Groundwater Quality Assessment Using Fuzzy-AHP in An Giang Province of Vietnam

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Abstract: Along with rapid population growth in Vietnam, there is an increasing dependence on groundwater for various activities. An Giang province is known to be one of the agricultural intensification areas of The Vietnamese Mekong Delta (VMD). This study aimed to evaluate the spatiotemporal variation of groundwater quality for a period of ten years from 2009 to 2018 in An Giang. The weighted groundwater quality index (GWQI) was developed based on the fuzzy analytic hierarchy process (Fuzzy-AHP) for assigning weighted parameters. The results show that that shallow wells in the Northeast and Southeast regions of An Giang were mostly categorized under "bad water" quality with high arsenic (As) concentration over the years partly due to huge amounts of sediment deposition in monsoon season. Overall, the reason for the poor groundwater quality in An Giang was the combined effect of both natural and human activities. On the other hand, we detected high values of GWQI links with high As concentration in areas where people extract more groundwater for irrigation. Temporal variation of GWQI suggested that groundwater quality at eight wells has improved from 2009 to 2018 in the wet season as compared to the dry season. The reason behind the improvement of groundwater quality during wet season was the decrease in river discharge, which causes less deposition of suspended solids near the flood plains. Moreover, the filling of unused wells can reduce the movement of pollutants from unused wells to groundwater aquifers. Although there was not sufficient evidence to show the relationship between As and sediment concentration, the temporal reduction trend in river discharge and suspended solids was detected in An Giang. The understanding of groundwater quality can help policymakers protect and manage limited water resources in the long-term.

Keywords: Fuzzy-AHP; weighted arithmetic index; groundwater quality; arsenic

1. Introduction

Multiple criteria decision-making (MCDM) is a selective model used for evaluation of many complex decisions [1,2]. The fuzzy analytic hierarchy process (Fuzzy-AHP), based on AHP under fuzzy environment, is one of the most robust and flexible MCDM tools in the evaluation procedure [2,3]. In 1980, a simple AHP method was introduced based on a ratio scale [2,4,5]. The method was commonly applied in previous studies with advantages such as its simple and flexible model with a wide range

of usages [6,7]. The disadvantages of the AHP method include the uncertainty and ambiguity in expressing opinions, as the method depends on the decision maker's knowledge and experiences during the decision-making process. Moreover, among other factors, the AHP method does not contain feedback loops [8].

The fuzzy set was first developed by Zadeh in 1965 [5] and combined with Saaty's priority theory to reduce human ambiguity [9,10]. Later, the Fuzzy-AHP was further developed in order to overcome the uncertainty and ambiguity of criteria weights in deterministic and inflexible classifications [11]. Using the Fuzzy-AHP can provide a fuzzy number—interval judgment values rather than fixed or exact values [2]. This approach reduces uncertainty in assigned relative weight. As a result, the Fuzzy-AHP has been successfully used in many actual decision situation making such as energy alternatives selection [12,13], supplier selection [14], environmental sustainability evaluation [7], and water quality assessment [15–17]. Baghapour et al. [15] conducted the Fuzzy-AHP with fuzzy ordered weighted averaging (FOWA) for developing of the groundwater quality index (GWQI). They further revealed that it could effectively calculate weights of groundwater quality parameters. Deng et al. [18] used fuzzy number scales with pair-wise comparisons for solving decision problems involving qualitative data very effectively in Australia. Two of the fuzzy pair-wise comparisons and FOWA were used for different water resource assessments, such as prioritizing the restoration strategies for Lake Urmia, Iran to avoid shrinkage [17], evaporation estimation [19,20], water consumption prediction [21], rainfall-runoff forecasting and modelling [22–24], and evaluation of groundwater pollution using GWQI [25].

To assess water quality, various multivariate statistical analyses were successfully applied in many previous studies, such as groundwater modelling using the principal component analysis (PCA) technique [26–28]. However, PCA can only reduce the dimensionality of large data sets based on the variation of variables in the new coordinate axis and the modelling approach required detailed data [28,29]. Whereas, the powerful water quality index (WQI) tool can be used to summarize a huge number of parameters into a single index [30]. The WQI method was first developed by Horton in 1965 [31] by using ten parameters of water quality. It has since been widely applied in Asian countries [32]. In 1970, Brown et al. [33] introduced a new WQI which is similar to the index of Horton. Later, many modifications were made for WQI such as the weight arithmetic water quality index (WAWQI), National Sanitation Foundation water quality index (NSFWQI), Oregon water quality index (GWQI), and WQI$_{al}$ for aquatic life recommended by Mekong River Commission (MRC) [32,34,35].

Many studies used the top-down approach, in which the fixed weight of groundwater quality is used to calculate the GWQI. For example, Asadi et al. [36] and Maheswaran et al. [37] used the weight of groundwater quality of WHO to calculate the GWQI in the Hyderabad and Salem districts of India. There are many water quality parameters that contribute to groundwater pollution in the study area, each with its own important value. In this study, we used the bottom-up approach at the local level in terms of weighted values of water quality parameters to find out the locally important groundwater parameters in An Giang province of the Vietnamese Mekong Delta (VMD). The Fuzzy-AHP technique, a systematic method, is an effective tool to weigh multiple parameters in classifying the clear groundwater quality based on GWQI. This study focused on estimating the groundwater WQI (GWQI) by using Fuzzy-AHP to assess groundwater quality in An Giang. The Fuzzy-AHP approach was used due to its computational effectiveness in weighted values of water quality selection and its ability to reduce uncertainty from experts' opinions [38,39]. The pair-wise comparison with triangular fuzzy numbers, along with the weighted arithmetic index methods were used to calculate the GWQI. Inverse distance weighted (IDW) interpolation was used to display variation in spatial and temporal parameters which was applied in many studies [28,40,41]. Clear understanding about groundwater quality changes in time and space is very essential in the VMD and An Giang.

In the Vietnamese Mekong Delta (VMD), people rely both on surface and groundwater resources not only for irrigation and aquaculture, but also for daily domestic usage. However, the poor surface water quality with high concentrations of nutrients in secondary canals was found in An Giang [29]. The reason is the release of untreated agricultural runoff from rice intensification inside the full-dike

protected area in An Giang [29]; as these pose serious health risks if the water is consumed without adequate treatment [42,43]. Therefore, groundwater sources serve as one of the main supplies for domestic water use, and partly for irrigation, due to surface water quality often exceeding the permission of the Vietnamese standards for domestic water supply in An Giang in recent years [44]. However, a few studies on groundwater quality assessment were conducted, such as Thu et al. [45] who investigated sources of As contamination in the groundwater in An Phu of An Giang province; Anh and Giao [46] evaluated the impact of water quality on the health risks in An Giang province, especially with regard to As concentration. To the best of our knowledge, there are no holistic studies that evaluate groundwater quality in the context of agricultural intensification and its evolution process using the GWQI using the fuzzy number in An Giang. Therefore, in this study, we clarify whether rice intensification has an effect on groundwater quality or not. The findings of this study can provide the status of groundwater quality at a spatiotemporal scale, which would be useful for decision-makers to design timely management plans for water resources and thus minimize any further adverse effect on human's health.

2. Study Area

An Giang province encompasses an area of about 3406 km^2 with a total population of about 2 million people in 2017. The province is located in the upper region of the VMD, which is comprised of a dense river network system (Figure 1). The wet season starts from May to November, and the dry season occurs during December through April (Figure 2). The mean annual rainfall is about 1400 mm, of which 90% occurs during the wet season. An Giang province is part of the agricultural intensification region of the VMD, where the water regime is mainly under human control through sluice gates, canals, and dike systems [47–49]. The land use/land cover (LULC) map of An Giang shows the percentage of various LULC classes in 2018 with a triple, double, and single rice crop, and other classes that cover 46.6%, 24.7%, 7.3%, and 21.4%, respectively [50]. The highly irrigated triple and double rice crops occur inside the dike system using surface water, which has negative impacts on surface and groundwater quality especially in the full-dike system [29,46]. Consequently, the health of two million people in An Giang may be at risk [45,46]. Moreover, the single rice cropping was dominant in the region that is far away from the main river with less of a river network system. The single rice cropping area also includes cultivation of rainfed rice in the wet season and vegetables in the dry season using groundwater. River discharge data show a decreasing trend in the wet season and a slightly increasing trend in the dry season from 2009 to 2017 (Figure 1).

An Giang province belongs to the Southern part of Vietnam with five main aquifers, namely, the Upper Pleistocene aquifer (qp$_3$), Middle-Upper Pleistocene aquifer (qp$_{2-3}$), Lower Pleistocene aquifer (qp1), Medium Pliocene aquifer (n$_{22}$), and Lower Pliocene aquifer (n$_{21}$) [51]. The groundwater was mostly extracted from the Pleistocene and Holocene aquifers since deposited outcrop on the surface was found, which is supplied for domestic and irrigation purposes [52]. In 2014, the total number of existing groundwater wells was 4746, which included 233 unused/discontinued wells. Out of 4513 wells, there were 553 wells used for irrigation and 3960 wells for the domestic water supply [53–55].
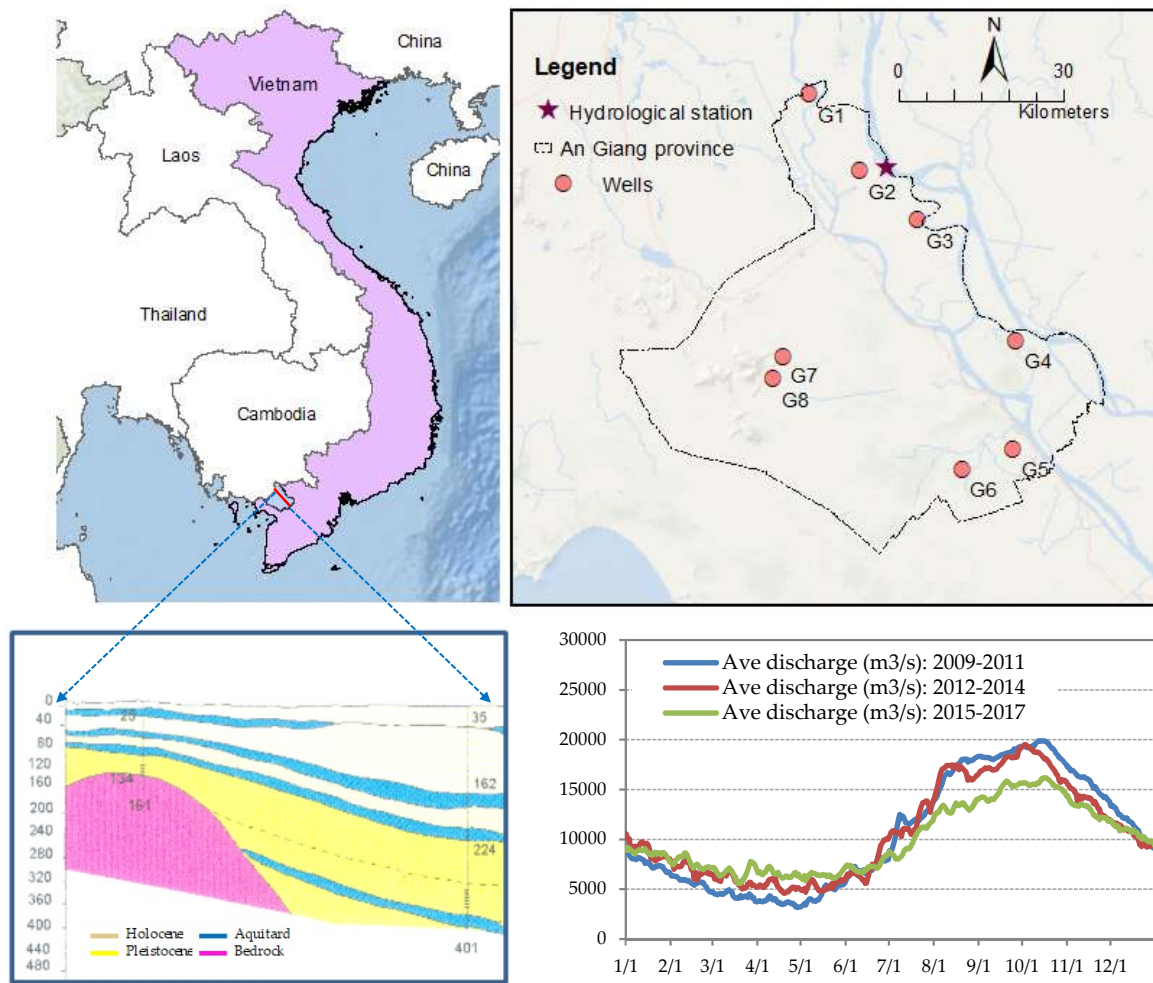
Figure 1. Study area map with the location of monitoring wells and Tan Chau hydrological station on Mekong River. Discharge data were collected from the Southern Regional Hydrometeorology Center in Vietnam, during 2009–2017. The geological data were collected from the Southern Geological Division geohydrology and Engineering Geology.
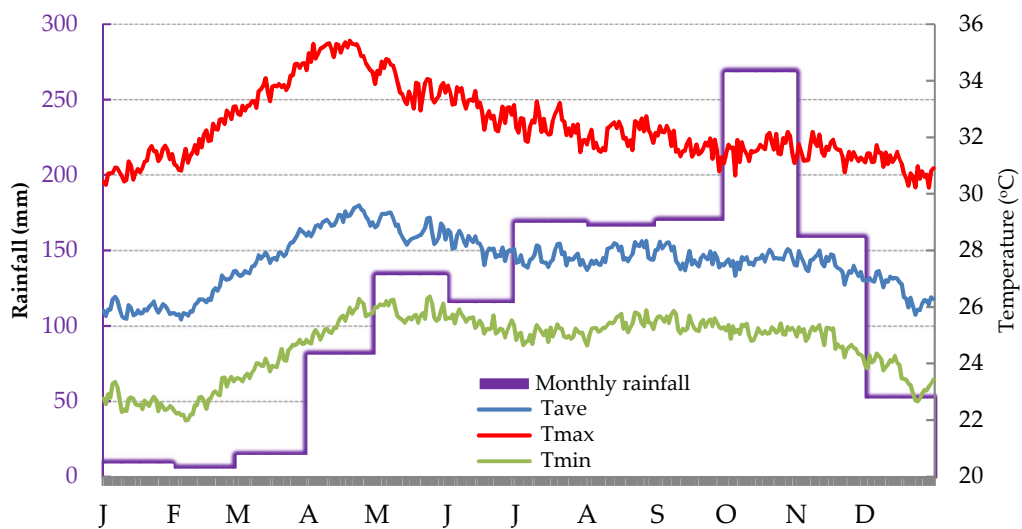

Figure 2. Average monthly rainfall, daily mean temperature (Tave), daily maximum temperature (Tmax), and daily minimum temperature (Tmin) during the period 1978–2016, collected from Southern Regional Hydrometeorology Center in Vietnam.

3. Data and Methodology

Groundwater quality data of eight wells were collected from 2009 to 2018 in the wet and dry seasons from the Department of Natural Resources and Environment of An Giang (DONRE) (Figure 1). The depth of G6 and G8 wells (deep well), which ranged from 80–300 m below ground level, lie in the Pleistocene aquifer. Whereas, depth of G1, G2, G3, G4, G5, and G7 wells (shallow wells), which are exploited at the average depth of 50 m, lie in the Holocene aquifer. Furthermore, wells G5 and G6 were exploited mainly for industrial zones, while the other six wells' supply was used for domestic uses and irrigation. Six groundwater quality parameters: As, NO_3, NH_4^+, $CaCO_3$, total Fe, and pH were analyzed in March and September each year using standard methodology [56].

Figure 3 shows the process to determine the relative weight for each groundwater quality parameter in order to calculate the GWQI. We conducted Fuzzy-AHP, which was developed to weight criteria in decision-making by using the output of the experts' opinions. The weighted value was assigned by pair-wise comparison for each of the six groundwater quality parameters, including As, NO_3, pH, NH_4^+, $CaCO_3$, and total Fe. Twenty experts were clustered in 4 groups and the experts in each group compared the parameters by pair-wise variables comparison using fuzzy triangular number scales and four scenarios of pair-wise were obtained. The Fuzzy-AHP process of weighting was accomplished in four steps and GWQI was then calculated [16]. We used inverse distance weighting (IDW) interpolation to display the results of the GWQI.
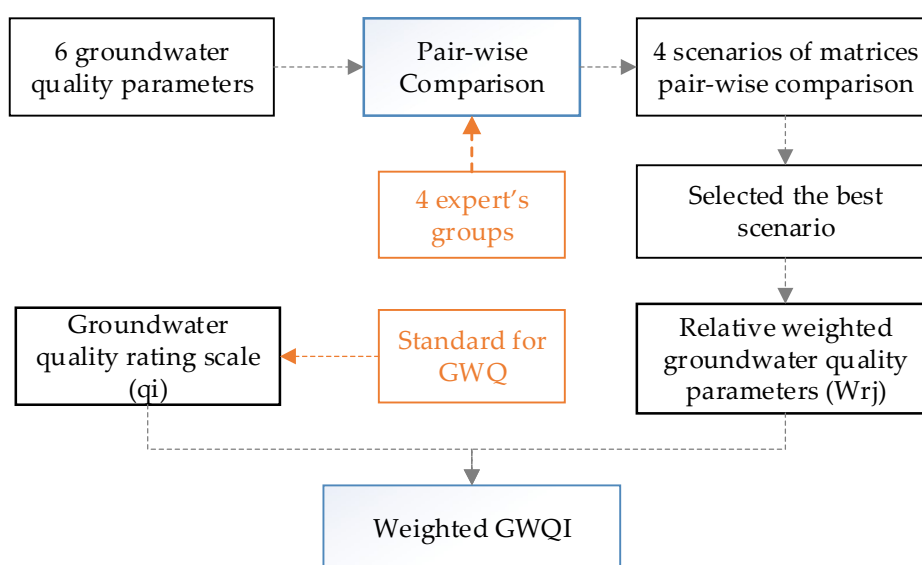
Figure 3. Flowchart of study progress of weighted groundwater quality indices by using the fuzzy analytic hierarchy process (Fuzzy-AHP).

3.1. The Fuzzy- AHP with Pair-Wise Comparison Approach

To achieve relative weights of groundwater quality parameters in the Fuzzy-AHP, the process was divided into four steps, namely: hierarchy construction development, pair-wise comparisons represented by fuzzy numbers, the fuzzy triangular number calculation, and normalized weights of parameters establishment.

3.1.1. Step 1: Hierarchy Construction Development

We conducted the hierarchy structure composed of three levels (Figure 4). The first level was the overall objective to determine the quantification of the potential of groundwater resources; the second level was the comparison of water quality parameters. We used a fuzzy triangular number scale which was transferred from linguistic terms corresponding to Saaty's scale (1980) in Table 1 through

pair-wise comparison matrices [4,57]. The higher weighting of a parameter shows high importance of that parameter. Finally, the groundwater quality was assessed based on five classes of GQWI.
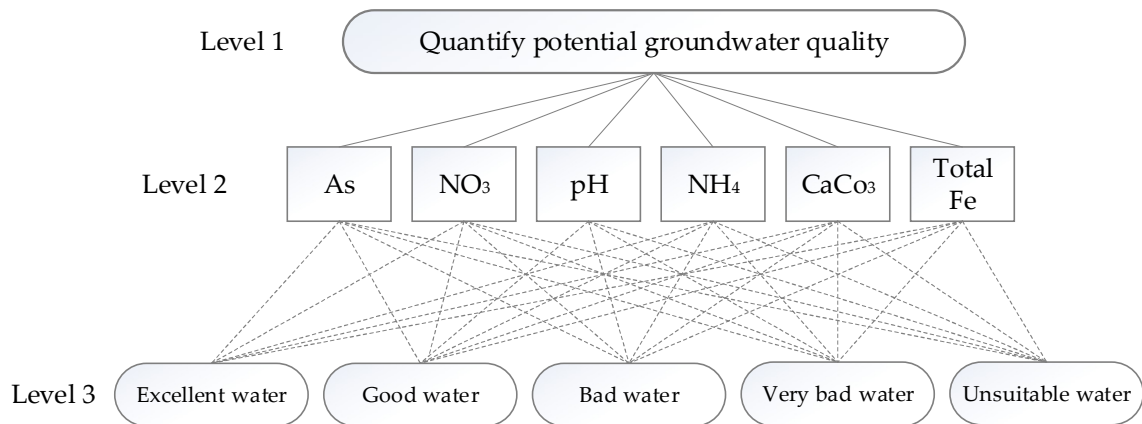
Figure 4. The hierarchy structure for the performance evaluation process of groundwater quality assessment.

Table 1. Linguistic terms and the corresponding triangular fuzzy scale.

Saaty's Scale	Linguistic Terms	Fuzzy Triangular Scale (TFN)
1	Equal importance	(1,1,1)
3	Moderate importance of one over another	(2,3,4)
5	Essential or strong importance	(4,5,6)
7	Demonstrated importance	(6,7,8)
9	Extreme importance	(9,9,9)
2, 4, 6, 9	Intermediate values between two adjacent judgments	(1,2,3), (3,4,5), (5,6,7), and (7,8,9)

3.1.2. Step 2: The Pair-Wise Comparisons Represented by Fuzzy Numbers

Decision making was based on the opinions of five experts in each group (professors in universities, government experts, nongovernment experts, and water supply companies) [15]. The fuzzy triangular number scales were used to compare between two parameters and find out the more important parameter. The parameters were compared by transferring them from linguistic terms to fuzzy number (Table 1). The pair-wise contribution matrix is expressed in Equation (Equation (1)). The sensitivity assessment was conducted to reduce the uncertainty of the experts' opinions by comparison of parameters' weights in four scenarios based on mean values and standard deviation (SD). The lowest SD in the weight change was selected as the optimal relative weight.

$$\widetilde{A}^k = \begin{bmatrix} \widetilde{d}^k_{11} & \widetilde{d}^k_{12} & \widetilde{d}^k_{13} & \widetilde{d}^k_{1n} \\ \widetilde{d}^k_{21} & \widetilde{d}^k_{22} & \widetilde{d}^k_{23} & \widetilde{d}^k_{2n} \\ \ldots & \ldots & \ldots & \ldots \\ \widetilde{d}^k_{n1} & \widetilde{d}^k_{n2} & \widetilde{d}^k_{n3} & \widetilde{d}^k_{mn} \end{bmatrix} \tag{1}$$

where, $\widetilde{A}^k$: Fuzzy triangle number, tilde (~): the triangular number; $\widetilde{d}^k_{ij}$ represents the *k*th decision maker's preference of *i*th criterion over *j*th criterion; $\widetilde{d}_{ij} = \frac{\sum_{k=1}^{K} \widetilde{d}^k_{ij}}{K}$ is the average decision-maker.

3.1.3. Step 3: Determine the Fuzzy Triangular Number

The geometric mean technique for computing the weights (W_i) was extended to the fuzzy positive reciprocal matrices [58] and comparison values of each parameter were calculated as shown in Equation (2).

$$\widetilde{r_i} = \left(\prod_{j=1}^{n} \widetilde{d}_{ij} \right)^{\frac{1}{n}} \tag{2}$$

where, $i = 1, 2, \ldots n$; $\widetilde{r_i}$: triangular values; later, replacing the fuzzy triangular number by (-1) power of summation vector and finally making it in increasing order.

3.1.4. Step 4: The Normalized Weights of Criteria

The normalized weight (N_i) was estimated by the corresponding normalized row mean (Equation (3)).

$$N_i = \frac{M_i}{\sum_{i=1}^{n} M_i} \tag{3}$$

where, $M_i = \frac{lw_i + mw_i + uw_i}{3}$; in which $\widetilde{w}_i = \widetilde{r}_i x (\widetilde{r}_1 + \widetilde{r}_2 + \widetilde{r}_3)$; M_i is non-fuzzy number i, $\widetilde{w}_i$ is the fuzzy weight of criterion i; $\widetilde{r}_i$ is reverse vector.

3.2. Groundwater Quality Index (GWQI)

The estimation of the GWQI was based on parameter weighting. A weighted value was used by using pair-wise comparison to each other and this assigned weighted value played a major role in the calculation of the index value. Table 2 shows that the limited threshold of quality values based on the National Technical Regulation on Groundwater Quality of Vietnam (standard number 09-MT:2015/BTNMT) [59], and this was used to calculate the quality rating scale. Due to the considered parameters having different units and ranges of values, all parameters had to be turned into sub-indices expressed on a single scale. Thus, we calculated based on the following: Relative weight (W_{rj}), quality rating scale (q_i), and GWQI.

Table 2. Groundwater quality parameters, units, and limited threshold values based on the Vietnamese standard.

Parameters	Units	Limited Threshold Values
As	mg/L	0.05
NO_3	mg/L	15
pH	-	5.5–8.5
NH_4^+	mg/L	1
$CaCO_3$	mg/L	500
Total Fe	mg/L	5

3.2.1. Relative Weight Calculation (W_{rj})

$$W_{rj} = \frac{W_j}{\sum_{j=1}^{n} W_j}, \tag{4}$$

where, W_{rj}: the relative weight for the nth parameters and $\sum W_{rj} = 1$; w_j: the weight of each parameter, a number between 0 to 1; n: number of parameters.

3.2.2. The Quality Rating Scale Calculation (q_j)

$$q_j = \frac{c_{mj}}{c_{sj}} \times 100, \tag{5}$$

where, q_j: the quality rating scale for the nth variable; c_{mj}: the concentration of each parameter in each sample (mg/L), c_{sj} groundwater threshold values were specified by [59] for each parameter (mg/L).

3.2.3. Groundwater Quality Index Calculation

In this study, we calculated the GWQI based on the weighted arithmetic index method recommended by the World Health Organization (WHO) using Equation (6). The weighted arithmetic index method can be used with different parameters. It is flexible for assessment and management of water quality because it was applied in many previous studies [16,60,61], Groundwater quality was classified based on rating values of GWQI in Table 3.

$$\text{GWQI} = \sum\nolimits_{j=1}^{n} \left(W_{rj} \times q_j \right), \tag{6}$$

where, GWQI: Groundwater quality index, a number between 0 to 300 units was divided into the five grade scales as shown in Table 3.

Table 3. Water quality classification based on the weighted groundwater quality index (GWQI) for human consumption.

GWQI	<50	50–100	100–200	200–300	>300
Quality classification	Excellent	Good	Bad	Very bad	Unsuitable for drinking

4. Results

4.1. The Fuzzy-AHP with Pair-Wise Comparison

To determine the weighted parameters, 20 experts in four groups were asked to enter into the pair-wise comparison matrix of AHP defined weighting parameters in four scenarios. Table 4 shows pair-wise comparisons in scenario 3, in which As, $CaCO_3$, and Fe were of "equal importance" while As was of "moderate importance" with NH_4^+, and NO_3, and "strong importance" with pH. Then, four scenarios were defined to estimate absolute group weights of different parameters obtained from pair-wise comparison as shown in Table 5. Next, the optimal status of decision-making powers were determined by sensitivity analysis. To reduce the uncertainty of the experts' opinions, we compared all the possible scenarios and calculated sensitivity values for all 6 paired scenarios. Table 6 shows the sensitivity comparison among six paired scenarios. It is clearly visible that scenario 3 performed best with the lowest SD (SD = 0.007) in the sensitivity analysis. Table 7 shows the relative weight factors of different water quality parameters. As together with Fe concentrations and pH values were the most and least important parameters, respectively (Table 7).

Table 4. Pair-wise comparisons in scenario 3.

Criteria	As	NO_3	pH	NH_4^+	$CaCO_3$	Fe
As	(1,1,1)	(2,3,4)	(4,5,6)	(2,3,4)	(1,1,1)	(1,1,1)
NO_3	(1/4,1/3,1/2)	(1,1,1)	(4,5,6)	(1,1,1)	(1/4,1/3,1/2)	(1/4,1/3,1/2)
pH	(1/6,1/5,1/4)	(1/6,1/5,1/4)	(1,1,1)	(1/6,1/5,1/4)	(1/6,1/5,1/4)	(1/6,1/5,1/4)
NH_4^+	(1/4,1/3,1/2)	(1,1,1)	(4,5,6)	(1,1,1)	(1,1,1)	(1/4,1/3,1/2)
$CaCO_3$	(1,1,1)	(2,3,4)	(4,5,6)	(1,1,1)	(1,1,1)	(1,1,1)
Fe	(1,1,1)	(2,3,4)	(4,5,6)	(2,3,4)	(1,1,1)	(1,1,1)

Table 5. Absolute group weights of parameters obtained from the pair-wise comparison.

Parameters	w_j			
	S_1	S_2	S_3	S_4
As	0.396	0.431	0.265	0.253
NO_3^-	0.075	0.095	0.111	0.136
pH	0.024	0.030	0.038	0.048
NH_4^+	0.068	0.106	0.131	0.125
$CaCO_3$	0.163	0.166	0.219	0.190
Total Fe	0.319	0.241	0.265	0.275
Sum	1.046	1.069	1.028	1.026
Mean	0.174	0.178	0.171	0.171
SD	0.151	0.143	0.093	0.085

Table 6. Scenarios' sensitivity analyses results.

Analysis	Compared Scenarios			Mean	SD
1	S_1	to	S_2	0.0037	0.008
2	S_1	to	S_3	0.0031	0.058
3	S_1	to	S_4	0.0034	0.065
4	S_2	to	S_3	0.0068	0.050
5	S_2	to	S_4	0.0071	0.058
6	S_3	to	S_4	0.0003	0.007
			Min	0.0003	0.007

Table 7. Relative weight factors of different water quality parameters.

Parameters	w_j
As	0.258
NO_3^-	0.107
pH	0.037
NH_4^+	0.127
$CaCO_3$	0.214
Total Fe	0.258

4.2. Groundwater Quality Index (GWQI)

The temporal average WQI changes and percentage of the groundwater samples for different categories for the period of 2009–2018 were presented in Figures 5–8, respectively. Observed groundwater quality was better in the dry season as compared to the wet season except for NH_4^+ (Tables A1 and A2 in Appendix A). Minh et al. [29] also found NH_4^+ concentration in surface water in the wet season was higher than in the dry season in An Giang. However, we found that the improvement in water quality took place since 2009. The wide range of the GWQI values from 2009 to 2018 show the best and the worst water quality observed at G_6 and G_4 in two seasons, respectively. In 2018, the GWQI values at G_6 were found to be 47 and 41 units in the dry and wet seasons respectively, while the GWQI at G_4 were detected to be 132 and 76 units in the dry season and the wet season, respectively. Net groundwater quality improvement occurred in most wells during the years 2009–2018 except G_1, G_3, and G_4 in the dry season. For example, in the dry season the "Very bad water" group of G_1, G_3, and G_4 experienced 202, 176, and 885 units, respectively in 2009, which improved to be "good" quality level in 2014 at G1 (GWQI = 67 units) compared to G_3 and G_4, with WQI values of 102 and 161 units to be "bad water" quality, respectively (Figure 6). However, G_3, an G_4 became "bad water" in 2016 and 2018 while G_1 became "unsuitable water for drinking" in 2018. Water quality was significantly improved from "very bad water" quality level at G_1, G_3, G_7, and G_8 in the wet season to the "good water" and "excellent water" quality level in 2016 and 2018 in the dry season. The water quality of G_4 had not been improved during 2009–2018 in both of the seasons. The water quality in

well G_5 in the wet season first declined from 2009 to 2012, but it has not recovered to achieve "good water" level in 2014 and in 2018. The well G_7 and G_8 improved significantly in quality during 2009 to 2018 in both of the seasons (Tables A3 and A4 in Appendix B). Figures 5 and 8 show the percentage of wells with groundwater quality based on GWQI in the dry and wet seasons, respectively. About 39% of wells were in the "bad water" group from 2009 to 2012, but only 25% of wells were considered to be at "bad water" level in the dry season in 2018.
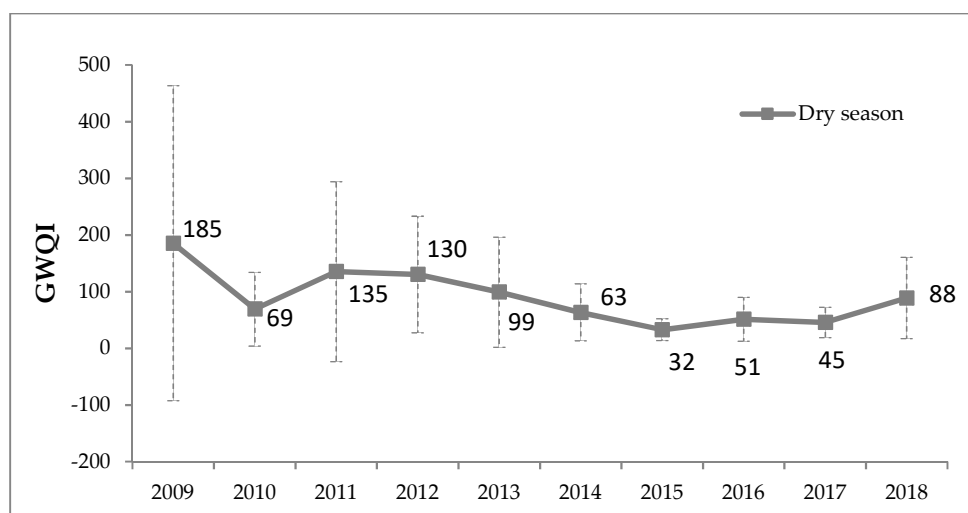
Figure 5. The average of GWQI units of eight wells in the dry season, during 2009–2018.

4.3. Groundwater Quality Assessment

GWQI spatial distribution maps were prepared for the years of 2009, 2013, 2014, and 2018, and classified in accordance with GWQI rating system shown in Figure 9. The spatial distribution maps for As are displayed in Figure A1 in Appendix C. The shallow wells (G_1, G_3, and G_4) had the lowest water quality in different years. Being shallow wells, these sampling locations encounter an aquitard (clay) layer as shown in Figure 1, which enhances the reducing condition inside the aquifer and triggers the arsenic mobilization when exposed to oxidative conditions by oxyhydroxide reduction theory [62]. However, the water quality of the G_1 well improved in the two seasons since 2014 and decreased again in the dry season of 2018. The shallow wells G_1 and G_3 are located at the Northeastern part of An Giang, while G_4 is located in the Southeastern part of An Giang. The wells G_1, G_3, and G4 are located between, and close to, the Mekong and Bassac Rivers. It is well-reported that mobilization is prominent in the river flood plain [63]. Therefore, As concentration was the highest for well numbers G_1 and G_4, as they are located in the vicinity of the river plain. Also, As had the highest weightage among all water quality parameters. High As concentration led to poor water quality.
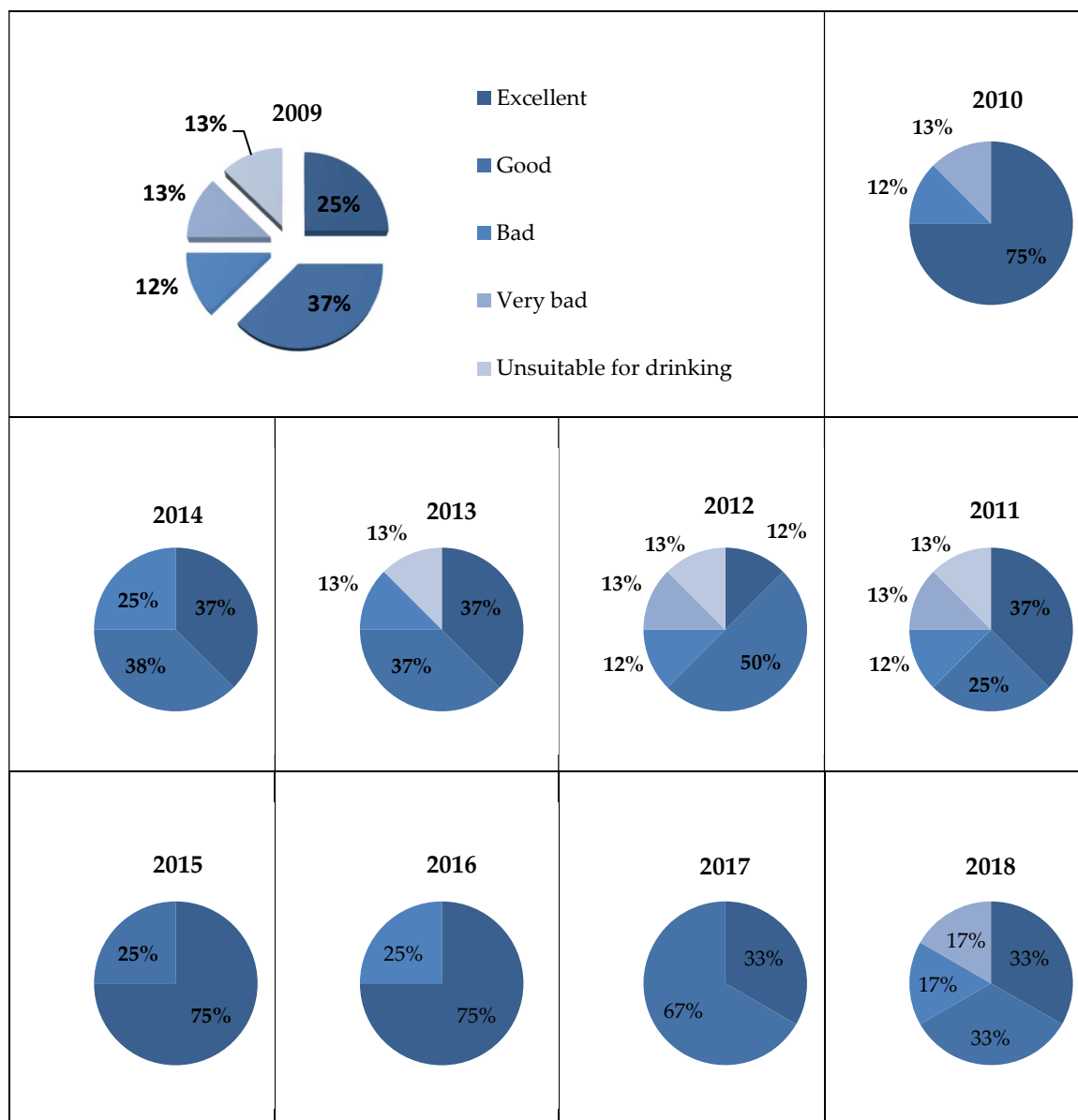
Figure 6. Percentage of groundwater quality at eight wells based on GWQI in the dry season during 2009–2018.
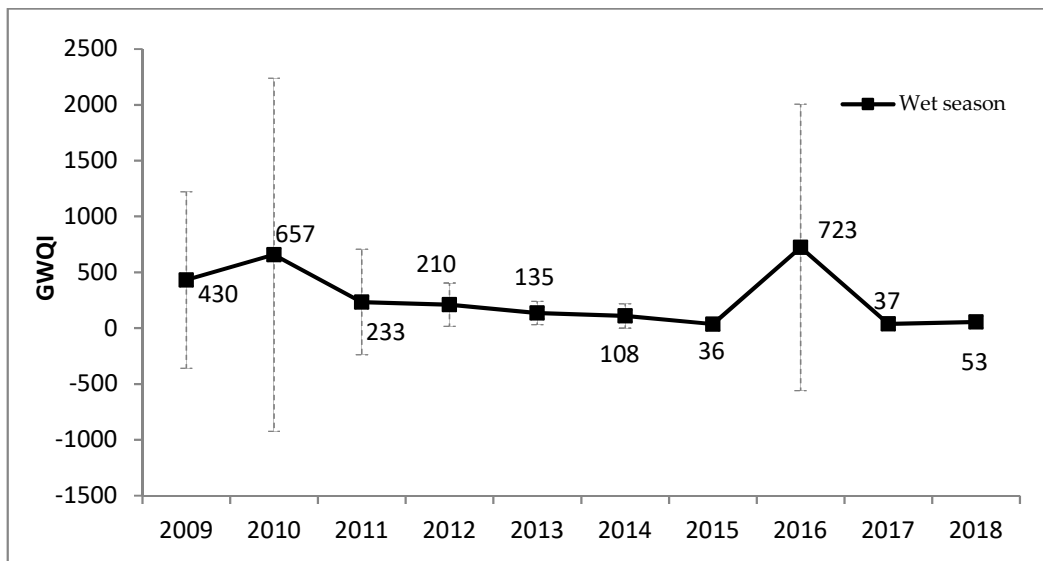
Figure 7. The average of GWQI unit of eight wells in the wet season, during 2009–2018.
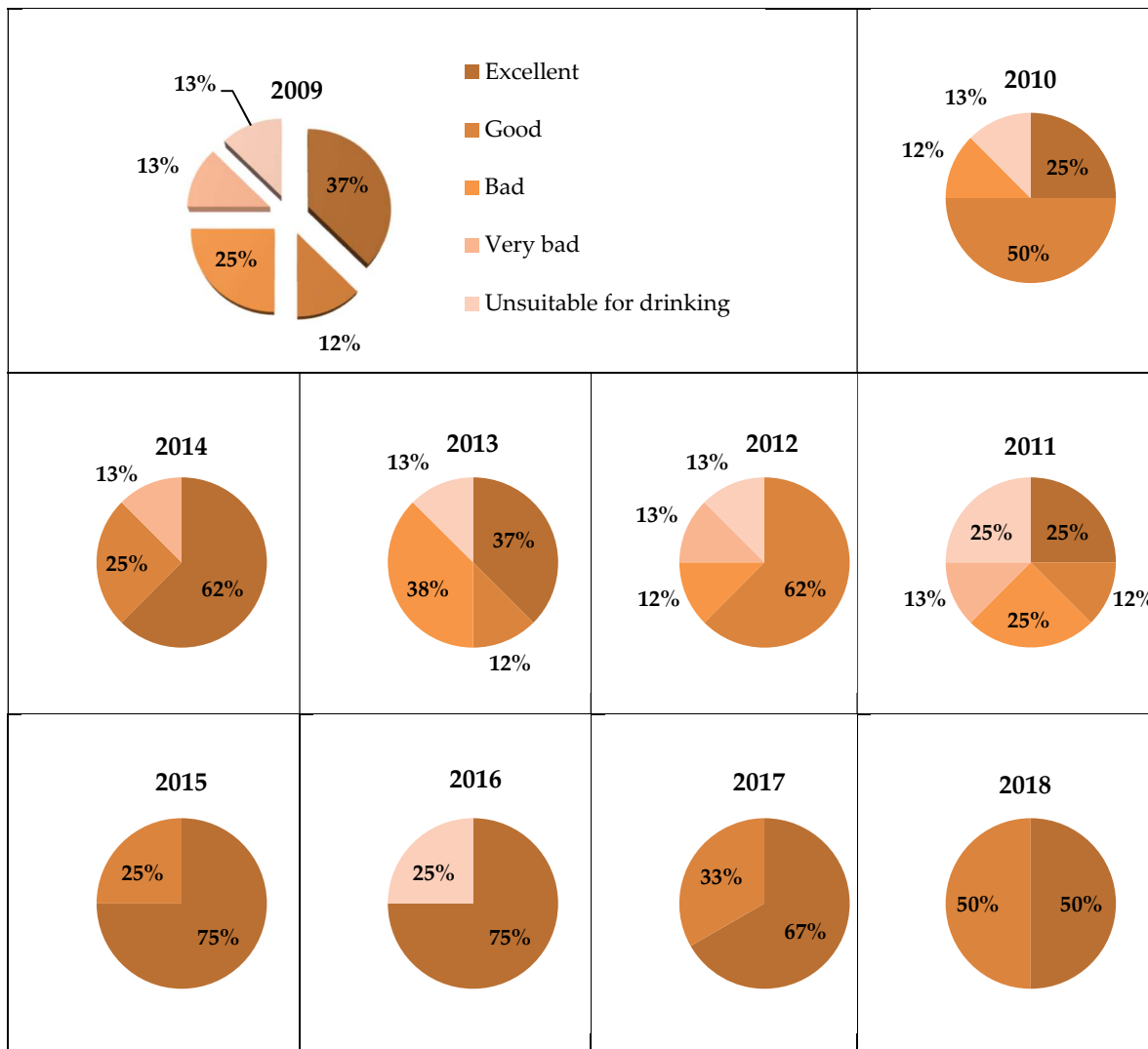

Figure 8. Percentage of groundwater quality at eight wells based on GWQI in the wet season, during 2009–2018.

In general, As values decreased from 2009 to 2018. The extremely high As concentrations were detected during 2009 and 2010 in both seasons. The highest concentration of As was found in G_4 in 2009 at Cho Moi district located in the Southeast region. The second-highest concentration levels of As were identified at well G_1 at An Phu district in the Northeast region. Thu et al. [45], Chakraborti et al. [64], and Anh et al. [46] also showed a high concentration of As in the VMD such as An Giang, Long An, and Dong Thap provinces. Arsenic above 0.01 mg/L was typically found in wells with aquifer of Holocene rather than Pleistocene aquifer [64].
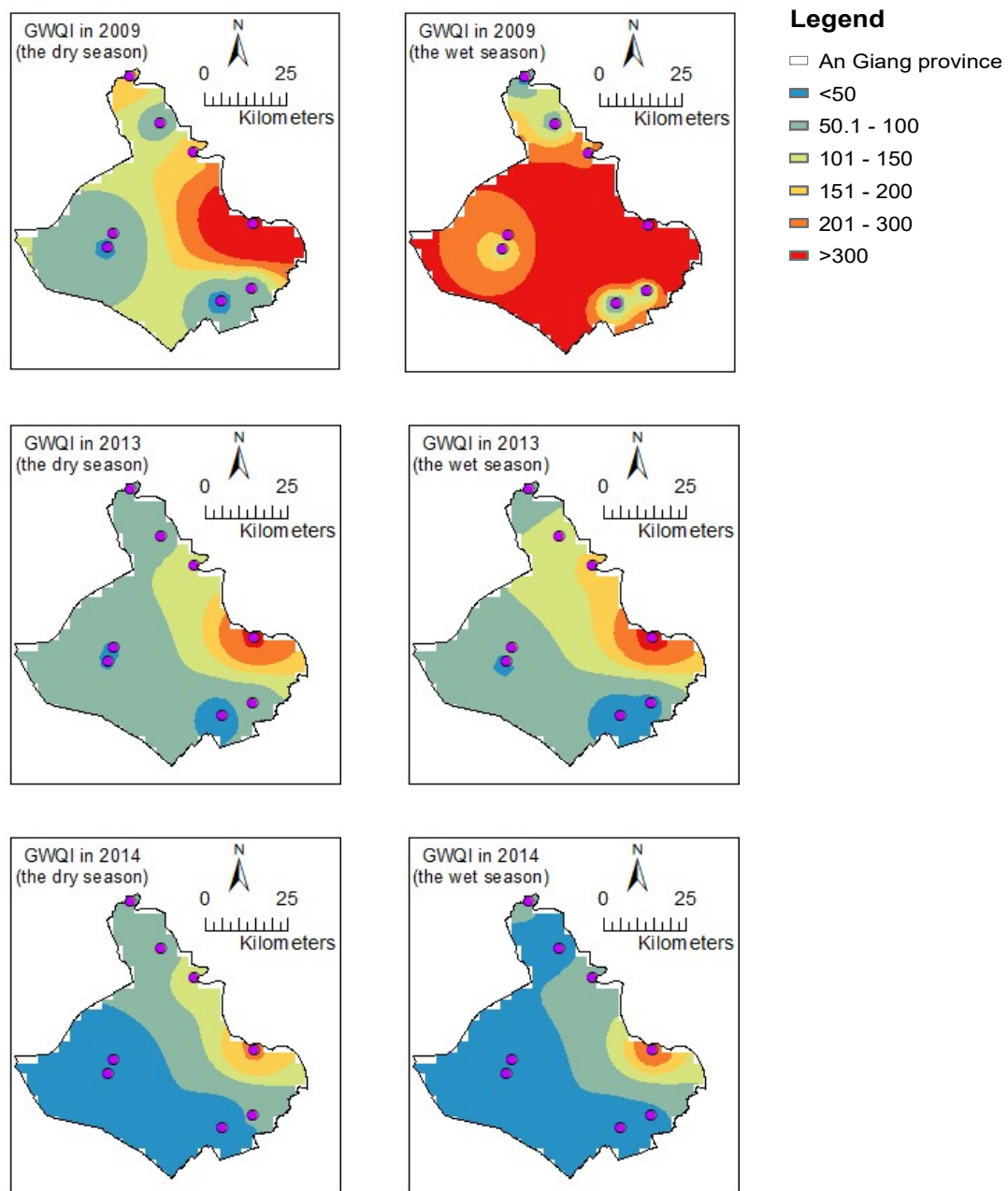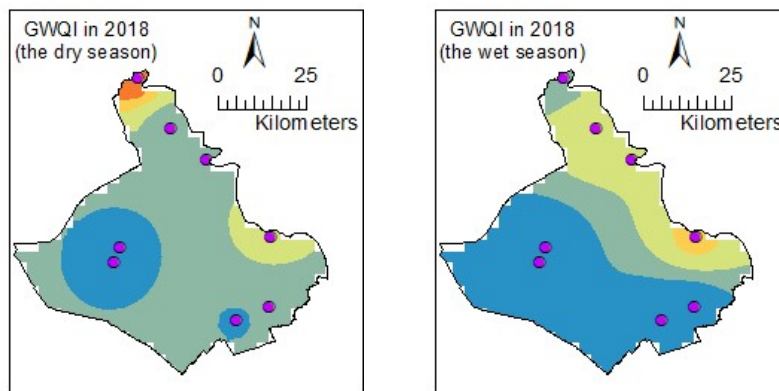

Figure 9. *Cont.*

Figure 9. The groundwater quality index (GWQI) in An Giang. Notes: <50 (Excellent water), 50–100 (Good water), 100–200 (Bad water), 200–300 (Very bad water), and >300 (Unsuitable for drinking).

5. Discussion

The groundwater sources serve as the main supply for domestic use, and partially for irrigation purposes in areas with less river networks. Understanding groundwater quality can serve policymakers to protect and effectively manage the limited water resources available in the region.

For this study, the Fuzzy-AHP considered As and total iron as the most important factors that affect the GWQI, with a weighted parameter of approximately $w_i = 0.258$. Temporal variation of GWQI suggests that the trend of groundwater quality at eight wells improved from 2009 to 2018 due to less sediment deposition and effective environmental management policies in An Giang. The construction of hydropower plants in the upper Mekong River basin caused a decrease in river discharge and sediment deposition in the study area [48,65]. Ngoc et al. [65] predicted the reduction of sediment at Tan Chau station in the future due to the expansion of hydropower plants. Arsenic contaminant is often found as a result of natural conditions and human activities in Asian countries [66]. Geogenically, As concentration in groundwater is found in young Quaternary deltaic and alluvial sediments and As concentration also related to sediment concentration [66,67]. Moreover, Chuan et al. [67] found the high relationship between sediment and As concentration in China. Figure 10 shows the decreasing trend of suspended solids in surface water from 2005 to 2017 at Long Xuyen station in An Giang. This result was consistent with the reduction of As concentration in the wet season. Besides the water resources, regulation related to water and environmental protection was effectively implemented in An Giang. The decision 1566/QĐ-UBND of "Environmental Protection Planning of An Giang Province up to 2020", which was issued in 2011 limits the use of chemical fertilizers in agricultural activities, and requires waste treatment systems for raising livestock, poultry and aquatic products basing on national standards and environmental sanitation.
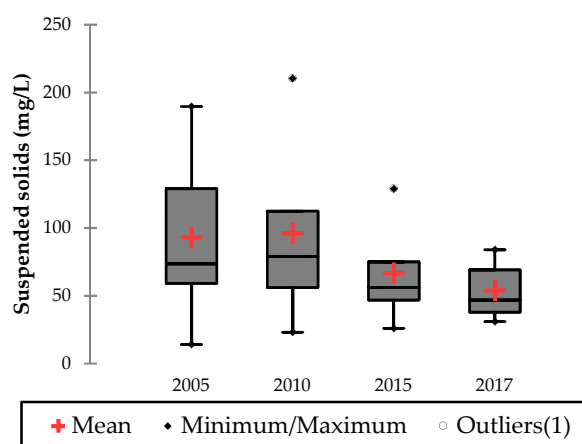
Figure 10. Box plot shows temporal changes in suspended solids in 2005, 2010, 2015, and 2017 at Long Xuyen station, An Giang. Data of suspended solids were obtained from the Southern Regional Hydrometeorological Center in Hochiminh city, Vietnam.

The quality of groundwater greatly improved from 2014 to 2018 as compared to 2009 to 2013. In 2009, four out of eight wells were identified as "bad water" quality to "unsuitable water for drinking" during wet season. In 2013, 75% and 50% of wells achieved "good water" to "excellent water" levels in the dry season and wet season, respectively. In 2018, water quality of six wells in the dry season and seven wells in the wet season achieved "good water" to "excellent water" quality (accounting for 66% in the dry season and 100% in the wet season). However, groundwater quality at well G_4 at Cho Moi district was considered mostly "unsuitable water for drinking" from 2009 to 2014 and became "good water" in the wet season in 2018. The shallow wells such as G_1, G_2, G_3, and G_4, which achieved a "bad water" quality with high As concentration, lie between the Mekong and Bassac Rivers, which have huge amounts of sediment deposition in monsoon season. This result agrees with other studies by Hoang et al. [64] and Vongphuthone et al. [63]. Arsenic deposition might be caused by huge amounts of sediment deposition during the monsoon season [64,66,67]. The An Giang government recommends treating As in groundwater taken from Holocene and the Upper Pleistocene aquifers before using, especially in the An Phu, Phu Tan, and Cho Moi districts.

The relationship between GWQI and agricultural intensification was not very clear. For example, high GWQI together with high As concentration was found in the Northeast and Southeast regions of An Giang province. However, the Southwestern side of An Giang province, including Thoai Son, and Chau Phu districts located on the left bank of the Bassac River, had slightly higher GWQI but lower As concentration. These regions mentioned above were not using groundwater for irrigation while a triple rice cropping model was mostly cultivated, which means excessive use of fertilizers caused water quality deterioration. On the other hand, only the regions near mountainous areas such as Tri Ton and Tinh Bien districts, where single rice cropping was often applied, extracted groundwater for irrigation. Furthermore, we detected the high values of GWQI links with high As concentration, where single and double rice crops were cultivated. In a nutshell, high As contamination in groundwater was found in agricultural land which used shallow groundwater for irrigation. Here, both the redox aquifer condition and use of phosphate-rich fertilizers lead to As enrichment in the groundwater. Here, microbial colony strongly absorb both As and phosphate while catalyzing reductive dissolution of iron-oxyhydroxide under the reducing condition. On the other hand, with excessive groundwater extraction, aquifer environment becomes oxidative in nature and microbial colonies start decomposing while releasing both As and phosphate. The other potential theory is with the application of phosphate-rich fertilizers, which enhance the competitive exchange of phosphate with As [62]. Findings from Thu et al. [45] supported high As concentration in wells in the Northeast and Southeast regions of An Giang province that are mainly concentrated in the riverside areas with depths of 15–36 m.

The groundwater quality improved in An Giang from 2009 to 2018 mainly due to the effective management of water resources by the An Giang government. Due to the high groundwater pollution observed during mid-2000, policymakers ordered the locals not to use 1460 unused wells during 2005–2009 because these wells were not covered and thus pollution sources may leach to the aquifer, especially in the wet season with deep water levels and inundation. There was effective implementation of decisions, which were issued by An Giang government, such as 69/2010/QD-UBND (in article 8 of chapter 3) and an updated decision version 38/2015/QD-UBND that specified protection of the quality of groundwater by filling unused wells. The government decided to fill the unused wells to prevent mixing of the pollutants from agricultural activities and human waste with the groundwater. Based on the preliminary data in An Giang, in 2017, less than 300 unused wells needed to be continuously filled.

The use of fuzzy logic seems to be the clearest innovation in the last decade, and its use is appropriate for an accurate GWQI. This approach allows for evaluating the impact of each variable on the final index of the quality of the water. However, it remains to establish weighting factors for specific water use. These weighting factors must be locally determined. Also, the weighting partly affects the final index obtained and can change significantly when changing the expert's awareness and perspective. Therefore, the sensitivity assessment was conducted to reduce the uncertainty by comparison of parameters' weights in four scenarios based on mean values and SD. The lowest SD in the weight change was selected as the optimal relative weight. However, the disadvantage of pair-wise comparison is the need to repeat calculations as it follows the same step as the pair-wise comparison of each water quality parameters.

Although As concentration in An Giang was under the permissible limits of the National technical regulation on groundwater quality (0.05 mg/L) since 2014, it still exceeds the WHO permissible limit of 0.01 mg/L [50]. There is insufficient evidence to conclude whether agricultural activity affects aquifer. However, we detected a high level of As in regions practicing agricultural production with the extraction of groundwater for irrigation. Many types of setup time reduction problems can be solved by using multiple criteria decision making (MCDM) techniques such as Fuzzy-AHP, but they must be utilized according to the suitability of the problem in order to develop the best decision.

6. Conclusions

This study applied of Fuzzy-AHP in weighing the parameters for calculating GWQI. The results indicated that the groundwater quality of some regions was bad in the year 2009. However, groundwater quality has improved over the years. One of the most important reasons for the poor water quality was the combined effect of both natural and human activities. The lithological structure (sediment deposition) and leaching of chemicals from agricultural runoff might go to groundwater aquifers. The As contaminant from sediment concentration was often found in the shallow aquifer. In recent years, there was less deposition of suspended solids near the flood plains which causes low As concentration. Besides, the An Giang government implemented the effective management of unused water wells by placing restrictions on filling the unused wells in these agricultural areas. Although enhancing the management of unused or abandoned groundwater wells did not eliminate the contamination it can improve aquifer water quality. Thus, the effective management of unused wells is one of the factors that improved groundwater quality in the periods of 2009–2018. The results of this study can help policymakers to make some future plans such as conducting suitability analyses of groundwater quality in different sectors. Another possible area of research is to look for the willingness of farmers to switch to different cropping patterns by growing crops with less water demand as mitigation measures in regions where groundwater has high As concentration.

The GWQI based on the Fuzzy-AHP was successfully applied to assess groundwater quality in An Giang. The weighted parameters via a bottom-up approach provided a better understanding of water quality issues at the local level. More monitoring wells should be installed for diligent and regular monitoring, which will give more reliable GWQI. It is necessary to consider more scenarios to reduce uncertainty at the first stage in terms of pair-wise comparison using Fuzzy-AHP method. Furthermore,

we need to consider degrees of confidence and attitudes of experts. Also, we should compare the weighted arithmetic index method to other methods in terms of the weighted parameter values.

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Conflicts of Interest: The authors declare no conflict of interest.

Appendix A

Table A1. Statistical summary of physico-chemical parameters for groundwater samples in the dry season in An Giang, during 2009–2018.

Well	As		NO_3^-		pH		NH_4^+		$CaCO_3$		Fe	
	Ave ± SD	Range	Ave ± SD	Range	Ave ± SD	Range	Ave ± SD	Range	Ave ± SD	Range	Ave ± SD	Range
G1	0.06 ± 0.09	0–0.25	0.43 ± 1.01	0–3.3	7 ± 0.51	6.6–8.4	2.01 ± 2.63	0–8.37	671 ± 316	83–975	1.53 ± 3.04	0–9.73
G2	0.01 ± 0.02	0–0.05	0.1 ± 0.11	0–0.25	6.9 ± 0.5	6.3–8.3	2.09 ± 1.96	0–4.80	279 ± 352	132–1265	0.34 ± 0.44	0–1.42
G3	0.02 ± 0.02	0–0.05	0.38 ± 0.52	0–1.45	7.1 ± 0.54	6.4–8.2	3.17 ± 2.7	0–6.28	958 ± 823	15–1788	0.19 ± 0.31	0–0.91
G4	0.01 ± 0.02	0–0.05	0.45 ± 0.79	0–2.5	6.7 ± 0.28	6.3–7.1	2.65 ± 2.53	0–7	851 ± 460	173–1485	0.55 ± 0.49	0–1.24
G5	0.01 ± 0.01	0–0.03	0.09 ± 0.19	0–0.62	7 ± 0.47	6.3–7.8	0.86 ± 0.67	0–2.13	648 ± 309	166–1063	0.28 ± 0.50	0–1.58
G6	0.01 ± 0.01	0–0.02	0.08 ± 0.17	0–0.56	7.1 ± 0.34	6.5–7.6	0.75 ± 0.63	0–1.8	284 ± 277	14–820	1.37 ± 2.62	0–8.31
G7	0.02 ± 0.01	0–0.03	3.96 ± 5.44	0–14.1	7.2 ± 0.25	6.5–7.5	0.43 ± 0.57	0–1.5	424 ± 201	172–722	0.09 ± 0.14	0–0.44
G8	0.01 ± 0.01	0–0.02	0.67 ± 1.17	0–3.47	7 ± 0.32	6.4–7.3	1.72 ± 1.67	0–4.43	234 ± 66	146–320	0.31 ± 0.38	0–0.94

Table A2. Statistical summary of physico-chemical parameters for groundwater samples in the wet season in An Giang, during 2009–2018.

Well	As		NO_3^-		pH		NH_4^+		$CaCO_3$		Fe	
	Ave ± SD	Range	Ave ± SD	Range	Ave ± SD	Range	Ave ± SD	Range	Ave ± SD	Range	Ave ± SD	Range
G1	0.26 ± 0.82	0–2.6	0.3 ± 0.38	0–1.08	6.9 ± 0.2	6.5–7.2	0.70 ± 0.88	0–2.86	788 ± 216	426–1056	0.22 ± 0.41	0–1.4
G2	0.54 ± 1.67	0–5.29	0.16 ± 0.31	0–1.06	6.9 ± 0.3	6.1–7.3	2.39 ± 2.66	0–7.08	365 ± 417	152–1504	0.14 ± 0.14	0–0.4
G3	0.01 ± 0.01	0–0.02	0.19 ± 0.26	0–0.81	6.6 ± 0.2	6.2–6.9	1.71 ± 2.22	0–5.25	1763 ± 1116	171–3966	0.16 ± 0.13	0–0.34
G4	0.84 ± 2.6	0–8.35	0.28 ± 0.43	0–1.29	6.5 ± 0.3	6.1–6.9	2.83 ± 3.01	0–10	1496 ± 1685	129–5996	0.52 ± 0.58	0–1.48
G5	0.55 ± 1.72	0–5.44	0.03 ± 0.04	0–0.11	6.7 ± 0.4	6.1–7.3	0.64 ± 0.95	0–2.7	651 ± 308	83–918	0.62 ± 0.94	0–2.41
G6	0.002 ± 0.003	0–0.01	0.08 ± 0.14	0–0.46	6.79 ± 0.3	6.2–7.2	0.23 ± 0.24	0–0.7	366 ± 280	66–808	1.17 ± 2.26	0–7.31
G7	0.006 ± 0.008	0–0.02	10 ± 14.78	0–38.5	6.9 ± 0.2	6.6–7.4	1.15 ± 1.93	0–4.7	521 ± 294	208–1228	0.42 ± 0.995	0–3.23
G8	0.005 ± 0.005	0–0.01	1.04 ± 2.14	0–6.26	6.8 ± 0.3	6.5–7.2	0.66 ± 0.88	0–2.32	270 ± 193	116–730	2.28 ± 4.99	0–14.6

Appendix B

Table A3. Annual Groundwater classification based on WQI in the dry season.

	2009	2010	2011	2012	2013	2014	2015	2016	2017	2018
G_1	Very bad	Very bad	Unsuitable for drinking	Good	Good	Good	Good	Excellent	Excellent	Very bad
G2	Good	Excellent	Very bad	Good	Good	Good	Excellent	Excellent	Good	Good
G3	Bad	Excellent	Excellent	Bad	Bad	Bad	Excellent	Bad	N/A	N/A
G4	Unsuitable for drinking	Bad	Excellent	Unsuitable for drinking	Unsuitable for drinking	Bad	Excellent	Bad	Good	Bad
G5	Good	Excellent	Good	Excellent	Good	Good	Excellent	Excellent	Good	Good
G6	Excellent	Excellent	Excellent	Good	Excellent	Excellent	Excellent	Excellent	Good	Excellent
G7	Good	Excellent	Bad	Very bad	Excellent	Excellent	Excellent	Excellent	Excellent	Excellent
G8	Excellent	Excellent	Good	Good	Excellent	Excellent	Good	Excellent	N/A	N/A

Table A4. Annual Groundwater classification based on WQI in the wet season.

|	2009	2010	2011	2012	2013	2014	2015	2016	2017	2018
G1	Excellent	Unsuitable for drinking	Good	Good	Good	Good	Good	Excellent	Excellent	Good
G2	Good	Good	Bad	Good	Bad	Excellent	Excellent	Unsuitable for drinking	Good	Good
G3	Bad	Good	Very bad	Bad	Bad	Good	Excellent	Excellent	N/A	N/A
G4	Unsuitable for drinking	Bad	Unsuitable for drinking	Unsuitable for drinking	Unsuitable for drinking	Very bad	Good	Excellent	Good	Good
G5	Excellent	Good	Unsuitable for drinking	Very bad	Excellent	Excellent	Excellent	Unsuitable for drinking	Excellent	Excellent
G6	Excellent	Excellent	Excellent	Good	Excellent	Excellent	Excellent	Excellent	Excellent	Excellent
G7	Very bad	Good	Bad	Good	Bad	Excellent	Excellent	Excellent	Excellent	Excellent
G8	Bad	Excellent	Excellent	Good	Excellent	Excellent	Excellent	Excellent	N/A	N/A

Appendix C
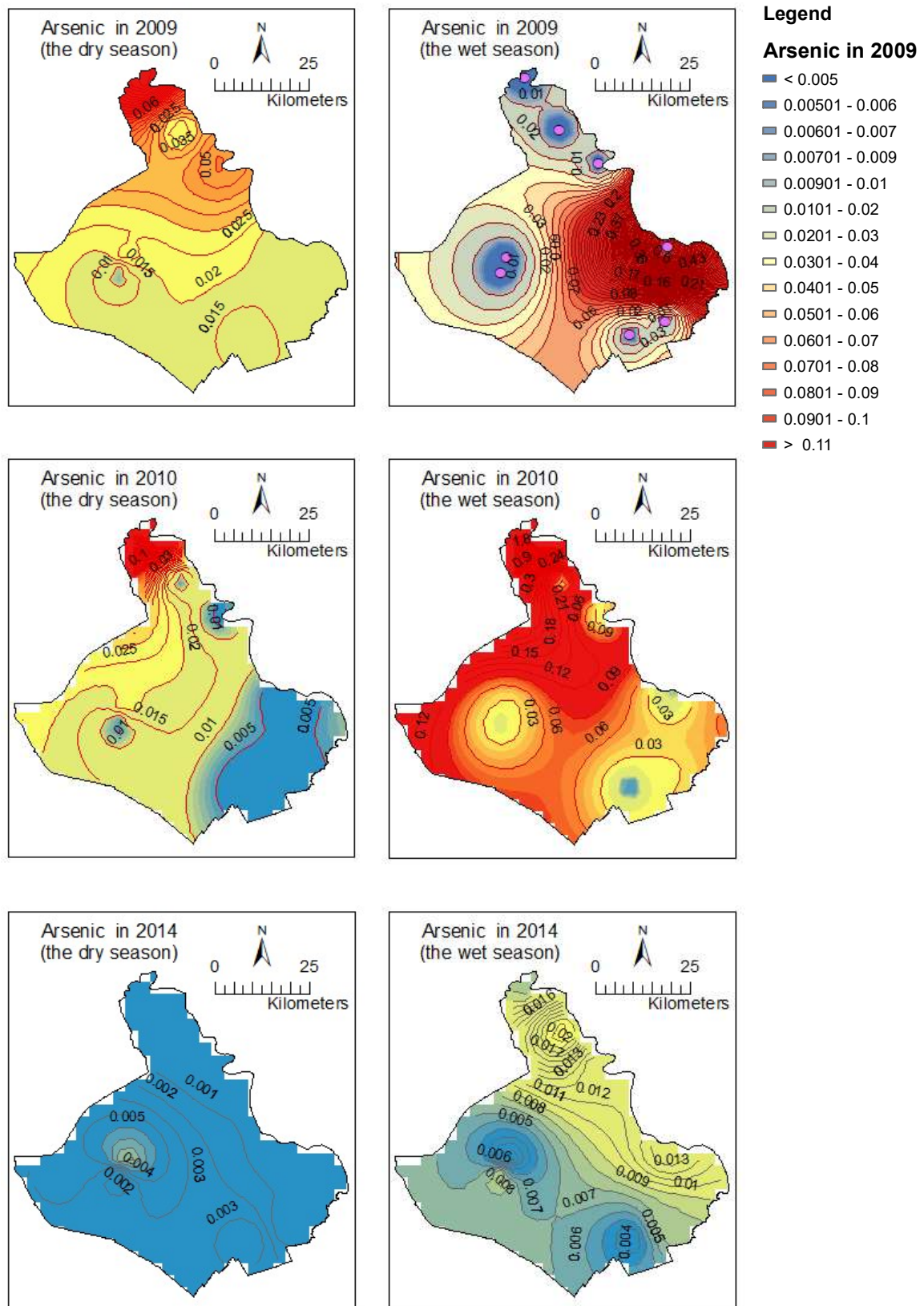
Figure A1. Concentrations of As (mg/L) were displayed by IDW in An Giang province.

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