

Article

Efficient Catalytic Reduction of Organic Pollutants Using Nanostructured CuO/TiO₂ Catalysts: Synthesis, Characterization, and Reusability

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Abstract: The catalytic reduction of organic pollutants in water is a critical environmental challenge due to the persistent and hazardous nature of compounds like azo dyes and nitrophenols. In this study, we synthesized nanostructured CuO/TiO₂ catalysts via a combustion technique, followed by calcination at 700 °C to achieve a rutile-phase TiO₂ structure with varying copper loadings (5–40 wt.%). The catalysts were characterized using X-ray diffraction (XRD), attenuated total reflectance-Fourier transform infrared (ATR-FTIR) spectroscopy, thermogravimetric analysis-differential thermal analysis (TGA-DTA), UV-visible diffuse reflectance spectroscopy (DRS), and scanning electron microscopy with energy-dispersive X-ray spectroscopy (SEM-EDS). The XRD results confirmed the presence of the crystalline rutile phase in the CuO/TiO₂ catalysts, with additional peaks indicating successful copper oxide loading onto TiO₂. The FTIR spectra confirmed the presence of all the functional groups in the prepared samples. SEM images revealed irregularly shaped copper oxide and agglomerated TiO₂ particles. The DRS results revealed improved optical properties and a decreased bandgap with increased Cu content, and 4-Nitrophenol (4-NP) and methyl orange (MO), which were chosen for their carcinogenic, mutagenic, and nonbiodegradable properties, were used as model organic pollutants. Catalytic activities were tested by reducing 4-NP and MO with sodium borohydride (NaBH₄) in the presence of a CuO/TiO₂ catalyst. Following the in situ reduction of CuO/TiO₂, Cu (NPs)/TiO₂ was formed, achieving 98% reduction of 4-NP in 480 s and 98% reduction of MO in 420 s. The effects of the NaBH₄ concentration and catalyst mass were investigated. The catalysts exhibited high stability over 10 reuse cycles, maintaining over 96% efficiency for MO and 94% efficiency for 4-NP. These findings demonstrate the potential of nanostructured CuO/TiO₂ catalysts for environmental remediation through efficient catalytic reduction of organic pollutants.

Keywords: titanium dioxide; catalysts; catalytic reduction; 4-nitrophenol; methyl orange; copper oxide

1. Introduction

Across the world, water and energy shortages represent urgent and increasingly severe global challenges. On a global level, over one billion people face the critical issue of not having access to clean drinking water, which is a necessity for human existence. Furthermore, more than two billion individuals lack proper sanitation facilities, resulting in almost two million annual fatalities due to diseases transmitted through impure water sources or insufficient sewage systems [1]. Water pollution occurs when the concentration of harmful chemicals or biological substances in a water body surpasses established standards, leading to adverse impacts on both human health and the natural environment [2].

Azo dyes and nitroaromatic compounds exhibit high stability and are known for their carcinogenic, mutagenic, and nonbiodegradable properties, posing a significant threat to human life, water quality, and the environment. These substances, such as Congo red, rhodamine B (RhB), and methyl orange (MO), are extensively used in numerous chemical sectors, including paper, textiles, paint, and plastics. These industrial processes produce significant quantities of dyes, leading to the emergence of dye-laden particles that are introduced directly into the environment [3,4]. MO, for example, is a highly toxic and nonbiodegradable azo. The adverse effects of this toxic dye may disrupt the balance of water within ecosystems and pose health risks, including symptoms such as vomiting, diarrhea, breathing difficulties, and nausea [5]. In particular, 4-Nitrophenol (4-NP) is a highly toxic compound that is challenging to degrade and treat effectively [6].

Compared with alternative methods, nitrophenol and dye reduction in the presence of a suitable catalyst and sodium borohydride (NaBH_4) is a recommended protocol, since it is relatively more affordable, secure, and environmentally friendly [7–9]. Therefore, the exploration of techniques such as chemical reduction to efficiently degrade diverse chemical pollutants is of considerable interest. However, this method commonly faces a challenge in achieving a fast degradation rate at lower concentrations of reducing agents, such as sodium borohydride. Consequently, to minimize the quantity of reducing agent required and increase the reaction rate, the development of an efficient catalyst for this chemical reduction becomes crucial [10,11]. Owing to their exceptional physicochemical properties, there has been significant interest in the field of catalysis regarding transition and noble metal nanoparticles in recent years [12–16]. Among these materials, copper nanoparticles (CuNPs) have gained recognition as excellent materials that exhibit novel physicochemical characteristics. This has prompted their investigation in various applications, including catalysis, sensors, photocatalysis, diverse biological activities, energy storage, and organic synthesis applications [17–21]. The catalytic activity is closely related to the degree of dispersion of the Cu(NPs), and a superior performance has been noted with smaller-sized nanoparticles [22,23].

One significant challenge in employing nanoparticles as catalysts involves their agglomeration and accumulation, resulting in a decrease in their catalytic efficiency. A viable solution to address this concern is the utilization of solid supports to stabilize the nanoparticles [24]. For this purpose, a series of solid supports, such as polymers, metal oxides, and carbon materials [25–28], have been adopted to avoid sintering, enhance stability, and facilitate the optimal mobilization/dispersion of nanoparticles, aiming to maximize catalytic activities across a wide variety of applications. Among these supports, the immobilization of CuNPs on TiO_2 , which is a representative n-type semiconductor material, has been extensively used as a photocatalyst, catalyst support, and cocatalyst because of its high oxidation ability, environmental friendliness, physicochemical stability, and cost effectiveness [29,30]. Research on the use of Cu/ TiO_2 for catalytic reduction has gained significant attention because of its promising applications in environmental remediation. One study focused on synthesizing Cu- TiO_2 nanoparticles from the extract of Phoenix

dactylifera. These catalysts effectively degraded dyes, such as RhB, in 11 min and MO in 25 min, achieving reductions of 89.8 and 95.3%, respectively, with 21.3 mg/mL of the catalyst [31]. Other investigations explored various shapes and sizes of Cu nanostructures and revealed that smaller particles and specific morphologies enhanced the catalytic activity for nitroaromatic reduction [32]. Additionally, Cu/TiO₂ catalyst synthesized from *Chimonanthus praecox* extract exhibited superior performance in degrading pollutants, such as 4-nitrophenol and other organic dyes, with 10 mg of catalyst [33]. CuO/TiO₂ composite created using *Tilia platyphyllos* extract also demonstrated high catalytic activity, reducing MO in 10 min and methyl blue in 9 min with 3 mg of the catalyst [34]. Recent studies of CuO/TiO₂ nanocomposite have further highlighted their exceptional catalytic performance, particularly under direct sunlight, where these photocatalysts effectively drive the selective hydrogenation of 4-NP to 4-AP in the presence of NaBH₄ [35]. Moreover, binary CuO/TiO₂ composites have demonstrated excellent reactivity in tandem hydrogenation processes for nitro compounds [36].

In the present study, different wt.% CuO/TiO₂ (rutile) were synthesized through the combustion technique followed by calcination at 700 °C to ensure a complete transition from anatase to rutile. The prepared catalysts were characterized via X-ray diffraction (XRD), UV-vis diffuse reflectance spectroscopy, attenuated total reflectance-Fourier transform infrared (ATR-FTIR) spectroscopy, thermogravimetric analysis-differential thermal analysis (TGA-DTA), and scanning electron microscopy with energy-dispersive X-ray spectroscopy (SEM-EDS). The catalytic performance of CuO/TiO₂ was evaluated afterwards via the reduction of MO and 4-nitrophenol as models of organic contaminants in the presence of NaBH₄. The influence of parameters, including NaBH₄ concentration, catalyst mass and reusability of the catalyst, were examined.

2. Results and Discussion

2.1. X-Ray Diffraction

The XRD patterns of the CuO/TiO₂ catalyst series are presented in Figure 1. The results obtained reveal that the materials are primarily composed of the rutile phase of TiO₂ in accordance with ICDD 21-1276. The presence of CuO in its monoclinic form was also verified in accordance with the ICDD 48-1548 reference pattern, confirming the effective loading of copper oxide in TiO₂. Figure 1 also reveals a correlation between the deposited quantity of Cu(II) ions and the intensities of the characteristic CuO peaks. These findings collectively indicate the successful preparation of the CuO/TiO₂ catalysts with varying percentages of CuO deposition. The average crystallite size of CuO was determined to be 26, 27, 32, 35, 36 nm for 40, 30, 20, 10 and 5 wt.% CuO/TiO₂ catalysts, respectively. The crystallite size was determined using the Scherrer equation [37]:

$$L = \frac{0.9\lambda}{\beta \cos \theta}$$

where λ is the X-ray wavelength (0.1540 nm), β is the full width at half maximum (FWHM), θ is Bragg's angle, and L represents the average crystallite size.

On the other hand, the absence of the anatase phase of TiO₂ is due to its transformation to rutile upon heating at 700 °C, as observed in Figure 2, which shows the XRD curves for the 40 wt.% CuO/TiO₂ catalysts subjected to various calcination temperatures. The primary difference observed was the phase transition of TiO₂ from anatase (ICDD 21-1272) to rutile (ICDD 21-1276), with complete transformation occurring at 700 °C. The same results were observed in many previous studies [38].

Following the reduction experiment using NaBH₄, the rutile structure of TiO₂ remained unchanged, indicating its stability under the reaction conditions. However, the monoclinic structure of CuO significantly decreased, leading to the formation of Cu(NPs), as depicted in Figure 3. Notably, two additional peaks emerged at 43.28° and 50.37°, corresponding to the (111) and (200) planes, respectively. These peaks are characteris-

tic of the face-centered cubic (FCC) crystal structure of metallic copper nanoparticles (ICDD No. 04-0836). The average Cu(NP) size is about 15 nm.

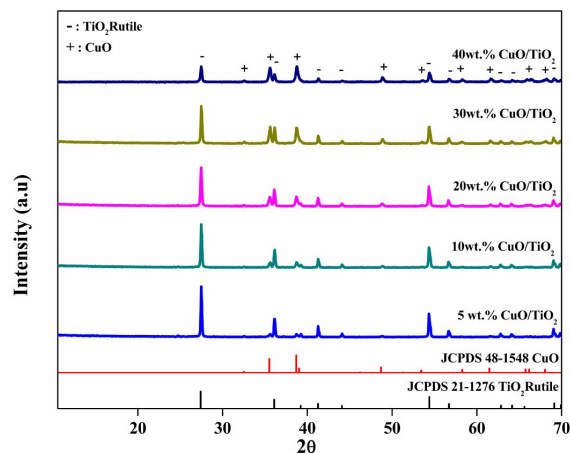


Figure 1. XRD patterns of different wt.% CuO/TiO₂ samples.

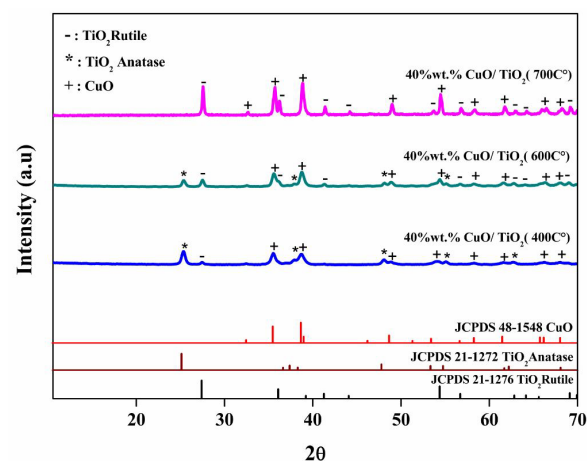


Figure 2. XRD patterns of 40 wt.% CuO/TiO₂ samples calcined at different temperatures.

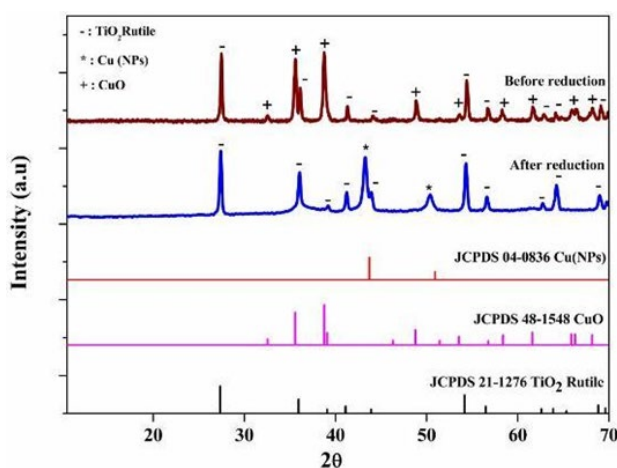


Figure 3. XRD patterns of 40 wt.% CuO/TiO₂ before and after the reduction experiment.

2.2. Thermal Analysis Measurement

To examine the thermal behavior of CuO/TiO₂, thermal experiments involving thermogravimetric and differential measurements were performed. The results are shown in Figure 4. The initial step, characterized by a weight loss of 3% within the temperature range

of 30 to 150 °C, can be attributed to the vaporization of water [39]. The notable weight reduction occurring between 200 and 350 °C, as evaluated at 30%, appears to be associated with the decomposition and combustion of the organic constituents. The latter was confirmed by an intense exothermic peak observed in the DTA curve and two intense peaks in the DTG curve [40]. On the other hand, no weight loss was observed at temperatures above 450 °C, and the anatase to rutile phase transition that should occur above 600 °C [38] was not observed in the DTA curve because of its reduced sensitivity compared with that of differential scanning calorimetry (DSC).

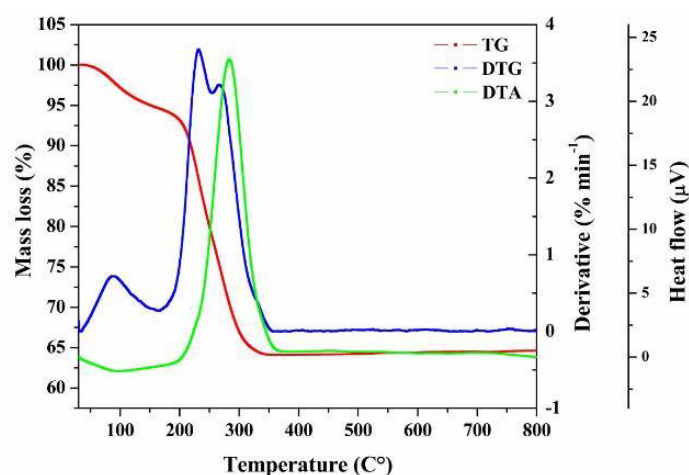


Figure 4. Thermal curves of the as-prepared 40 wt.% CuO/TiO₂ without calcination under air flow at 10 °C min⁻¹.

2.3. ATR-FTIR Spectroscopy

ATR-FTIR spectroscopy was performed on the catalysts calcined at 700 °C. The measurements were conducted across the spectrum ranging from 400 to 4000 cm⁻¹, and the results are shown in Figure 5. The analysis revealed similar results for the different catalysts with different percentages of CuO. The absence of water molecules and other contaminants in the catalyst was confirmed by the observation of strong bands in the infrared spectrum just below 1000 cm⁻¹. These bands typically correspond to lattice vibrations within metal oxides. The peak observed at approximately 650 cm⁻¹ is associated with the presence of the Ti-O-O stretching vibration bond. Additionally, the spectra revealed two other distinct bands below 500 cm⁻¹ that were attributed to metal-oxygen (M-O) bond vibrations. All the observed bands suggest the existence of both Ti-O vibrational stretching and Cu-O stretching vibration bonds [41], which confirms the successful incorporation of CuO into TiO₂.

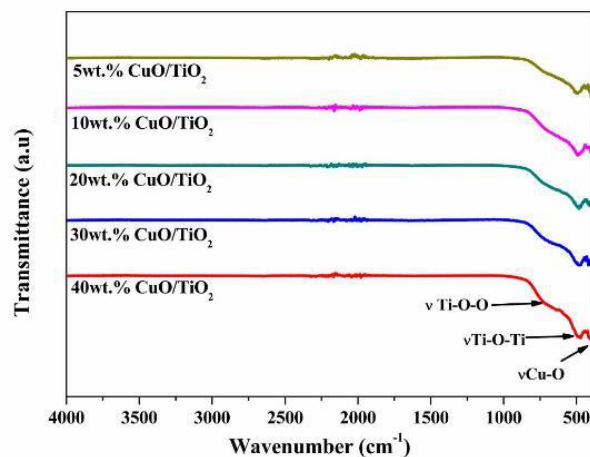


Figure 5. Infrared spectra of the wt.% CuO/TiO₂ samples.

2.4. UV–Vis Diffuse Reflectance Spectroscopy

Figure 6a presents the diffuse reflectance spectroscopy (DRS) results for TiO₂ and the various wt.% CuO/TiO₂ catalysts, revealing their optical properties in the 200–800 nm range. TiO₂ exhibits a pronounced absorption peak below 400 nm, which is attributed to its intrinsic interband absorption [42]. In contrast, the spectrum of the wt.% CuO/TiO₂ catalysts displays additional absorption bands in the 400–800 nm range. The absorption band between 400 and 600 nm is attributed to charge transfer from the valence band (VB) of TiO₂ to CuO. Additionally, the lower absorption band within the 500–800 nm range is associated with the intrinsic exciton band of CuO and the d–d transition of Cu²⁺ species [43]. The band gap energies (E_g) of TiO₂, CuO, and the different wt.% CuO/TiO₂ catalysts were determined by identifying the intersection of the linear portion of the $(\alpha h\nu)^2$ vs. energy (eV) curves, as detailed in Figure 6b. The E_g values for TiO₂ and wt.% CuO/TiO₂ are summarized in Table 1. These results indicate that increasing the CuO content leads to a progressive narrowing of the band gap. This trend can be attributed to the incorporation of Cu, which introduces additional energy states within the band structure, thereby lowering the bandgap. The reduced bandgap in the CuO/TiO₂ catalysts enhances their visible light absorption, which is beneficial for photocatalytic applications [44].

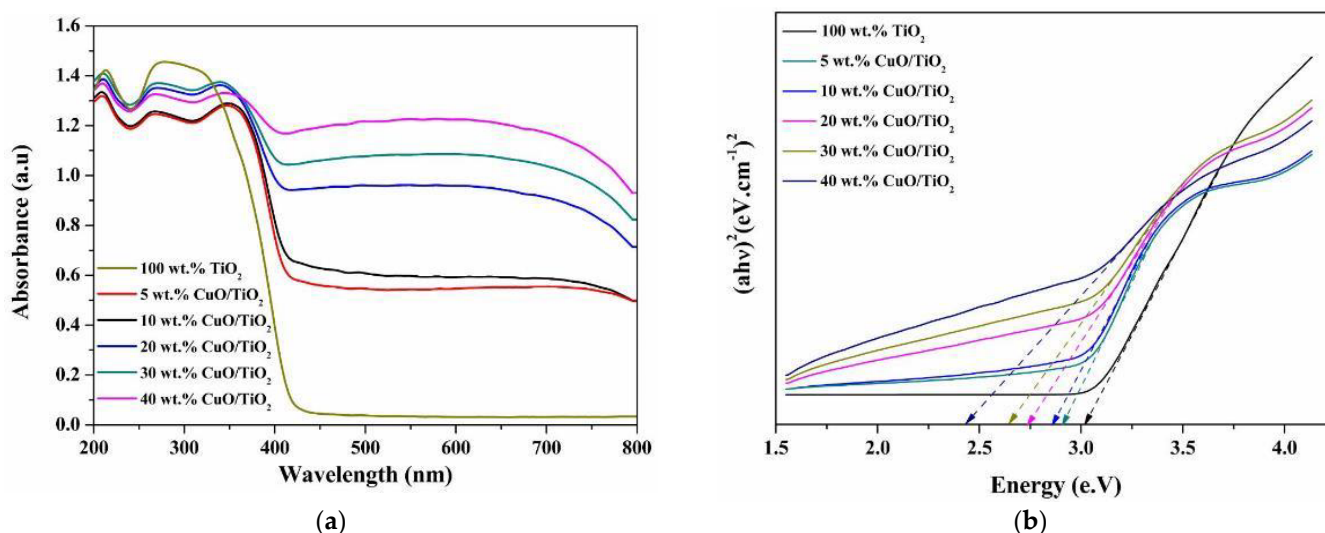


Figure 6. (a) UV–vis diffuse reflectance spectra of TiO₂ and wt.% CuO/TiO₂ and (b) Tauc plot.

Table 1. Bandgaps of different catalysts.

| Catalyst | TiO ₂ | 5 wt.% CuO | 10 wt.% CuO | 20 wt.% CuO | 30 wt.% CuO | 40 wt.% CuO |
|---------------|------------------|------------|-------------|-------------|-------------|-------------|
| Band gap (eV) | 3.02 | 2.91 | 2.86 | 2.73 | 2.65 | 2.43 |

2.5. SEM–EDS Analysis

Morphological studies via SEM–EDS were carried out on the CuO/TiO₂ catalyst. Figure 7 shows representative SEM images of synthesized CuO/TiO₂, which reveal a homogeneous spherical morphology with varying diameters and numerous micrometer-sized particles. Additionally, the catalyst has a porous structure and some aggregation of spheres. The analysis of the CuO/TiO₂ catalyst via energy dispersive X-ray (EDX) spectroscopy aimed to examine the presence and distribution of copper, titanium, and oxygen within the material. The EDX spectrum of CuO/TiO₂ is shown in Figure 8, and the results indicate the successful incorporation of copper into the TiO₂ sample. Additionally, the elemental mapping analysis revealed a uniform distribution of copper (depicted in yellow), titanium (represented in pink), and oxygen (shown in blue) in the samples, as illustrated in Figure 9.

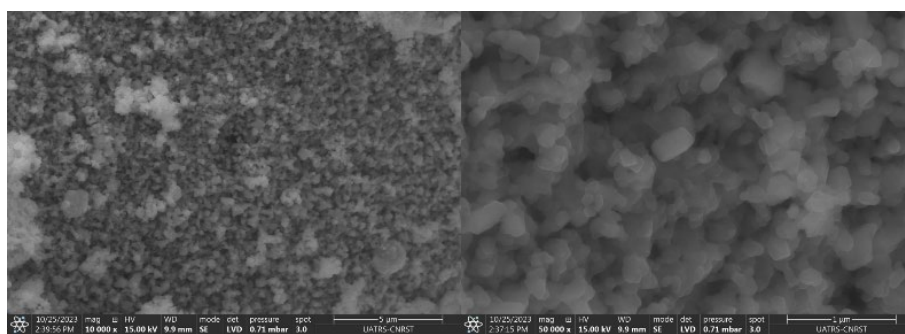


Figure 7. SEM micrographs of 40 wt.% CuO/TiO₂.

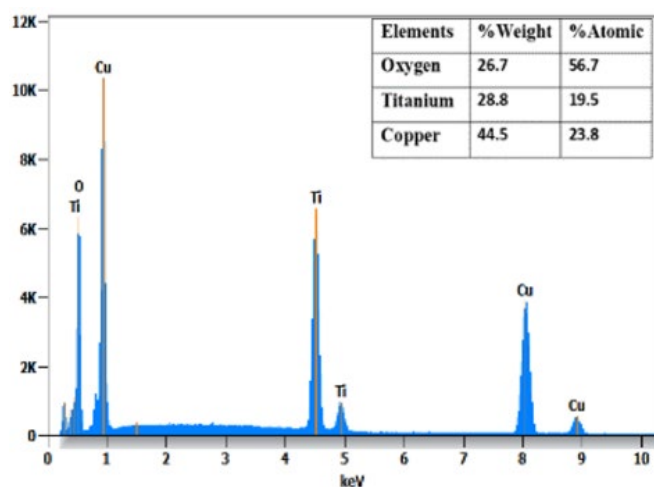


Figure 8. EDS Spectrum of 40 wt.% CuO/TiO₂.

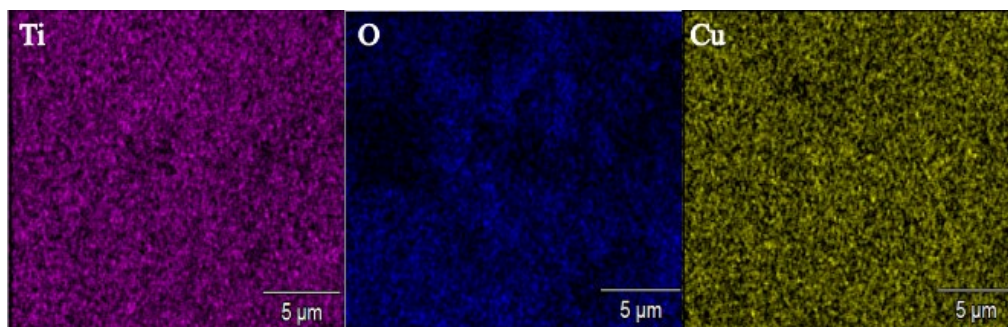


Figure 9. EDX mapping of the 40 wt.% CuO/TiO₂ catalyst.

2.6. Catalytic Reduction of 4-NP and MO by a CuO/TiO₂ Catalyst

The catalytic performance of the CuO/TiO₂ samples with different CuO loadings (5 to 40 wt.%) was evaluated for the reduction of organic pollutants via NaBH₄, as shown in Figure 10a,b. As shown in Figure 10a, which represents the reduction of 4-nitrophenol (4-NP), the 40 wt.% CuO/TiO₂ catalyst exhibited the fastest catalytic activity, reaching nearly 100% removal within a shorter reaction time than the other catalysts and pure CuO. Similarly, in Figure 10b, which shows the reduction of MO, 40 wt.% CuO/TiO₂ again demonstrates the fastest catalytic performance. The pure CuO sample exhibited lower catalytic activity than the 40 wt.% CuO/TiO₂ sample did, likely due to the agglomeration of copper oxide nanoparticles, which reduced the number of accessible active sites. On the other hand, at lower CuO loadings (e.g., 5 wt.%), the catalytic activity is the slowest, which can be attributed to the lower number of active CuO/TiO₂ sites available for the reduction process.

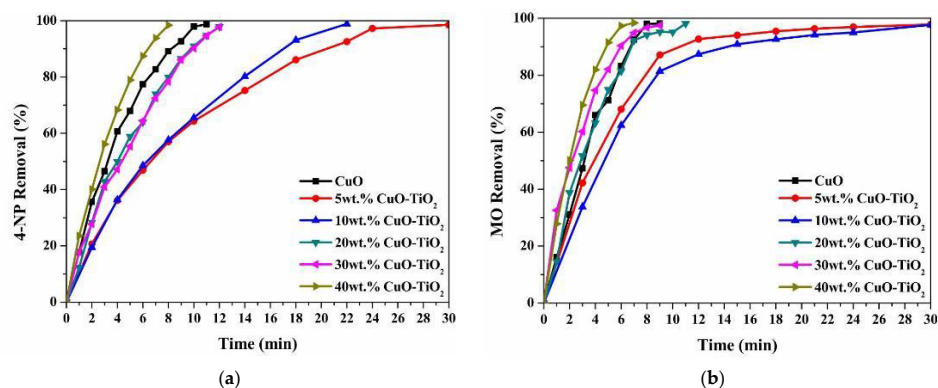


Figure 10. Reduction performance of the wt.% CuO/TiO₂ catalysts for 4-NP (a) and MO (b).

Figure 11a,b display the absorption spectra of 4-NP and MO in the presence of NaBH₄ but without the use of a catalyst. Under these conditions, the peak intensities show little change even after 180 min, indicating a negligible reduction rate, as no significant degradation occurred during this time. However, when CuO/TiO₂ is introduced alongside NaBH₄ (Figure 11c,d) and forming Cu(NPs)/TiO₂, a rapid and noticeable reduction in the characteristic absorption peaks of both 4-NP and MO is observed. For 4-NP, the peak at ~400 nm diminished progressively, vanishing completely within 480 s (Figure 11c). Similarly, the absorption peak of MO at approximately 464 nm disappears within 420 s (Figure 11d). This demonstrates a much faster and more efficient reduction process in the presence of the CuO/TiO₂ catalyst than in the presence of NaBH₄ alone. Simultaneously, a pair of new peaks emerged at 297 nm and 231 nm for the reduction of 4-NP and at 242 nm during the reduction of MO. These new peaks can be attributed to the characteristic absorption bands of the colorless 4-aminophenol (4-AP) from 4-NP and sulfanilic acid along with dimethyl-4-phenylenediamine from OM [45,46].

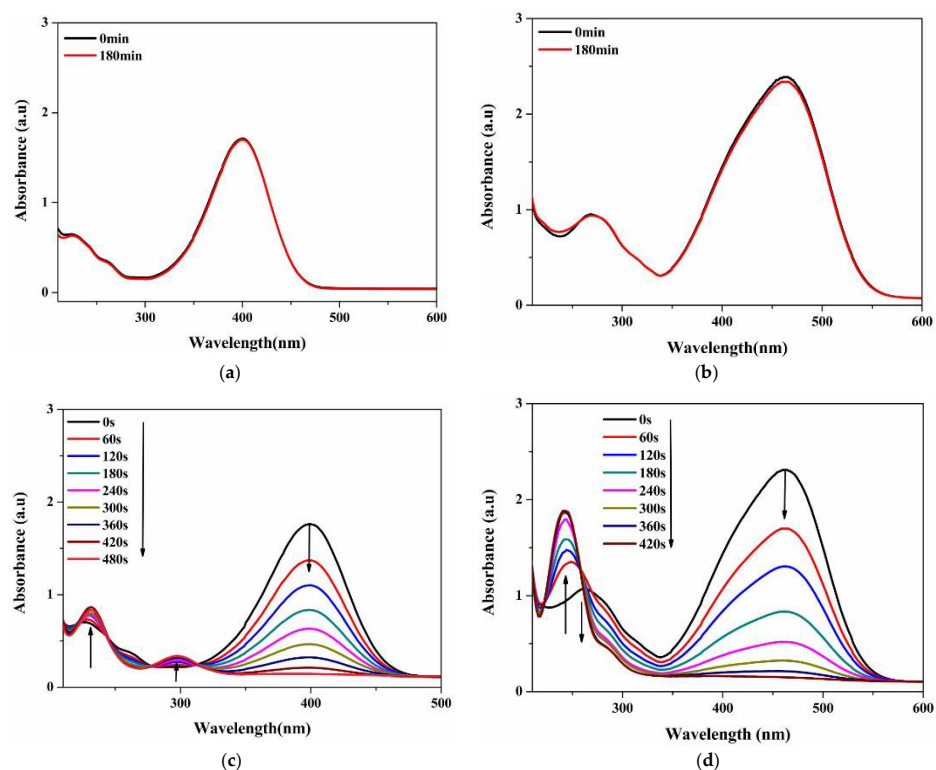


Figure 11. UV-vis spectra of the reduction of (a) 4-NP and (b) MO by NaBH₄ without the catalyst and absorbance spectra of (c) 4-NP and (d) MO in the presence of NaBH₄ and CuO/TiO₂.

2.6.1. Kinetic Study

The pseudo-first-order kinetic model of Langmuir–Hinshelwood was employed to analyze the kinetics data due to the substantial excess of NaBH_4 compared with the organic dyes, denoted as $([\text{NaBH}_4]/[\text{Pollutant}]) = 40$. The rate constants (k_{app}) governing the conversion reactions for both 4-NP and MO were determined via the following equation:

$$\ln \frac{C_t}{C_0} = \ln \frac{A_t}{A_0} = -k_{\text{app}}t \quad (1)$$

As depicted in Figure 12a,b, the $\ln(A_t/A_0)$ vs. time of reaction was determined by analyzing the absorption peak intensities at 400 nm (4-NP) and 464 nm (MO). Remarkably, the reduction process was initiated promptly, devoid of any induction time requirement. The graphical representation of $\ln(A_t/A_0)$ against time exhibited a robust linear correlation, aligning well with the expectations of pseudo-first-order kinetics. Consequently, the calculated rate constants were determined to be 0.00517 s^{-1} for 4-NP and 0.00727 s^{-1} for MO. To enable a comparative assessment of CuO/TiO_2 performance, an activity factor denoted as k' was computed via the following equation:

$$k' = \frac{k_{\text{app}}}{m} \quad (2)$$

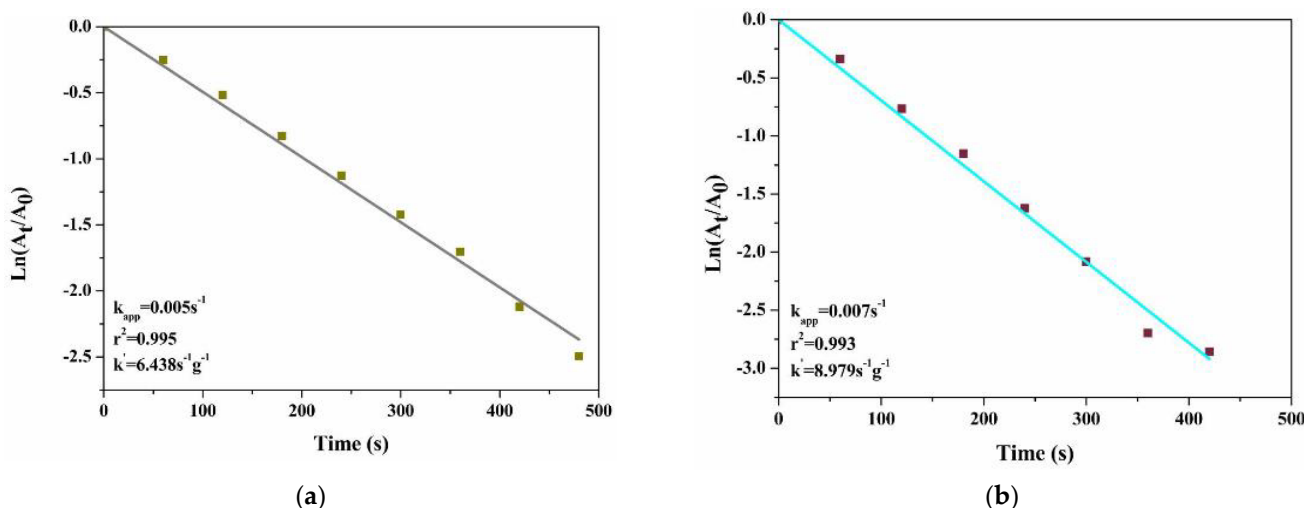


Figure 12. First-order plot for the catalytic reduction of (a) 4-NP and (b) MO. Experimental conditions: 40 wt.% CuO/TiO_2 dose = 0.04 g/L, $[\text{NaBH}_4] = 14 \text{ mM}$, and $[4\text{-NP or MO}] = 0.35 \text{ mM}$.

In this equation, $m(\text{g})$ represents the mass of the copper in the catalyst involved in the reaction.

In this context, the derived k' values were $6.46 \text{ s}^{-1} \text{ g}^{-1}$ for 4-NP and $8.98 \text{ s}^{-1} \text{ g}^{-1}$ for MO. For a comprehensive comparison, Table 2 illustrates the varying efficacies of $\text{Cu}(\text{NPs})$ catalysts documented in the literature.

Tables 3 and 4 show an increase in the observed rate constants (k_{app}) with increasing CuO percentage in the catalyst. The 40 wt.% CuO/TiO_2 catalyst exhibits the highest rate constant (0.005 s^{-1}), indicating that a higher copper loading favors faster reduction kinetics for 4-NP and MO. However, when considering the mass of active copper in the catalyst, the normalized values (k') indicate that the 5 wt.% CuO/TiO_2 catalyst exhibits the highest activity ($17.5 \text{ s}^{-1} \text{ g}^{-1}$). This suggests that, although increasing the CuO loading accelerates the overall reaction, the efficiency per unit mass of copper is maximized at lower loadings.

After the in situ reductive reaction of CuO/TiO_2 to $\text{Cu}(\text{NPs})/\text{TiO}_2$ with NaBH_4 , an underlying mechanism enables the catalytic reduction of 4-NP and MO using $\text{Cu}(\text{NPs})/\text{TiO}_2$ in the presence of NaBH_4 (Figure 13). In this process, the BH_4^- ions adhere to the surface of the catalyst, thereby initiating the creation of BO_2^- through the self-hydrolysis

of NaBH_4 . Moreover, BH_4^- interacts with the catalyst, facilitating the transfer of active hydrogen species, leading to the formation of an energetically charged hydrogen layer on the catalyst surface [47]. The organic pollutants subsequently attach themselves to the catalyst surface, undergoing reduction into 4-AP for 4-NP and sulfanilic acid along with dimethyl-4-phenylenediamine for MO during the step that governs the reaction rate. The produced compounds are then desorbed from the surface of the Cu(NPs)/ TiO_2 catalyst. As a result, Cu(NPs)/ TiO_2 exhibits highly efficient catalytic reduction due to the supply of electrons to the catalyst by BH_4^- ions, a mechanism that enables 4-NP and MO to bond with the catalyst surface, consequently yielding enhanced catalytic activity [48].

Table 2. Comparison of the catalytic properties of catalysts with those of Cu(NPs) reported in the literature and this work.

| Pollutant | Catalyst | k' ($\text{s}^{-1} \text{g}^{-1}$) | Time (min) | Conditions | | | Reference |
|----------------------|-------------------------------------|---|------------|---------------------|---------------|--------------------|-----------|
| | | | | NaBH_4 (M) | Pollutant (M) | Catalyst Mass (mg) | |
| 4-Nitrophenol (4-NP) | Cu10/MZ | 5.00 | 10 | 0.0045 | 0.0000675 | 1 | [22] |
| | Cu10/ZSM-5 | 1.17 | 50 | 0.0045 | 0.0000675 | 1 | [22] |
| | CuO/ TiO_2 | 6.46 | 8 | 0.014 | 0.00035 | 2 | This work |
| | Cu NPs@ Fe_3O_4 -LS | 1.87 | 3 | 0.125 | 0.00125 | 7 | [4] |
| | MnO@Cu/C | 17.33 | 1.5 | 0.0193 | 0.0008 | 4 | [16] |
| | C@Cu | 59.00 | 1 | 0.033 | 0.00014 | 1 | [49] |
| | Cu-Ag/GP | 0.40 | 10 | 0.0714 | 0.000857 | 10 | [50] |
| | Cu-Ni/GP | 0.60 | 7 | 0.0714 | 0.000857 | 10 | [50] |
| | CuVOS@ SiO_2 -3 | 1.57 | 2 | 0.00528 | 0.00014377 | 5 | [51] |
| | CuVOS-3 | 8.20 | 2 | 0.00582 | 0.000138 | 5 | [52] |
| Methyl orange (MO) | C@Cu | 62.00 | 1 | 0.033 | 0.000061 | 1 | [49] |
| | CuO/ TiO_2 | 8.98 | 7 | 0.014 | 0.00035 | 2 | This work |
| | Cu-Ag/GP | 0.77 | 4 | 0.0714 | 0.00004286 | 10 | [50] |
| | Cu-Ni/GP | 0.38 | 5 | 0.0714 | 0.00004286 | 10 | [50] |
| | CuVOS@ SiO_2 -3 | 1.37 | 4 | 0.00528 | 0.0003055 | 5 | [51] |
| | CuVOS-3 | 6.47 | 2 | 0.00582 | 0.00029 | 5 | [52] |

Table 3. Rate constants obtained for 4-NP reduction by wt.% CuO/ TiO_2 .

| Catalyst | k_{app} (s^{-1}) | k' ($\text{s}^{-1} \text{g}^{-1}$) |
|-----------------------------|--------------------------------------|--|
| 5 wt.% CuO/ TiO_2 | 0.00175 | 17.5 |
| 10 wt.% CuO/ TiO_2 | 0.00175 | 8.7 |
| 20 wt.% CuO/ TiO_2 | 0.00328 | 8.2 |
| 30 wt.% CuO/ TiO_2 | 0.00308 | 5.1 |
| 40 wt.% CuO/ TiO_2 | 0.00512 | 6.4 |

Table 4. Rate constants obtained for MO reduction by wt.% CuO/ TiO_2 .

| Catalyst | k_{app} (s^{-1}) | k' ($\text{s}^{-1} \text{g}^{-1}$) |
|-----------------------------|--------------------------------------|--|
| 5 wt.% CuO/ TiO_2 | 0.00150 | 15.0 |
| 10 wt.% CuO/ TiO_2 | 0.00126 | 6.3 |
| 20 wt.% CuO/ TiO_2 | 0.00479 | 12.0 |
| 30 wt.% CuO/ TiO_2 | 0.00546 | 9.1 |
| 40 wt.% CuO/ TiO_2 | 0.00727 | 8.9 |

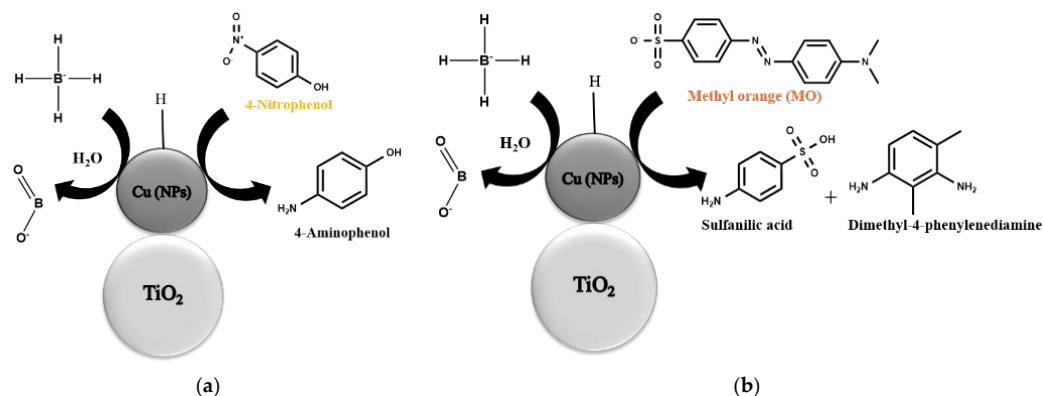


Figure 13. Schematic reactions of (a) 4-NP and (b) MO, by NaBH_4 in the presence of $\text{Cu(NPs)}/\text{TiO}_2$.

2.6.2. Effects of Parameters and Reusability

Multiple factors were investigated to understand their influence on the catalyzed reduction of 4-NP and MO using 40 wt.% CuO/TiO_2 catalysts. The parameters examined included the catalyst dosage and concentration of NaBH_4 . Figure 14a,b show the impact of varying catalyst amounts on the rate of pollutant reduction. By making more active sites accessible for the reaction, the large amount of catalyst led to improved efficiency. The performance constantly increased as the catalyst amount increased from 0.02 g/L to 0.06 g/L, which confirms the principle of Sabatier [53]. The variation in NaBH_4 concentration had a pronounced influence on the catalytic reduction of 4-NP and MO (Figure 14c,d). Because of the limited quantity of hydrogen emitted, the catalytic reduction rate decreased as the NaBH_4 concentration decreased [54].

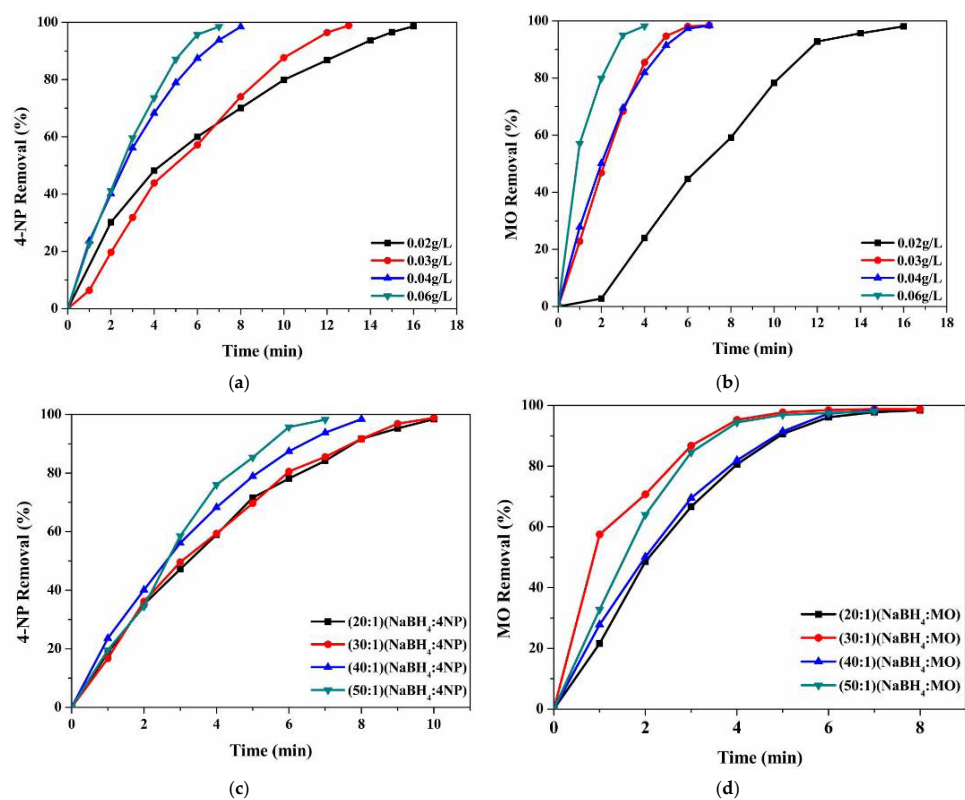


Figure 14. Pollutant removal versus time plots for catalytic reduction using 40 wt.% CuO/TiO_2 , showing the effect of catalyst dose on the reduction of (a) 4-NP and (b) MO, and the effect of NaBH_4 concentration on the reduction of (c) 4-NP and (d) MO.

An essential characteristic of catalysts is their stability and reusability. To assess this, a fresh solution containing pollutants was introduced after each catalytic reduction cycle. The initial concentration remained constant at 0.35 mM throughout. As depicted in Figure 15a,b, the catalyst consistently maintained over 96% MO capacity and 94% 4-NP capacity for reduction of the dyes across 10 successive reaction cycles. This outcome highlights the exceptional reusability of the CuO/TiO₂ catalyst. This excellent reusability can be attributed to the even distribution and robust stability of Cu on the TiO₂ surface, which in turn provides a greater number of active sites for catalytic processes.

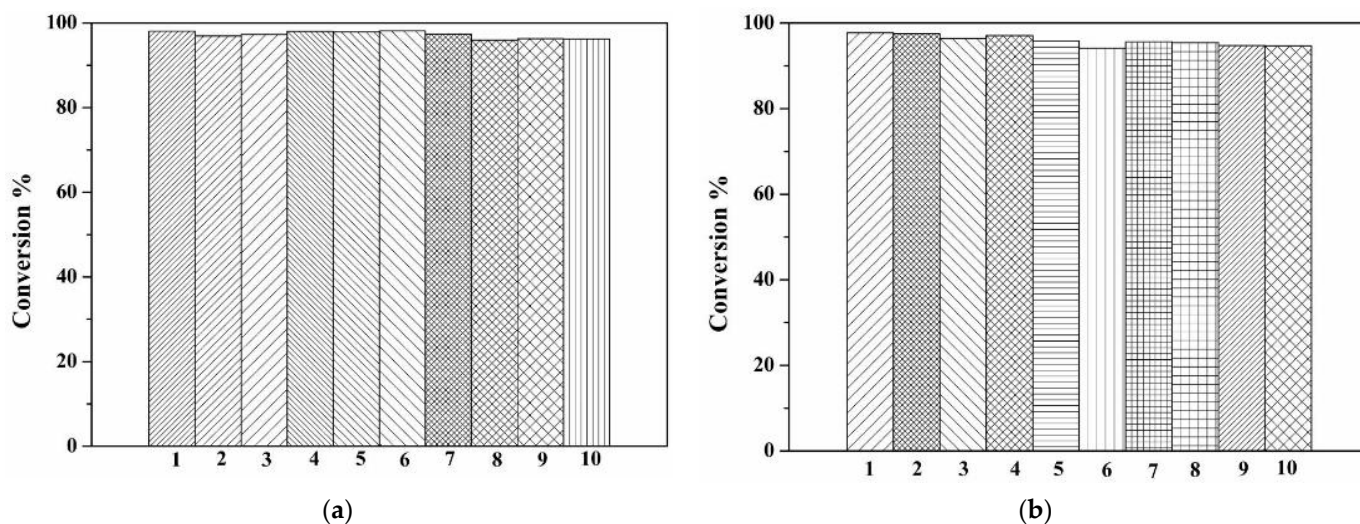


Figure 15. Recycling of CuO/TiO₂ in the hydrogenation of (a) 4-NP and (b) MO.

3. Materials and Methods

3.1. Synthesis of the CuO/TiO₂ Catalysts

The combustion technique was used to prepare different percentages of wt.% CuO/TiO₂ (wt = 5, 10, 20, 30, and 40%). Analytical grade Cu(NO₃)₂·6H₂O (Sigma Aldrich, ≥99.9%), along with TiO₂-P25 (Degussa Aeroxide P25) and citric acid as a fuel (Sigma Aldrich, ≥99.5%), served as the starting materials. Initially, these materials were added to deionized water with a fixed weight percentage of metal on TiO₂ and an excess amount of citric acid. The resulting mixtures were then stirred to achieve a well-dispersed suspension. The suspension subsequently underwent evaporation from 80 to 130 °C to yield a thick gel. This gel was subsequently converted into a dry form. Finally, the dry gel was calcined in an air atmosphere at 700 °C for 4 h with a ramp rate of 10 °C/min. Figure 16 represents the synthesis method for the catalysts.

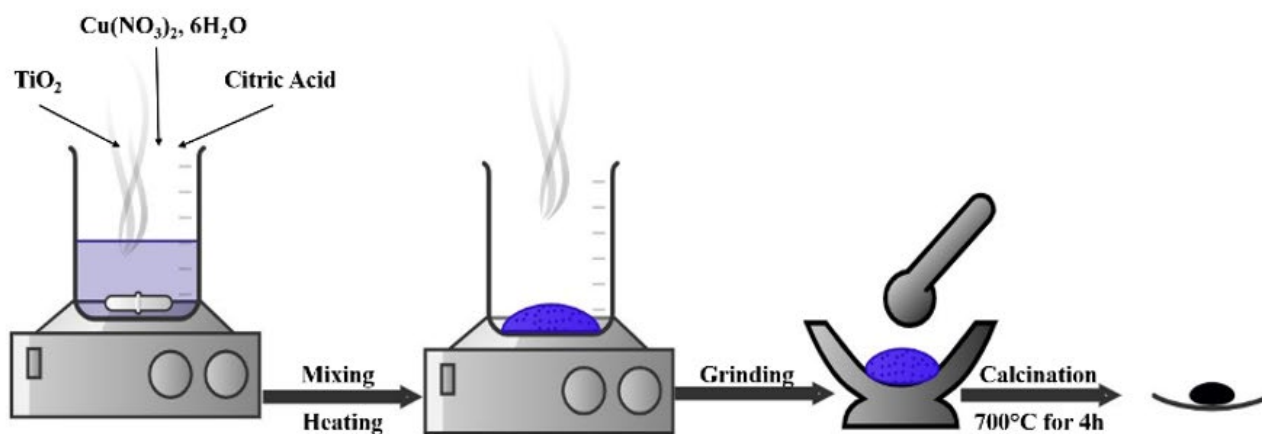


Figure 16. Synthesis procedure for the catalysts.

3.2. Catalytic Reduction of Organic Pollutants

The catalytic activity of CuO/TiO₂ was evaluated by studying the reduction of MO (Sigma Aldrich, ≥98.0%) and 4-nitrophenylphenol (Sigma Aldrich, ≥99.0%) as the target pollutants for the experiments. The experimental setup involved a beaker (50 mL) placed at room temperature. The catalytic reaction proceeded as follows: 2 mg of CuO/TiO₂ was uniformly dispersed in 35 mL of deionized water, followed by the addition of 10 mL of freshly prepared aqueous NaBH₄ (Sigma Aldrich, ≥99.0%) (70 mM). The solution was sonicated for 10 min, resulting in a color change from a black CuO/TiO₂ suspension to a gray color, indicating the formation of Cu(NPs)/TiO₂. Next, 5 mL of MO (3.5 mM) or 4-NP (3.5 mM) was added to the mixture. The overall concentrations of the organic dyes and NaBH₄ were 0.35 mM and 14 mM, respectively. The progress of the reaction was monitored by recording the time-dependent UV–vis absorption spectra of the reaction mixture in the wavelength range between 210 and 600 nm via a spectrophotometer (Techcomp UV 2300). The catalytic reduction experiment is shown in Figure 17.

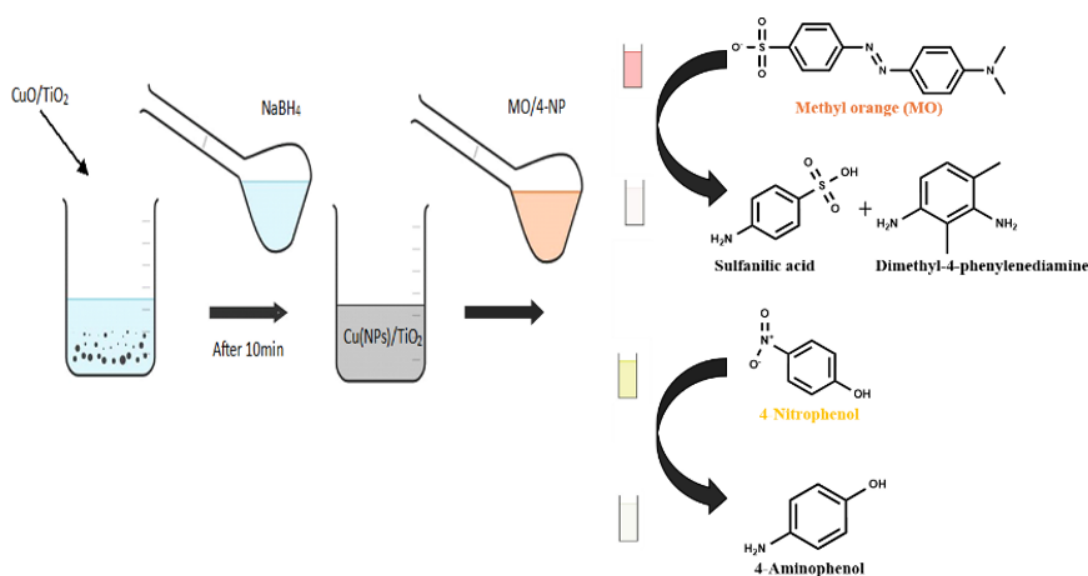


Figure 17. Catalytic reduction experiments of organic pollutants.

3.3. Characterization Techniques

The identification of the mineral crystalline phase of the catalysts was carried out via X-ray diffraction, via a Shimadzu 6100 powder diffractometer with a monochromatic beam ($\lambda_{\text{CuK}\alpha} = 1.541838^\circ$). These measurements were taken at room temperature over a 2θ range of 10° to 70° , with a scanning rate of $2^\circ/\text{min}$. A UV–vis spectrophotometer, namely, a PerkinElmer Lambda 900 UV/Vis/NIR spectrometer, was used to obtain diffuse reflectance spectra of the catalysts in the wavelength range of 200–800 nm. To identify the main functional groups of the catalyst, ATR–FTIR spectroscopy was used with a Nicolet iS50 instrument with a resolution of 4 cm^{-1} in the spectral range of $400\text{--}4000\text{ cm}^{-1}$. The thermal stability and degradation behavior of the catalysts were verified via simultaneous TGA and DTA under an air flow rate of $30\text{ mL}/\text{min}$ via a LabsysTM (1F) Setaram instrument. An 8 mg sample was placed in an alumina crucible and heated from 30°C to 800°C at a heating rate of $10^\circ\text{C}/\text{min}$. The surface morphology and chemical analysis of the samples were performed by SEM–EDX on a Quattros S-FEG-ThermoFisher scientific instrument.

4. Conclusions

In this study, CuO/TiO₂ catalysts were prepared through a combustion technique. The catalysts were characterized via XRD, ATR–FTIR spectroscopy, TGA–DTA, UV–vis DRS, and SEM–EDX techniques. Compared with TiO₂ and CuO, the presence of Cu in the CuO/TiO₂ catalysts significantly enhanced the reduction of 4-NP and MO. The

fastest dye reduction was obtained with 40 wt.% CuO, followed by 40 > 30 > 20 > 10 > 5 wt.% CuO/TiO₂ in the presence of NaBH₄ and 2 mg of catalyst. The kinetic data were successfully modeled via the pseudo-first-order Langmuir–Hinshelwood mechanism, yielding rate constants of 0.000517 s⁻¹ for 4-NP and 0.00727 s⁻¹ for MO. The efficiency of 4-NP and MO reduction with the CuO/TiO₂ catalyst improved as the catalyst dosage increased from 0.02 g/L to 0.06 g/L, offering more active sites. However, a decrease in NaBH₄ concentration led to a decrease in the catalytic reduction rate due to reduced hydrogen production. The presence of TiO₂ rutile functions as a protective agent to slow catalyst damage and agglomeration of particles. This material can be reused for several cycles without significant loss of catalytic activity.

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