


## Article

# Composition, Source Apportionment, and Health Risk of PM<sub>2.5</sub>-Bound Metals during Winter Haze in Yuci College Town, Shanxi, China

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**Abstract:** The composition, source, and health risks of PM<sub>2.5</sub>-bound metals were investigated during winter haze in Yuci College Town, Shanxi, China. The 24-h PM<sub>2.5</sub> levels of 34 samples ranged from 17 to 174  $\mu\text{g}\cdot\text{m}^{-3}$ , with a mean of  $81 \pm 35 \mu\text{g}\cdot\text{m}^{-3}$ . PM<sub>2.5</sub>-bound metals ranked in the following order: Zn > Cu > Pb > As > Ni > Cr (VI) > Cd > Co. The concentrations of 18% As and 100% Cr (VI) exceeded the corresponding standards of the Ambient Air Quality Standards set by China and the WHO. Subsequently, positive matrix factorization analyses revealed that the three major sources of metals were combustion (37.91%), traffic emissions (32.19%), and industry sources (29.9%). Finally, the non-carcinogenic risks for eight metals indicated that only 2.9% of the samples exceeded a threshold value of one, and As accounted for 45.31%. The total carcinogenic risk values for six metals (As, Cd, Co, Cr (VI), Ni, and Pb) were in the range from  $10^{-6}$  to  $10^{-4}$ , with Cr (VI) and As accounting for 80.92% and 15.52%, respectively. In conclusion, winter haze in Yuci College Town was characterized by higher metal levels and health risks; among the metals, As and Cr (VI) were probably the main contributors.



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**Keywords:** heavy metals; atmosphere; PMF

## 1. Introduction

Atmospheric haze has attracted considerable attention, especially during winter. Atmospheric fine particulate matter (PM<sub>2.5</sub>) plays an important role in hazy episodes and is the fifth leading cause of death globally after high blood pressure, smoking, diabetes, and hypercholesterolemia [1]. It is significantly associated with the incidence and mortality of bronchitis, asthma, and lung cancer [2]. PM<sub>2.5</sub> can be attached to a wide variety of chemical contaminants, such as polycyclic aromatic hydrocarbons (PAHs), heavy metals, nitrogen oxides, and emerging pollutants [3]. Some toxicants can be absorbed by the human respiratory system and can affect human health.

Studies on PM<sub>2.5</sub>-bound metals have been performed throughout the world; in some areas, the concentrations of PM<sub>2.5</sub>-bound metals have exceeded the threshold range of the WHO global air quality guidelines [4], such as in Isfahan of Iran [5], Saudi Arabia [6], Kitakyushu of Japan [7], as well as Beijing–Tianjin–Hebei [8], Xi'an [9], Guangzhou [10], and Taiyuan in China [11]. The source apportionment of heavy metals has helped to establish targeted pollution control strategies; for instance, coal burning, industrial pollution, and traffic often have been identified as the major contributors of metals [12–14]. In addition, PM<sub>2.5</sub>-bound metals pose health risks to humans [15]. Some metals have been identified as toxic and hazardous air pollutants in China [16] and some metals (e.g., As, Cd, Co, Cr, Ni, and Pb) have been defined by the WHO as carcinogenic to humans [17]. Therefore, a better understanding of the composition, source apportionment, and carcinogenic risk of metals in PM<sub>2.5</sub> is crucial to protect human health, especially during haze periods.

Yuci College Town is located in the Yuci District, Jinzhong City, Shanxi Province, China, adjacent to the provincial capital of Taiyuan City. It covers an area of approximate 12 km<sup>2</sup>.

It is a campus of provincial colleges and universities built by Shanxi Province. Currently, there are nearly ten colleges and universities, such as Taiyuan University of Technology, Shanxi Medical University, Taiyuan Normal University, and Jinzhong University. The total number of teachers and students in Yuci College Town is about 150,000. The campus has experienced frequent heavy air pollution in the past, particularly in winter. For instance, from January 2016 to December 2018, heavy air pollution events occurred 23 times for a total of 88 days in Jinzhong, Shanxi, China [18]. The predominant source of pollution was  $PM_{2.5}$ , which accounted for 78.3% of the total pollution.

Shanxi Province is well-known for its coal resources in China, and coal mining is considered to be one of the most significant sources of heavy metal contamination. As previously reported, heavy metal pollution from As and its carcinogenic risk to humans have been reported in Shanxi' mines [19]. Atmospheric  $PM_{2.5}$  pollution levels are very important to the health of teachers and students. However, investigations of  $PM_{2.5}$ -bound heavy metal pollution are lacking in Yuci College Town, Shanxi, China (YCT of China), especially during winter haze periods.

The objectives of the present study were (1) to measure the concentrations of heavy metals (including As, Cd, Co, Cr, Cu, Ni, Pb, and Zn) in  $PM_{2.5}$  in YCT of China; (2) to analyze the source apportionment; and (3) to assess the health risks (non-carcinogenic and carcinogenic) of exposure to eight heavy metals, via inhalation exposure, during winter haze periods.

## 2. Experiments and Methods

### 2.1. $PM_{2.5}$ Sample Collection

Figure 1 describes the location of the sampling site in Jinzhong University (Yuci district, Jinzhong City, Shanxi Province, China). This site is located at the southwest of Yuci College Town as well as multiple campuses and residential areas, representing an urban area. From 3 November 2020 to 9 December 2020, a total of 34 daily  $PM_{2.5}$  samples were collected using a medium flow particle sampler and quartz fiber filters at a gas flow rate of  $100 L \cdot min^{-1}$ . The sampler was placed 15 m from the ground and surrounded by pollution-free emission sources. The daily 24-h mean of  $PM_{2.5}$  was calculated using the gravimetric method [20]. The filters were all baked at  $450\text{ }^{\circ}C$  for 4 h before sampling to remove organic contaminants. Then, the filters were placed in boxes with constant humidity ( $50\% \pm 5\%$ ) and temperature ( $24 \pm 1\text{ }^{\circ}C$ ) for use. The filters were weighed after sampling, and the sensitivity of the balance was 0.01 mg. Finally, the filters were stored at  $-20\text{ }^{\circ}C$ , and pollutants were extracted within 2 months.

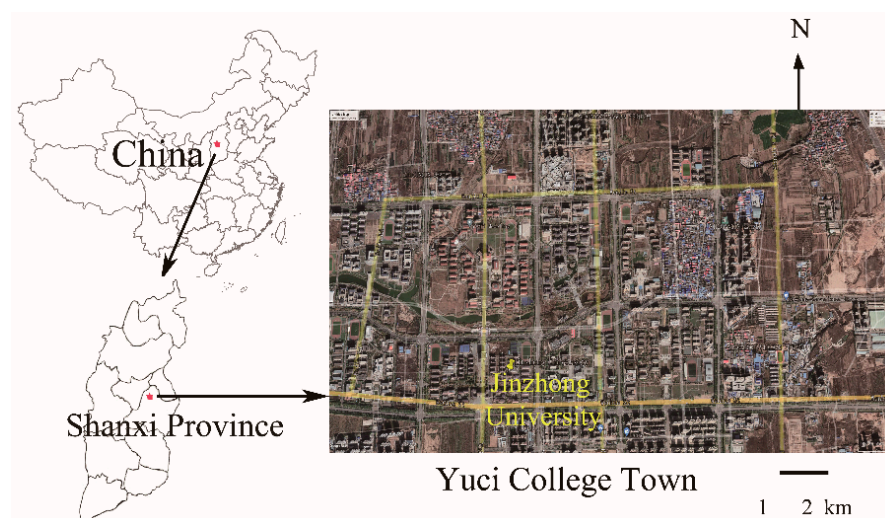


Figure 1. Map of the sampling area in Yuci College Town, Shanxi, China.

## 2.2. Metals Analysis

The PM<sub>2.5</sub> filters were digested on mixed acid, dissolved using a microwave digestion system, and the concentrations of eight metals (As, Cd, Co, Cr, Cu, Ni, Pb, and Zn) were analyzed by internal calibration using an inductively coupled plasma–mass spectrometer (ICP-MS, Agilent 7700), according to the method of HJ 657-2013 in China [21]. The concentration of Cr (VI) was calculated to be 1/7 of the measured Cr concentration according to the US EPA regional screening levels [22]. Detailed information on the digestion, internal standards (Table S1), and detection limit for metal analysis are shown in the Supplementary Information (SI).

## 2.3. PMF Model

The positive matrix factorization (PMF) model is widely used for source apportionment of PM<sub>2.5</sub>. It can simply transform the input data (multiple pollutant data arranged in matrix form) into a factor profile matrix and a factor contribution matrix. Detailed information was obtained according to the US EPA PMF 5.0 Fundamentals and User Guide [23]. Uncertainties ( $U_{nc}$ ) were calculated using the following equation:

If the concentration was higher than the method detection limit (MDL):

$$U_{nc} = \sqrt{(\text{Error fraction} \times \text{concentration})^2 + (0.5 \times \text{MDL})^2} \quad (1)$$

If the concentration was less than or equal to the MDL:

$$U_{nc} = \frac{5}{6} \times \text{MDL} \quad (2)$$

In addition, the rate of  $Q_{\text{robust}}/Q_{\text{true}}$  was calculated to determine the optimal number of factors, and bootstrap (BS) and displacement (DISP) analyses were performed to estimate the uncertainties of the PMF model.

## 2.4. Human Health Risk Assessment

Health risks of heavy metals in PM<sub>2.5</sub> were estimated via the inhalation pathway, as proposed by the US EPA [11]. The exposure concentration (EC) was calculated to assess the carcinogenic risk using Equation (3):

$$\text{EC} = (\text{CA} \times \text{ET} \times \text{EF} \times \text{ED}) / \text{AT} \quad (3)$$

where EC is the exposure concentration ( $\mu\text{g} \cdot \text{m}^{-3}$ ), CA is the contaminant concentration in air ( $\mu\text{g} \cdot \text{m}^{-3}$ ), ET is the exposure time ( $24 \text{ h} \cdot \text{d}^{-1}$ ), EF is the exposure frequency ( $180 \text{ d} \cdot \text{y}^{-1}$ ), ED is the exposure duration (24 y for adults), and AT is the averaging lifetime: for non-carcinogens ( $24 \times 365 \times 24 \text{ h}$ ) and for carcinogens ( $70 \times 365 \times 24 \text{ h}$ ).

The hazard index (HI) is traditionally used to assess the overall non-carcinogenic risk posed by multiple chemicals, and it was hypothesized that all metal risks were additive effects despite existing synergistic effects, the equation for HI is defined as follow:

$$\text{HI} = \sum \text{HQ}_i = \sum \text{EC}_i / \text{RfC}_i \quad (4)$$

where HQ is the hazard quotient (unitless) and  $\text{RfC}_i$  is the reference concentration of  $i$ th heavy metal ( $\mu\text{g} \cdot \text{m}^{-3}$ ) for inhalation (Table 1). According to the US EPA [24], if the HI value is less than one, the exposed population is unlikely to experience obvious adverse health effects. In contrast, if the HI value exceeds one, an adverse effect may occur for a specific population.

**Table 1.** Toxicological parameters of the heavy metals used for health risk assessment via the inhalation route.

	Contaminants	Class <sup>a</sup>	RfC <sub>i</sub> <sup>b</sup>	IUR <sup>b</sup>
		WHO	µg·m <sup>-3</sup>	(µg·m <sup>-3</sup> ) <sup>-1</sup>
As	Arsenic	1	0.015	0.0043
Cd	Cadmium	1	0.01	0.0018
Co	Cobalt	2B	0.006	0.009
Cr (VI)	Chromium	1	0.1	0.084
Cu	Copper	- <sup>c</sup>	1000 <sup>d</sup>	-
Ni	Nickel	2B	0.014	0.00026
Pb	Lead	2B	0.15 <sup>e</sup>	0.000012
Zn	Zinc	-	300	-

<sup>a</sup> Class: Agents classified by IARC the monographs adapted with permission from Ref. [17]. 2021, International Agency for Research on Cancer. <sup>b</sup> RfC<sub>i</sub>: reference concentration of *i*th heavy metal; IUR: inhalation unit risk, with their values from the US EPA: Regional screening levels (RSLs)—Generic Tables [22]. <sup>c</sup> Not reported. <sup>d</sup> Data from the literature [25]. <sup>e</sup> Data from the literature [26].

Carcinogenic risk (CR) is defined as the probability of an individual developing any type of cancer throughout their lifetime owing to exposure to carcinogenic hazards. CR was summarized by inhalation for an individual over a lifetime according to the following equation:

$$CR = \sum CR_i = \sum IUR \times EC_i \quad (5)$$

where IUR is the inhalation unit risk (µg·m<sup>-3</sup>)<sup>-1</sup> (Table 1). According to the US EPA [24], a CR lower than 10<sup>-6</sup> indicates an acceptable level, a CR range of 10<sup>-6</sup> to 10<sup>-4</sup> is generally considered a potential risk level, and a CR above 10<sup>-4</sup> is likely to be harmful to the human body.

### 2.5. Air Mass Backward Trajectory

Backward trajectory was performed using the hybrid single-particle Lagrangian integrated trajectory (HYSPLIT) online model from the National Oceanic and Atmospheric Administration [27]. The start altitude was chosen at three different heights (50, 500, and 1000 m), representing the low, middle, and upper atmosphere, respectively. The global data assimilation system (GDAS) data of 72-h backward trajectories on 25 November 2020 are presented in Figure S1.

### 2.6. Network Data Collection

The consumption of end-use energy, the number of days that reached the air quality standards, and the annual mean concentrations of PM<sub>2.5</sub> in Shanxi, China, were collected from the Shanxi Statistics Yearbook [28].

## 3. Results and Discussion

### 3.1. Mass Levels of PM<sub>2.5</sub>

The levels of PM<sub>2.5</sub> are presented in Table S2, and their descriptive statistics are presented in Table 2. The daily PM<sub>2.5</sub> concentrations of 34 samples ranged from 17 to 174 µg·m<sup>-3</sup>, with a median of 74 µg·m<sup>-3</sup> and an average concentration of 81 ± 35 µg·m<sup>-3</sup>; 94% of the samples exceeded Grade I (35 µg·m<sup>-3</sup>) of the Chinese ambient air quality standards [29]; none of the samples reached the WHO's AQC level (15 µg·m<sup>-3</sup>) in 2021 [30]; however, 53% of the samples reached the Grade II standard (75 µg·m<sup>-3</sup>) [29] and Interim Target-1 (75 µg·m<sup>-3</sup>) of the WHO global air quality guidelines [30]. During the sampling period, there were two heavy haze episodes, including 11 to 16 November and 24 to 26 November. During haze episodes, the PM<sub>2.5</sub> concentrations averaged over 100 µg·m<sup>-3</sup> and peaked at 127 and 174 µg·m<sup>-3</sup>, respectively. These results indicate that air pollution during winter was severe in this area.

**Table 2.** Statistical description of the daily PM<sub>2.5</sub> mass ( $\mu\text{g}\cdot\text{m}^{-3}$ ) and the concentrations of its metals ( $\text{ng}\cdot\text{m}^{-3}$ ) during winter in 2020 at Yuci College Town, Shanxi, China ( $n = 34$ ).

	PM <sub>2.5</sub>	As	Cd	Co	Cr (VI)	Cu	Ni	Pb	Zn	Sum
Min	17	0.43	0.04	0.01	0.39	0.69	0.22	1.04	0.71	27.92
Median	74	4.22	0.68	0.15	0.99	12.82	1.48	15.19	164.02	196.46
Max	174	11.36	4.88	1.61	2.91	67.15	6.82	40.52	823.39	905.26
Mean	81	4.71	0.89	0.29	1.31	20.04	1.82	14.95	191.87	235.87
SD <sup>a</sup>	35	2.70	0.86	0.38	0.77	17.35	1.48	9.09	145.92	161.88
WHO guideline value <sup>b</sup>	25	6.6	5	-	0.25	-	25	500	-	-
Grade II threshold <sup>c</sup>	75	6	5	- <sup>d</sup>	0.025	-	-	500	-	-

<sup>a</sup> SD: standard deviation. <sup>b</sup> WHO global air quality guidelines adapted with permission from Ref. [4]. 2005, World Health Organization. <sup>c</sup> Annual average concentrations of the Chinese ambient air quality standards [29]. <sup>d</sup> Not reported.

### 3.2. Concentrations of Heavy Metals in PM<sub>2.5</sub>

The levels of eight PM<sub>2.5</sub>-bound heavy metals are presented in Table S2 and their descriptive statistics are presented in Table 2. The total concentration of the eight metals was  $235.87 \pm 161.88 \text{ ng}\cdot\text{m}^{-3}$ . The daily mean levels of the heavy metals were ranked in the order Zn > Cu > Pb > As > Ni > Cr (VI) > Cd > Co. Zn was the most abundant metal with a mean of  $191.87 \pm 145.92 \text{ ng}\cdot\text{m}^{-3}$ , followed by Cu ( $20.04 \pm 17.35$ ), Pb ( $14.95 \pm 9.09$ ), As ( $4.71 \pm 2.70$ ), Ni ( $1.82 \pm 1.48$ ), Cr (VI) ( $1.31 \pm 0.77$ ), Cd ( $0.89 \pm 0.86$ ), and Co ( $0.29 \pm 0.38$ )  $\text{ng}\cdot\text{m}^{-3}$ . Moreover, the concentrations of Cd, Ni, and Pb in all samples were lower than the annual values recommended by the WHO [4] and China [29] (Table 2). However, the concentrations of 18% As and 100% Cr (VI) exceeded the Grade II standard of the Chinese Ambient Air Quality Standards [29]. There were no corresponding standards for Co, Cu, and Zn. These results are similar to those found in other areas; for instance, the levels of As and Cr (VI) greatly exceeded their corresponding threshold values in 60 cities in China [31].

In recent years, the mean concentrations of the metals in PM<sub>2.5</sub> have been obtained from China and abroad, and the results of other studies are summarized in Table 3.

For As, the average concentration during the 2020 winter haze periods in YCT of China was lower than that in Changzhi, Shanxi, China [32]. For other provinces in China, the average concentration of As, during the 2020 winter haze episodes in YCT of China, were higher than those in Beijing [33], Chengdu [34], Guangzhou [10], and Lanzhou [35]. Worldwide, it was higher than those in T<sup>~</sup>rinec-Kosmos of Czech Republic [36] and Kitakyushu of Japan [7]. In contrast, the average concentration of As in the 2020 winter haze episodes in YCT of China was lower than those in Yuci and Taiyuan [11] in 2017 within Shanxi Province. For other provinces in China, it was lower than those in Chongqing [37], Handan [38], Xi'an [9], and Xuanwu [10]. For worldwide, it was lower than those in Iasi of Romania [39], Isfahan of Iran [5], Karaj of Iran [40], and Saudi Arabia [6].

For Cr (VI), since total chromium concentrations were only available from the literature in some regions, the concentrations of Cr (VI) were also calculated to be 1/7 of the Cr concentration according to the US EPA regional screening levels [22]. For comparison, the mean concentration of Cr (VI) during the 2020 winter haze periods in YCT of China was higher than those in T<sup>~</sup>rinec-Kosmos of Czech Republic [36], Iasi of Romania [39], Koldata of India [41], Kitakyushu of Japan [7], Saudi Arabia [6], Los Angeles of USA [12], Beijing [33], and Chongqing [37] of China. Conversely, the concentrations of Cr (VI) in the remaining regions were higher than that of YCT in China. It has been suggested that Cr (VI) in PM<sub>2.5</sub> is very common and is a common air pollution problem worldwide. More stringent measures are required to control Cr pollution in the future.

Overall, the PM<sub>2.5</sub>-bound heavy metal contents were generally high in the YCT of China, with severe As and Cr (VI) contamination in PM<sub>2.5</sub> during winter haze periods.

**Table 3.** A comparison of the results from other studies regarding the mean concentrations ( $\text{ng}\cdot\text{m}^{-3}$ ) of  $\text{PM}_{2.5}$ -bound heavy metals in urban regions.

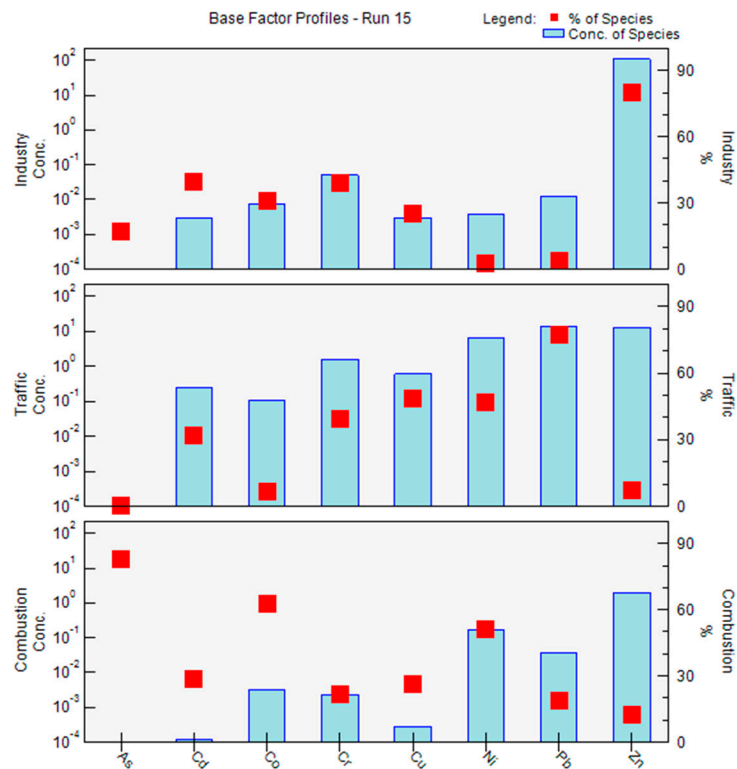
Country	Areas	Year	$\text{PM}_{2.5}$	As	Cd	Co	Cr (VI)	Cu	Ni	Pb	Zn	References
Czech Republic	Třinec-Kosmos	2020	28	1.3	0.26	0.05	1.1/7 <sup>d</sup>	4.6	0.81	11	34	[36]
Romania	Iasi	2016	20	<b>5.70</b>	0.33	0.09	1.78/7	7.70	24.3	1.99	33.9	[39]
India	Kolkata	2019	111.7	-	1.2	0.3	6.9/7	15	8.1	36	370	[41]
Iran	Isfahan	2015	- <sup>a</sup>	<b>32.47<sup>d</sup></b>	5.77	-	<b>57.4/7</b>	13.76	7.43	46.72	-	[5]
Iran	Karaj	2019	67	<b>32.1</b>	84	-	<b>49.5/7</b>	203	60.8	133	242	[40]
Japan	Kitakyushu	2019	21.3	1.4	-	-	3.0/7	3.6	3.3	10.5	29.5	[7]
Saudi Arabia		2020	-	<b>83</b>	17	-	8/7	9	10	119	31	[6]
USA	Los Angeles	2018	13.8	-	6	1	3/7	10	3	5	10	[12]
China	Beijing	2019	-	4.02	-	-	1.79/7	7.37	0.77	21.13	78.99	[33]
	Chengdu	2018	113.2	4.5	-	-	-	7.5	7.7	21.9	60.8	[34]
	Chongqing	2019	97.1	<b>7.56</b>	-	-	4.29/7	15.83	1.39	37.93	94.22	[37]
	Guangzhou	2017	55	4.39	0.74	0.53	<b>10.1/7</b>	16.37	5.72	25.52	127.31	[10]
	Guilin (haze)	2017	144	-	19.0	-	<b>11.5/7</b>	17.4	-	78.8	300.7	[42]
	Handan	2017	-	<b>11.94</b>	2.74	-	<b>11.1/7</b>	23.17	2.11	104.3	286.9	[43]
	Hefei	2017	81	-	-	-	<b>10/7</b>	11.29	-	12.64	273.5	[44]
	Lanzhou	2018	73	3	1	1.3	-	29	-	407	-	[35]
	Xi'an	2016	50.1	<b>117.2</b>	16.3	-	<b>343/7</b>	-	11.3	35.0	267.1	[9]
	Xuanwu	2016	61.1	6.44	1.88	0.29	<b>77.5/7</b>	20.99	3.73	54.72	212.76	[10]
Shanxi, China	Changzhi	2018	56.1	4.9	0.7	0.2	<b>14.3/7</b>	7.8	4.2	30.8	82.3	[32]
	Taiyuan	2017	-	<b>8.15</b>	1.07	1.20	<b>29.9/7</b>	29.56	12.69	94.36	230.57	[11]
	Yuci	2017	-	<b>9.45</b>	1.12	0.70	<b>11.7/7</b>	14.66	3.56	91.29	263.26	[11]
	Yuci	2020	80.65	4.71	0.89	0.29	1.31	20.04	1.82	14.95	191.87	This study
WHO guideline value <sup>b</sup>			25	6.6	5	-	0.25	-	25	500	-	
Grade II threshold <sup>c</sup>			75	6	5	-	0.025	-	-	500	-	

<sup>a</sup> Not reported. <sup>b</sup> WHO global air quality guidelines adapted with permission from Ref. [4]. 2005, World Health Organization. <sup>c</sup> Annual average concentrations of the Chinese ambient air quality standards [29]. <sup>d</sup> /7: The concentration of Cr (VI) was calculated to be 1/7 of the Cr concentration according to the US EPA regional screening levels [22]. Values higher than those in Yuci College Town are shown in bold.

### 3.3. Source Apportionment of $\text{PM}_{2.5}$ -Bound Elements

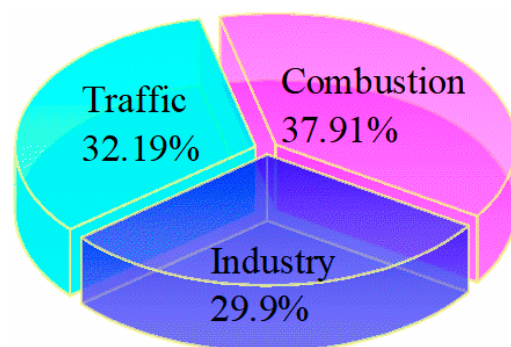
The source apportionment of eight heavy metals in  $\text{PM}_{2.5}$  was conducted using the PMF model [23]. Two to six factors were used for each dataset to determine the optimal solutions. Finally, the optimum result with a three-factor solution ( $Q_{\text{robust}} = 1141.5$ ,  $Q_{\text{true}} = 1255.4$ ,  $Q_{\text{robust}}/Q_{\text{true}} = 0.91$ , 92–100% of BS runs, and no swaps of DISP runs) was selected based on the interpretability of the source profiles and the results of the modeling diagnostics. The correlation coefficient ( $R^2$ ) values between the predicted data and the input data ranged from 0.69 to 0.92, meaning a good fit of the model.

As shown in the factor profile in Figure 2, the PMF resolved three factor profiles, namely, industry, traffic, and combustion, as the three anthropogenic sources contributing to the total  $\text{PM}_{2.5}$ -bound metals. Factor 1 was associated with high loadings of Cd (39.9%), Cr (VI) (39.2%), and Zn (80.3%). A previous study reported that Cr (VI) originated from the glassmaking industry, while Ni originated from steelworks [45]. Therefore, Factor 1 was identified as an industrial source. Factor 2 accounted for 48.8% and 77.6% of the Cu and Pb concentrations, respectively. Cu and Zn have been documented as surrogates for brake wear [46]. Pb emissions may result from the use of leaded gasoline. Thus, Factor 2 was labeled as a traffic source. Factor 3 demonstrated high loadings of the metals As (82.7%), Co (62.6%), and Ni (50.8%). These elements are all related to coal combustion [47], and Ni is also typical of oil combustion [14]. Therefore, Factor 3 was regarded as “combustion sources”, which included emissions from oil and coal combustion.



**Figure 2.** Source profiles obtained by the PMF analyses for Yuci College Town, Shanxi, China. Bars represent mass concentrations and red squares represent contribution percentages from each source factor.

Overall, three major sources of PM<sub>2.5</sub>-bound heavy metals were characterized based on the PMF analyses. As shown in Figure 3, combustion was still the largest contributor to PM<sub>2.5</sub>-bound heavy metals (37.91%) in Jinzhong, China. The contribution of traffic emissions (32.19%) was ranked second, just higher than the industry source (29.9%).

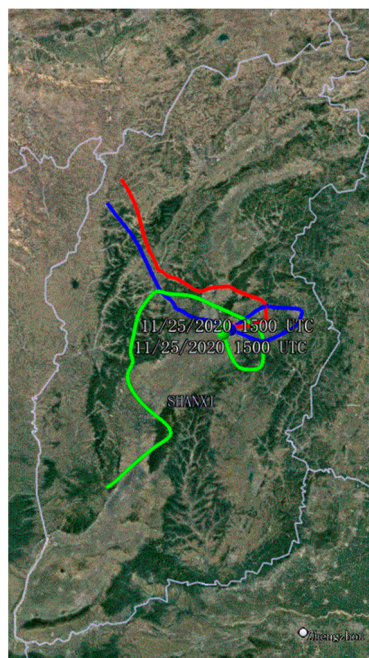


**Figure 3.** Contributions of identified sources to PM<sub>2.5</sub>-bound metal based on the PMF model in Yuci College Town, Shanxi, China.

Coal combustion was a main contributor that cannot be ignored. Shanxi is a famous coal province in China [48], and coal mining is considered one of the most important sources of heavy metal pollution [49]. More attention should be paid to the burning of coal for heating in winter, and the use of clean energy should be further increased.

The second contribution of traffic sources resulted from the increase in car ownership year by year. For example, the number of cars in Shanxi Province increased from 3.76 million in 2013 to 7.61 million in 2020 [28]. However, in the following years, this situation is expected to improve because of the continuous investment of new energy vehicles [50].

In fact, the formation of  $PM_{2.5}$  is very complex, and they involve adverse meteorological conditions, local emission accumulation, and regional transport, etc. [51]. Industrial sources of heavy metals in  $PM_{2.5}$  were likely to be regional transport because there were almost no industrial pollution sources around the sampling sites. Evidence of regional transport resulted from the analysis of the reverse trajectory. As shown, 72-h air mass backward trajectories occurred in haze episodes on 25 November 2020 (Figure 4); the results of 50 and 100 m were basically the same, and the air masses passed through Xinzhou-Taiyuan and reached Yuci after a roundabout in Yangquan. At 1000 m, the air mass passed through Linfen-Lvliang-Taiyuan, made a detour through Yangquan before reaching Yuci. These results indicated that most of the long-distance air masses were influenced by northwest winds and reached YTC of China through several industrial cities. This also indirectly suggested that the industrial pollution in YTC might come from regional transport rather than local sources. Due to the implementation of air pollution control policies and the shift of industrial production, industrial pollution is no longer the main source of pollution in YTC of China.



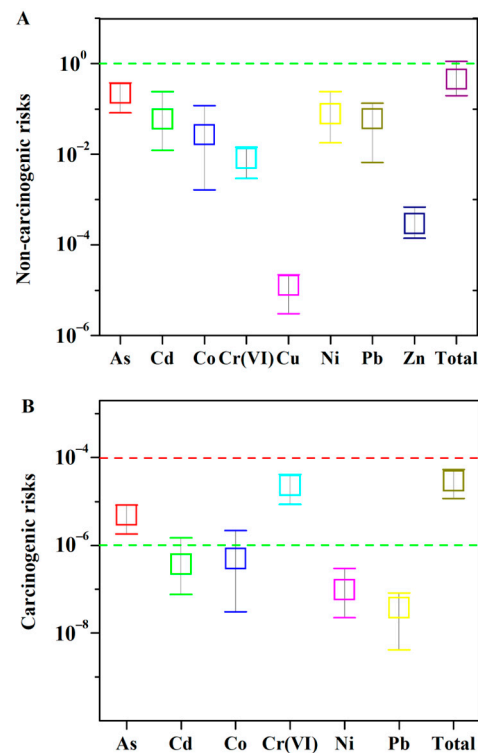
**Figure 4.** The 72-h back trajectories on 25 November 2020 in Yuci College Town, Shanxi, China. The red color line represents the height of 50 m; blue line: 500 m; green line: 1000 m.

### 3.4. Human Health Risk Assessment

#### 3.4.1. Non-Carcinogenic Risk Assessment

The non-carcinogenic risks of the eight  $PM_{2.5}$ -bound metals were calculated via the inhalation route in YCT of China (Figure 5). HI values ranged from 0.052 to 1.12 in 34 samples with a median of 0.30 and a mean of  $0.34 \pm 0.21$ . In this study, 2.9% of the samples exceeded an acceptable threshold of one. Among the eight metals, As accounted for 45.31% of the entire HI value, which demonstrated that the metal As was probably the main contributor to non-carcinogenic risk, while the contributions of the remaining metals were negligible. It can be concluded that heavy metals in  $PM_{2.5}$  have low non-carcinogenic risks to the public in YCT of China.





**Figure 5.** Non-carcinogenic (A) and carcinogenic (B) risks of exposure to PM<sub>2.5</sub>-bound metals estimated via the inhalation route in Yuci College Town, Shanxi, China ( $n = 34$ ).

### 3.4.2. Cancer Risk Assessment

The carcinogenic risk of exposure to six metals (As, Cd, Co, Cr, Ni and Pb) was estimated via the inhalation route and are presented in Figure 4. They ranged from  $7.56 \times 10^{-6}$  to  $5.36 \times 10^{-5}$ , with an average value of  $(2.29 \pm 1.30) \times 10^{-5}$ . The carcinogenic risk values all ranged from  $10^{-6}$  to  $10^{-4}$ . Among the six metals, Cr (VI) was the main contributor to carcinogenic risk, accounting for  $80.92 \pm 6.20\%$  of the total CR, and As accounted for  $15.52 \pm 5.42\%$  of the total CR. These results demonstrated that the metals As and Cr (VI) were probably the main contributors to carcinogenic risk, while the contribution of the remaining metals was negligible. These values are lower than the total CR of metals in PM<sub>2.5</sub> reported in Taiwan, China [52], but are higher than the total CR reported in Changzhi, China ( $10.31 \times 10^{-6}$ ) [32]. These results are consistent with previous studies that indicated the CR also mainly resulted from the contribution of Cr in Shenzhen [53], Taiyuan and Yuci [11], and Changzhi [32] in China. In addition, Cr (VI) also contributed to the highest potential years of life lost in most cities, with a proportion of 72.7% across 60 cities in China [31].

In summary, the non-carcinogenic risks of heavy metals were negligible in PM<sub>2.5</sub> in YCT of China. However, more attention should be paid to the carcinogenic risks of heavy metals, especially As and Cr (VI).

### 3.5. Policy Implication

A series of strict control measures have been implemented to prevent and control air pollution in China, and the PM<sub>2.5</sub> levels have declined since 2013. For instance, China's "Action Plan for the Prevention and Control of Air Pollution" was proclaimed in 2013; the "Blue Sky Protection" campaign was enacted in 2018; the 14th "Five-Year Plan for Modern Energy System" was issued in 2022; and the "Opinions on Further Strengthening the Prevention and Control of Heavy Metal Pollution" was issued in March 2022, which set two goals for 2025 and 2035. By 2025, the emissions of key heavy metal pollutants from key industries will be reduced by 5 percent as compared with those in 2020. By 2035, a heavy metal pollution prevention and control system and long-term mechanism

will be established to comprehensively improve the ability to monitor environmental pollution, to control heavy metal pollution, and to prevent environmental risks. In addition, Shanxi has implemented many policies, such as promoting coal energy transformation, developing emerging industries, limiting the traffic volume in winter by odd or even days, and operating new energy vehicles.

The implementation of coal banning was very effective in controlling air pollution during winter. For instance, the proportion of coal consumption in Shanxi Province decreased annually, while the proportion of electricity and natural gas increased. As shown in Figure 6 (original data were presented in Table S3), from 2013 to 2020, the proportion of coal consumption decreased from 27.92% to 16.39%, while electricity increased from 32.90% to 39.33% and natural gas increased from 13.89% to 22.72%.

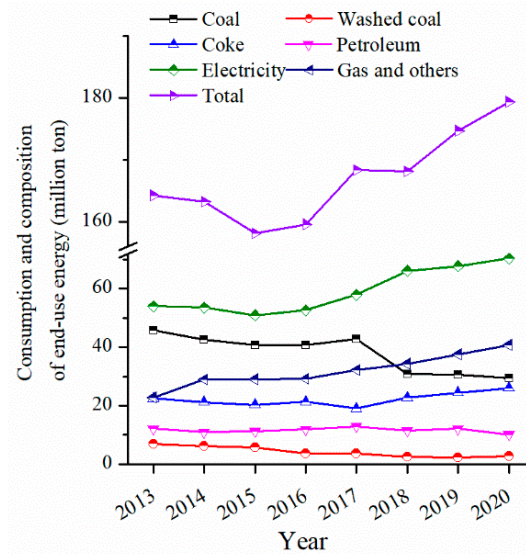


Figure 6. The composition of energy consumption in recent years in Shanxi Province, China.

Air quality has improved significantly, owing to a series of strict control policies in China. The number of days that reached the air quality standards gradually increased, while the average annual PM<sub>2.5</sub> level gradually decreased. As shown in Figure 7 (the original data are listed in Table S4), from 2014 to 2020, in Jinzhong, Shanxi Province, China, the number of days that reached the air quality standards increased from 241 to 267 days, while the annual mean concentrations of PM<sub>2.5</sub> decreased from 64 to 42  $\mu\text{g}\cdot\text{m}^{-3}$ . These results indicate that Shanxi’s energy transformation has achieved initial effects and will continue to control coal consumption and further increase the use of clean energy to win the battle for a blue sky in the future.

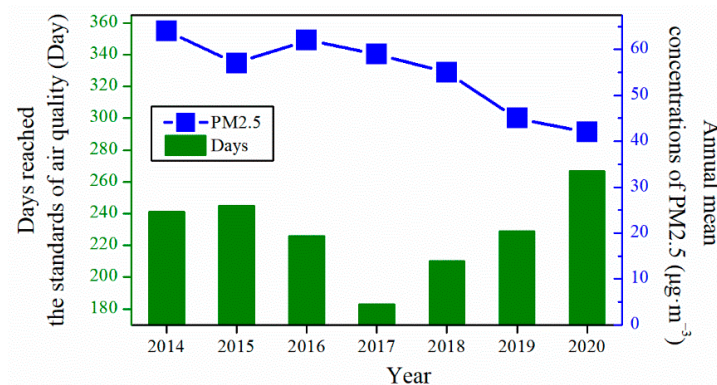


Figure 7. The number of days that reached the air quality standards in recent years and the annual mean concentrations of PM<sub>2.5</sub> in Jinzhong, Shanxi Province, China.

### 3.6. Limitations

First, there are two main types of health risk assessments of PM<sub>2.5</sub>-bound heavy metals. One type of health risk assessment is based solely on one exposure route, i.e., the inhalation route, while the other type of health risk assessments considers three exposure routes: inhalation, dermal, and oral intake. The former type was applied for health risk assessments in the industrial cities of Iran [40], Kolkata of India [41], Beijing of China [13], and Tianjin of China [38]. The latter type was used in Saudi Arabia [6], Hebei of China [54], and Shenzhen of China [53]. However, which is more representative? The models and parameters of health risk assessment should be optimized for specific populations and regions. Second, the traditional approach to heavy metals inherently hypothesizes that the carcinogenic risks of all metals are additive effects in a mixture. However, the synergistic or antagonistic effects may also occur during metabolism. For instance, three joint effects (synergistic, antagonistic, or additive effects) of the toxicity were all observed in the different component metal mixtures (Cd, Cr, Cu, Hg, Mn, Ni, Pb, and Zn) [55]. The mixtures of Cd + Pb [56], Cd + Ni [57], and Zn + Al [46] have synergistic effects [58], while Cd + Cu [59] and Cd + Zn [60] mixtures have antagonistic effects. In addition, heavy metals are only a small part of the PM<sub>2.5</sub>, which is also composed of other substances, such as organic matter, nitrate, sulfate, ammonium, elemental carbon, and chloride [61]. In future research, the above insufficiency should be overcome to assess the risk to human health more accurately.

### 4. Conclusions

In this study, the levels of eight metals in PM<sub>2.5</sub> were detected during the 2020 winter haze periods in YCT of China. The 24-h PM<sub>2.5</sub> levels of 34 samples ranged from 17 to 174  $\mu\text{g}\cdot\text{m}^{-3}$ , with a mean of  $81 \pm 35 \mu\text{g}\cdot\text{m}^{-3}$ . The PM<sub>2.5</sub>-bound heavy metals ranked in the following order: Zn > Cu > Pb > As > Ni > Cr (VI) > Cd > Co. A total of 18% and 100% of the concentrations assessed in the samples for As and Cr, respectively, exceeded the corresponding threshold values of the Chinese Ambient Air Quality Standards and the WHO Global Air Quality Guidelines; the levels of As and Cr (VI) were also higher than those in some areas worldwide. Overall, higher PM<sub>2.5</sub> levels were found in this area, suffering from severe As and Cr (VI) contamination during winter haze periods.

Based on the PMF model, combustion was the largest contributor to PM<sub>2.5</sub>-bound heavy metals (37.91%), followed by traffic emissions (32.19%) and industrial sources (29.9%). Finally, the potential health risks were estimated through exposure by the inhalation route. The non-carcinogenic risks of PM<sub>2.5</sub>-bound heavy metals were negligible, but the carcinogenic risk values were all within the potential level ( $10^{-6}$ – $10^{-4}$ ), and both As and Cr (VI) were the main contributors. Therefore, more attention should be paid to the prevention and control of As and Cr (VI) pollution in the environment. By reducing coal burning and using clean energy in winter, we will continue to reduce industrial source pollution, thereby making a positive contribution to reduce the occurrence of winter haze periods.

**Supplementary Materials:** The following supporting information can be downloaded at: <https://www.mdpi.com/article/10.3390/toxics10080467/s1>. Text: Detailed description of chemical analyses; Table S1: Isotopes of heavy metals, corresponding internal standards, and detection limit for ICP-MS analysis; Table S2: PM<sub>2.5</sub> mass ( $\mu\text{g}\cdot\text{m}^{-3}$ ) and its metals concentrations ( $\text{ng}\cdot\text{m}^{-3}$ ) during winter in 2020 in the college town of Shanxi Province, China ( $n = 34$ ); Table S3: The consumption, composition, and their proportion (%) of end-use energy (million ton) in recent years in Shanxi Province, China; Table S4: The days reached the standards of air quality in recent years and the annual mean concentrations of PM<sub>2.5</sub> ( $\mu\text{g}\cdot\text{m}^{-3}$ ) in Jinzhong, Shanxi Province, China; Figure S1. The 72-h back trajectories on 25 November 2020 in Yuci College Town, Shanxi, China. References [21,62–69] are cited in the supplementary materials.

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## References

1. Cohen, A.J.; Brauer, M.; Burnett, R.; Anderson, H.R.; Frostad, J.; Estep, K.; Balakrishnan, K.; Brunekreef, B.; Dandona, L.; Dandona, R.; et al. Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: An analysis of data from the Global Burden of Diseases Study 2015. *Lancet* **2017**, *389*, 1907–1918. [[CrossRef](#)]
2. Turner, M.C.; Krewski, D.; Diver, W.R.; Pope, C.A.; Burnett, R.T.; Jerrett, M.; Marshall, J.D.; Gapstur, S.M. Ambient air pollution and cancer mortality in the cancer prevention study II. *Environ. Health Perspect.* **2017**, *125*, 087013. [[CrossRef](#)] [[PubMed](#)]
3. Landrigan, P.J.; Fuller, R.; Acosta, N.J.R.; Adeyi, O.; Arnold, R.; Basu, N.; Balde, A.B.; Bertollini, R.; Bose-O'Reilly, S.; Boufford, J.I.; et al. The Lancet Commission on pollution and health. *Lancet* **2018**, *391*, 462–512. [[CrossRef](#)]
4. WHO. *Air Quality Guidelines: Global Update 2005: Particulate Matter, Ozone, Nitrogen Dioxide, and Sulfur Dioxide*; World Health Organization: Geneva, Switzerland, 2005.
5. Soleimani, M.; Amini, N.; Sadeghian, B.; Wang, D.; Fang, L. Heavy metals and their source identification in particulate matter (PM<sub>2.5</sub>) in Isfahan City, Iran. *J. Environ. Sci.* **2018**, *72*, 166–175. [[CrossRef](#)] [[PubMed](#)]
6. Abdulaziz, M.; Alshehri, A.; Yadav, I.C.; Badri, H. Pollution level and health risk assessment of heavy metals in ambient air and surface dust from Saudi Arabia: A systematic review and meta-analysis. *Air Qual. Atmos. Health* **2022**, *15*, 799–810. [[CrossRef](#)]
7. Zhang, X.; Eto, Y.; Aikawa, M. Risk assessment and management of PM<sub>2.5</sub>-bound heavy metals in the urban area of Kitakyushu, Japan. *Sci. Total Environ.* **2021**, *795*, 148748. [[CrossRef](#)]
8. Li, Z.Y.; Ma, H.Q.; Fan, L.; Zhao, P.; Wang, L.; Jiang, Y.J.; An, C.X.; Liu, A.Q.; Hu, Z.S.; Jin, H. Size distribution of inorganic elements in bottom ashes from seven types of bio-fuels across Beijing-Tianjin-Hebei Region, China. *Aerosol Air Qual. Res.* **2017**, *17*, 2450–2462. [[CrossRef](#)]
9. Wu, T.; Liu, P.; He, X.; Xu, H.; Shen, Z. Bioavailability of heavy metals bounded to PM<sub>2.5</sub> in Xi'an, China: Seasonal variation and health risk assessment. *Environ. Sci. Pollut. Res.* **2021**, *28*, 35844–35853. [[CrossRef](#)] [[PubMed](#)]
10. Xie, J.; Jin, L.; Cui, J.; Luo, X.-S.; Li, J.; Zhang, G.; Li, X.-D. Health risk-oriented source apportionment of PM<sub>2.5</sub>-associated trace metals. *Environ. Pollut.* **2020**, *262*, 114655. [[CrossRef](#)] [[PubMed](#)]
11. Liu, X.; Wang, Z.; Bai, H.; Zhang, S.; Mu, L.; Peng, L. Characteristics and health risk assessments of heavy metals in PM<sub>2.5</sub> in Taiyuan and Yuci College Town, China. *Air Qual. Atmos. Health* **2020**, *13*, 909–919. [[CrossRef](#)]
12. Farahani, V.J.; Soleimani, E.; Pirhadi, M.; Sioutas, C. Long-term trends in concentrations and sources of PM<sub>2.5</sub>-bound metals and elements in central Los Angeles. *Atmos. Environ.* **2021**, *253*, 118361. [[CrossRef](#)]
13. Yang, X.; Zheng, M.; Liu, Y.; Yan, C.; Liu, J.; Liu, J.; Cheng, Y. Exploring sources and health risks of metals in Beijing PM<sub>2.5</sub>: Insights from long-term online measurements. *Sci. Total Environ.* **2022**, *814*, 151954. [[CrossRef](#)] [[PubMed](#)]
14. Sofowote, U.M.; Su, Y.; Dabek-Zlotorzynska, E.; Rastogi, A.K.; Brook, J.; Hopke, P.K. Sources and temporal variations of constrained PMF factors obtained from multiple-year receptor modeling of ambient PM<sub>2.5</sub> data from five speciation sites in Ontario, Canada. *Atmos. Environ.* **2015**, *108*, 140–150. [[CrossRef](#)]
15. Swietlik, R.; Trojanowska, M. Chemical fractionation in environmental studies of potentially toxic particulate-bound elements in urban air: A critical review. *Toxics* **2022**, *10*, 124. [[CrossRef](#)]
16. MEEC (Ministry of Ecology and Environment of China). *List of Toxic and Hazardous Air Pollutants (2018)*; Ministry of Ecology and Environment of China and National Health Commission of China: Beijing, China, 2019; (In Chinese). Available online: [https://www.gov.cn/xinwen/2019-2002/2005/content\\_5364004.htm](https://www.gov.cn/xinwen/2019-2002/2005/content_5364004.htm) (accessed on 16 April 2022).
17. IARC (International Agency for Research on Cancer). *Agents Classified by the IARC Monographs, Volumes 1–130*. 2021. Available online: <https://monographs.iarc.who.int/agents-classified-by-the-iarc/> (accessed on 18 April 2022).
18. Dong, Y.; Zhou, H.; Fu, Y.; Li, X.; Geng, H. Wavelet periodic and compositional characteristics of atmospheric PM<sub>2.5</sub> in a typical air pollution event at Jinzhong city, China. *Atmos. Pollut. Res.* **2021**, *12*, 245–254. [[CrossRef](#)]
19. Li, Z.; Ma, Z.; van der Kuijp, T.J.; Yuan, Z.; Huang, L. A review of soil heavy metal pollution from mines in China: Pollution and health risk assessment. *Sci. Total Environ.* **2014**, *468*, 843–853. [[CrossRef](#)]
20. MEEC (Ministry of Ecology and Environment of China). *Determination of Atmospheric Particles PM10 and PM2.5 in Ambient Air by Gravimetric Method*; China Environmental Press: Beijing, China, 2011. (In Chinese)
21. NEPSC (National Environmental Protection Standards of China). *Ambient Air and Stationary Source Emission: Determination of Metals in Ambient Particulate Matter—Inductively Coupled Plasma/Mass Spectrometry (ICP-MS)*; China Environmental Science Press NEPSC: Beijing, China, 2013. (In Chinese)

22. US EPA. Regional Screening Levels (RSLs)-Generic Tables as of November 2021. 2021. Available online: <https://www.epa.gov/risk/regional-screening-levels-rsls-generic-tables> (accessed on 16 April 2022).
23. Norris, G.; Duvall, R.; Brown, S.; Bai, S. *EPA Positive Matrix Factorization (PMF) 5.0 Fundamentals and User Guide*; EPA/600/R-14/108; U.S. Environmental Protection Agency: Washington, DC, USA, 2014; pp. 1–136.
24. US EPA. *Risk Assessment Guidance for Superfund: Volume I: Human Health Evaluation Manual (Part F, Supplemental Guidance for Inhalation Risk Assessment)*; U.S. Environmental Protection Agency, Office of Emergency and Remedial Response: Washington, DC, USA, 2009.
25. Oosthuizen, M.; Wright, C.Y.; Matoane, M.; Phala, N. Human health risk assessment of airborne metals to a potentially exposed community: A screening exercise. *Clean Air J.* **2015**, *25*, 51–57. [[CrossRef](#)]
26. State of Michigan, USA. Chemical Update Worksheet. Michigan Department of Environmental Quality. 2016. Available online: [https://www.michigan.gov/documents/deq/deq-rrd-chem-LeadDatashet\\_527864\\_527867.pdf](https://www.michigan.gov/documents/deq/deq-rrd-chem-LeadDatashet_527864_527867.pdf) (accessed on 16 April 2022).
27. NOAA (National Oceanic and Atmospheric Administration). Available online: <https://www.ready.noaa.gov/HYSPLIT.php> (accessed on 20 July 2022).
28. SSY (Shanxi Statistics Yearbook). (In Chinese). Available online: <http://www.shanxi.gov.cn/sj/tjnj/> (accessed on 18 April 2022).
29. MEPC (Ministry of Environmental Protection of China). *Ambient Air Quality Standards (GB 3095-2012), National Standards of the People's Republic of China*; China Environmental Press: Beijing, China, 2012. (In Chinese)
30. WHO. *WHO Global Air Quality Guidelines: Particulate Matter (PM<sub>2.5</sub> and PM<sub>10</sub>), Ozone, Nitrogen Dioxide, Sulfur Dioxide and Carbon Monoxide: Executive Summary*; WHO Regional Office for Europe: Copenhagen, Denmark, 2021. Available online: <https://www.who.int/publications/i/item/9789240034433> (accessed on 18 April 2022).
31. Liu, J.; Cao, H.; Zhang, Y.; Chen, H. Potential years of life lost due to PM<sub>2.5</sub>-bound toxic metal exposure: Spatial patterns across 60 cities in China. *Sci. Total Environ.* **2022**, *812*, 152593. [[CrossRef](#)]
32. Duan, X.; Yan, Y.; Li, R.; Deng, M.; Hu, D.; Peng, L. Seasonal variations, source apportionment, and health risk assessment of trace metals in PM<sub>2.5</sub> in the typical industrial city of Changzhi, China. *Atmos. Pollut. Res.* **2021**, *12*, 365–374. [[CrossRef](#)]
33. Zhao, S.; Tian, H.; Luo, L.; Liu, H.; Wu, B.; Liu, S.; Bai, X.; Liu, W.; Liu, X.; Wu, Y.; et al. Temporal variation characteristics and source apportionment of metal elements in PM<sub>2.5</sub> in urban Beijing during 2018–2019. *Environ. Pollut.* **2021**, *268*, 115856. [[CrossRef](#)]
34. Qu, Y.; Gao, T.; Yang, C. Elemental characterization and source identification of the near-road PM<sub>2.5</sub> using EDXRF in Chengdu, China. *X-ray Spectrom.* **2019**, *48*, 232–241. [[CrossRef](#)]
35. Li, Y.; Zhao, B.; Duan, K.; Cai, J.; Niu, W.; Dong, X. Contamination characteristics, mass concentration, and source analysis of metal elements in PM<sub>2.5</sub> in Lanzhou, China. *Elem. Sci. Anthr.* **2021**, *9*, 00125. [[CrossRef](#)]
36. Seibert, R.; Nikolova, I.; Volná, V.; Krejčí, B.; Hladký, D. Air pollution sources' contribution to PM<sub>2.5</sub> concentration in the northeastern part of the Czech Republic. *Atmosphere* **2020**, *11*, 522. [[CrossRef](#)]
37. Han, Y.; Wang, Z.; Zhou, J.; Che, H.; Tian, M.; Wang, H.; Shi, G.; Yang, F.; Zhang, S.; Chen, Y. PM<sub>2.5</sub>-bound heavy metals in southwestern China: Characterization, sources, and health risks. *Atmosphere* **2021**, *12*, 929. [[CrossRef](#)]
38. Chen, R.; Jia, B.; Tian, Y.; Feng, Y. Source-specific health risk assessment of PM<sub>2.5</sub>-bound heavy metals based on high time-resolved measurement in a Chinese megacity: Insights into seasonal and diurnal variations. *Ecotoxicol. Environ. Saf.* **2021**, *216*, 112167. [[CrossRef](#)] [[PubMed](#)]
39. Galon-Negru, A.G.; Olariu, R.I.; Arsene, C. Size-resolved measurements of PM<sub>2.5</sub> water-soluble elements in Iasi, north-eastern Romania: Seasonality, source apportionment and potential implications for human health. *Sci. Total Environ.* **2019**, *695*, 133839. [[CrossRef](#)] [[PubMed](#)]
40. Kermani, M.; Jonidi Jafari, A.; Gholami, M.; Arfaenia, H.; Shahsavani, A.; Fanaei, F. Characterization, possible sources and health risk assessment of PM<sub>2.5</sub>-bound heavy metals in the most industrial city of Iran. *J. Environ. Health Sci. Eng.* **2021**, *19*, 151–163. [[CrossRef](#)]
41. Mitra, S.; Das, R. Health risk assessment of construction workers from trace metals in PM<sub>2.5</sub> from Kolkata, India. *Arch. Environ. Occup. Health* **2022**, *77*, 125–140. [[CrossRef](#)]
42. Zhong, S.; Zhang, L.; Jiang, X.; Gao, P. Comparison of chemical composition and airborne bacterial community structure in PM<sub>2.5</sub> during haze and non-haze days in the winter in Guilin, China. *Sci. Total Environ.* **2019**, *655*, 202–210. [[CrossRef](#)]
43. Cai, A.; Zhang, H.; Wang, L.; Wang, Q.; Wu, X. Source apportionment and health risk assessment of heavy metals in PM<sub>2.5</sub> in Handan: A typical heavily polluted city in North China. *Atmosphere* **2021**, *12*, 1232. [[CrossRef](#)]
44. Xue, H.; Liu, G.; Zhang, H.; Hu, R.; Wang, X. Similarities and differences in PM<sub>10</sub> and PM<sub>2.5</sub> concentrations, chemical compositions and sources in Hefei City, China. *Chemosphere* **2019**, *220*, 760–765. [[CrossRef](#)] [[PubMed](#)]
45. Ledoux, F.; Kfoury, A.; Delmaire, G.; Roussel, G.; El Zein, A.; Courcot, D. Contributions of local and regional anthropogenic sources of metals in PM<sub>2.5</sub> at an urban site in northern France. *Chemosphere* **2017**, *181*, 713–724. [[CrossRef](#)] [[PubMed](#)]
46. Hulskotte, J.H.J.; Roskam, G.D.; Denier van der Gon, H.A.C. Elemental composition of current automotive braking materials and derived air emission factors. *Atmos. Environ.* **2014**, *99*, 436–445. [[CrossRef](#)]
47. Xu, M.; Yan, R.; Zheng, C.; Qiao, Y.; Han, J.; Sheng, C. Status of trace element emission in a coal combustion process: A review. *Fuel Process. Technol.* **2004**, *85*, 215–237. [[CrossRef](#)]
48. Cao, X. Policy and regulatory responses to coalmine closure and coal resources consolidation for sustainability in Shanxi, China. *J. Clean. Prod.* **2017**, *145*, 199–208. [[CrossRef](#)]

49. Liu, X.; Bai, Z.; Shi, H.; Zhou, W.; Liu, X. Heavy metal pollution of soils from coal mines in China. *Nat. Hazards* **2019**, *99*, 1163–1177. [[CrossRef](#)]
50. Lin, B.; Shi, L. Do environmental quality and policy changes affect the evolution of consumers' intentions to buy new energy vehicles. *Appl. Energy* **2022**, *310*, 118582. [[CrossRef](#)]
51. Zhang, Q.; Zheng, Y.; Tong, D.; Shao, M.; Wang, S.; Zhang, Y.; Xu, X.; Wang, J.; He, H.; Liu, W.; et al. Drivers of improved PM<sub>2.5</sub> air quality in China from 2013 to 2017. *Proc. Natl. Acad. Sci. USA* **2019**, *116*, 24463–24469. [[CrossRef](#)]
52. Tsai, P.J.; Young, L.H.; Hwang, B.F.; Lin, M.Y.; Chen, Y.C.; Hsu, H.T. Source and health risk apportionment for PM<sub>2.5</sub> collected in Sha-Lu area, Taiwan. *Atmos. Pollut. Res.* **2020**, *11*, 851–858. [[CrossRef](#)]
53. Yan, R.-H.; Peng, X.; Lin, W.; He, L.-Y.; Wei, F.-H.; Tang, M.-X.; Huang, X.-F. Trends and challenges regarding the source-specific health risk of PM<sub>2.5</sub>-bound metals in a Chinese megacity from 2014 to 2020. *Environ. Sci. Technol.* **2022**, *56*, 6996–7005. [[CrossRef](#)] [[PubMed](#)]
54. Li, X.; Yan, C.; Wang, C.; Ma, J.; Li, W.; Liu, J.; Liu, Y. PM<sub>2.5</sub>-bound elements in Hebei Province, China: Pollution levels, source apportionment and health risks. *Sci. Total Environ.* **2022**, *806*, 150440. [[CrossRef](#)]
55. Lin, X.; Gu, Y.; Zhou, Q.; Mao, G.; Zou, B.; Zhao, J. Combined toxicity of heavy metal mixtures in liver cells. *J. Appl. Toxicol.* **2016**, *36*, 1163–1172. [[CrossRef](#)] [[PubMed](#)]
56. Xu, X.; Li, Y.; Wang, Y.; Wang, Y. Assessment of toxic interactions of heavy metals in multi-component mixtures using sea urchin embryo-larval bioassay. *Toxicol. Vitro.* **2011**, *25*, 294–300. [[CrossRef](#)] [[PubMed](#)]
57. Alsop, D.; Wood, C.M. Metal and pharmaceutical mixtures: Is ion loss the mechanism underlying acute toxicity and widespread additive toxicity in zebrafish? *Aquat. Toxicol.* **2013**, *140*, 257–267. [[CrossRef](#)] [[PubMed](#)]
58. Gebara, R.C.; Alho, L.d.O.G.; Rocha, G.S.; da Silva Mansano, A.; Melão, M.d.G.G. Zinc and aluminum mixtures have synergic effects to the algae *Raphidocelis subcapitata* at environmental concentrations. *Chemosphere* **2020**, *242*, 125231. [[CrossRef](#)] [[PubMed](#)]
59. Zhu, B.; Wu, Z.-F.; Li, J.; Wang, G.-X. Single and joint action toxicity of heavy metals on early developmental stages of Chinese rare minnow (*Gobiocypris rarus*). *Ecotoxicol. Environ. Saf.* **2011**, *74*, 2193–2202. [[CrossRef](#)] [[PubMed](#)]
60. Mebane, C.A.; Dillon, F.S.; Hennessy, D.P. Acute toxicity of cadmium, lead, zinc, and their mixtures to stream-resident fish and invertebrates. *Environ. Toxicol. Chem.* **2012**, *31*, 1334–1348. [[CrossRef](#)] [[PubMed](#)]
61. Liu, Z.; Gao, W.; Yu, Y.; Hu, B.; Xin, J.; Sun, Y.; Wang, L.; Wang, G.; Bi, X.; Zhang, G. Characteristics of PM<sub>2.5</sub> mass concentrations and chemical species in urban and background areas of China: Emerging results from the CARE-China network. *Atmos. Chem. Phys.* **2018**, *18*, 8849–8871. [[CrossRef](#)]
62. SSY. Table 6–15: Consumption and Composition of End-Use Energy in 2013; Shanxi Statistics Yearbook. 2014. Available online: <http://tjj.shanxi.gov.cn/tjsj/tjnj/nj2014/html/njcx.htm> (accessed on 25 June 2022).
63. SSY. Table 6–13: Consumption and Composition of End-Use Energy in 2014; Shanxi Statistics Yearbook. 2015. Available online: <http://tjj.shanxi.gov.cn/tjsj/tjnj/nj2015/indexch.htm> (accessed on 25 June 2022).
64. SSY. Table 6–13: Consumption and Composition of End-Use Energy in 2015; Shanxi Statistics Yearbook. 2016. Available online: <http://tjj.shanxi.gov.cn/tjsj/tjnj/nj2016/indexch.htm> (accessed on 25 June 2022).
65. SSY. Table 6–13: Consumption and Composition of End-Use Energy in 2016; Shanxi Statistics Yearbook. 2017. Available online: <http://tjj.shanxi.gov.cn/tjsj/tjnj/nj2017/indexch.htm> (accessed on 25 June 2022).
66. SSY. Table 6–13: Consumption and Composition of End-Use Energy in 2017; Shanxi Statistics Yearbook. 2018. Available online: <http://tjj.shanxi.gov.cn/tjsj/tjnj/nj2018/indexch.htm> (accessed on 25 June 2022).
67. SSY. Table 6–13: Consumption and Composition of End-Use Energy in 2018; Shanxi Statistics Yearbook. 2020. Available online: <http://tjj.shanxi.gov.cn/tjsj/tjnj/nj2019/zk/indexch.htm> (accessed on 25 June 2022).
68. SSY. Table 6–14: Consumption and Composition of End-Use Energy in 2019; Shanxi Statistics Yearbook. 2020. Available online: <http://tjj.shanxi.gov.cn/tjsj/tjnj/nj2020/zk/indexch.htm> (accessed on 25 June 2022).
69. SSY. Table 6–13: Consumption and Composition of End-Use Energy in 2020; Shanxi Statistics Yearbook. 2021. Available online: <http://tjj.shanxi.gov.cn/tjsj/tjnj/nj2021/zk/indexch.htm> (accessed on 25 June 2022).