing from Amazonian rivers and wetlands as a large tropical source of atmospheric CO<sub>2</sub>. *Nature* **2002**, *416*, 617–620. [[CrossRef](#)] [[PubMed](#)]
22. Regnier, P.; Friedlingstein, P.; Ciais, P.; MacKenzie, F.T.; Gruber, N.; Janssens, I.; Laruelle, G.; Lauerwald, R.; Luysaert, S.; Andersson, A.J.; et al. Anthropogenic perturbation of the carbon fluxes from land to ocean. *Nat. Geosci.* **2013**, *6*, 597–607. [[CrossRef](#)]
23. Richey, J.E. Pathways of Atmospheric CO<sub>2</sub> through Fluvial Systems. *Scope-Sci. Comm. Probl. Environ. Int. Counc. Sci. Unions* **2004**, *62*, 329–340.
24. Abril, G.; Bouillon, S.; Darchambeau, F.; Teodoru, C.R.; Marwick, T.R.; Tamooh, F.; Ochieng Omengo, F.; Geeraert, N.; Borges, A.V. Large overestimation of pCO<sub>2</sub> calculated from pH and alkalinity in acidic, organic-rich freshwaters. *Biogeosciences* **2015**, *12*, 67–78.
25. Duvert, C.; Butman, D.E.; Marx, A.; Ribolzi, O.; Hutley, L.B. CO<sub>2</sub> evasion along streams driven by groundwater inputs and geomorphic controls. *Nat. Geosci.* **2018**, *11*, 813–818. [[CrossRef](#)]
26. Ran, L.; Lu, X.X.; Richey, J.E.; Sun, H.; Han, J.; Yu, R.; Liao, S.; Yi, Q. Long-term spatial and temporal variation of CO<sub>2</sub> partial pressure in the Yellow River, China. *Biogeosciences* **2015**, *12*, 921–932. [[CrossRef](#)]
27. Zeng, F.W.; Masiello, C.A. Sources of CO<sub>2</sub> evasion from two subtropical rivers in North America. *Biogeochemistry* **2010**, *100*, 211–225. [[CrossRef](#)]
28. Butman, D.; Raymond, P.A. Significant efflux of carbon dioxide from streams and rivers in the United States. *Nat. Geosci.* **2011**, *4*, 839–842. [[CrossRef](#)]
29. Eccles, R.; Zhang, H.; Hamilton, D. A review of the effects of climate change on riverine flooding in subtropical and tropical regions. *J. Water Clim. Chang.* **2019**, *10*, 687–707. [[CrossRef](#)]
30. Cherchi, A.; Ambrizzi, T.; Behera, S.; Freitas, A.C.V.; Morioka, Y.; Zhou, T. The Response of Subtropical Highs to Climate Change. *Curr. Clim. Chang. Rep.* **2018**, *4*, 371–382. [[CrossRef](#)]

31. Osland, M.J.; Stevens, P.W.; Lamont, M.M.; Brusca, R.C.; Hart, K.M.; Waddle, J.H.; Langtimm, C.A.; Williams, C.M.; Seminoff, J.A. Tropicalization of temperate ecosystems in North America: The northward range expansion of tropical organisms in response to warming winter temperatures. *Glob. Chang. Biol.* **2021**, *27*, 3009–3034. [[CrossRef](#)] [[PubMed](#)]
32. Foley, J.A.; Kucharik, C.J.; Twine, T.E.; Coe, M.T.; Donner, S.D. Land use, land cover, and climate change across the Mississippi Basin: Impacts on selected land and water resources. *Ecosyst. Land Use Chang.* **2006**, *15*, 249–261. [[CrossRef](#)]
33. Tian, H.; Ren, W.; Yang, J.; Tao, B.; Cai, W.J.; Lohrenz, S.E.; Hopkinson, C.S.; Liu, M.; Yang, Q.; Lu, C.; et al. Climate extremes dominating seasonal and interannual variations in carbon export from the Mississippi River Basin. *Glob. Biogeochem. Cycles* **2015**, *29*, 1333–1347. [[CrossRef](#)]
34. Tavakol, A.; Rahmani, V.; Harrington, J. Evaluation of hot temperature extremes and heat waves in the Mississippi River Basin. *Atmospheric Res.* **2020**, *239*, 104907. [[CrossRef](#)]
35. Tang, C.; Dennis, R.; Cooter, E. *Water Temperature Changes in the Mississippi River Basin*; Internal. Presentation, RTP, NC; Environmental Protection Agency: Washington, DC, USA, 2015.
36. Raymond, P.A.; Oh, N.-H.; Turner, R.E.; Broussard, W. Anthropogenically enhanced fluxes of water and carbon from the Mississippi River. *Nature* **2008**, *451*, 449–452. [[CrossRef](#)]
37. Munoz, S.E.; Giosan, L.; Therrell, M.D.; Remo, J.W.F.; Shen, Z.; Sullivan, R.M.; Wiman, C.; O'Donnell, M.; Donnelly, J.P. Climatic control of Mississippi River flood hazard amplified by river engineering. *Nature* **2018**, *556*, 95–98. [[CrossRef](#)]
38. Tao, B.; Tian, H.; Ren, W.; Yang, J.; Yang, Q.; He, R.; Cai, W.; Lohrenz, S. Increasing Mississippi river discharge throughout the 21st century influenced by changes in climate, land use, and atmospheric CO<sub>2</sub>. *Geophys. Res. Lett.* **2014**, *41*, 4978–4986. [[CrossRef](#)]
39. Reiman, J.; Xu, Y.J. Dissolved carbon export and CO<sub>2</sub> outgassing from the lower Mississippi River—Implications of future river carbon fluxes. *J. Hydrol.* **2019**, *578*, 124093. [[CrossRef](#)]
40. Cai, Y.; Guo, L.; Wang, X.; Aiken, G. Abundance, stable isotopic composition, and export fluxes of DOC, POC, and DIC from the Lower Mississippi River during 2006–2008. *J. Geophys. Res. Biogeosci.* **2015**, *120*, 2273–2288. [[CrossRef](#)]
41. Joshi, S.; Xu, Y.J. Assessment of Suspended Sand Availability under Different Flow Conditions of the Lowermost Mississippi River at Tarbert Landing during 1973–2013. *Water* **2015**, *7*, 7022–7044. [[CrossRef](#)]
42. Rosen, T.; Xu, Y.J. Estimation of sedimentation rates in the distributary basin of the Mississippi River, the Atchafalaya River Basin, USA. *Hydrol. Res.* **2015**, *46*, 244–257. [[CrossRef](#)]
43. Vaughn, D.R.; Kellerman, A.M.; Wickland, K.P.; Striegl, R.G.; Podgorski, D.C.; Hawkings, J.R.; Nienhuis, J.H.; Dornblaser, M.M.; Spencer, R.G. Anthropogenic landcover impacts fluvial dissolved organic matter composition in the Upper Mississippi River Basin. *Biogeochemistry* **2021**, 1–25. [[CrossRef](#)]
44. Wohl, E.; Hall, R.O., Jr.; Lininger, K.B.; Sutfin, N.A.; Walters, D.M. Carbon dynamics of river corridors and the effects of human alterations. *Ecol. Monogr.* **2017**, *87*, 379–409. [[CrossRef](#)]
45. Repasch, M. Unexpected Consequences of River Engineering on the Carbon Cycle. *AGU Adv.* **2021**, *2*, e2021AV000402. [[CrossRef](#)]
46. Wang, B.; Xu, Y.J. Dynamics of 30 large channel bars in the Lower Mississippi River in response to river engineering from 1985 to 2015. *Geomorphology* **2018**, *300*, 31–44. [[CrossRef](#)]
47. Cai, W.-J.; Wang, Y. The chemistry, fluxes, and sources of carbon dioxide in the estuarine waters of the Satilla and Altamaha Rivers, Georgia. *Limnol. Oceanogr.* **1998**, *43*, 657–668. [[CrossRef](#)]
48. Dubois, K.D.; Lee, D.N.; Veizer, J. Isotopic constraints on alkalinity, dissolved organic carbon, and atmospheric carbon dioxide fluxes in the Mississippi River. *J. Geophys. Res. Biogeosci.* **2010**, *115*, G02018. [[CrossRef](#)]
49. Weiss, R. Carbon dioxide in water and seawater: The solubility of a non-ideal gas. *Mar. Chem.* **1974**, *2*, 203–215. [[CrossRef](#)]
50. Shen, Y.; Fichot, C.; Benner, R. Floodplain influence on dissolved organic matter composition and export from the Mississippi-Atchafalaya River system to the Gulf of Mexico. *Limnol. Oceanogr.* **2012**, *57*, 1149–1160. [[CrossRef](#)]
51. Soria-Reinoso, I.; Alcocer, J.; Sánchez-Carrillo, S.; García-Oliva, F.; Cuevas-Lara, D.; Cortés-Guzmán, D.; Oseguera, L.A. The Seasonal Dynamics of Organic and Inorganic Carbon along the Tropical Usumacinta River Basin (Mexico). *Water* **2022**, *14*, 2703. [[CrossRef](#)]
52. Dagg, M.J.; Bianchi, T.S.; Breed, G.A.; Cai, W.-J.; Duan, S.; Liu, H.; McKee, B.A.; Powell, R.T.; Stewart, C.M. Biogeochemical characteristics of the lower Mississippi River, USA, during June 2003. *Estuaries* **2005**, *28*, 664–674. [[CrossRef](#)]
53. Reiman, J.H.; Xu, Y.J. Diel Variability of pCO<sub>2</sub> and CO<sub>2</sub> Outgassing from the Lower Mississippi River: Implications for Riverine CO<sub>2</sub> Outgassing Estimation. *Water* **2018**, *11*, 43. [[CrossRef](#)]
54. Turner, R.E.; Rabalais, N.N.; Alexander, R.B.; McIsaac, G.; Howarth, R.W. Characterization of nutrient, organic carbon, and sediment loads and concentrations from the Mississippi River into the northern Gulf of Mexico. *Estuaries Coasts* **2007**, *30*, 773–790. [[CrossRef](#)]
55. Goulder, R. Seasonal variation in heterotrophic activity and population density of planktonic bacteria in a clean river. *J. Ecol.* **1980**, *68*, 349–363. [[CrossRef](#)]
56. Servais, P.; Garnier, J. Contribution of heterotrophic bacterial production to the carbon budget of the river Seine (France). *Microb. Ecol.* **1993**, *25*, 19–33. [[CrossRef](#)] [[PubMed](#)]
57. Voss, B.M.; Wickland, K.P.; Aiken, G.R.; Striegl, R.G. Biological and land use controls on the isotopic composition of aquatic carbon in the Upper Mississippi River Basin. *Glob. Biogeochem. Cycles* **2017**, *31*, 1271–1288. [[CrossRef](#)]
58. Cai, Y.; You, C.-F.; Wu, S.-F.; Cai, W.-J.; Guo, L. Seasonal variations in strontium and carbon isotope systematics in the Lower Mississippi River: Implications for chemical weathering. *Chem. Geol.* **2020**, *553*, 119810. [[CrossRef](#)]

- 
59. Calabrese, S.; Parolari, A.J.; Porporato, A. Hydrologic Transport of Dissolved Inorganic Carbon and Its Control on Chemical Weathering. *J. Geophys. Res. Earth Surf.* **2017**, *122*, 2016–2032. [[CrossRef](#)]
  60. Ford, T.W.; Chen, L.; Schoof, J.T. Variability and Transitions in Precipitation Extremes in the Midwest United States. *J. Hydrometeorol.* **2021**, *22*, 533–545. [[CrossRef](#)]